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RECEIVED 08 September 2023

ACCEPTED 26 October 2023

PUBLISHED 14 November 2023

CITATION

Ouyang H, Wang L, Sapkota D, Yang M,
Morán J, Li L, Olson BA, Schwartz M,
Hogan CJ Jr and Torremorell M (2023) Control
technologies to prevent aerosol-based disease
transmission in animal agriculture production
settings: a review of established and emerging
approaches.

Front. Vet. Sci. 10:1291312.

doi: 10.3389/fvets.2023.1291312

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Control technologies to prevent aerosol-based disease transmission in animal agriculture production settings: a review of established and emerging approaches

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Transmission of infectious agents via aerosols is an ever-present concern in animal agriculture production settings, as the aerosol route to disease transmission can lead to difficult-to-control and costly diseases, such as porcine respiratory and reproductive syndrome virus and influenza A virus. It is increasingly necessary to implement control technologies to mitigate aerosol-based disease transmission. Here, we review currently utilized and prospective future aerosol control technologies to collect and potentially inactivate pathogens in aerosols, with an emphasis on technologies that can be incorporated into mechanically driven (forced air) ventilation systems to prevent aerosol-based disease spread from facility to facility. Broadly, we find that control technologies can be grouped into three categories: (1) currently implemented technologies; (2) scaled technologies used in industrial and medical settings; and (3) emerging technologies. Category (1) solely consists of fibrous filter media, which have been demonstrated to reduce the spread of PRRSV between swine production facilities. We review the mechanisms by which filters function and are rated (minimum efficiency reporting values). Category (2) consists of electrostatic precipitators (ESPs), used industrially to collect aerosol particles in higher flow rate systems, and ultraviolet C (UV-C) systems, used in medical settings to inactivate pathogens. Finally, category (3) consists of a variety of technologies, including ionization-based systems, microwaves, and those generating reactive oxygen species, often with the goal of pathogen inactivation in aerosols. As such technologies are typically first tested through varied means at the laboratory scale, we additionally review control technology testing techniques at various stages of development, from laboratory studies to field demonstration, and in doing so, suggest uniform testing and report standards are needed. Testing standards should consider the cost-benefit of implementing the technologies applicable to the livestock species of interest. Finally, we examine economic models for implementing aerosol control technologies, defining the collected infectious particles per unit energy demand.

KEYWORDS

bioaerosol control, airborne pathogen transmission, swine health, bioaerosol control technology, airborne infection control, bioaerosol, livestock biosecurity, agricultural biosecurity

1. Introduction

Airborne transmission of infectious agents has been documented or suspected for many years (1), and airborne pathogens cause some of the most devastating, costly, and difficult-to-control diseases. Examples of airborne pathogens in humans include *Mycobacterium tuberculosis*, SARS-CoV-2, respiratory syncytial virus (RSV), and influenza virus (2–4). Airborne transmitted pathogens in livestock include porcine reproductive respiratory syndrome virus (PRRSV), influenza A virus (IAV), high pathogenic avian influenza (HPAI), foot and mouth disease (FMD), *Mycoplasma hyopneumoniae*, classical swine fever (CSF), and porcine epidemic diarrhea (PED) viruses (5–8). Each of these animal pathogens has been responsible for causing devastating losses, in particular when causing disease in regions with high farm densities, due to the pathogens' abilities to spread rapidly and, in some instances, cause zoonotic infections. The importance of having effective disease control strategies will continue as population growth expands (9), demand for animal protein increases (10), and feeding the world in a sustainable manner is central to food security. The cost of diseases to food animal production is significant and it cannot be ignored. For example, the 2014–2015 outbreak of HPAI was one of the most devastating foreign animal diseases in U.S. animal history, and containing it cost nearly \$850 million dollars in obligated response activities and indemnity payments (11). In swine, PRRSV costs the North American swine industry more than \$581 million dollars a year (12, 13).

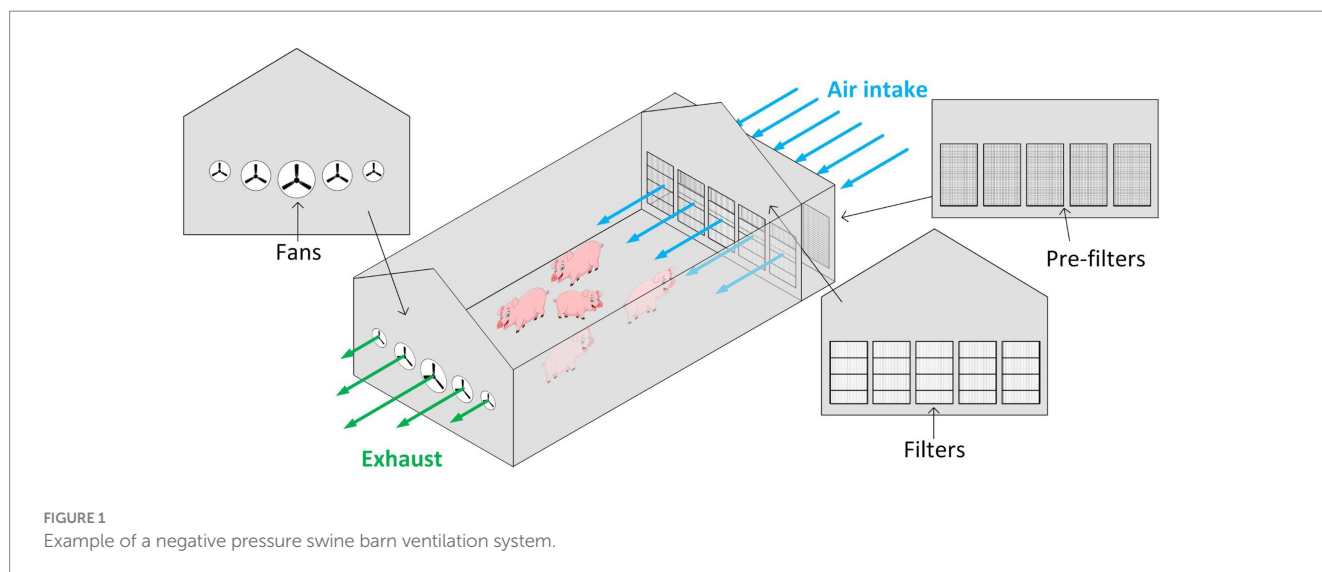
There is a need to have effective strategies to control the spread of bioaerosols, i.e., aerosols containing viable pathogens, in modern animal agriculture production settings. Animals are commonly raised in confined, naturally or mechanically ventilated buildings (see Figure 1), with high population densities. Over time, producers have developed measures to prevent disease introduction and disease spread to protect the health of animal populations. However, most of those measures have focused on minimizing disease risks through vaccination programs or biosecurity protocols that mitigate risks of contaminated fomites or infected animals (14, 15). Examples of such measures include employees showering in and out of farms, cleaning and disinfection of transport equipment, disinfection of materials prior to farm entry, and isolation and testing of replacement animals. These measures, although effective against many diseases, have not been sufficient to prevent the introduction (bioexclusion) and spread of airborne viruses (biocontainment).

There is also evidence of long-distance airborne transmission of viruses. In pigs, for example, PRRSV can be transported over extended distances by virus-laden aerosols (16–18). Evidence of long-distance airborne transport of PRRSV out to 4.7 km and 9.1 km has been reported. (19) In FMD, airborne transmission was the probable cause of several historic outbreaks where the FMD virus was suspected to travel over distances of 60 to 500 km over bodies of water (20) and land (21). Influenza A virus, which is a threat to both animal and public health because of its ability to change, reassort and cause pandemics (22), has also been detected in air samples from rooms of experimentally infected pigs (23), in air samples inside animal buildings and in the exhaust air from infected farms and at 1 mile from infected farms (24), highlighting the potential for aerosol transmission in pigs and between farms. Furthermore, the airborne spread of the HPAI virus was also implicated in the spread of avian influenza under certain conditions (25–27).

Given the evidence of long-distance airborne transmission of diseases in animals, in particular for PRRSV in pigs, filtration of incoming air to swine facilities has been proposed and implemented as a means to reduce the risk for PRRSV introduction (28–31). Filtration involves the passage of incoming air through fibrous filters, which mechanically (in most circumstances) collect aerosol particles, with the collection efficiency dependent upon both the sizes of the particles and the properties of the filter. The use of air filtration has been tested under experimental conditions and over the course of a 4-year study period involving a model of a swine production region. Under the study conditions, airborne transmission of PRRSV to susceptible populations housed in filtered facilities was prevented 100% of the time (19). As a result, the use of air filtration in breeding herds has become widespread in the Midwestern US (30, 32). Farms with air filtration had decreased incidence of PRRSV infections and improved wean pig quality (29, 33).

Although useful to prevent the introduction of airborne pathogens, air filtration implementation in large commercial animal settings is costly and has challenges. Costs associated with filtration include not only the purchase and replacement of the filters themselves, but also the increased energy requirements and increased energy consumption by the blower, to drive flow through filters. HEPA and high MERV (minimum efficiency reporting value) rated filters are widely considered the gold standard of air treatment technologies for pathogen control. However, their high capital cost, need for regular maintenance, and creation of large pressure drops in air systems pose a barrier to industrial application. For example, the capital cost of a HEPA filtration system has been estimated at \$1,500–\$2000 per boar/sow (34).

For this reason, the types of filters employed need to be judiciously selected, to ensure that they are not only effective in disease spread mitigation but also that costs are not overly cumbersome. Outside the animal production industry, e.g., in hospital environments as well as in industrial environments, there are alternative control technologies, such as electrostatic precipitators (ESPs) and ultraviolet type C (UV-C) light-based devices implemented in both removing aerosol particles from the air and inactivating pathogens in aerosols. Furthermore, over the last several decades and additionally intensified during the COVID-19 pandemic, there are newer control technologies in the developmental stage that may eventually be scalable toward farm-scale airborne disease transmission mitigation. Previous reviews have focused on general biosecurity measures (35) without much emphasis on aerosols. Some exceptions include identifying aerosol pollutants and the design of ventilation systems (35, 36) in poultry and swine farms. In addition, La et al. (37), reviewed different methods for aerosol research in virus transmission including PRRSV and African swine fever virus. However, a comprehensive assessment of existing and emerging virus aerosol control technologies is missing in the literature. With the already available wide variety of filtration options and the growing number of alternative technologies, we presently intend to review technologies directed at removing or inactivating airborne pathogens with a focus on those technologies that aim to prevent transmission among animals and from animals to people. We present the working principles of both existing and emerging technologies and discuss them in the context of using them in food animal production settings, particularly through incorporation with mechanically ventilated systems. We also discuss methods to test the effectiveness of the technologies to remove or inactivate viable



pathogens and identify opportunities for further development. Finally, we discuss the installation and maintenance costs associated with various technologies. Overall, we aim to spark interest in developing and implementing technologies that can help protect animal populations and people and that contribute to long-term disease control, sustainable animal production systems and food security.

2. Control technologies overview

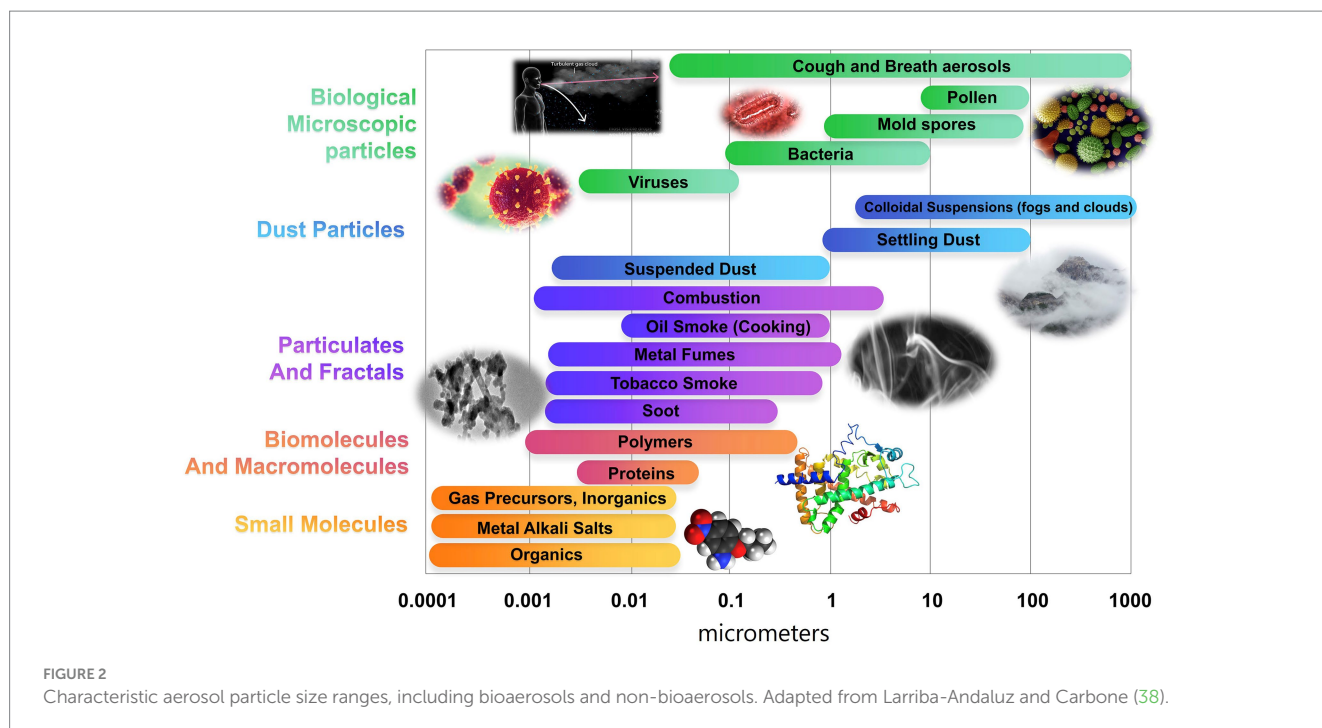
A variety of technologies are available as engineering controls to remove and/or inactivate bioaerosols. The performance of these technologies depends on environmental and operational factors, including but not limited to the size and type of aerosol particles (which vary over orders of magnitude, as noted in Figure 2), operation period, temperature, and relative humidity. We focus specifically on discussing the capabilities of technologies applicable as engineering controls (i.e., applicable at the room scale or higher, and not as personal protective equipment or at the single-animal level) and largely technologies that can be incorporated into mechanically ventilated systems. We subdivide these technologies into three classes:

- i. Established and implemented technologies. Solely consists of fibrous filters, which are ubiquitously employed in reducing the spread of PRRSV, influenza, and a variety of other respiratory pathogens. We review the performance and ratings of filters as a baseline to compare to other technologies.
- ii. Industrially- and medically-implementable technologies. Consisting of electrostatic precipitators and ultraviolet-C systems (both in-duct and upper room), they are commonly implemented in other industries but are less prevalent in livestock management. Nonetheless, because they have demonstrated scalable performance in aerosol particle removal and/or bioaerosol inactivation, they merit discussion and consideration toward agricultural biosecurity.
- iii. Emerging technologies. Consisting of more recently developed technologies, typically at the laboratory or bench scale. Technologies in this category, which may include photocatalytic

and plasma technologies, have not been tested at the scales (building sizes and flow rates) required for agricultural biosecurity and are hence several developmental steps away from direct implementation.

2.1. Established and implemented technologies – filters

If produced and installed properly, filters are very effective in removing a wide variety of particle types (dust, smoke, soot, bioaerosols) from flowing air streams, and are hence the most widely used technology in aerosol control. Filters are manufactured with multiple layers of non-woven fibers or with other materials, which are often cylindrical and randomly arranged inside the filter. As shown in Figure 3, the diameters of fibers often vary, with a filter typically described by the mean fiber diameter and a span of the fiber size distribution; given the random arrangement of fiber, the opening areas between cross-linked fibers vary in size as well within a filter (39, 43–45). The performance of filters varies greatly between filters and is described by the collection efficiency (η), which is defined as $\eta = 1 - C_{down} / C_{up}$, where C_{down} is the downstream particle concentration and C_{up} is the upstream particle concentration. When the filtration efficiency reaches 1 (100%), all particles flowing through the filter are collected by it. The operational principle of filters is depicted in Figure 4A at the single filter fiber level. With a uniform incoming flow, the trajectory of air around the filter fiber is depicted by the streamlines. When the flow enters the filter, streamlines are curved and compressed near the surface of the fiber. Particles initially on streamlines approaching a filter fiber can be collected by several disparate mechanisms. First, as streamlines bend, particles with high mass-to-drag ratios (larger size and higher density), will not bend in trajectory with streamlines. Such particles inertially impact with fibers (impaction) (46). While impaction acts upon larger particles, particle diffusion coefficients increase with decreasing particle diameter (47), and high diffusivity particles have fluctuating motion, moving from streamline to streamline stochastically. Such diffusive motion can also lead to particle-fiber collisions and hence collection (diffusion)



(48–50). Even particles that do not deviate from gas streamlines may collide with fibers through interception (51), which occurs for particles traveling along the streamline that approach a filter fiber at a distance less than the radius of the particle. As depicted in Figure 4C, impaction, diffusion, and interception are commonly the three mechanisms contributing the most to filter collection and enabling collection across orders of magnitude in particle diameter. However, as depicted in Figure 4D, because these three mechanisms are size-dependent, all filters have a collection efficiency curve with the most penetrating particle diameter (MPPD, the particle size of lowest collection efficiency) (52). The MPPD is typically in the 100 nm–400 nm diameter range for fibrous filters, as such particles have reduced diffusion coefficients but also insufficient inertia and size for collection by impaction or interception. The collection efficiency curve is affected by the filter face velocity (the speed of the flow), the diameter of the fibers and their size distribution (53), and other microstructural influences, such as solidity/porosity. A common assumption, which is reasonably valid in filters for aerosols, is that particles adhere irreversibly to fibers upon collision. Additional filtration mechanisms can also serve to collect particles in fibrous filters, including gravitational sedimentation for extremely large particles, sieving (Figure 4B), which occurs for particles too large to traverse through openings between fibers, and electrostatic effects, both for charged particles and uncharged particles (dielectrophoresis) within electret filters (54, 55), i.e., filters with intentionally introduced charge distributions on them.

To model collection by different mechanisms, it is common to invoke single fiber collection efficiency theories (56), which derive from more general collector theories for porous media and packed beds (57). However, while modeling filters can lead to predictions of efficiency that are qualitatively consistent with their performance; measurements are consistently required to properly define filter function. To evaluate the filter performance, ANSI and ASHRAE (American National Standards Institute and the American Society of

Heating, Refrigerating and Air Conditioning Engineers) have established a standardized test that defines the Minimum Efficiency Reporting Values (MERV) rating for filters (the ANSI/ASHRAE 52.2–2017 standard). Table 1 shows how MERV ratings are assigned to filters based on their average filtration efficiencies in three different size ranges: 0.3 μm – 1 μm, 1 μm – 3 μm, and 3 μm – 10 μm. Note that smaller particles are not measured in the test, as doing so requires more specialized equipment than standard optical particle spectrometers, and the expectation is that below 0.3 μm, as particle diameters decrease, collection efficiency increases with a “U-shape” collection efficiency as shown in Figure 4D. As the MERV rating increases, the filtration efficiencies in all three ranges increase as well. For example, MERV 16 filters must reach 95% efficiency across all three ranges, while MERV 10 filters need only achieve 85% in the range of 3 μm – 10 μm, 50% in the range of 1 μm – 3 μm and have a negligible efficiency in the range of 0.3 μm – 1 μm. In general, using filters for bioaerosol removal, the higher the MERV rating, the better the removal efficiency. When a filter has filtration efficiencies in the three different size ranges which would lead to three different assigned MERV ratings, it is assigned to the lowest MERV, hence the MERV defines the “minimum efficiency” for the filter. Further, high-efficiency particulate air filters, known as HEPA filters, and Ultra-low particulate air (ULPA) filters meet requirements beyond the MERV rating scale. For example, HEPA filters must have a measured collection efficiency of 99.97% or greater for particles of 0.3 μm diameter and even higher collection efficiencies for particles smaller and larger in diameter.

While at first glance it would appear higher MERV rated filters are always preferable, it is important to note that filters must be incorporated with forced air ventilation systems which are implemented to drive the pollutant air through the filter, and adding the filter into the ventilation system increases the pressure drop. The product of the pressure drop and air flow rate, divided by the electrical-to-mechanical energy efficiency of the blower used, gives the

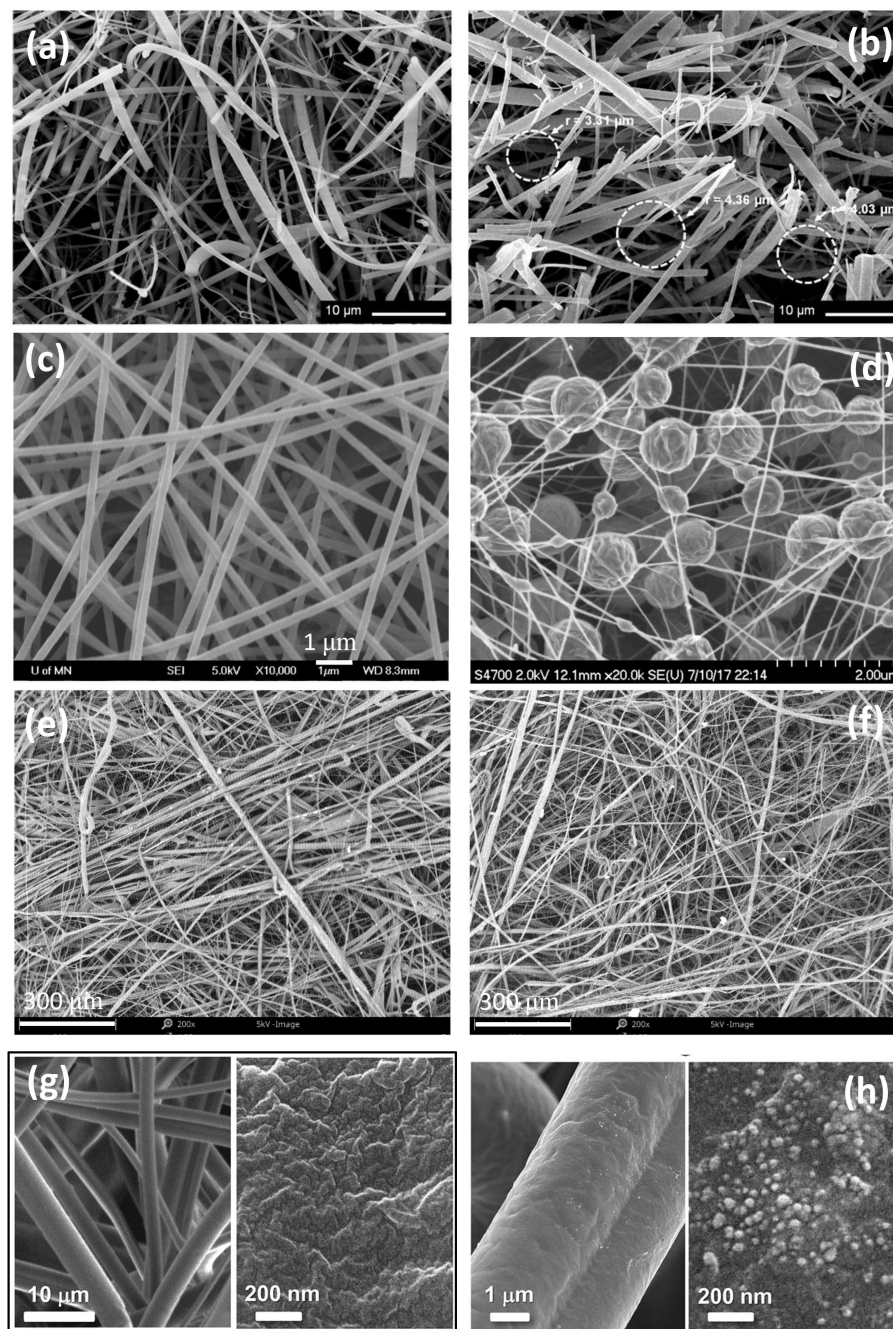
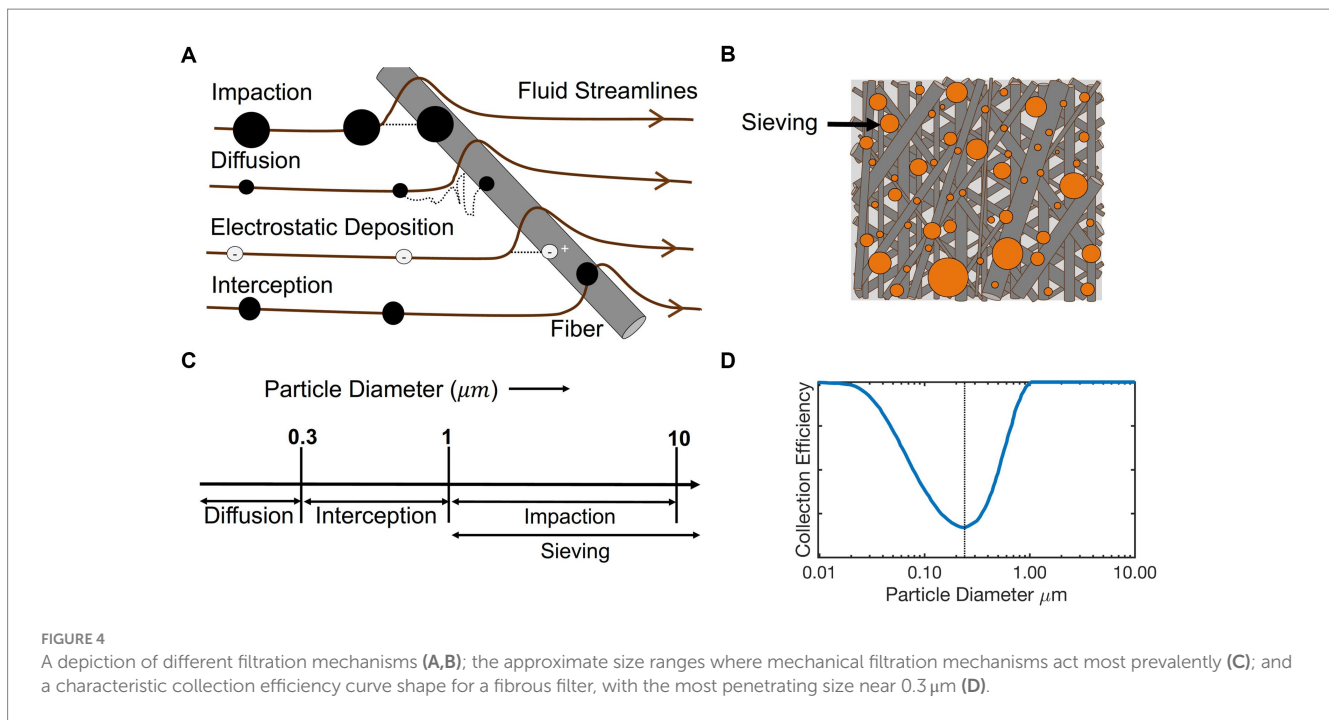


FIGURE 3 SEM images of typical quartz fiber filters with varying fiber diameters and varying “void” space dimensions [(A,B) adapted from Suárez-Peña et al. (39)]; a smooth nanofiber filter (C); a beaded nanofiber filter (D) [(C,D) adapted from Kim et al. (40)], to be consistent with other images. TiO₂ treated fiber filters [(E,F) adapted from Lou et al. (41)]; bare HEPA filter (left) with a zoom-in image of the fiber surface (G), a tannic acid-treated HEPA filter (left) with a zoomed-in image of the fiber surface (right) (H) [(G,H) adapted from Kim et al. (42)].

additional power (rate of energy use) associated with using filters. Therefore, a higher pressure drop requires a higher energy consumption to drive the same amount of flow. Filters with a higher MERV rating are typically made with smaller diameter fibers or made of higher solidity, often resulting in higher pressure drop (reduced permeability) (60), which increases energy cost or even requires replacement with a high-power blower/compressor to maintain the ventilation rate.

To quantify the filter’s overall performance using both the pressure drop (ΔP) and the collection efficiency (CE), the figure of merit (FOM) (43, 61) is introduced with the units of Pa⁻¹ and calculated via the equation:

$$FOM = \frac{\ln\left(\frac{1}{1-CE}\right)}{\Delta P} \tag{1}$$



The FOM, also commonly referred to as the quality factor, is a metric that enables a comparison of the energy costs associated with a filter use, in comparison to the collection it provides, and in general higher FOM filters are preferred in all applications; a lower pressure drop and high collection efficiency will result in a higher FOM. If two filters have the same collection/filtration efficiency, the filter with a higher FOM, thus a lower pressure drop, will be preferred. In practice, multiple banks are commonly installed along side walls, or multiple filter modules are installed in the attics in animal production facilities. In these installations, it is common to place two filters in series; the first a pre-filter of low MERV rating (to collect large dust particles and prevent the loading of the second filter) and a higher MERV-rated second filter, whose purpose is to collect micrometer-to-submicrometer particles. The FOM can also be computed for two filters used in series; in this case, the collection efficiency is the net efficiency considering both filters (upstream before the first filter, and downstream after the second filter), and the pressure drop is the sum of the pressure drops across the individual filters. Different products have also been marketed to improve FOM (improve collection efficiency and/or decrease the pressure drop). For example, electrostatic-enhanced (electret) filters can raise the figure of merit with a lower pressure drop and have been implemented in N95 respiratory face masks (62, 63). However, degradation of the electric charges over time will lower the filtration efficiency. Filters on the market are almost always pleated (54, 64) to increase the surface area of the filter, thus increasing the capacity for the collection of pollution. While pleating generally reduces the pressure drop, filters can be over pleated leading to an increase in pressure drop causing a lower FOM.

Unsurprisingly, as the collection efficiency is a function of particle diameter, the FOM is also size-dependent. In addition, in practice, the FOM varies with time, as the filter performance changes with loading [particles deposit onto and stay on the filter (65)]. While loading often leads to an increase in filtration efficiency, it typically leads to a proportionally greater increase in pressure drop (66). Thus, the

replacement of filters periodically is required and the replacement frequency depends on the quality of the filter, flow condition, and the level of particle pollution in the air stream. It also depends on the filter arrangement; for example, in the noted case of a low MERV rated pre-filter and a higher MERV-rated main filter downstream, the pre-filter loads faster and needs more frequent replacement than the main filter. In general, for the same type of filter (same MERV rating and initial pressure drop), if the FOM of a filter decreases with time rapidly due to higher particle concentrations, the filter needs to be replaced more frequently.

For bioaerosol control and removal, even though the filtration efficiency is not strictly equivalent to bioaerosol removal efficiency (67), the filter still follows the same collection principles for bioaerosols as it does for other types of aerosols. For this reason, in collecting particles, the filter material is less significant than its structure and for aerosol filtration, it is typically not necessary to design filters out of materials promoting specific interactions with biomolecules. Bioaerosols which are typically larger than the size of bare viruses, fungi, and bacteria due to attachments of other materials, range from a few tens of nanometers to a few micrometers. Nevertheless, filters already target this range of particle diameters. Often, high MERV rating filters are needed to achieve high efficiency in bioaerosol removal from the air stream (68–70). At the same time, for bioaerosols, filters alone may not be sufficient: collected bacteria and fungi on filters may grow and eventually be resuspended into the air stream (71). In many instances for viruses, the half-lives on surfaces such as filter media have been measured to be hours to days (72, 73). The inactivation of collected bioaerosols upon capture by the filters is therefore also important to consider. To enhance the inactivation rate and prevent bacterial or fungal growth, one approach is to coat filters with different antimicrobial materials targeting collected bioaerosols. Numerous researchers have worked with a variety of coatings to this end, though with variable results on their effectiveness, and without a universal, commercialized approach

TABLE 1 Collection efficiency requirements for minimum efficiency reporting value (MERV) rated filters.

ASHRAE standard 52.2 minimum efficiency reporting value (MERV) rating	Average particle collection efficiency in the size range of			Typical filter types
	Range 1 (0.3–1 μm)	Range 2 (1–3 μm)	Range 3 (3–10 μm)	
MERV 1–4	n/a	n/a	< 20%	Fiberglass and aluminum mesh
MERV 5	n/a	n/a	≥ 20%	Pre-filters and dust filters
MERV 6	n/a	n/a	≥ 35%	
MERV 7	n/a	n/a	≥ 50%	
MERV 8	n/a	≥ 20%	≥ 70%	
MERV 9	n/a	≥ 35%	≥ 75%	Home and commercial building filter
MERV 10	n/a	≥ 50%	≥ 80%	
MERV 11	≥ 20%	≥ 65%	≥ 85%	
MERV 12	≥ 35%	≥ 80%	≥ 90%	
MERV 13	≥ 50%	≥ 85%	≥ 90%	Hospital and clean room filter
MERV 14	≥ 75%	≥ 90%	≥ 95%	
MERV 15	≥ 85%	≥ 90%	≥ 95%	
MERV 16	≥ 95%	≥ 95%	≥ 95%	
MERV 17	99.97%	≥ 99%	≥ 99%	HEPA and ULPA
MERV 18	99.997%	≥ 99%	≥ 99%	
MERV 19	99.9997%	≥ 99%	≥ 99%	
MERV 20	99.99997%	≥ 99%	≥ 99%	

The MERV rating for a filter is determined by the ASHRAE/ANSI 52.2–2017 standard (58) and is based on “Range” corresponding to the filter’s lowest MERV value (59).

presently available. Nanoparticle-coated filters with antimicrobial materials such as silver (Ag), copper (Cu), and carbon nanotubes. Among others have been recently examined. (74–77) Natural products including *Melaleuca alternifolia* (tea tree oil) coated filters have also been shown to inactivate filter-collected bioaerosols (78–81). For efficient bioaerosol control using surface-coated filters, the use of pre-filters and regular maintenance [to remove collected aerosols and avoid dust-loading (82)] and replacement would be required with a higher frequency, but this may prove prohibitively costly.

Despite the cost of high-quality filters and frequent replacements, filters are the most commonly used method to control bioaerosols, mainly because of their simple implementation, and proven effectiveness. Laboratory and field tests have been performed to evaluate the performance of filter-based air filtration systems in controlling bioaerosol transmission in livestock. Specifically, a group of researchers in Minnesota, USA conducted a series of studies confirming that air filtration can reduce the occurrence of PRRSV infections (30, 31, 83, 84). In laboratory studies, Wenke et al. (85) tested four filters for removing different bioaerosols, including equine arteritis virus (EAV), PRRSV, bovine enterovirus (BEV), *Actinobacillus pleuropneumoniae* (APP), and *Staphylococcus aureus*, and found that air filtration effectively reduces pathogen concentrations. Alonso et al. (29) traced outbreaks of PRRSV in sow farms and concluded that air filtration could significantly reduce the risk of PRRSV aerosol transmission by about 80%. Wenke et al. (86) implemented three different types of filters for existing ventilation systems in pig farms and suggested that recirculating air filter modules resulted in improved pig lung health.

Although there are multiple metrics quantifying the quality of filters in controlled measurement settings (MERV rating), the

performance of filters installed on-site still needs further assessment. The installed efficiency in the field varies from that measured in controlled conditions (87); the installed efficiency depends on filter type, airflow (face velocity), the HVAC system, and even the installation as air can bypass the filter if there are gaps at the edge of the filter (87). Furthermore, buildings where air filtration systems are installed may permit some degree of air leakage, given that the building may have construction gaps. Thus, more field tests are needed to better define the effectiveness of filters in farms and provide a better estimate of the cost-to-benefit ratio for the implementation of filters and developing a suitable protocol for maintenance.

2.2. Industrially and medically implementable technologies

2.2.1. Electrostatic precipitators

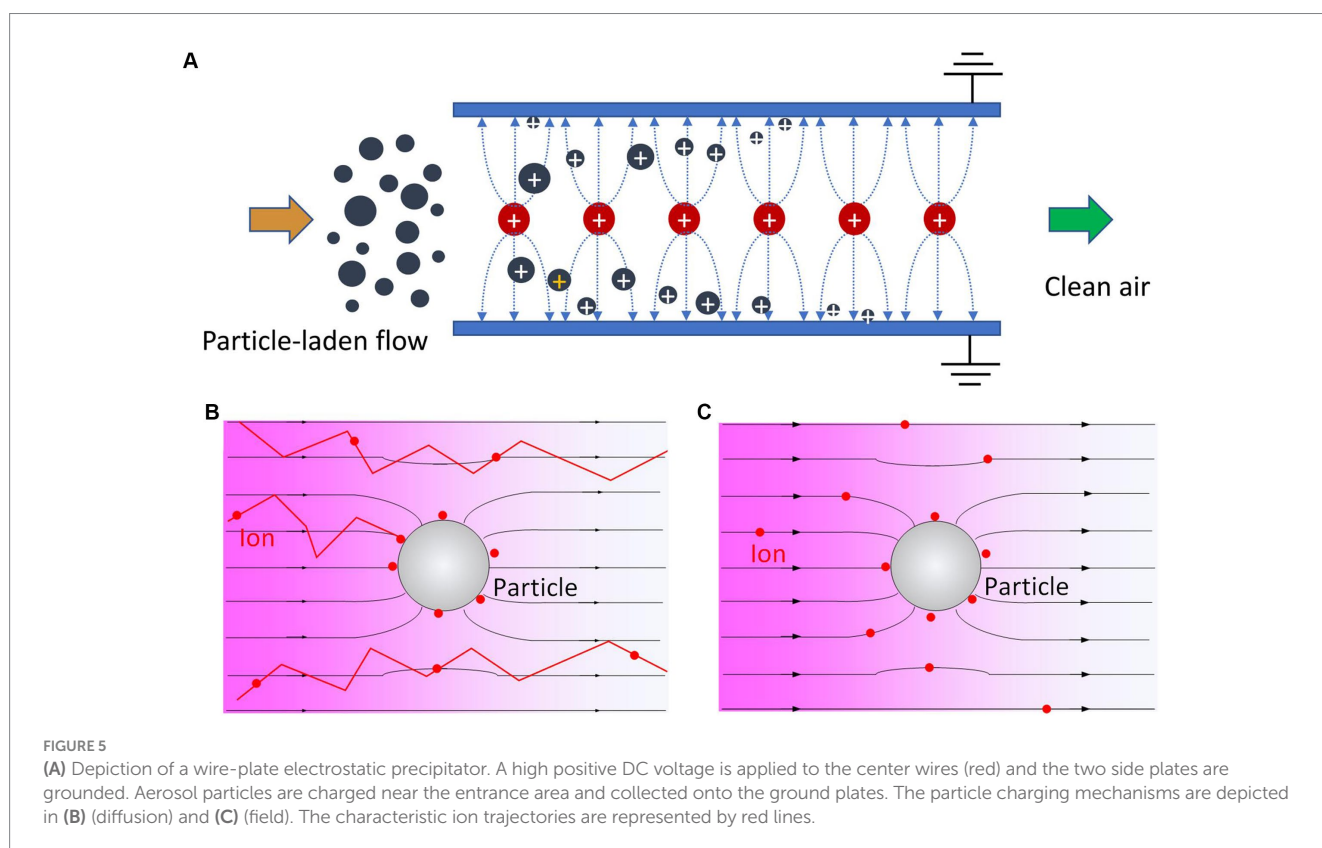
Electrostatic precipitators (ESPs) are an alternative to filters, primarily used in instances where filter loading and high-pressure drop need to be avoided (often high flow rates with high particle concentrations) (88, 89). An ESP typically consists of two types of electrodes with configurations commonly referred to as wire-plate, wire-cylinder, needle-plate, and spike-plate [see Figure 4 and Table 2 from Qu et al. (90)]. Here, the aerosol flows between both electrodes. A positive or negative high voltage (in the order of several kilovolts) is applied to the wire or needle electrodes, while the plates or cylinder are grounded, resulting in the formation of a corona discharge (91, 92) at the wire or needles (the location of highest electric field intensity due to the small radius of curvature). Ions are generated in a small zone around the high-voltage electrode where the electric field exceeds

the critical value for gas breakdown. These ions then move toward the grounded electrodes by Coulomb drift. Positive applied voltage yields positive ions, and negative applied voltage yields negative ions. Aerosol particles enter the ESP with the flow and become charged (93) via random collisions with ions [by a combination of diffusion (94) and field charging mechanisms (95, 96)], thus moving toward the grounded electrodes, provided the residence time in the ESP is sufficient for particle charging and electrophoretic deposition. While diffusion charging (Brownian with Coulomb drift motion, see Figure 5B) is a more relevant charging mechanism of sub-micrometer-sized particles, field charging (charges following electric field lines, see Figure 5C) is commonly the main mechanism of supermicrometer aerosol particle charging. Mainly due to these differences in charging efficiency combined with size-dependent particle electrical mobility, ESPs commonly exhibit a size-dependent collection efficiency with a minimum efficiency in the 0.3–0.6 μm size range and are highly efficient in removing particles below or above this range. The latter has been shown in both experiments and numerical simulations (93). Since the electrodes can vary in configuration, ESPs can be designed in different configurations (e.g., multi-stage, wet or dry, wire-plate, tubular) for various applications. Figure 5A depicts a common wire-plate ESP with positive high voltage applied at the center wires; polydisperse aerosol particles enter the ESP and become positively charged and are collected onto the ground plates at various locations. ESPs, unlike filters, have minimal pressure drop; therefore, they are often well-suited for large-scale applications. For example, high-flow rate (millions of cubic feet per minute) ESPs have been widely used in power plants to remove fly ash, soot, and other pollutants from coal combustion exhaust. (97–99) Gradually, ESPs have also been implemented in indoor air quality control applications, including

incorporation in stand-alone recirculating air cleaners and in-duct units in HVAC systems. Afshari et al. (100) have provided a systematic review of ESP as indoor air cleaners.

ESPs have also been used in various embodiments as bioaerosol samplers (101–104) or bioaerosol control devices (105–107). ESPs can control airborne bioaerosol by the collection effects (bioaerosols are collected onto the electrodes) and possibly ionization or reaction effects [viability of bioaerosols can decrease due to ozone, produced in some ESPs, or collision with high-energy ions (93, 108)]. The effectiveness of bioaerosol control is hence a combination of bioaerosol collection and inactivation, which in turn depends upon the particle's physical efficiency, electric field strength, flow rate, relative humidity, and the design of the ESP. (109) Though a proven technology for particle collection, due to highly variable geometries, lab and field tests are needed to investigate the efficiency of each newly designed ESP device for bioaerosol control application. Also, as ESPs produce ozone (110), therefore, the level of ozone needs to be monitored in each device.

Thus far, to our knowledge, there has been no monitored and reported implementation of ESPs on farms. Nonetheless, like filters, ESPs have the potential to be a usable bioaerosol control technology in farms, because they enable high flow rate without high pressure drop, can often be added to HVAC systems or barn ventilation systems without substantial remodeling, and the electrical power requirements are typically low in comparison to the increase blower power requirements for filters. Furthermore, while filters require regular replacement, ESPs only require electrode cleaning. Cleaning can be obviated as well; wet ESPs (111–113) have been successfully designed and implemented in the power industry, which remove collected particles continuously by a thin water film covering the



collection plates. Combining ESPs with filters may also improve the collection efficiency (114) while extending the filter life, thus lowering the cost. In summary, ESP technology is potentially a viable alternative or augmentation to filter systems in farms: both remove bioaerosols by the collection, yet ESP requires less energy due to low-pressure drop. However, ESPs would need to be further tested for such applications.

2.2.2. Ultraviolet light sources

Ultraviolet (UV) light has a shorter wavelength (100–400 nm) compared to visible light and thus has higher energy; it can penetrate into cells, damaging pathogens and acting as a means of disinfection. There are three types of UV sources based on wavelength range: UV-A (315 nm to 400 nm), UV-B (280 nm to 315 nm), and UV-C (100 nm to 280 nm), with energy level inversely proportional to the wavelength. UV-A has less energy and thus is safer for people and livestock; it has been used in activating semiconductor photocatalysts for the mitigation of pathogens (115). UV-C, often referred to as ultraviolet germicidal irradiation (UVGI) (116), has the highest energy among all UV lights. UVGI produced by mercury arc lamps is emitted at 253.7 nm (117), making it extremely effective in damaging nucleic acids (117–120) because the maximum absorption wavelength of nucleic acids occurs near 260 nm.

UVGI has a more than 100-year history in application toward inactivating airborne pathogens (121), and as it is reviewed in detail by Reed (122), we provide only a brief overview here. The effectiveness in inactivation depends on the exposure time and the intensity of irradiation, and hence on system design and operating parameters. UV-C sources can be installed in stand-alone devices in the near-upper region of a room (123) or in HVAC ducts. However, because UVGI acts indiscriminately on nucleic acids, UV-C irradiation exposure poses adverse health effects to humans and livestock, in addition to pathogen inactivation. UVGI should only be in operation when humans and animals are not exposed directly to it, or measures need to be taken so exposure is well below exposure limits. Often, exposure avoidance is accomplished by incorporating UVGI into current HVAC duct systems. The effectiveness of this approach has been demonstrated more than 50 years ago by Jensen (124), where a ducted UV-C source led to greater than 99.9% inactivation for Coxsackie, influenza, Sindbis, and vaccinia viruses, and more than 90% inactivation for a variety of other pathogens. The ducted system in question was also scaled up and installed in the HVAC system of a hospital, with demonstrated as-installed bacteriophage inactivation rates above 99% for flow rates beyond 30,000 m³ h⁻¹ (125). More recently, in light of the COVID-19 pandemic, Qiao et al. tested a smaller-scale UV-C duct unit and concluded that the unit can achieve greater than 2-log reduction in viable coronavirus inactivation with a flow rate of 2,439 L min⁻¹ (126). Since UV-C light itself does not capture or remove pathogens from the fluid stream, it is often used as an additional device to augment current HVAC systems with filters (127); it is commonly available commercially for this purpose. Importantly, during filter disinfection, the filters still collect pathogens and the time of UV-C exposure is increased for collected pathogens, thus the UV-C intensity can be reduced as compared to that needed for standalone ducted systems. Incorporating UV technology into HVAC systems has been recently reviewed by Thornton et al. (128) and by Luo et al. (129). A more recent emerging technology is the far-UV-C (222 nm) source. Though of lower wavelength, because it is

farther from 260 nm, it has fewer adverse health effects on humans and animals. Far UV-C application has been demonstrated capable of inactivating airborne pathogens in room-size chambers (130, 131). In addition, as a final note, hand-held UV-C devices typically are not effective in air disinfection. As they can lead to human exposure to UV-C, they are typically of much lower fluence than in-duct units, and hence require significantly longer times for pathogen inactivation. Furthermore, their “coverage” areas are limited, also reducing efficacy.

UV irradiation systems are certainly promising in farms for bioaerosol control due to their high effectiveness in inactivation and ease of implementation. UV-C disinfection can also be used on surfaces (132) while transporting livestock, lowering the risk of introducing viruses during transport. (133) Unlike filters, the implementation of UV-C can also mitigate odor and other gaseous emissions by driving chemical reactions of such species, potentially improving overall air quality on farms and the lung health of animals (134–136). Eisenloffel et al. (137) added a recirculating UV-C module to a barn where air flowed first through a filter module and then passed into a ducted UV-C module. The module effectively reduced bioaerosol concentration in the barn but notably with variable concentrations and levels of effectiveness at different locations, specifically near the exit of the recirculating module; the sampled bioaerosol concentration was much lower than elsewhere. Overall, with high inactivation rates (>99%) UV-C may be installed in the air ductways as a cost-effective alternative to high performance filters in farm-scale applications (138). However, there is presently no information on the use of UV-C in farms on the long-term reduction of disease incidence and the overall functionality of UV-C systems under farm conditions; field testing of UV devices is rare. A final note on the implementation of UV-C for disinfecting air is that measurement of the UV dose is required at the beginning and during operation periodically to make sure the light bulbs are installed correctly and functioning properly (129, 139).

2.3. Emerging technologies

Filtration, ESPs, and UV sources are all either actively used in mitigating bioaerosols in farms, or have been scaled to the size and flow rates needed for implementation. At the same time, there are numerous emerging technologies that may be potential candidates to minimize bioaerosol concentrations and aerosol-based disease transmission in farms if demonstrated at scale in the future. These include reactive air disinfection technologies (140) such as ionization, microwave source, photocatalytic systems, non-thermal plasma systems, ozone generators (141) and chemical technologies, as well as devices incorporating these technologies in combination with more traditional technologies (e.g., filters). The majority of these technologies remain under development at the laboratory scale, though some have been successfully commercialized. While it is not within the scope of this review to describe all such technologies in detail, here, we briefly overview several technologies and their potential for implementation on farms.

2.3.1. Ionization technologies – bipolar and unipolar

Ionization technologies have been widely used in clean air applications by adding them in-duct to HVAC systems (142–144) or

by incorporating them into stand-alone air purifiers (145, 146). Aerosol particles can deposit on walls both in rooms and HVAC systems; in the absence of other forces, deposition is primarily due to diffusive motion or sedimentation. However, when particles are ionized, they will respond to electric fields (as in ESPs), and simply ionizing particles may enhance deposition through naturally or inadvertently present electric fields in indoor spaces, as recently shown by Kolarž et al. (147). Ionization technologies have also been demonstrated for bioaerosol control (148), and may additionally inactivate pathogens through the reactions taking place in them. As in ESPs, ionization technologies often consist of a high voltage electrode (usually a needle) with an applied positive or negative voltage, producing a corona discharge in the air. In unipolar ionization a single polarity is used and particles attain only positive or only negative charge. In bipolar ionization, either two distinct electrodes are used of opposite polarities, or an alternating voltage is applied to a single electrode, resulting in the formation of both positive and negative ions, with the particles becoming both positively and negatively charged. It has been shown that placing an ionizer in line with a filter can enhance filtration efficiency (143, 149), and there are studies noting that ionizers can disinfect bioaerosol-contaminated filter surfaces (150–153). However, the efficacies of devices with ionization technologies vary greatly from device to device and still need further investigation. In addition, the performance of corona ionizers is subject to ambient conditions, including temperature and relative humidity, and thus these parameters should be monitored and reported for repeatable testing and comparison with other filtration technologies (154–156). Zeng et al. (157) tested a commercially available in-duct bipolar ionization device both in chamber tests and in the field and concluded that the device had minimal impact on particle removal within the short duct length in its current setting. The performance of ionizers also depends on ventilation conditions such as flow velocity, air temperature, and relative humidity (158, 159). Byproducts may also form during the ionization process, such as volatile organic components (VOCs) and ozone (O₃) which need to be monitored during operation (157).

2.3.2. Microwave technologies

Microwaves can inactivate pathogens through both thermal (radiation-induced heating) and non-thermal (microwave irradiation) effects (160, 161). Water absorbs microwave irradiation particularly well, leading to rapid heating. Traditionally, microwaves use heating of water to directly destroy microorganisms, making them suitable for disinfecting, particularly waterborne bacteria and fungi. For airborne bacteria and fungi, microwave systems can still facilitate inactivation but it is less clear whether this is due to heating or non-thermal effects. Several studies suggest that the non-thermal effect is dominant in the inactivation of airborne bacteria (160, 162, 163). For airborne viruses, microwaves can also facilitate inactivation via damaging proteins (164), with several studies suggesting that here too the non-thermal effect is dominant in inactivating airborne viruses (164, 165). In implementation, often, the microwave generator will be installed in the walls of a chamber or duct. The effectiveness of disinfection will depend on the microwave power, bioaerosol type, the flow, system volume (exposure time), and relative humidity (161, 163, 166). Although less studied for bioaerosols, overall, microwaves are an established technology, and hence may find application in bioaerosol mitigation, e.g., in inactivating pathogens on filters (167).

2.3.3. Photocatalytic systems

Photocatalytic materials facilitate chemical reactions when exposed to radiation, such as the generation of reactive oxygen species (ROS), or toxic metal ions if the photocatalytic material contains metal. (168) Such species, in turn, can inactivate pathogens through chemical oxidation reactions, toxic exposure, and even physical damage (169). Photocatalytic disinfection has been applied to waterborne and foodborne viruses in numerous instances (170, 171) and is more established in these applications. However, since the pathogens need to have direct contact with the photocatalytic material to achieve disinfection, or the species generated need to be transported to pathogens, there are significant challenges in adapting photocatalysis alone for airborne pathogens (169).

Most photocatalytic systems are hence coupled with filters to collect bioaerosol particles, with the photocatalytic reaction occurring for collected particles on the filter surface. The most popular photocatalysis material for bioaerosol control is TiO₂, due to its low cost, nontoxicity, strong oxidizing ability, and long durability (172). Recirculating air purifiers with TiO₂-based photocatalytic technology have been commercially developed and tested for indoor air quality control including volatile organic components (VOCs) and airborne viruses (173–176). For in-duct HVAC applications, TiO₂-coated filters or TiO₂ filters can be utilized for bioaerosol control. (177–179) However, the effectiveness of photocatalysis on disinfection depends on the flow rate, relative humidity, air temperature, catalyst structure, irradiation source, and power (169, 180). The catalyst's lifetime is limited (181) and requires regeneration or replacement. By-products (182) can also form during oxidation reactions.

2.3.4. Non-thermal plasma (NTP)

Plasmas are a state of matter containing elevated densities of free electrons; such electrons move at extremely high speeds (and hence have high temperatures). However, if the surrounding gas remains at a low temperature, non-equilibrating with the high-speed electrons, it is called a non-thermal plasma (NTP). While NTPs refer broadly to plasmas where the gas is below the electron temperature (and hence the gas may still be elevated in temperature), it is possible to generate the NTP in air where the air temperature remains close to room values. NTPs are highly reactive environments, and hence have been widely investigated for sterilization purposes as they can inactivate pathogens without producing waste. NTP devices have been successfully used to disinfect surfaces and liquid solutions. NTPs have also been used to treat wounds (183), cancer (184), and in a variety of other medical applications (185–187). Several recent review papers (188–192) have been published to examine the current application state of NTP in the decontamination of pathogens, with a focus on viruses (motivated by the COVID-19 pandemic).

With specific regard to airborne pathogens, NTPs generate ions and free electrons, as well as radicals, high-energy photons, and many other reactive oxygen species (ROS) and nitrogen species (RNS), all of which can effectively damage pathogens (188). NTPs can effectively sterilize the air locally (193, 194). Researchers continue to develop NTP technologies for indoor air bioaerosol control. For example, a volumetric dielectric barrier discharge (DBD) (195, 196) non-thermal plasma has been developed toward installation inside an HVAC duct. Nayak et al. demonstrated that an NTP can achieve inactivation of airborne viruses in flight within several milliseconds (197–200). Zhang et al. (201) tested the effects

of relative humidity and flow rate for in-duct DBD systems with various electrode arrangements. However, even though multiple studies have demonstrated that DBDs can effectively mitigate viruses in small wind tunnels, field tests are lacking, and further developments are needed to work with large-scale realistic HVAC systems. Scalability remains an issue in NTP application, as does ensuring ROS species are not produced at concentrations too high for humans and animals.

3. Testing approaches of technology effectiveness

As noted in the prior section, while there are promising results for a number of emerging bioaerosol control technologies, further testing of such technologies is paramount. Selecting appropriate testing methods for bioaerosol control technologies is challenging, as there are currently no universally applied standards specific to bioaerosols as there are for the characterization of the physical collection efficiency of filters. The choice of testing methods depends largely on the development stage, working principles, and application of the technology. Bioaerosol control technologies generally operate using two principles: inactivation and removal. The testing approaches for technologies that can inactivate pathogens in bioaerosols vary according to their development stage. Typically, small-scale laboratory tests are performed in the early stages of development to demonstrate the technology's ability to inactivate pathogens on a surface, i.e., after collection. This is often done because of the costs associated with more direct aerosol testing as well as the need for specialized equipment and facilities for aerosol research. Subsequently, chamber or flow tube tests (including larger-scale wind tunnels) are conducted to determine the efficiency of bioaerosol control with tunable operating variables, such as the exposed bioaerosol size distribution, system and control technology flow rate, and relative humidity. Finally, field tests or near-field tests are conducted on-site or in a simulated field environment to evaluate control technology *in-situ* effectiveness for bioaerosol control. In contrast, if the technology's primary working principle is to collect and remove bioaerosols from the air stream only, small-scale surface inactivation tests are not applicable. Instead, flow tube or chamber tests are more common in initial development, followed by wind tunnel or large room testing. As such devices typically incorporate or function akin to filters, the single-pass wind tunnel, used in ASHRAE testing to determine filter MERV ratings, is ideal for control technologies that couple to the mechanical ventilation systems. Finally, of note is that bioaerosol control devices often integrate multiple technologies to enhance their overall effectiveness. For instance, UV light may be combined with filters to inactivate and remove pathogens from the air in series with one another. Testing may take place on components individually, or on the integrated unit, depending on the state of development.

In this section, we review research relevant to testing control technologies at different scales and different development stages—small-scale testing, including surface inactivation, smaller flow tubes, chambers and wind tunnels, as well as simulated field or true field tests. We then proceed to discuss parameters that must be considered in designing control tests, including bioaerosol generation, sampling, and assays.

3.1. Laboratory testing

3.1.1. Small-scale plate test

Without aerosolizing pathogens, the small-scale plate test often is used to examine the inactivation mechanisms and effectiveness of bioaerosol inactivation technologies over various pathogens, commonly incorporating UV (202, 203), ozone, photocatalysis, non-thermal plasma (200), and ion generation (204) or to demonstrate technologies for surface decontamination applications (205). Figure 6 displays a representative test setup from Noyce et al. (204), used in determining bacterial inactivation rates due to ion bombardment in a high electric field environment. An agar plate with viable pathogens is placed in a biosafety chamber and exposed directly to the inactivation technology. The number of viable pathogens or TCID₅₀ (tissue culture infectious dose) is measured with and without the technology after pre-defined exposure times. Because the setup is simple to implement, the test is very useful in earlier technological development stages toward proof-of-concept. However, demonstration of surface inactivation is difficult to extrapolate to performance at large scale and for airborne bioaerosols, as the collection step (if needed) is obviated and the inactivation mechanism acts upon surface-bound pathogens. Therefore, surface tests alone should not be considered complete as validation tests for bioaerosol control technology performance.

3.1.2. Flow-tube tests

In addition to the small-scale plate test, tests are needed to demonstrate the inactivation or removal of airborne pathogens. Small-scale flow-tube laboratory tests are likely the most suitable for this purpose. As shown in Figure 7, adapted from Kettleson et al. (107) and used in examining the performance of a soft X-ray enhanced electrostatic precipitator in collecting and inactivating MS2 bacteriophages (107, 206), the test setup typically includes three parts: bioaerosol generation, the in-line technology, and the bioaerosol sampler (for subsequent assay) (119, 124, 194, 207–209). We discuss bioaerosol generation approaches subsequently; in the figure, bioaerosols were generated using an atomizer or nebulizer with a suspension of bacteriophages in culture broth and passed through the technology (ESP in the figure) in operation with control over the volumetric flow rate of air passing through the ESP. Bioaerosols will be either collected (with possible inactivation), not collected but inactivated, or both not collected and not inactivated during passage through the technology. Downstream, the pathogen-containing particles are collected using a sampler, a liquid impinger (BioSampler), and a filter in the diagram, with samples extracted for titration and RT-qPCR (reverse transcriptase quantitative polymerase chain reaction) assays. While flow-tube tests can be run in principle with any flow rate, it is commonplace to first use low or medium flow rates (1–200 L min⁻¹) to demonstrate the working mechanism of the technology. Frequently, this leads to laminar flow within the device, while during actual application at scale, flows are turbulent. Smaller flow tubes can also permit larger concentrations of reactive species (e.g., ozone, ions, radicals) than are achievable at a larger scale. Therefore, similar to the surface test, flow-tube tests are important in terms of proof-of-concept and technology development, but should not be extrapolated to larger systems without further testing or without carefully ensuring the conditions tested do not bias results.

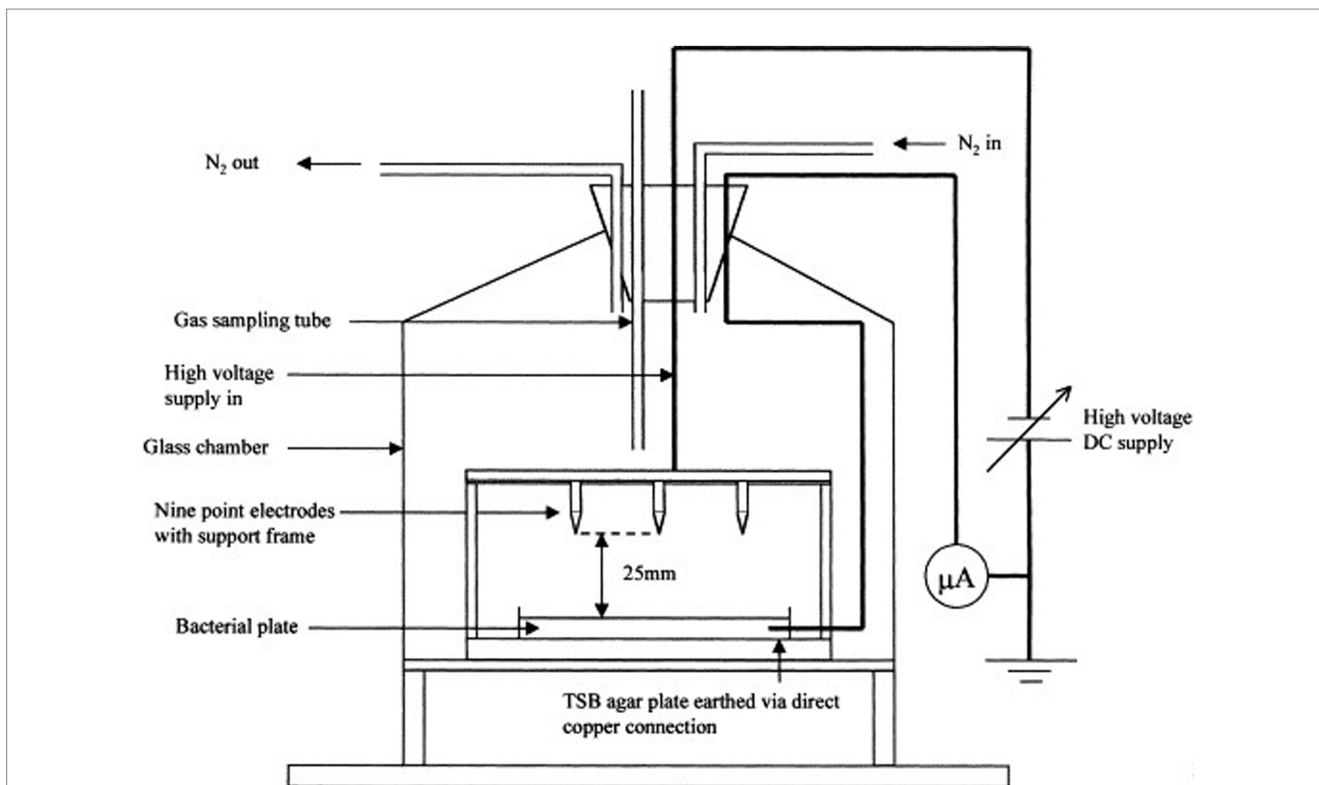


FIGURE 6
A schematic diagram of an experimental setup for bacterial exposure to negative or positive ions in nitrogen, Adapted from Noyce et al. (204).

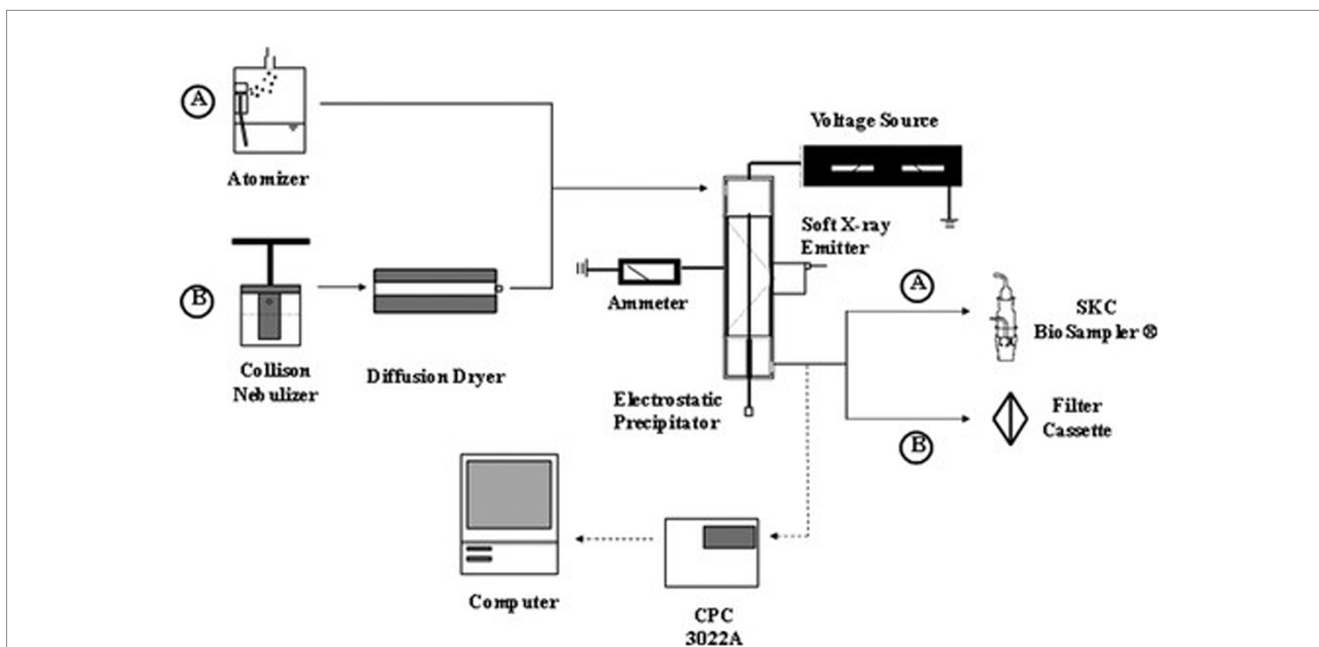


FIGURE 7
Schematic diagram of an experimental setup for the aerosolization and electrostatic precipitation of bacteriophages T3 (path A) and MS2 (path B). Total particle concentration measurements of the ESP effluent air stream are depicted by dashed lines. Adapted with permission from Kettleon et al. (107).

Large-scale wind tunnel or chamber tests are hence needed to examine technologies with controlled settings (flow rate, relative humidity, room volume) that are closer to application conditions in terms of flow behavior.

3.1.3. Large-scale wind tunnel test

Toward implementation, bioaerosol control technologies need to be designed to operate with high flow rates relative to those used in laboratory tests, namely in excess of 10^3 Lmin^{-1} and exceeding

10^5 L min^{-1} for high flow rate HVAC applications. A single-pass wind tunnel test, adapted from ANSI/ASHRAE 52.2 standard, is suitable for both high-flow in-duct technologies as well as recirculating technologies if the inlet and outlet of air can be separated and can operate in this flow rate range. Figures 8A,B depicts a recirculating wind tunnel capable of operating at a flow rate of up to $56,000 \text{ L min}^{-1}$ toward use in the ANSI/ASHRAE 52.2 standard test. Such wind tunnels can be used in bioaerosol testing as well; for example, Farnsworth et al. (210) demonstrated the feasibility of using this wind tunnel for filter tests directly with bacterial aerosols, collecting particles with BioSamplers upstream and downstream of the tested filter, as well as extracting particles for assay from the filter (211, 212).

Wind tunnel tests, however, can also be designed for “intermediate scale testing,” i.e., beyond what is commonplace in flow tube systems, but slightly below flow rates encountered in the application ($\sim 280 \text{ L min}^{-1}$ to $8,000 \text{ L min}^{-1}$) (195, 196, 201). It can be necessary to choose intermediate flow rates because of the large costs and safety concerns associated with bioaerosol testing with specific pathogens. For example, Qiao et al. (126, 176) and Ouyang et al. (213) describe an intermediate-scale wind tunnel (depicted in Figures 8C–E) that can be housed within an isolation facility for biosafety level class 2 testing with both alpha- and beta-coronaviruses. As described in Ouyang et al., the flow profiles in these wind tunnels can be controlled to be uniform, as in filter testing, and they enable the installation of both in-duct and recirculating control technologies. While larger in scale, humidity control is nonetheless also possible, as is simultaneous upstream and downstream bioaerosol collection. We do remark that testing at the intermediate wind tunnel scale is less common than surface or smaller flow tube tests. However, because conditions can better match field conditions, we do argue it is an appropriate part of control technology development prior to field testing.

3.1.4. Chamber tests

A standard chamber has been developed to assess the CADR (Clean Air Delivery Rate) as part of the Association of Home Application Manufacturers/American National Standards Institute (AHAM/ANSI) testing method (214), used largely for the testing of recirculating air cleaners. While the test is not specific to bioaerosols, in it, three different test particles (tobacco smoke, pollen, or dust) are introduced into a sealed room with the size shown in Figure 9, and particle concentration decay is measured with an appropriate real-time particle counter with and without the operation of the air cleaner device. To achieve bioaerosol control, new types of recirculating air cleaners are continuously being developed and tested with a similar procedure to the CADR test, but modified for bioaerosol sampling rather than continuous monitoring of particle concentration (215). In such tests, the chamber is free of bioaerosol at the beginning. Bioaerosol is then delivered to or generated inside the chamber without the operation of the test technology. Right before the technology starts to operate, bioaerosol is sampled from the chamber for the initial concentration assay. The technology starts to operate for a period of time. During operation, bioaerosol continues to be sampled at pre-defined intervals of time for assay. The bioaerosol removal rate is reported as a function of time (216–218) and compared to a control test where the same procedure is repeated without the control technology. Such tests are particularly useful in testing upper-room UV devices (215, 217), as they do not need to be confined in a duct HVAC system in order to operate. The chamber test is straightforward

from the adaptation of the CADR test for technologies built upon recirculating air, which can utilize small chambers up to entire room-scale environments, and serves as a near-field test, particularly for upper-room mounted UV devices. However, there is a loss of control of experimental variables in chamber tests in comparison to flow tube and wind tunnel tests. The actual performance of any control technology will be affected by room dimensions and geometry, ventilation system layout (or whether the chamber is sealed), control technology mounting, and even the arrangement of furniture (219). The chamber test is also affected by chamber design, hence greater inter-laboratory variability is expected than in wind tunnel tests. Also noteworthy, analysis in the chamber test implicitly assumes that the aerosol is “well-mixed,” i.e., its spatial concentration variability minimally affects results. It is challenging to keep generated aerosol well-mixed in such tests.

3.2. Near-field chamber and field testing

Used more sparingly because of the time, safety level, and cost required, near-field chamber testing (84, 85, 220–222) has been set up to examine the effectiveness of control technologies in preventing airborne pathogens transmission between animals. As shown in Figure 10, near-field chamber tests often have three components: a pathogen source chamber, the control technology (connecting two chambers), and the recipient chamber. Chambers are mechanically ventilated and air flows from the source chamber (upstream of the control technology), passing through the control technology, and exiting to the receiving chamber (downstream of the control technology). There are typically two kinds of experiments, those with and those without animals in the chambers. Without animals in the chamber, pathogens are aerosolized with specific aerosolization devices, similar to flow tube tests and bioaerosols are sampled in both chambers with and without the working control technology under various conditions, such as different flow rates. The collected bioaerosol concentrations are then compared to infer the effectiveness of the control technology. To understand control technology influences on bioaerosol-based transmission, animals (termed sentinels) have been placed in the receiving chamber only (84, 220) or in both chambers. When animals are in both chambers, the source chamber will host infectious seeder animals while the receiving chamber will host naive sentinel animals. Bioaerosols are periodically sampled (typically multiple times over multiple days). At the same time, samples representing body fluids (e.g., blood, saliva, nasal secretions, etc.) are collected to monitor whether the animals are infected with the pathogen. As an example of this type of test, Dee et al. (223) conducted a near-field test for filters as the control technology and reported that all pigs in a chamber downstream of a filter were free of pathogen infection while 30% of the pigs in the no-filter chamber were infected during the experimental period, with PRRS virus as the tested pathogen.

To fully evaluate the benefits and performance of technology and to anticipate issues related to durability and how different factors interact together, technologies do need to be tested under field conditions. There are no specific guidelines on how to evaluate the performance of technologies under field conditions. Measurements not only related to the proper functioning of the technology but also

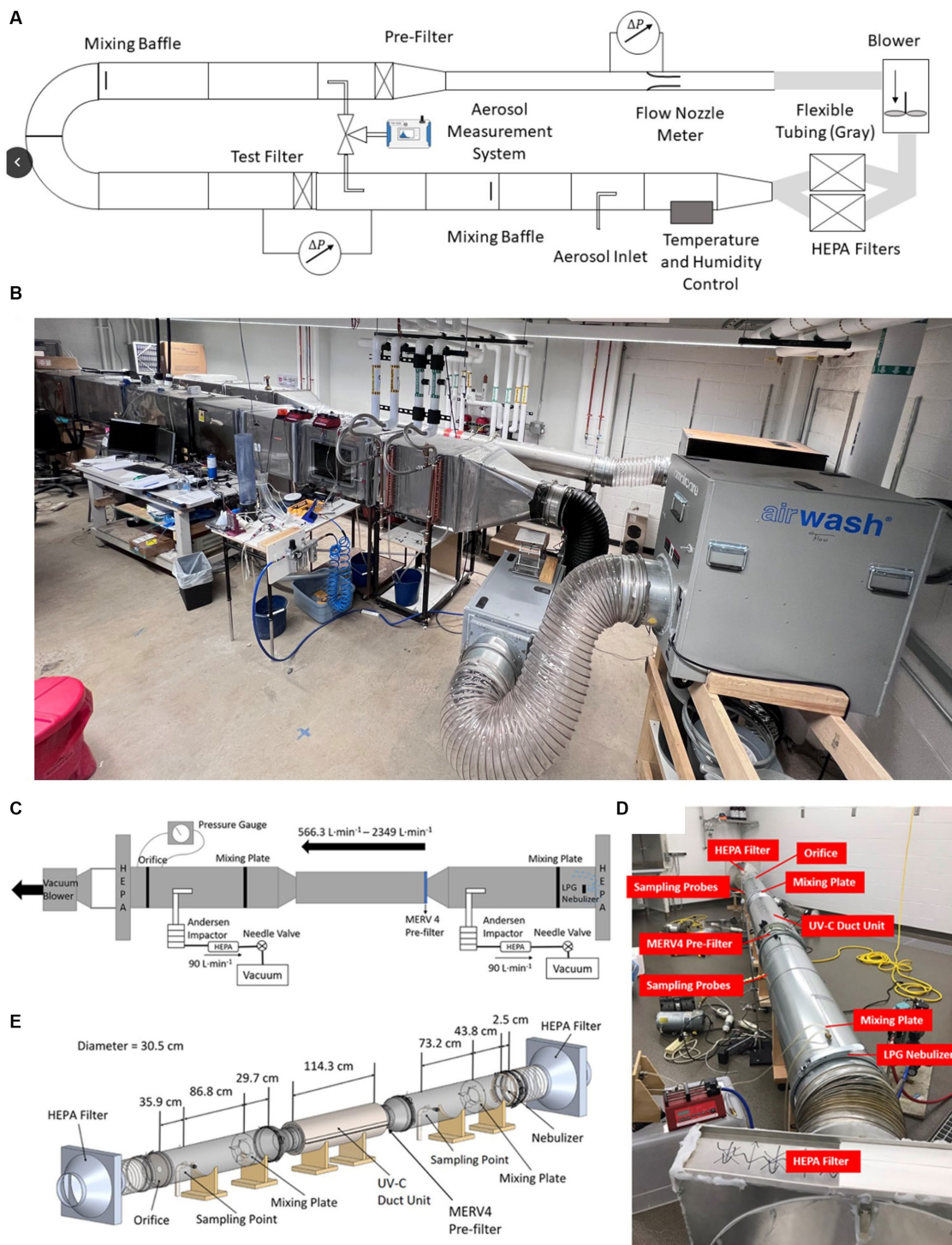


FIGURE 8 Schematic diagram of an intermediate scale wind tunnel (A–C) (126) and regular single-pass wind tunnel (D,E) used for testing of bioaerosol collection efficiency of bioaerosol control technologies. The photographs correspond to systems at the University of Minnesota.

the impact of the technology on decreasing disease incidence and/or prevalence are desirable. Unfortunately, such studies need to be carried out over long periods of time, and hence are difficult to conduct, and costly. Nonetheless, they ultimately provide a more comprehensive assessment of the technology. Such studies allow the identification of

other components of using the technology besides the technology itself that are critical to the long-term benefit of using the technology. For example, as noted in prior sections, the installed efficiency of filters can be much lower than expected for a filter’s MERV rating because of bypass flow, arising from poor sealing and leaks during

installation. The implications of this may only become clear during a field test.

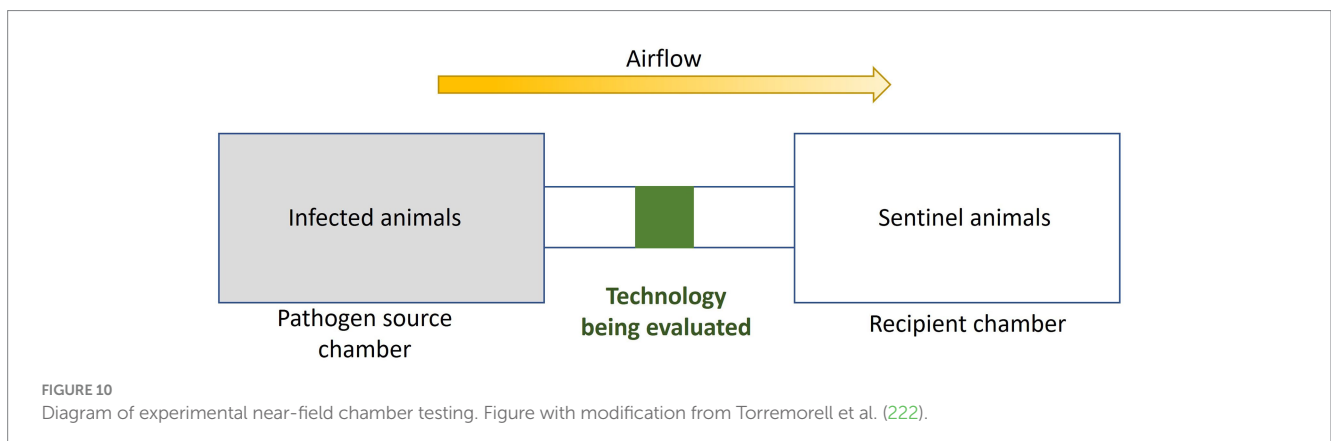
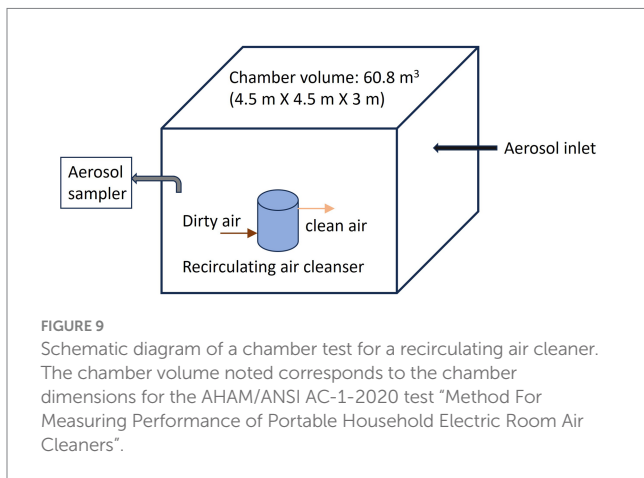
Examples of comprehensive field testing using air filtration in swine farms are the studies by Dee et al. (31, 83, 84). Figure 11 depicts a field test configuration where building 1 is about 120 meters from other downwind buildings and hosts 300 pigs as a source of pathogens. The filtered buildings (mechanical and antimicrobial filtration) were free of infected pigs and conversely, about 50% of the pigs were infected in the nonfiltered building. Further in this study, a production system model was designed with infected animals and buildings retrofitted with and without filtration technologies, allowing direct measurements of disease incidence while controlling for other routes of disease transmission. The results from this particular study represented a turning point for the swine industry, highlighting the importance of field study results. The study showed not only that air filtration could work in swine farms but also spoke of the benefits of disease reduction due to the technology. Field testing analysis has continued with the ongoing monitoring of farms and disease incidence against PRRSV. Such studies have provided additional cost–benefit analysis to assist in the decision-making process (28, 29, 33, 224). Field testing has also been helpful in identifying areas of improvement outside of filtration itself but related to using filtration. For instance, in the case of air filtration in swine farms, field testing was necessary to fully identify new air leakage happening at the farm, the durability of construction components, and the impact of the pressure drop as a

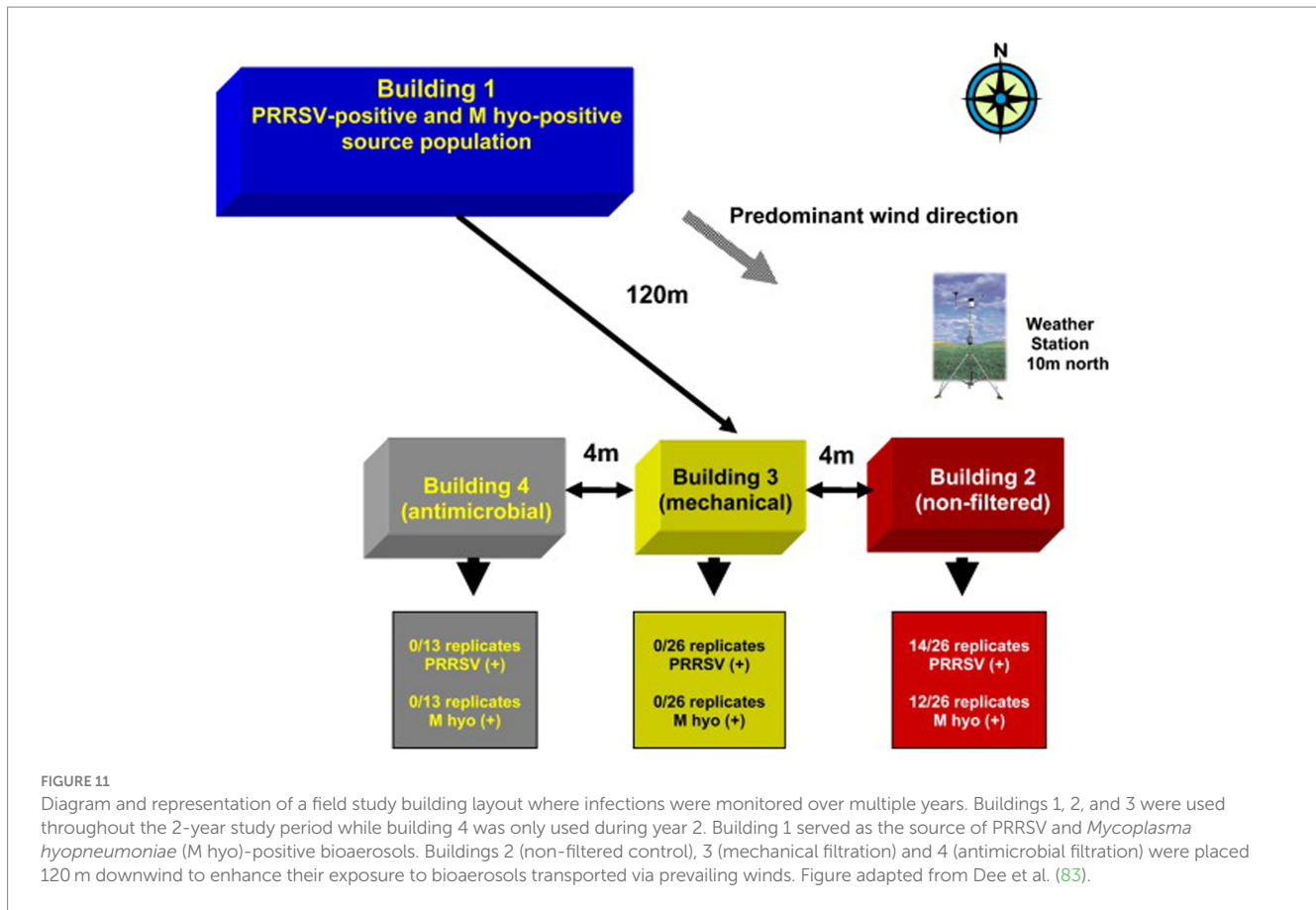
result of using air filters, which in turn required the development of standard operating procedures and protocols to ensure the technology would work long term.

It is also worth mentioning that testing equivalent to field testing has also been performed for bioaerosol control technologies for humans. In these instances, it is largely through bioaerosol and infection concentration sampling (225) or infection rate monitoring (226) in healthcare settings (227), with and without the presence of tested control technologies. Such studies are time-consuming, and because of the limited ability to control study variables, and to eliminate other routes of transmission, it is often difficult to draw clear conclusions with regards to reduction in incidence of infection. Viable pathogen bioaerosol sampling also remains a challenge, as concentrations are typically low, and viability decays over time for collected pathogens.

3.3. Designing test protocols

In all airborne bioaerosol tests, except for small-scale surface tests, biological particles are collected from an aerosol. In laboratory testing, assuming researchers are intentionally producing surrogate pathogen particles, they have control over both the bioaerosol size distribution and the method of collection. Not only are there no standard procedures on control technology performance testing for bioaerosols, but also there are no standardized bioaerosol generation and collection methods, which makes it harder to compare testing results across research groups. Those interested in testing bioaerosol control technologies and in understanding test results need to judiciously evaluate how the results are affected by the bioaerosol generation method and sampling method. In terms of generation, there are two prevailing methods: suspension atomization/nebulization, and fluidization of solid particles. The more commonly employed former method involves utilizing a continuous output, pneumatic nebulizer (often a commercial Collision (228) or BLAM nebulizer) to aerosolize a suspension of the pathogens of interest. Aerosolization has a crucial effect on test results. First, each nebulizer will produce a specific size distribution of particles, affected by the nebulizer geometry and operating conditions (backing pressure) and also the composition of the nebulized media (229). Pathogens are typically incorporated into particles of all sizes generated, and tend to follow the volumetric size distribution unless aerosolization itself





leads to size-specific inactivation. It is common for the media to contain solutes that promote pathogen viability in solution. It is common for such solutes to lead to “foaming” during nebulizer operation, which needs to be addressed by adding anti-foaming agents (230, 231). After generation, it is often necessary to charge neutralize the aerosol to mitigate the influence of high charge levels attainable during spray-based aerosol generation (232). It is also necessary to dry the generated droplets, which can be accomplished with driers in smaller systems, but requires time in larger wind tunnels. The nebulizer operating conditions, anti-foaming agents, and drying, may all influence generated pathogen aerosol viability, though there are limited studies directly examining these effects, given the wide variability of possible choices in aerosolization. Recently, it has been shown that the size distribution of virus-carrying particles (with bovine coronavirus as a model system) can be tuned via nebulization to mimic the size distribution of particles measured from human respiratory activities (213, 233), demonstrating that the bioaerosol size distribution can and should be controlled in testing, and then reported as part of testing.

As a final note in aerosolization, it is typically desirable to nebulize suspensions with as high a titer as possible for each pathogen, in order to improve signal-to-noise in downstream sampling. Practitioners should make efforts to calculate the pathogen aerosolization rate, i.e., the product of the nebulized suspension titer and the nebulizer liquid feed rate. The input pathogen aerosol concentration is the ratio of this feed rate to the operating air flow rate. This concentration then needs to be considered in sampling, where the upper limit in sampled pathogens is the product of the input pathogen concentration, the

sampler flow rate, and the sampling time. Common bioaerosol sampling devices, used to collect bioaerosols from the air, include filters, impactors, impingers, wet-wall cyclones, and electrostatic precipitators. Numerous reviews on bioaerosol sampling, including both active and passive samplers, are available (234–244). We therefore elect not to discuss samplers in greater detail, as collectively such reviews discuss a wide range of sampling approaches. Overall, however, it is important to note that the collection efficiency and biological function recovery efficiency vary among sampling methods (245) due to different sampling mechanisms (206, 246), sampling flow rates, pathogens (245), bioaerosol size distributions, sampling period, and environmental conditions (247) such as relative humidity and temperature. For example, Figure 12 notes the variation in RNA copies as a function of particle size emitted by infected pigs. A sampler with a strongly size-dependent collection efficiency might bring bias or error to the measurements if used in assessing pathogen concentrations from swine. Well-characterized samplers (with known collection efficiencies) used with clearly reported operating conditions are needed when testing control technologies such as to not bias results (for example, through a poor collection of particles larger than 10 micrometers in diameter).

Once collected, follow-up genetic (i.e., RNA/DNA) or viability (the ability of the pathogen to replicate) analysis is performed to evaluate the effectiveness of technologies to remove and/or inactivate pathogens. Here too, multiple reviews have been published on these methods (236, 240, 249). Genetic material tests sampling with and without, or sampling before and after the technology, can reveal the bioaerosol removal efficiency, and viability tests can provide a

combined efficiency resulting from both removal and inactivation. Polymerase chain reaction (PCR) assays (250) are the most common methods to detect a pathogen’s genetic material (DNA/RNA). PCR assays detect the genetic material of both viable and non-viable organisms. Viability tests are often culture-based methods for viruses (126) and bacteria (251). Typically, the effectiveness of technologies to remove airborne particles should also be reported in addition to the removal/inactivation of pathogens to provide better indoor air quality. To identify the removal efficiency of the technology, total particle concentrations are measured with and without or before and after the technology. Several instruments exist to measure the total particle concentration and size distribution, including optical particle counter (OPC), aerodynamic particle sizer (APS), and scanning mobility particle sizer (SMPS). These instruments can provide particle concentration with respect to different size bins independently of their nature, shape, or viability.

Finally, the extent of removal or inactivation can be quantified in different manners. Most directly, the collection efficiency is the fraction of pathogens removed through a single pass in the device, which may be determined as a function of particle diameter in more detailed tests. However, for extremely efficient devices, because the difference between collection efficiencies of 99, 99.9, and 99.99% appears small, yet leads to downstream particle and pathogen concentrations that span two orders of magnitude, the penetration or log reduction is often preferred. The penetration is simply equal to (1- collection efficiency), and hence varies by orders of magnitude as the collection efficiency varies from 99 to 99.99%. The log reduction is the base-ten logarithm of the upstream sampled concentration (C_{up})

to the downstream concentration (C_{down}) in flow tube tests ($\log_{10}(C_{up} / C_{down})$). Therefore, a log reduction 2 system has a penetration that is ten times lower than a log reduction of 1 system. Note in defining these parameters, we specifically refer to single-pass flow-through systems. In chamber tests, the log reduction is commonly reported as well, but will always increase with time and is often a multi-pass log reduction. We therefore suggest that chamber test results also be converted into single-pass log reductions, to better intercompare results. A method to do this is provided by Ouyang et al. (213). As examples of penetration and log reduction, Figure 13A provides plots of particle size-dependent penetration adapted from Qiao et al. (126), in flow tube tests of a filter-based control technology. Figure 13B displays plots of the log-reduction on physical removal, qRT-PCR, and on virus titration from Ouyang et al. (213), for three different control technologies, all measured in the same flow tube system. Evident from the comparison of log-reduction, technology 1 uniquely has a low log reduction based on physical removal measurements yet a high log-reduction based on virus titer, suggesting it primarily acts through virus inactivation rather than virus removal. Unsurprisingly, this technology is a UV-C based system, while technologies 2 and 3 are filtration systems.

4. Implementation and cost analysis

It is established that mechanical ventilation systems can be implemented in farms along with control technologies that can prevent farm-to-farm aerosol pathogen transmission and improve

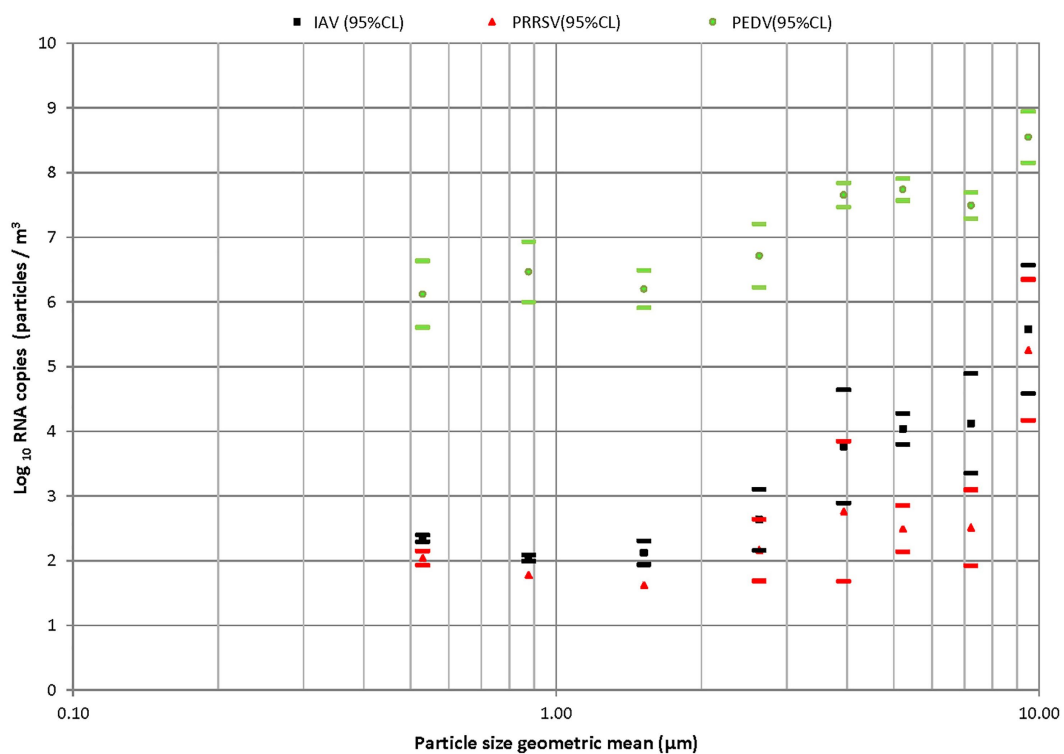


FIGURE 12 Particle size distributions measured by size-dependent collection device (cascade impaction) and RT-qPCR for influenza, porcine reproductive, and respiratory syndrome and porcine epidemic diarrhea viruses for aerosols generated by infected pigs. Figure adapted from Alonso et al. (248).

animal health (86). For example, the installation of filters into mechanically ventilated farms has successfully prevented PRRSV transmission from infected farms. (19, 28, 252) Nonetheless, filters may be costly due to the number of filters needed, which can be quite large depending on farm size (i.e., from a few hundred to thousands), and typically filters are replaced every three to seven years, and prefilters are replaced every 6–12 months. High-efficiency filters perform better in collecting bioaerosols yet the cost of a single filter increases as performance increases (shown in Figure 14). Filters will also cause an extra pressure drop in the airstream, thus requiring extra power and adding to the energy cost. It is therefore important to consider costs in implementing effective aerosol biosecurity measurements.

Here, we focus on the cost–benefit of implementing control technologies into mechanical ventilation systems, assuming mechanical ventilation systems have already been installed. Specifically, we discuss how to lower the cost of control technologies either by improving or replacing current filtration systems with other technologies to lower the risk due to farm-to-farm aerosol pathogen transmission. The costs and benefits of implementing HVAC filters on residential or commercial buildings to improve indoor air quality have been modeled by several studies (70, 253–258). For human-occupied residential or commercial buildings, the benefit is typically characterized by the modeled reduced risk, and the risk of airborne transmission pathogen is often quantified by a simplified steady-state Wells-Riley (W-R) model (259) where the risk is associated with pathogen generation rates (carrier emitting pathogens), exposure time, the volume of supplied clean air (260), and the infection probability (how many pathogens are needed to successfully infect a person). Here, we mainly focus on preventing airborne transmission of external pathogens from neighboring farms. Thus, we quantify benefits by looking at the total number of collected infectious particles ($N_{i,c}$) and the associated cost with it (C).

To implement bioaerosol control technologies on a farm, several parts of the cost need to be considered, including a fixed cost to purchase and install the technology, a maintenance cost, and an operation cost. The cost of implementing a technology can vary substantially across facilities and factors such as building design, type

of construction (new or retrofit), climate, environment (moisture, temperature) and presence of gases (ammonium, hydrogen sulfur, etc.) may affect the longevity and functionality of a particular technology. The operation cost is mainly due to personnel overseeing the ongoing proper functioning of the technology and energy consumption, assuming the system is continuously in operation as turning off or removing the technologies will not protect the animals, which can be represented by the total energy consumed since the final operating cost is equal to the price per energy (dollar/Joule) times the total energy consumed for a given period. While different bioaerosol control technologies have different effectiveness in removing or inactivating bioaerosols, not only the collection efficiency is size dependent, but the quantity of pathogens is also size dependent (261).

Rigorously, for a control technology (e.g., a filter) with a known collection efficiency as a function of particle diameter $CE(d_p)$,

exposed over time to aerosol with a size distribution $\frac{dn}{dd_p}$ (the number

concentration of particles per unit change in diameter) and an effective number of pathogens per particle $\omega(d_p)$ (certainly the greatest unknown in calculations), the number of collected or removed pathogens ($N_{i,c}$) can be calculated as:

$$N_{i,c} = N_u \bar{Q}_f t \int_0^\infty \frac{dn}{dd_p} CE(d_p) \omega(d_p) dd_p \quad (2)$$

where N_u is the number of bioaerosol control technology units, \bar{Q}_f is the incoming flow rate per unit, t is the operation time, and d_p is the bioaerosol particle diameter. In equation (2), for simplicity, we have assumed that the collection efficiency does not change with time, though as noted in prior sections loading typically leads to collection efficiency changes. The total cost of controlling the incoming airborne pathogens (collection and/or disinfection) is the sum of installation cost, technology purchase cost, energy cost during operation for driving the flow, energy cost during operation for the technology, and maintenance and operating labor cost.

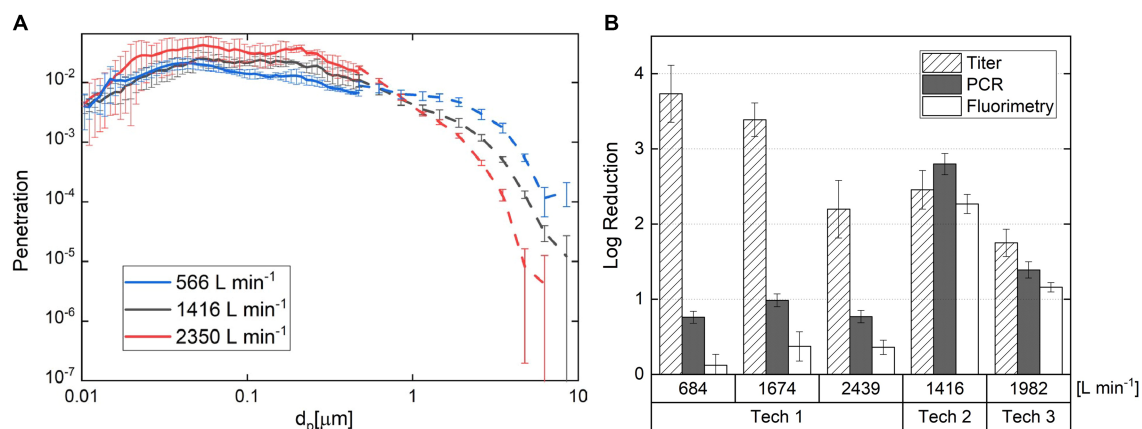


FIGURE 13 (A) Particle penetration versus particle diameter for an air purification unit, measured in a wind tunnel. (B) Log reductions of porcine respiratory coronavirus (PRCV) concentration by each of the three tested technologies by virus titration, RT-qPCR, and fluorimetry. Results adapted from Qiao et al. and Ouyang et al. (126, 213).

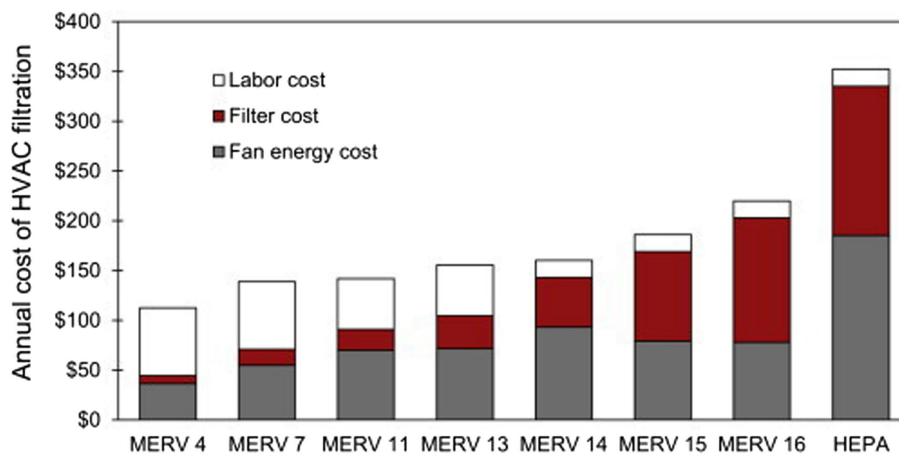


FIGURE 14 Estimated annual costs of filtration (a single filter) in 2013 in a hypothetical office environment. Adapted from Azimi et al. (70).

$$C = N_u N_t c_{inst} + N_u N_t c_{tech} + c_{elec} N_u \frac{\Delta P \bar{Q}_f t}{\eta_{pump}} + N_u I V c_{elec} + N_u c_{mmtc} \tag{3}$$

where N_t is the number of turn-overs (replacement instances), c_{inst} is the one-time installation cost, c_{tech} is the purchase cost, c_{elec} is the electricity cost for providing the flow through the technology, ΔP is the pressure drop across the technology (again, assumed constant for simplicity), I is the electric current and V is the electric voltage for operating the technology (also assumed constant), and c_{mmtc} is the maintenance labor cost.

Thus the benefit of implementing a control technology is quantified by the ratio of collected/disinfected pathogens over the total cost:

$$\frac{N_{i,c}}{C} = \frac{\int_0^\infty \frac{dn}{dd_p} CE(d_p) \omega(d_p) dd_p}{\frac{(c_{inst} + c_{tech}) N_t}{\bar{Q}_f t} + c_{elec} \frac{\Delta P}{\eta_{pump}} + \frac{IVc_{elec}}{\bar{Q}_f} + \frac{c_{mmtc}}{\bar{Q}_f}} \tag{4}$$

For simplicity, the integral can be equated with $CE n_{path}$, where CE is the average collection or removal efficiency over the incoming aerosol size distribution, and n_{path} is the incoming pathogen concentration in the aerosol (on average). This leads to:

$$\frac{N_{i,c}}{C} = \frac{CE n_{path}}{\frac{(c_{inst} + c_{tech}) N_t}{\bar{Q}_f t} + c_{elec} \frac{\Delta P}{\eta_{pump}} + \frac{IVc_{elec}}{\bar{Q}_f} + \frac{c_{mmtc}}{\bar{Q}_f}} \tag{5}$$

From Equation 5, even without knowledge of the pathogen concentration, individual terms affecting the cost of implementation become clear. To maximize the benefit, technology should be highly efficient on collecting/disinfecting incoming pathogens with a low total cost (summation of the four costs at the denominator). The first installation and technology cost $\left(\frac{(c_{inst} + c_{tech}) N_t}{\bar{Q}_f t} \right)$ could be very

high if the replacement is frequent, as in filters; the flow power energy cost $\left(c_{elec} \frac{\Delta P}{\eta_{pump}} \right)$ will also decrease if the pressure drop decreases,

as in UV sources and ESP; the cost to run the electric components of the technology $\frac{IVc_{elec}}{\bar{Q}_f}$ could be low for an ESP with low current, but

may be higher for UV lights (UV power increases with increasing disinfection efficiency), and is zero for regular filters. Technologies that require minimal operator monitoring and maintenance also become more cost-effective. However, the overwhelming factors reducing cost are to (1) use the highest collection efficiency devices possible and (2) the highest ventilation rates (\bar{Q}_f) possible, as these

two terms appear in all except the “ $c_{elec} \frac{\Delta P}{\eta_{pump}}$ ” term. Stated

differently, cost analysis suggests it is beneficial to operate with higher ventilation rates if pathogens are effectively removed, and the pressure drop is not too high.

4.1. Cost–benefit analysis considerations from the field

A main outcome of implementing a bioaerosol control technology in animal production is to prevent pathogen entry into a population and cause disease or to reduce disease incidence altogether. Thus, in addition to the aforementioned cost of the technology related to purchase, installation, maintenance and operating costs, the economic feasibility of implementing the technology needs to be measured against the reduction in losses associated with the disease outbreak. Economic losses from a disease outbreak can be measured taking into consideration the performance parameters relevant to the animal species and the production system impacted. Examples of performance parameters to consider include: mortality rates, average daily rate of gain, feed conversion, and number of animals or products sold. Other parameters

may include antimicrobial use, additional labor to treat and care for the well-being of the animals, and veterinary cost. The cost/benefit analysis is then measured against the historical frequency of the disease outbreaks caused by the pathogen of interest, the expected severity of the losses and additional costs and compared with the expected parameters should the disease be prevented. An example is the case of PRRS virus infections (224), where the payback period of implementing filtration in swine farms was estimated at 5.35 years or 7.13 years depending on the type of filtration systems used and based on breeding herd productivity but the payback was reduced to 2.1 and 2.8 years based on whether the pigs sold had a premium of \$5 to be PRRSV negative. The cost–benefit analysis is often further complicated by the fact that pathogens often spread through multiple routes, making it difficult to know whether the benefit of the technology could not be materialized due to the technology itself or the fact introduction of disease through non-airborne routes could not be prevented.

In summary, the cost–benefit analysis of a given technology should include not only laboratory and modeled data but also a measure of the impact of the technology in preventing disease and productivity losses to production systems.

5. Summary

Despite the many biosecurity intervention protocols in farms directed at minimizing the risk of pathogen dissemination through people, animals, and fomites, airborne pathogens can still be transported into, within, and from farms due to the limited use of technologies to mitigate aerosol-based disease transmission. In this report, we review bioaerosol control technologies that can be implemented to prevent airborne transmission of pathogens between farms. Mechanical ventilation systems with filters can remove pathogens from the incoming air and thus have been effective at preventing the spread of the PRRS virus. However, filters are costly and other technologies are hence proposed to either replace or assist filters (prolong the filter's life) to lower the cost and reach the same or even better decontamination efficiency. At present, ESPs and UVC devices are the most promising technologies for farm applications which can possibly replace and/or complement filters and work with mechanical ventilation systems alone, as they have been proven to work in industrial environments and in healthcare environments, respectively. However, those technologies have not been adapted yet to farms and it is unknown whether their cost/benefit analysis would support their implementation. Other emerging technologies may find application in farms in the future but require additional testing at multiple levels first.

For this reason, this report also focuses on reviewing testing setups for bioaerosol control technologies, in an effort to harmonize testing approaches, as emerging control technologies are often tested inconsistently from one another. We find many testing parameters need to be considered when comparing technology effectiveness, such as the method of bioaerosol generation, collection for analysis, experimental setup (flow tube versus chamber), and in-file performance among others, as they all affect the resulting data.

Lastly, preventing disease introduction into farms is a priority to ensure food security. We note the choice of technologies depends on the specific configuration of the farm, the cost–benefit analysis, and the ability of technologies to operate under specific environmental

conditions determined by the farm type. Airborne control technologies should be used in conjunction with all other on-farm biosecurity protocols to comprehensively limit the introduction and spread of pathogens through all possible routes. Furthermore, airborne control technologies should be considered as part of the package of biosecurity measures employed by farmers to prevent disease introduction, in particular for animals housed in confinement in mechanically ventilated buildings with high densities and for farms located in regions where proximity to other farms is a risk. Lastly, to ensure food security, preventing disease introduction into and spread from farms should be a priority thus, further research in aerosol technologies is warranted.

Author contributions

HO: Investigation, Methodology, Visualization, Writing – original draft, Writing – review & editing. LW: Investigation, Writing – review & editing. DS: Visualization, Writing – review & editing. MY: Writing – review & editing, Investigation. JM: Visualization, Writing – review & editing, Investigation. LL: Writing – review & editing. BO: Writing – review & editing, Conceptualization, Investigation. MS: Writing – review & editing, Writing – original draft. CH: Conceptualization, Methodology, Writing – original draft, Writing – review & editing, Funding acquisition, Investigation, Project administration. MT: Conceptualization, Methodology, Writing – original draft, Writing – review & editing, Funding acquisition, Investigation, Project administration.

Funding

The author(s) declare financial support was received for the research, authorship, and/or publication of this article. This review was funded by the Swine Health Information Center (grant # 21-109).

Acknowledgments

The authors acknowledge the travel support from University of Texas at Dallas ASPIRE² mentorship program.

Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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