



OPEN ACCESS

EDITED BY

Calvin A. Henard,
University of North Texas, United States

REVIEWED BY

Doug Bartlett,
University of California, San Diego,
United States
Thomas Smith,
Sheffield Hallam University, United Kingdom

*CORRESPONDENCE

Nora Richter
✉ nora.richter@eawag.ch

PRESENT ADDRESS

Nora Richter,
Surface Water Research and Management,
Dübendorf, Switzerland

RECEIVED 22 November 2024

ACCEPTED 21 January 2025

PUBLISHED 07 February 2025

CITATION

Richter N, Villanueva L, Hopmans EC, Bale NJ,
Sinninghe Damsté JS and Rush D (2025)
Methanotroph-methylotroph lipid
adaptations to changing environmental
conditions.
Front. Microbiol. 16:1532719.
doi: 10.3389/fmicb.2025.1532719

COPYRIGHT

© 2025 Richter, Villanueva, Hopmans, Bale,
Sinninghe Damsté and Rush. This is an
open-access article distributed under the
terms of the [Creative Commons Attribution
License \(CC BY\)](https://creativecommons.org/licenses/by/4.0/). The use, distribution or
reproduction in other forums is permitted,
provided the original author(s) and the
copyright owner(s) are credited and that the
original publication in this journal is cited, in
accordance with accepted academic
practice. No use, distribution or reproduction
is permitted which does not comply with
these terms.

Methanotroph-methylotroph lipid adaptations to changing environmental conditions

Nora Richter^{1*†}, Laura Villanueva^{1,2}, Ellen C. Hopmans¹,
Nicole J. Bale¹, Jaap S. Sinninghe Damsté^{1,3} and Darci Rush¹

¹Department of Marine Microbiology and Biogeochemistry, NIOZ Royal Netherlands Institute for Sea Research, Den Burg, Netherlands, ²Department of Biology, Faculty of Science, Utrecht University, Utrecht, Netherlands, ³Department of Earth Sciences, Faculty of Geosciences, Utrecht University, Utrecht, Netherlands

Methanotrophs, in particular methane-oxidizing bacteria (MOB), regulate the release of methane from lakes, and often co-occur with methylotrophs that may enhance methane-oxidation rates. Assessing the interaction and physiological status of these two microbial groups is essential for determining the microbial methane buffering capacity of environmental systems. Microbial membrane lipids are commonly used as taxonomic markers of specific microbial groups; however, few studies have characterized the changes of membrane lipids under different environmental conditions. For the case of methane-cycling microorganisms, this could be useful for determining their physiological status and potential methane buffering capacity. Here we investigated the changes in membrane lipids, bacteriohopanepolyols (BHPs) and respiratory quinones, produced by MOB and methylotrophs in an enrichment co-culture that primarily consists of a methanotroph (*Methylobacter* sp.) and a methylotroph (*Methylotenera* sp.) enriched from a freshwater lake under different methane concentrations, temperatures, and salinities. To assess whether the lipid response is similar in methanotrophs adapted to extreme environmental conditions, we also characterize the BHP composition and respiratory quinones of a psychrotolerant methanotroph, *Methylovulum psychrotolerans*, isolated from an Arctic freshwater lake and grown under different temperatures. Notably, in the *Methylobacter-Methylotenera* enrichment the relative abundance of the BHPs aminobacteriohopanepentol and aminobacteriohopanepolyols with additional modifications to the side chain increased at higher temperatures and salinities, respectively, whereas there was no change in the distribution of respiratory quinones. In contrast, in the *Methylovulum psychrotolerans* culture, the relative abundance of unsaturated BHPs increased and ubiquinone 8:8 (UQ_{8:8}) decreased at lower temperatures. The distinct changes in lipid composition between the *Methylobacter-Methylotenera* enrichment and the psychrotolerant methanotroph at different growth temperatures and the ability of the *Methylobacter-Methylotenera* enrichment to grow at high salinities with a singular BHP distribution, suggests that methane-cycling microbes have unique lipid responses that enable them to grow even under high environmental stress.

KEYWORDS

bacteriohopanepolyols, methane-oxidizing bacteria, respiratory quinones, membrane lipids, methylotroph

1 Introduction

Methane-oxidizing bacteria (MOB) modulate the natural release of methane (CH₄), a potent greenhouse gas, through aerobic methane oxidation (see [Hanson and Hanson, 1996](#) for a review). MOB belong to the phylum Proteobacteria in the classes Gammaproteobacteria (Type I and Type X methanotrophs) and Alphaproteobacteria (Type II methanotrophs; [Hanson and Hanson, 1996](#); [Bowman, 2006](#)), and were also identified in the phyla Verrucomicrobia ([Dunfield et al., 2007](#); [Pol et al., 2007](#); [Islam et al., 2008](#)) and NC10 ([Raghoebarsing et al., 2006](#); [Ettwig et al., 2009](#)). Type I members of Gammaproteobacteria are widespread in both terrestrial and marine environments ([Knief, 2015](#)). In lakes, for instance, Type I methanotrophs are the primary methane-oxidizers in both the water column and surface sediments ([Hanson and Hanson, 1996](#)). Members of the methylotrophic (i.e., microorganisms that consume single carbon compounds) *Methylotenera* genus are known to co-occur with methanotrophs, and under nitrate-rich conditions are thought to play a role in enhancing methane oxidation rates by the removal of toxic products (e.g., methanol and formaldehyde) that inhibit methanotrophy ([Mustakhimov et al., 2013](#); [Krause et al., 2017](#); [Yu and Chistoserdova, 2017](#); [van Grinsven et al., 2020](#)). The capacity for methanotrophs to regulate methane emissions from lakes is, thus, linked to microbial interactions with methylotrophs, as well as the physiological ability of both methanotrophs and methylotrophs to cope with environmental stress. In microbes, the physiological response to external stress is regulated by membrane lipids, such as bacteriohopanepolyols (BHPs; see [Belin et al., 2018](#) and [Newman et al., 2016](#) for a review) and respiratory quinones (see [Franza and Gaudu, 2022](#) for a review); therefore, membrane lipids are crucial for understanding the functional potential of methane-cycling microbes.

In gram-negative bacteria, BHPs are found in the inner and outer membrane (e.g., [Jahnke et al., 1992](#); [Jürgens et al., 1992](#); [Doughty et al., 2009](#); [Wu et al., 2015](#)), and play an important physiological role in regulating the permeability and rigidity of the cell membrane in response to external environmental stress ([Welander et al., 2009](#); [Doughty et al., 2011](#); [Schmerk et al., 2011](#)). BHPs are structurally diverse compounds with unique side-chain modifications that are thought to be specific to certain microbes ([Rohmer et al., 1984](#); [Talbot et al., 2003](#); [Talbot and Farrimond, 2007](#); [Kusch and Rush, 2022](#)). For instance, 35-aminobacteriohopane-30,31,32,33,34-pentol (aminopentol from herein) and 35-aminobacteriohopane-31,32,33,34-tetrol (aminotetrol from herein) are considered specific to Type I and Type II methanotrophs, respectively ([Neunlist and Rohmer, 1985](#); [Cvejic et al., 2000](#); [Talbot et al., 2001](#)). Incubation experiments with methanotrophs further suggest that concentrations and relative abundances of aminotetrol, aminopentol, and their unsaturated counterparts vary in relation to temperature ([Jahnke et al., 1999](#); [Osborne, 2015](#); [Osborne et al., 2017](#); [Bale et al., 2019](#); [van Winden et al., 2020](#)), and, therefore, might be involved in maintaining the fluidity of the cell membrane. Further work is needed, however, to determine whether the BHP response in methanotrophs is the same across different species and various environmental factors.

Respiratory quinones are isoprenoidal-based membrane lipids associated with metabolic processes in eukaryotes, bacteria, and archaea, as they are essential components of electron transport chains involved in electron and proton shuttling within the cytoplasmic membrane ([Anraku, 1988](#)). Respiratory quinones are characterized by

a polar cyclic headgroup and isoprenoid side chain that imparts a redox potential and can be adapted to certain metabolic processes ([Nowicka and Kruk, 2010](#)). As such, respiratory quinones have been used as metabolic markers for redox processes (e.g., [Dupont et al., 2014](#); [Becker et al., 2018](#)), quantitative measures of bacterial biomass ([Hiraishi et al., 1998](#); [Saitou et al., 1999](#)), and as chemotaxonomic biomarkers ([Collins and Green, 1985](#); [Hiraishi, 1999](#)). Further studies suggest that respiratory quinones might also be involved in regulating membrane fluidity under osmotic stress ([Sévin and Sauer, 2014](#); [Eriksson et al., 2019](#)), oxidative stress ([Søballe and Poole, 1999](#)), and at low temperatures ([Seel et al., 2018](#)). So far, the role of respiratory quinones in stress resistance for methane-cycling microbes has not been evaluated.

Continuous advancements in analytical techniques and their application to environmental samples and cultures has led to the discovery of many new membrane lipids (e.g., [Talbot et al., 2016](#); [Hopmans et al., 2021](#)). For instance, the analysis of underivatized BHPs using ultra high pressure liquid chromatography (UHPLC) coupled to electrospray ionization (ESI)-high resolution dual-stage mass spectrometry (HRMS²) led to the identification of many novel BHPs ([Hopmans et al., 2021](#)). Similarly, sample analyses for respiratory quinones using an UHPLC system equipped with ESI revealed a wide array of respiratory quinones in environmental samples (e.g., [Becker et al., 2018](#)). The application of these analytical techniques to study membrane lipids, has the potential to uncover novel lipids that could provide new insights into environmental stress resistance.

In this study, we use UHPLC-HRMS² to characterize how BHP and respiratory quinone distributions in an enrichment culture, consisting of a dominant methanotroph (*Methylobacter* sp.) and a methylotroph (*Methylotenera* sp.) that metabolically interact and were obtained from a eutrophic lake ([van Grinsven et al., 2020](#)), vary in response to changing environmental conditions. We use an enrichment co-culture to assess how the microbial community responds to external environmental stress and provide context for how this might influence microbial interactions in a lake environment. Further, we evaluate BHPs for their biomarker potential in methanotrophs, and we assess changes in respiratory quinones to understand how the redox status of the cells varies in response to external stress. To determine whether methanotrophs adapted to extreme environments have a similar physiological response to changing temperatures, we also analyzed the lipid composition of a psychrotolerant methanotroph, *Methylovulum psychrotolerans*, isolated from an Arctic freshwater lake ([Oshkin et al., 2016](#); [Bale et al., 2019](#)).

2 Methods

2.1 *Methylobacter-Methylotenera* enrichment culture

The enrichment co-culture was previously isolated from a hypereutrophic, monomictic lake (Lacamas Lake, WA, United States; [van Grinsven et al., 2020](#)). Prior to setting up the incubation experiments the enrichment was grown at 15°C in oxic and dark conditions with nitrate mineral salts (NMS) media ([Whittenbury et al., 1970](#)). Incubation experiments were set up to observe how methane concentration, temperature, and salinity affect the BHP lipidome of the enrichment co-culture (see [Supplementary Table S1](#) for experimental conditions). All experiments were set up in triplicate in 580 mL

acid-washed and autoclaved glass bottles with butyl rubber stoppers and total volume of 250 mL of NMS medium. Each incubation was inoculated with the same amount of concentrated enrichment culture, closed and crimp sealed. For abiotic controls (heat killed), the inoculated enrichment co-culture was autoclaved. All bottles (except for the unamended experiments) were supplemented with methane (CH₄, 99.99% pure) corresponding to a percentage of the headspace volume (0.5, 5, 10%). Time zero samples were also taken at the start of each experiment and filtered onto muffled glass fiber filters (GF/F 47 mm diameter with 0.3 μm pore size; Whatman) and frozen at −80°C until analysis. The bottles were then shaken for 10 s to establish an equilibrium between the gas and water phase. The resulting CH₄ concentration in the headspace was measured by piercing the butyl septum with a syringe to retrieve a gas sample that was analyzed with gas chromatography-flame ionization detection (GC-FID; Thermo Scientific Focus GC). The bottles were incubated under oxic conditions in the dark at 15°C unless other temperatures are specified. For the salinity experiments, the initial enrichment co-culture was gradually adapted to higher sodium chloride (NaCl) additions to NMS media over a period of 6 months. The resulting experiments were conducted as previously described with unamended controls (CH₄ 0%). Methane concentrations in the headspace were regularly monitored throughout the experiment. The incubations were ended when the CH₄ concentration in the headspace was <10% of the initial CH₄ concentration, and incubations reached a stationary phase based on the reduced rate of decrease in CH₄ concentrations. The incubations lasted from 6 days to 48 days depending on the growth conditions. The methane oxidation rates for each individual treatment were calculated from fitted slopes spanning the linear phase of methane removal after the initial lag phase. Experiments were ended by filtering the cultures onto muffled glass fiber filters (GF/F 47 mm diameter with 0.3 μm pore size; Whatman) and were immediately frozen at −80°C until lipid and/or DNA extractions.

2.2 *Methylotenera mobilis*

Biomass of *Methylotenera mobilis* (DSM 17540) from the Deutsche Sammlung von Mikroorganismen und Zellkulturen (DSMZ) was previously analyzed for intact polar lipids (Richter et al., 2023). Existing UHPLC-HRMS data was used in this study for the identification of quinones.

2.3 *Methylovulum psychrotolerans*

Methylovulum psychrotolerans (Sph56, NCBI Accession number MH701868) was grown in NMS media at different temperatures (4, 10, and 20°C) as described by Bale et al. (2019). Existing UHPLC-HRMS data for intact polar lipids was used for the identification of BHPs and quinones in this study.

2.4 Lipid extractions

All filters from the *Methylobacter-Methylotenera* enrichments were divided in half, freeze-dried, and extracted using a modified Bligh-Dyer method (Bligh and Dyer, 1959; Bale et al., 2021). The filters were ultrasonically extracted twice using methanol (MeOH),

dichloromethane (DCM), and phosphate buffer (2:1:0.8, v:v:v). DCM and phosphate buffer was added to the resulting solvent in a separate flask for a new volume ratio of 1:1:0.9 (v:v:v). The DCM layer was collected and the remaining aqueous layer was washed twice using DCM. The filters were ultrasonically extracted two more times using MeOH:DCM:aqueous trichloroacetic acid (TCA) solution (2:1:0.8, v:v:v). The same procedure as described above was used to collect the DCM layers. The combined DCM layers were then dried using N₂ gas and stored at −20°C. Before analysis, an internal standard (deuterated diacylglyceryltrimethylhomoserine; DGTS-d9; Avanti® Polar Lipids, United States) was added to the total lipid extracts (TLEs). The samples were dissolved in MeOH:DCM (9:1, v:v) and filtered through 0.45 μm regenerated cellulose syringe filter (4 mm diameter, Grace Alltech, Deerfield, IL).

2.5 Lipid analysis

TLEs for the *Methylobacter-Methylotenera* incubation experiments were analyzed after Hopmans et al. (2021) to identify BHPs. All samples were analyzed on an Agilent 1290 Infinity I UHPLC coupled to a quadrupole-orbitrap (Q-Exactive) HRMS equipped with an Ion Max source and heated electrospray ionization (HESI) probe (ThermoFisher Scientific, Waltham, MA). An Acquity C18 BEH column (2.1 × 150 mm, 1.7 μm particle; Waters) and pre-column was used for separation with a solvent system of (A) MeOH:H₂O (85:15) and (B) MeOH:isopropanol (1:1) containing 0.12% (v/v) formic acid and 0.04% (v/v) aqueous ammonia in both solvents. A positive ion monitoring mode of *m/z* 350–2,000 (resolution 70,000 ppm at *m/z* 200) and an inclusion list of 357 calculated exact masses of BHPs was used for lipid detection. We used a data dependent MS² with an isolation window 1 *m/z*; resolution 17,500 ppm at *m/z* 200 of the 10 most abundant ions for a total cycle of ca. 1.2 s and dynamic exclusion (6 s) with a 3 ppm mass tolerance to identify BHPs. A stepped normalized collision energy of 22.5 and 40 was used for optimal fragmentation of BHPs. Every 48 h a mass calibration using a Thermo Scientific Pierce LTQ Velos ESI Positive Ion Calibration Solution was performed. BHPs for *Methylovulum psychrotolerans* (Sph56) and quinones for all samples were analyzed using the same method as described above, but with a stepped normalized collision energy of 15, 22.5, and 30 (Bale et al., 2019).

All peak areas were corrected for matrix effects and variability in MS performance with an internal standard, DGTS-d9. Authentic standards to determine absolute BHP and quinone concentrations currently do not exist, therefore all lipids are reported using their relative peak area as response units (RU). Lipids from the co-culture enrichments and the *Methylotenera mobilis* biomass sample are additionally normalized to liters of media filtered and grams of freeze-dried biomass, respectively. Therefore, all reported lipid concentrations are expressed as response units per liter (RU/L) or response units per gram (RU/g).

2.6 DNA extraction and 16S rRNA gene amplicon sequencing

A subset of samples was extracted for DNA to confirm that *Methylobacter* sp. and *Methylotenera* sp. were the primary bacteria present in the enrichment culture and that there were no major changes in its composition under different incubation conditions. Time 0 samples

were taken at the start of the different experimental set ups and extracted for DNA. Triplicate incubations that were grown under standard conditions (CH₄ 5%, 15°C, 0 g/L NaCl in NMS media and in the dark), as well as triplicate incubations with large changes in lipid composition (i.e., temperature 30°C and 10 g/L NaCl addition) were also extracted.

A quarter of the filters were extracted for DNA using the DNeasy PowerSoil Pro Kit (Qiagen). A negative extraction blank and mock culture sample (ZymoBIOMICS® Gut Microbiome Standard, Zymo Research Copr.) were included in the extractions as negative and positive controls, respectively. The universal (bacterial and archaeal) primer pairs, 515F and 806RB, were used to target the V4 region of the small subunit ribosomal RNA region (Caporaso et al., 2011; Apprill et al., 2015; Parada et al., 2016). PCR reactions were performed in Phusion buffer (Qiagen) with dNTPs and Invitrogen™ Platinum™ SuperFi™ Polymerase (Thermo) with the following conditions: 98°C for 30 s, 98°C for 10 s, 30 cycles of 98°C for 10 s, 50°C for 15 s, and 72°C for 30 s, followed by 72°C for 5 s, and 4°C for 5 s. The resulting PCR products were pooled in equimolar amounts and loaded on a 1% agarose gel. The target bands were cut out and purified using QIAquick® PCR gel extraction kit (Qiagen). Samples were sent to the University of Utrecht Sequencing Facility (USEQ, the Netherlands) for Truseq DNA nano library preparation and sequencing on an Illumina NextSeq2000 (Illumina, San Diego, CA) 2 × 300 bp sequencing platform. All sequences are available at the sequence read archive under the BioProject PRJNA1149959.

16S rRNA gene amplicon sequences were analyzed using the Cascabel pipeline (Abdala Asbun et al., 2020). This included quality assessment using FastQC (Andrews, 2010), paired-end reads assembly with PEAR (Zhang et al., 2014), library demultiplexing using QIIME (Caporaso et al., 2010), and picking and taxonomy assignment of amplicon sequence variants (ASVs) using DADA2 (Callahan et al., 2016). Taxonomy was assigned using Silva 138.1 as a reference database (Quast et al., 2013; Yilmaz et al., 2014).

2.7 Statistical analyses

We applied a Hellinger transformation and scaling to the BHP dataset prior to performing a principal component analysis (PCA) to determine what drives the largest variations in BHPs under different treatments for the *Methylobacter-Methylotenera* co-culture experiments. All analyses were performed in R (version 4.3.2; R Core Team, 2023) using the vegan package (version 2.5–7; Oksanen et al., 2022), ggplot2 (version 3.4.0; Wickham et al., 2023), FactoMineR (version 2.7; Lê et al., 2008), and factoextra (version 1.0.7; Kassambara and Mundt, 2020). Methane oxidation rates in the experimental conditions were statistically compared to those of the unamended and heat killed incubations using a one-way ANOVA test and Tukeys Honest Significant Differences (HSD).

3 Results

3.1 Methane-oxidation rates in *Methylobacter-Methylotenera* enrichment incubations

Methylobacter-Methylotenera enrichments were grown in triplicate under different methane concentrations, temperatures, and salinities. Methane concentrations in the headspace were regularly

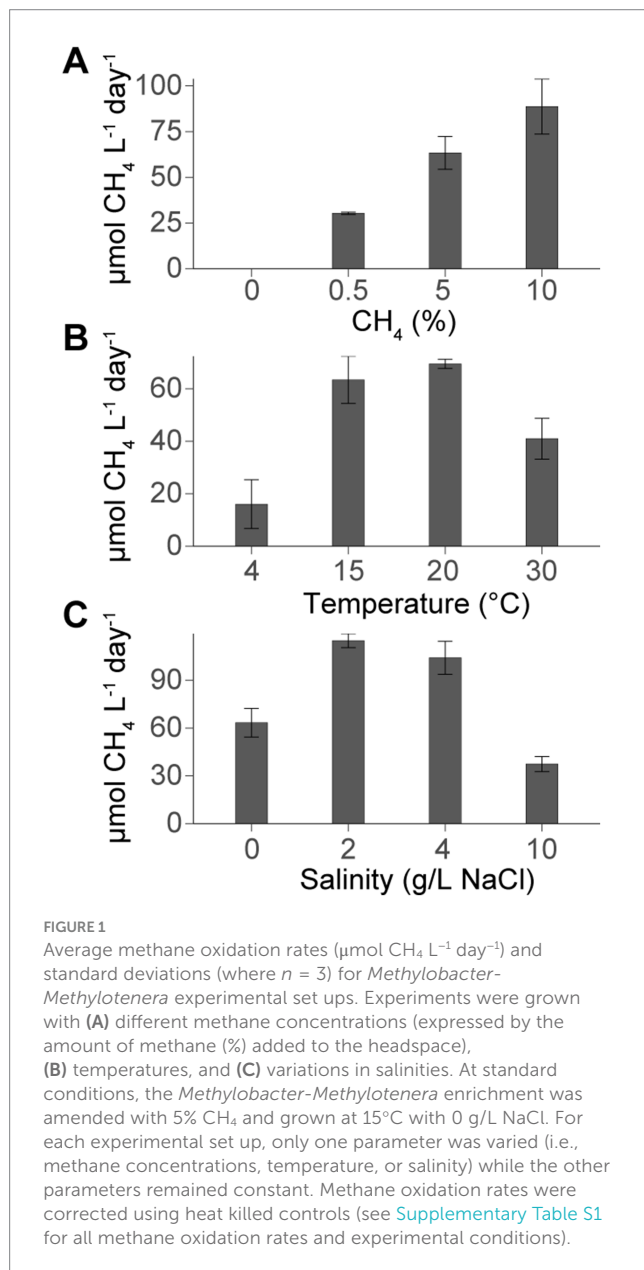
measured until the enrichment reached the stationary phase when the rate of methane consumption slowed. The resulting methane-oxidation rates were calculated over the period of maximum methane consumption. The methane-oxidation rates for the experimental conditions were corrected using the heat killed controls to account for any potential loss of methane during the experiments through the rubber stoppers (Figure 1; Supplementary Table S1). Methane-oxidation rates in the unamended controls were close to zero, indicating there was no methane production or consumption without the addition of methane. In general, we observe the highest methane-oxidation rates at 10% methane concentrations, temperatures of 15 and 20°C, and lower salinities (0–4 g/L NaCl).

3.2 Verification of the *Methylobacter-Methylotenera* enrichment culture composition

The original co-culture of *Methylobacter* sp. and *Methylotenera* sp. was enriched from a lake (van Grinsven et al., 2020). The 16S rRNA gene amplicon results confirmed that *Methylobacter* spp. remains the primary methanotroph present in the time 0 samples (i.e., representing 79% of the total 16S rRNA gene reads; Table 1) and the incubations grown at 5% CH₄ and 15°C with 0 g/L NaCl (84% of the total 16S rRNA gene reads). However, the relative abundance of *Methylobacter* spp. decreased in the incubations grown at 30°C (61% of the total 16S rRNA gene reads) and at 10 g/L NaCl (63% of the total 16S rRNA gene reads). *Methylomonas* spp. was the only other methanotroph detected but was present in low abundance (0.1% of the total 16S rRNA gene reads) at standard conditions (i.e., 5% CH₄ and 15°C with 0 g/L NaCl). The methylophilus, *Methylotenera* spp., increased in relative abundance at 30°C (26% of the total 16S rRNA gene reads), but decreased at higher salinities (0.1% of the total 16S rRNA gene reads). *Methylophilus* spp., a methanol-utilizing bacteria, increased in relative abundance (15% of the total 16S rRNA gene reads) at 10 g/L of NaCl relative to experiments with 0 g/L of NaCl (0.3% of the total 16S rRNA gene reads). In addition, 16S rRNA gene sequences attributed to other non-methanotrophs increased in the experiments grown with 10 g/L of NaCl (22% of the total 16S rRNA gene reads) relative to the experiments grown with 0 g/L of NaCl (10% of the total 16S rRNA gene reads) with the most abundant 16S rRNA gene reads being attributed to the families: *Chitinophagaceae*, *Flavobacteriaceae*, *Devisiaceae*, and *Optiutaceae*.

3.3 Variations in respiratory quinone distributions

Respiratory quinone distributions for the *Methylobacter-Methylotenera* co-culture experiments are shown as averages of the triplicate incubations grown at different temperatures (Figure 2A) and salinities (Figure 2B). Note, that respiratory quinones were not detected in *Methylobacter-Methylotenera* enrichments grown with 0, 5, and 10% methane additions at temperatures of 15°C and salinities of 0 g/L NaCl. We speculate this might be related to the lipid extraction procedure, but we are unsure why this was the case. To determine which respiratory quinones are likely being produced by *Methylotenera* spp. in the enrichment experiments, the quinone distributions of a



pure culture from a closely related strain to that of *Methylotenera* sp. in the enrichment, i.e., *Methylotenera mobilis*, were also analyzed ([Figure 2C](#)).

To evaluate whether methanotrophs adapted to extreme environmental conditions have a similar metabolic response to environmental stress, we also report changes in respiratory quinones for a psychrotolerant methanotroph, *Methylovulum psychrotolerans*, grown at different temperatures ([Figure 2D](#)). Note, *Methylovulum psychrotolerans* were not grown in triplicate, so lipid results reflect single incubation experiments.

In all cultures, ubiquinones with 8 isoprenoid units and 8 unsaturations ($\text{UQ}_{8:8}$), followed by $\text{UQ}_{7:7}$ were the most abundant quinones. In the *Methylobacter-Methylotenera* enrichment, methylene-ubiquinone ($\text{MQ}_{8:7}$) was also detected. A $\text{UQ}_{8:8}$ with an additional methoxy group on the unsaturated isoprenoidal chain ($\text{UQ}_{8:8} + \text{OCH}_3$) was identified in both the enrichment and *Methylotenera mobilis*. Further, a $\text{UQ}_{8:8}$ with an additional methoxy group and additional hydroxyl group on the side chain ($\text{UQ}_{8:8} + \text{OCH}_3 + \text{OH}$) was observed in the enrichment (see [Supplementary material S3.2](#)). $\text{UQ}_{9:9}$ and $\text{UQ}_{10:10}$ were identified in all samples, but in low relative abundance. Menaquinones (i.e., $\text{MK}_{6:6}$, $\text{MK}_{7:7}$, and $\text{MK}_{8:8}$) were only found in the *Methylobacter-Methylotenera* enrichment incubations, but in minor proportions (i.e., <1%). The full list of quinones identified in the *Methylobacter-Methylotenera* enrichments, *Methylotenera mobilis*, and *Methylovulum psychrotolerans* are listed in [Supplementary Tables S3–S5](#).

3.4 BHP distributions in methanotroph culture experiments

BHP distributions are reported as averages of triplicates for the *Methylobacter-Methylotenera* enrichment grown under different methane concentrations, temperatures, and salinities ([Figures 3A–C](#)). *Methylotenera mobilis* (DSM 17540), a phylogenetically closely related species of the methylotroph identified in the enrichment co-culture, was previously analyzed to confirm that the methylotroph does not produce any BHPs ([Richter et al., 2023](#)). The BHP distributions for *Methylovulum psychrotolerans* grown under

TABLE 1 Results from 16S rRNA gene amplicon sequencing reported as average relative abundance (% of total) and standard deviation ($n = 3$) of 16S rRNA gene reads for methanotrophs and methylotrophs detected in the *Methylobacter-Methylotenera* enrichment (where Temp. = temperature and n.d. = not detected).

Growth conditions				Relative abundance of 16S rRNA gene reads (% of total)				
Experiment	CH ₄ (%)	Temp. (°C)	Salinity (g/L NaCl)	<i>Methylobacter</i> spp.	<i>Methylomonas</i> spp.	<i>Methylotenera</i> spp.	<i>Methylophilus</i> spp.	Other
Time 0*	–	15	0	78.8 ± 2.3	0.1 ± 0.1	9.4 ± 2.5	1.2 ± 1.2	10.5 ± 2.2
Methane	5	15	0	83.6 ± 0.9	0.1 ± 0.1	2.6 ± 0.4	0.3 ± 0.1	13.3 ± 0.7
Temp.	5	30	0	60.7 ± 5.6	n.d.	25.6 ± 2.6	2.6 ± 1.1	11.0 ± 3.5
Salinity	5	15	10	62.9 ± 5.4	n.d.	0.1 ± 0.0	15.3 ± 3.3	21.7 ± 2.2

Total reads per sample are reported in [Supplementary Table S2](#).

*Time 0 samples were grown under standard conditions (15°C with 5% CH_4 and 0 g/L NaCl) and taken at the start of each experiment except for the salinity cultures, as these were slowly adapted to higher salinities over several months and time 0 samples could not be grown at standard conditions.

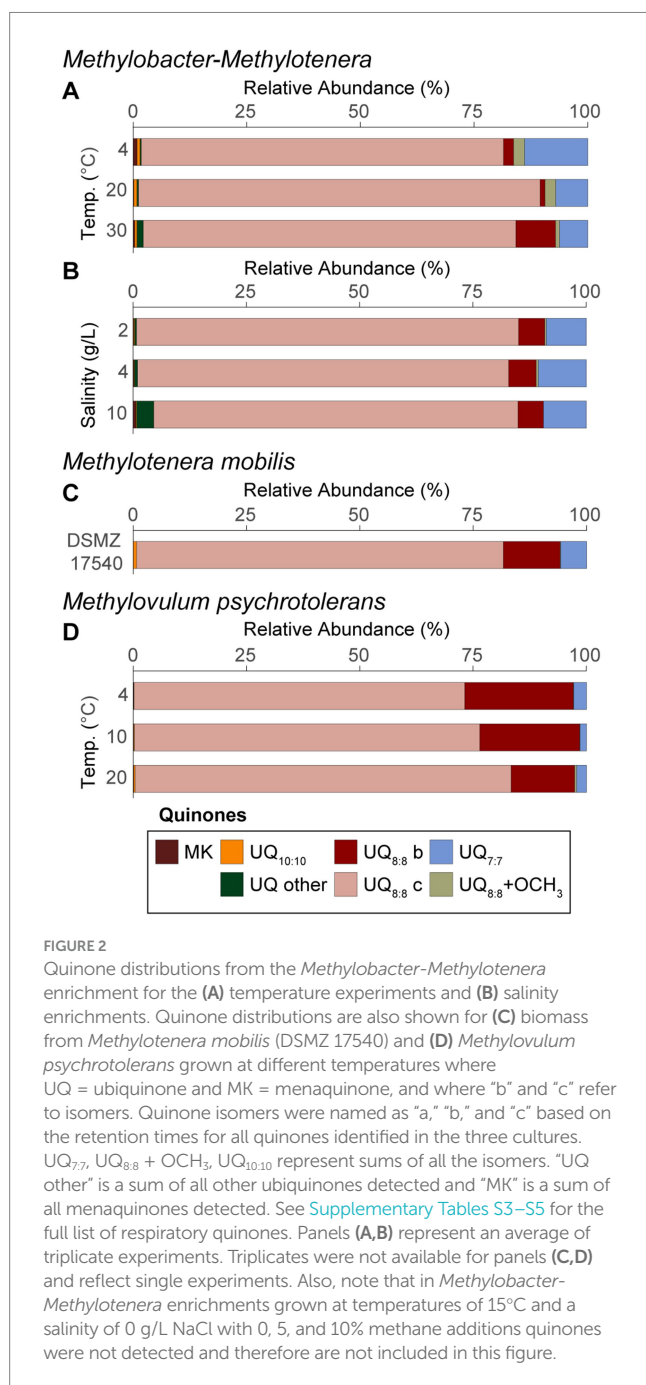


FIGURE 2

Quinone distributions from the *Methylobacter-Methylotenera* enrichment for the (A) temperature experiments and (B) salinity enrichments. Quinone distributions are also shown for (C) biomass from *Methylobacter mobilis* (DSMZ 17540) and (D) *Methylovulum psychrotolerans* grown at different temperatures where UQ = ubiquinone and MK = menaquinone, and where “b” and “c” refer to isomers. Quinone isomers were named as “a,” “b,” and “c” based on the retention times for all quinones identified in the three cultures. UQ_{7,7}, UQ_{8,8} + OCH₃, UQ_{10,10} represent sums of all the isomers. “UQ other” is a sum of all other ubiquinones detected and “MK” is a sum of all menaquinones detected. See [Supplementary Tables S3–S5](#) for the full list of respiratory quinones. Panels (A,B) represent an average of triplicate experiments. Triplicates were not available for panels (C,D) and reflect single experiments. Also, note that in *Methylobacter-Methylotenera* enrichments grown at temperatures of 15°C and a salinity of 0 g/L NaCl with 0, 5, and 10% methane additions quinones were not detected and therefore are not included in this figure.

different temperatures are also reported (Figure 3D). All the BHPs identified in this study are listed in [Supplementary Tables S6, S7](#).

In the *Methylobacter-Methylotenera* enrichment experiments, 35-aminobacteriohopane-32,33,34-triol (aminotriol from herein) and aminopentol were the most abundant BHPs (Figures 3A–C). The PCA of the BHP data confirms that the BHPs extracted from our triplicates were consistent with each other and cluster together in the biplot, with the exception of one of the triplicates grown at a temperature of 30°C (Figure 4). Further, this highlights that the largest variability observed in the BHP distributions occurred in incubations grown at higher salinities (i.e., PC1 42.6%).

Using new analytical techniques, we were able to detect a much broader diversity of BHPs in our samples (Hopmans et al., 2021;

Figures 3A–C) than observed using previous analytical techniques (Schulenberg-Schell et al., 1989; Moreau et al., 1995; Talbot et al., 2001, 2016; Kusch et al., 2018). Notably, BHPs with an ethyl group attached to the carbamate moiety of the side chain were detected in the *Methylobacter-Methylotenera* enrichments grown at higher salinities (Figure 3C; [Supplementary Figure S8](#)). We refer to these BHPs as ethylcarbamate (EC)-aminoBHPs: EC-aminotriol, EC-aminotetrol, and EC-aminopentol (see [Supplementary material S4.2.1](#)). In addition, an unknown composite BHP was also detected in *Methylobacter-Methylotenera* enrichments adapted to higher salinities ([Supplementary material S4.2.2](#); [Supplementary Figure S7](#)).

The most abundant BHPs detected in *Methylovulum psychrotolerans* were aminotriol and Δ^{11} -aminotriol, followed by aminotetrol, aminopentol, and their unsaturated versions (Figure 3D). In addition, methylcarbamate-aminoBHPs, ethenolamine-BHPs, N-formylated-aminoBHPs, and a novel unknown composite BHP were also detected in the *Methylovulum psychrotolerans* cultures ([Supplementary material S4.2.3](#); [Supplementary Figure S8](#)).

4 Discussion

4.1 Lipid response of *Methylobacter-Methylotenera* to increasing methane concentrations

In the *Methylobacter-Methylotenera* enrichment experiments, methane oxidation rates increased with increasing methane concentrations (Figure 1). In good agreement with these results, methane concentrations in lakes are usually positively correlated with methane oxidation rates (Kankaala et al., 2006; Guérin and Abril, 2007; Martinez-Cruz et al., 2015), suggesting that methanotrophs are sensitive to changes in methane availability and thereby actively regulate the amount of methane emitted from lakes.

The 16S rRNA gene amplicon sequencing results showed that *Methylobacter* spp. was the most abundant methanotroph in the 5% CH₄ amended experiments (83% of the total 16S rRNA gene reads; Table 1). The BHP distributions of all methane amended enrichment experiments are characterized by a high abundance of aminotriol and aminopentol (Figure 3A); this is consistent with the BHP distributions observed in other *Methylobacter* strains (Osborne, 2015; Rush et al., 2016). We observed no major changes in the overall BHP distributions in the 0, 0.5, 5, and 10% CH₄ amended enrichment experiments (Figure 3A). This is confirmed by a PCA, where the methane concentration experiments all clustered together with our time zero samples (Figure 4). Overall, these results point to the utility of aminotriol and aminopentol as chemotaxonomic biomarkers for *Methylobacter* methanotrophs, which are an important component of the methanotrophic community contributing to methane oxidation in freshwater systems (Knief, 2015).

The total BHP concentrations in the enrichments (not corrected for cell density) increased with increasing CH₄ and higher methane oxidation rates, which reflects an increase in aminotriol, aminotetrol, aminopentol in the CH₄ amended experiments relative to the unamended CH₄ experiments (Figure 3A). In previous enrichment incubations for methanotrophs, only aminopentol concentrations increased with increasing methane concentrations, likely reflecting increased growth or activity of *Methylobacter* sp.

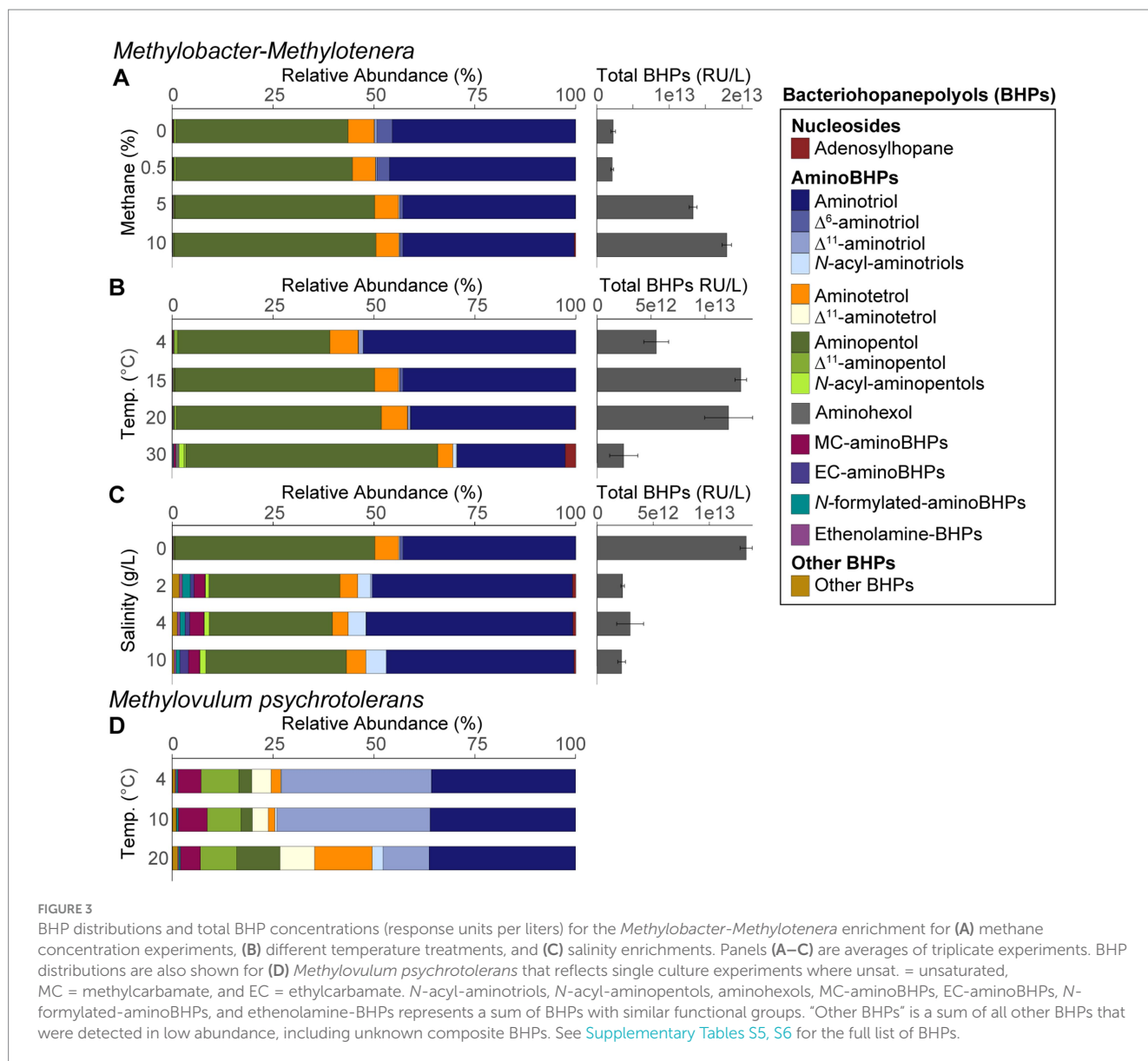


FIGURE 3

BHP distributions and total BHP concentrations (response units per liters) for the *Methylobacter-Methylotenera* enrichment for (A) methane concentration experiments, (B) different temperature treatments, and (C) salinity enrichments. Panels (A–C) are averages of triplicate experiments. BHP distributions are also shown for (D) *Methylovulum psychrotolerans* that reflects single culture experiments where unsat. = unsaturated, MC = methylcarbamate, and EC = ethylcarbamate. *N*-acyl-aminotriols, *N*-acyl-aminopentols, aminohexols, MC-aminoBHPs, EC-aminoBHPs, *N*-formylated-aminoBHPs, and ethenolamine-BHPs represents a sum of BHPs with similar functional groups. “Other BHPs” is a sum of all other BHPs that were detected in low abundance, including unknown composite BHPs. See [Supplementary Tables S5, S6](#) for the full list of BHPs.

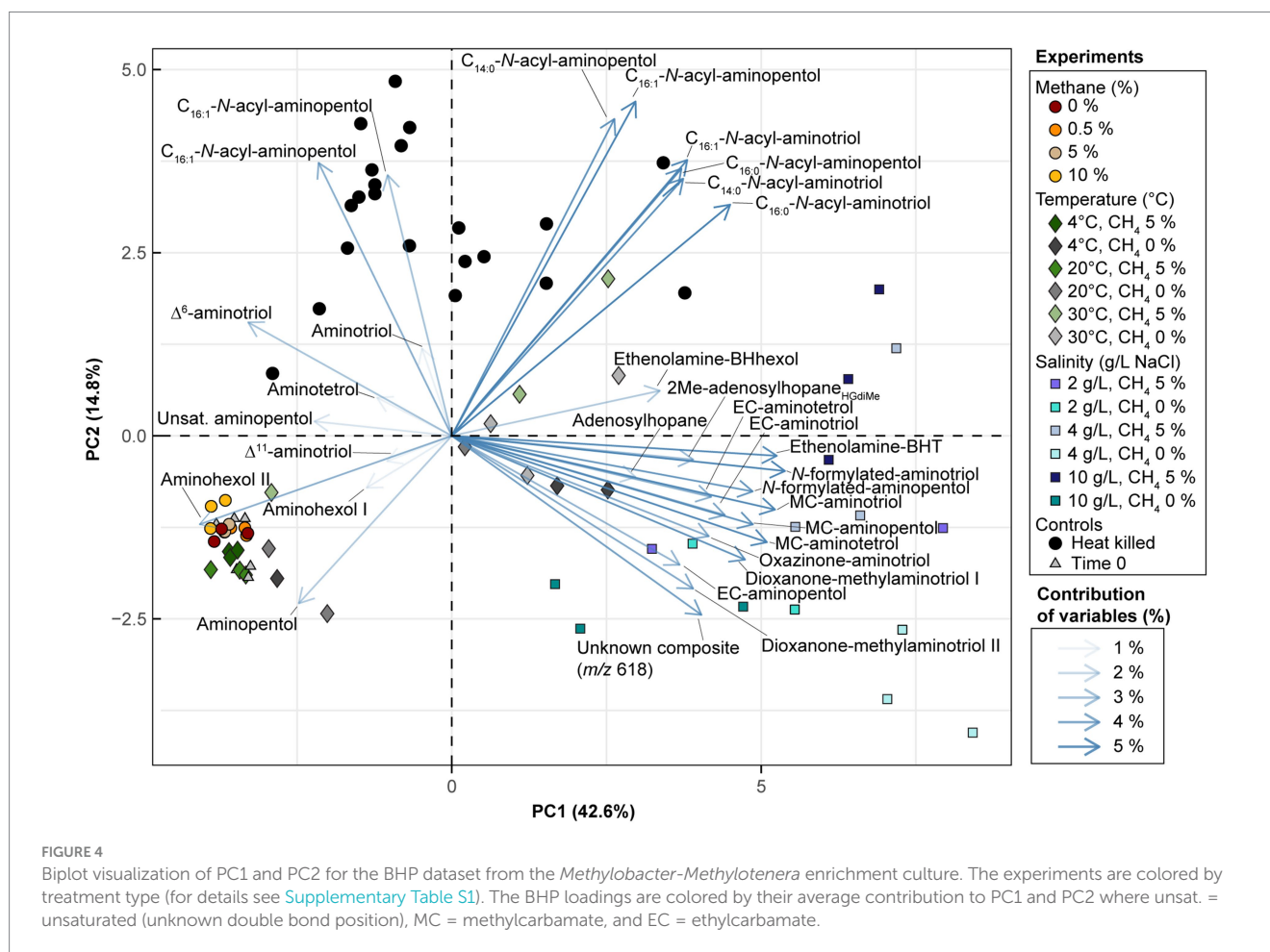
present in the microcosm experiments (Osborne, 2015; Sherry et al., 2016). An increase in aminoBHP abundance is often interpreted as evidence for increased methane-oxidation by MOB in the environment, particularly in paleo-records (Coolen et al., 2008; Blumenberg et al., 2013; Talbot et al., 2014). Further work with pure MOB cultures and exact concentrations of BHPs per cell are needed to confirm whether the increase in the total absolute abundance of aminoBHPs at higher methane concentrations reflects an increase in cells oxidizing methane or a higher concentration of aminoBHPs in a small percentage of cells performing methane oxidation.

4.2 Methanotroph lipid response to temperature

The highest methane oxidation rates for the *Methylobacter-Methylotenera* enrichments occurred at 15 and 20°C (Figure 1). This

reflects the optimal growth temperatures for the mesophilic members of the *Methylobacter* and *Methylotenera* genus at 23–35°C and 18–21°C, respectively (Afshin et al., 2021; Collins et al., 2017). *Methylovulum psychrotolerans* is considered psychrotolerant and can grow at temperatures from 2 to 36°C, but with optimal growth also occurring between 20 and 25°C (Oshkin et al., 2016).

Methylovulum psychrotolerans and *Methylotenera mobilis* both have a similar quinone distribution to that of the *Methylobacter-Methylotenera* enrichment, where UQ_{8.8} is the most abundant quinone, followed by UQ_{7.7} (Figure 2). Proteobacteria, including Betaproteobacteria and Gammaproteobacteria, are known to produce UQ_{8.8} and UQ_{7.7} in high abundance (Collins and Green, 1985; Hiraishi, 1999; Kersters et al., 2006). In *Methylovulum psychrotolerans*, the relative abundance of UQ_{8.8} increases with temperature from 4 to 20°C. Similarly, the relative abundance of UQ_{8.8} increases at 20°C in the *Methylobacter-Methylotenera* co-culture compared to the incubations at 4 and 30°C (Figure 2A). Increased relative abundance of UQ_{8.8} at 20°C suggests that production of this quinone is associated



with optimal growth conditions due to either an increase in biomass or an increase in UQ_{8,8} production per cell. In the *Methylobacter-Methylotenera* enrichment both menaquinones and UQ_{8,8} with methoxy groups increase at lower temperatures (Figure 2A); however, this does not occur in *Methylovulum psychrotolerans*. It was previously demonstrated that in *Listeria monocytogenes* the menaquinone content in the cell membrane increases at lower temperatures to improve membrane fluidity (Seel et al., 2018; Flegler et al., 2021). This might also explain the increase in menaquinones in the *Methylobacter-Methylotenera* enrichments. Alternatively, the changes in quinone distributions observed in the *Methylobacter-Methylotenera* enrichment could be attributed to a shift in the microbial community composition at lower temperatures.

In the *Methylobacter-Methylotenera* enrichment, *Methylobacter* spp. is the most abundant methanotroph in the 15 and 30°C temperature experiments (Table 1) and also the primary BHP producer, as *Methylotenera* spp. does not produce BHPs (Richter et al., 2023). In addition, all other species detected in the culture are present in low relative abundance (Table 1). In the BHP distributions, aminotriol decreases at higher temperatures, whereas the relative abundances of aminopentol and adenosylhopane both increase (Figure 3C). In the Type I methanotroph, *Methylovulum psychrotolerans*, the relative abundance of aminotetrol and aminopentol increase with temperature from 4 to 20°C (Figure 3D). A similar increase in aminoBHPs at higher

temperatures was previously observed in the River Tyne microcosm experiments where the highest concentrations of aminotriol and aminotetrol occurred between 4 and 21°C and aminopentol increased from 4 to 40°C (Osborne et al., 2017). However, community succession of different species of *Methylobacter* sp. in the River Tyne experiments explained some of the differences in aminoBHP production at different temperatures (Sherry et al., 2016; Osborne et al., 2017). Pure culture experiments with another Type I methanotroph (strain CEL 1923) did not show the same trends observed in our cultures, where only aminotetrol increased with temperature and aminopentol decreased at higher temperatures (Jahnke et al., 1999). This could indicate species-specific adaptations of BHP production at different temperatures. In mesocosm experiments in *Sphagnum* peat bogs that were likely dominated by Type II methanotrophs, aminotriol, aminotetrol, and aminopentol concentrations also increased in response to increasing temperatures from 5 to 25°C (van Winden et al., 2020). These studies suggest that changes in aminoBHPs could reflect either community succession or a physiological adaptation at different temperatures. In the case of the *Methylobacter-Methylotenera* enrichment and in *Methylovulum psychrotolerans*, the increase in aminotetrol and aminopentol reflects an increase in hydroxyl groups likely promoting increased hydrophilic interactions among lipids in the lipid membrane. An increase in the number of hydroxylations at higher growth temperatures was

previously observed in membrane lipids in *Rhizobium tropici* and *Burkholderia cepacia* to increase the lateral interactions between lipid molecules in response to stress (Taylor et al., 1998; Vences-Guzmán et al., 2011). Thus, an increase in aminoBHPs with more hydroxyl groups could be a stress related response in the *Methylobacter-Methylotenera* enrichment and in *Methylovulum psychrotolerans*.

PCA results from the *Methylobacter-Methylotenera* enrichment experiments show that Δ^6 - and Δ^{11} -aminotriol have a minor contribution to temperatures below 30°C (Figure 4). We observed a slight increase in Δ^6 -aminotriol at lower temperatures. However, this response in the *Methylobacter-Methylotenera* enrichment was much smaller than that observed in *Methylovulum psychrotolerans*, where Δ^{11} -aminotriol, Δ^{11} -aminotetrol, Δ^{11} -aminopentol, unsat. MC-aminotriol, unsat. ethenolamine-bacteriohopanetetrol, and unsat. *N*-formylated-aminotriol all increase with decreasing temperatures relative to their saturated counterparts (Figure 3D). So far, the increase in unsaturated BHPs at colder temperatures has not been observed in other methanotroph cultures (Jahnke et al., 1999; Osborne et al., 2017; van Winden et al., 2020). In the River Tyne enrichment experiments, for instance, unsaturated BHPs were only observed at high temperatures (i.e., 40 and 50°C) and corresponded to a shift in the methanotroph community to a different species of *Methylobacter* and the thermophilic genus *Methylocaldum* (Sherry et al., 2016; Osborne et al., 2017). This suggests that mono-unsaturated aminoBHPs, in particular Δ^{11} -aminopentol, could be a unique temperature adaptation in certain methanotrophs, such as *Methylovulum* spp. and *Methylocaldum* spp., for maintaining membrane homeostasis under colder conditions (Cvejic et al., 2000; van Winden et al., 2012; Bale et al., 2019). The differences in BHP response to temperature in the *Methylobacter-Methylotenera* enrichment compared to *Methylovulum psychrotolerans* and previous culture studies (i.e., Jahnke et al., 1999; Osborne et al., 2017; van Winden et al., 2020) suggests species specific adaptations in BHPs to changes in temperature. In general, this highlights the need for more environmental and culture studies to understand how temperature affects BHP distributions in MOB.

4.3 Salinity effects on methanotroph-associated BHPs and quinones

Adapting the *Methylobacter-Methylotenera* enrichment to different salinities resulted in a notable decrease in methane oxidation rates at 10 g/L NaCl and total BHP concentrations (not corrected for cell density; Figure 1C), likely because this enrichment was originally isolated from a freshwater lake and not accustomed to large changes in salinity (van Grinsven et al., 2020). In these experiments, *Methylobacter* spp. remains the primary methanotroph at 10 g/L NaCl and other methanotrophs present at 0 g/L NaCl are absent from our 16S rRNA gene amplicon results. For the methylotrophs, however, *Methylophilus* spp. increased in abundance, whereas *Methylotenera* spp. is largely absent indicating a change in the microbial interactions (Table 1). In pure culture experiments with a haloalkaliphilic Type I methanotroph, *Methylovulum alcaliphilum*, growth rates and hopanoid absolute abundance decreased at higher salinities (Cordova-Gonzalez et al., 2021), suggesting that an increase in salinity could

also lead to lower methane oxidation rates and BHP abundance in a *Methylovulum*-rich community. Similarly, previous enrichment experiments with River Tyne sediments led to a decrease in methane oxidation rates at higher salinities; however, there was no corresponding decrease in total BHP concentrations (Osborne, 2015; Sherry et al., 2016). This lack of change in BHP concentration in the River Tyne enrichments, could be explained by a shift in the methanotroph community from *Methylobacter* to *Methylomicrobium* spp., which was better adapted to growing at higher salinities (Sherry et al., 2016). These results seem to suggest that methane oxidation rates generally decrease at higher salinities, although it is unclear whether this is driven by a decrease in the number of cells or in methanotroph activity.

In the *Methylobacter-Methylotenera* enrichments, the only notable change in quinone distribution with increasing salinity is an increase in UQ_{10:10}. Previous studies demonstrated that UQ_{10:10} is directly involved in regulating membrane permeability and elasticity under osmotic stress in an *Escherichia coli* strain (Eriksson et al., 2019). It is likely that UQ_{10:10} plays a similar role in the enrichment culture. However, the increase in UQ_{10:10} could also be linked to non-methanotrophs at 10 g/L NaCl relative to the enrichments grown at 0 g/L NaCl, as shown by the change in the 16S rRNA gene read distribution and a change in dominance in the methylotroph from *Methylotenera* spp. to *Methylophilus* spp. (Table 1).

Salinity accounts for the largest variability of BHPs in the *Methylobacter-Methylotenera* experiments (PC1 42.6%; Figure 4). We observe an overall decrease in the relative abundance of aminopentol relative to 0 g/L NaCl (Figure 3C). The relative abundances of MC-aminotriol, MC-aminotetrol, and MC-aminopentol, as well as their ethylcarbamate forms all increase at higher salinities (Figure 3C). In contrast, there are no significant changes in individual BHP concentrations in the River Tyne incubations at salinities from 1 to 70 g/L NaCl, and only at 120 g/L NaCl do individual BHP concentrations significantly decrease (Osborne, 2015). Incubations at different salinities with the haloalkaliphilic *Methylovulum alcaliphilum*, led to either a decrease or no change in individual BHP concentrations at higher salinities (Cordova-Gonzalez et al., 2021). As we are working with an enrichment co-culture (see van Grinsven et al., 2020), the individual changes in BHP abundances could either be explained as a physiological adaptation by *Methylobacter* or a change in the microbial community. As demonstrated by the 16S rRNA gene sequencing results, the relative abundance of 16S rRNA gene reads of non-methanotrophs, such as *Methylophilus* spp., increase at 10 g/L NaCl in the *Methylobacter-Methylotenera* co-culture (Table 1); suggesting that other microbes might also be responsible for BHP production in the enrichments. A protein blast search (NCBI, National Center for Biotechnology Information) for the squalene-hopene cyclase gene of *Bradyrhizobium japonicum* (accession no. WP_038942977.1) did not yield any hits for *Methylophilus* spp. Although this suggests that the *Methylophilus* spp. present in the enrichment might not be responsible for BHP production in the enrichment, it does not rule out the possibility of BHP production by other microorganisms present. As a physiological adaptation, the increase in BHPs with additional modifications to the sidechain was previously proposed to promote intracellular associations that might lead to lipid raft-like domains and tighter packing of saturated phospholipids in a liquid-ordered phase (Sáenz, 2010). More work,

however, is needed to test whether this holds true for BHPs with amine functional groups. To simulate natural methanotroph adaptations to changes in salinity, culture experiments with halophilic methanotrophs and/or enrichments from marine sites are needed, and further modifications in membrane lipids should be evaluated.

5 Conclusion

Lipid analyses of *Methylobacter-Methylotenera* enrichment from a freshwater lake and from preliminary experiments with an extremophile, *Methylovulum psychrotolerans*, highlight several distinct membrane lipid responses of MOBs to environmental stress. In the *Methylobacter-Methylotenera* enrichments grown at higher methane concentrations both methane-oxidation rates and total BHP concentrations (not corrected for cell density) increased. The increase in BHPs was driven by an increase in aminoBHPs, which should be explored further as potential biomarkers for methane oxidation. In the enrichments and *Methylovulum psychrotolerans*, the relative abundance of aminopentol increased at higher temperatures. In contrast, the relative abundances of unsaturated BHPs increased in the *Methylovulum psychrotolerans* culture at lower temperatures. For *Methylovulum psychrotolerans* and the *Methylobacter-Methylotenera* enrichment, UQ_{8,8} increased in relative abundance at 20°C, corresponding to the optimal growth temperatures of both cultures. This highlights potential specific BHP adaptations in different methanotrophs to variations in temperature, but a more consistent response in ubiquinone distributions. Our salinity incubations account for the largest variance in the BHP dataset for the *Methylobacter-Methylotenera* enrichment, but only a slight increase in UQ_{10,10} and no other major changes in the quinone distributions. The BHP variability at higher salinities is mainly explained by increases in the relative abundance of aminoBHPs with additional modifications to the side chain (e.g., MC-aminoBHPs, EC-aminoBHPs, and N-formylated-aminoBHPs). However, more work is needed to test whether this is an adaptation to changes in salinity by methanotrophs. This work highlights the potential of using a combined biomarker approach by analyzing both respiratory quinones and BHPs to understand how environmental changes influence methanotroph activity and lipid membrane adaptations in modern settings.

Data availability statement

All sequence data generated in this study were deposited in the NCBI sequence read archive (SRA) under the BioProject number PRJNA1149959: <https://www.ncbi.nlm.nih.gov/bioproject/PRJNA1149959>. All other datasets generated in this study are included in the article/Supplementary material.

Author contributions

NR: Conceptualization, Data curation, Formal analysis, Investigation, Visualization, Writing – original draft, Writing – review & editing. LV: Conceptualization, Funding acquisition, Validation, Writing – review & editing. EH: Formal analysis, Validation,

Writing – review & editing. NB: Conceptualization, Formal analysis, Writing – review & editing. JS: Conceptualization, Funding acquisition, Writing – review & editing. DR: Conceptualization, Funding acquisition, Writing – review & editing, Resources.

Funding

The author(s) declare that financial support was received for the research, authorship, and/or publication of this article. This work was supported by the NWO gravitation grant for Soehngen Institute of Anaerobic Microbiology (024.002.002). This project received funding from the European Research Council (ERC) under the European Union's Horizon 2020 research and innovation program (grant agreement no. 694569), and from Utrecht University.

Acknowledgments

We would like to thank M. Verweij, D. Dorhout, W.I.C. Rijpstra, Ossebaar, R. van Bommel, M. van der Meer, I. Posthuma, A. Nordeloos, M. Grego, M. Brouwer, W. Reitsma, S.E. Belova, and I.Y. Oshkin for technical support and advice. We would like to thank S.N. Dedysh for input on the manuscript. We would also like to thank S. van Grinsven for advice on the enrichment co-culture.

Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

The author(s) declared that they were an editorial board member of Frontiers, at the time of submission. This had no impact on the peer review process and the final decision.

Generative AI statement

The authors declare that no Generative AI was used in the creation of this manuscript.

Publisher's note

All claims expressed in this article are solely those of the authors and do not necessarily represent those of their affiliated organizations, or those of the publisher, the editors and the reviewers. Any product that may be evaluated in this article, or claim that may be made by its manufacturer, is not guaranteed or endorsed by the publisher.

Supplementary material

The Supplementary material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmicb.2025.1532719/full#supplementary-material>

References

- Abdala Asbun, A., Besseling, M. A., Balzano, S., van Bleijswijk, J. D. L., Witte, H. J., Villanueva, L., et al. (2020). Cascabel: a scalable and versatile amplicon sequence data analysis pipeline delivering reproducible and documented results. *Front. Genet.* 11:489357. doi: 10.3389/fgene.2020.489357
- Afshin, Y., Delherbe, N., and Kalyuzhnaya, M. G. (2021). "Methylotenera" in Bergey's manual of systematics of Archaea and Bacteria. ed. W. B. Whitman (Wiley), 1–11. Available at: <https://onlinelibrary.wiley.com/doi/book/10.1002/9781118960608>
- Andrews, S. (2010). Babraham bioinformatics-FastQC a quality control tool for high throughput sequence data. Available at: <https://www.bioinformatics.babraham.ac.uk/projects/fastqc/> (Accessed January 26, 2024).
- Anraku, Y. (1988). Bacterial Electron transport chains. *Annu. Rev. Biochem.* 57, 101–132. doi: 10.1146/annurev.bi.57.070188.000533
- Apprill, A., McNally, S., Parsons, R., and Weber, L. (2015). Minor revision to V4 region SSU rRNA 806R gene primer greatly increases detection of SAR11 bacterioplankton. *Aquat. Microb. Ecol.* 75, 129–137. doi: 10.3354/ame01753
- Bale, N. J., Ding, S., Hopmans, E. C., Arts, M. G. I., Villanueva, L., Boschman, C., et al. (2021). Lipidomics of environmental microbial communities. I: visualization of component distributions using untargeted analysis of high-resolution mass spectrometry data. *Front. Microbiol.* 12:659302. doi: 10.3389/fmicb.2021.659302
- Bale, N. J., Rijpstra, W. I. C., Sahonero-Canavesi, D. X., Oshkin, I. Y., Belova, S. E., Dedysh, S. N., et al. (2019). Fatty acid and Hopanoid adaption to cold in the Methanotroph Methylotulvum psychrotolerans. *Front. Microbiol.* 10:589. doi: 10.3389/fmicb.2019.00589
- Becker, K. W., Elling, F. J., Schröder, J. M., Lipp, J. S., Goldhammer, T., Zabel, M., et al. (2018). Isoprenoid Quinones resolve the stratification of redox processes in a biogeochemical continuum from the photic zone to deep anoxic sediments of the Black Sea. *Appl. Environ. Microbiol.* 84, e02736–e02717. doi: 10.1128/AEM.02736-17
- Belin, B. J., Busset, N., Giraud, E., Molinaro, A., Silipo, A., and Newman, D. K. (2018). Hopanoid lipids: from membranes to plant–bacteria interactions. *Nat. Rev. Microbiol.* 16, 304–315. doi: 10.1038/nrmicro.2017.173
- Bligh, E. G., and Dyer, W. J. (1959). A rapid method of total lipid extraction and purification. *Can. J. Biochem. Physiol.* 37, 911–917. doi: 10.1139/y59-099
- Blumenberg, M., Berndmeyer, C., Moros, M., Muschalla, M., Schmale, O., and Thiel, V. (2013). Bacteriohopanepolyols record stratification, nitrogen fixation and other biogeochemical perturbations in Holocene sediments of the Central Baltic Sea. *Biogeochemistry* 10, 2725–2735. doi: 10.5194/bg-10-2725-2013
- Bowman, J. (2006). "The Methanotrophs — the families Methylococcaceae and Methylocystaceae" in *The Prokaryotes: Volume 5: Proteobacteria: Alpha and Beta subclasses*. eds. M. Dworkin, S. Falkow, E. Rosenberg, K.-H. Schleifer and E. Stackebrandt (New York, NY: Springer), 266–289.
- Callahan, B. J., McMurdie, P. J., Rosen, M. J., Han, A. W., Johnson, A. J. A., and Holmes, S. P. (2016). DADA2: high-resolution sample inference from Illumina amplicon data. *Nat. Methods* 13, 581–583. doi: 10.1038/nmeth.3869
- Caporaso, J. G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F. D., Costello, E. K., et al. (2010). QIIME allows analysis of high-throughput community sequencing data. *Nat. Methods* 7, 335–336. doi: 10.1038/nmeth.f.303
- Caporaso, J. G., Lauber, C. L., Walters, W. A., Berg-Lyons, D., Lozupone, C. A., Turnbaugh, P. J., et al. (2011). Global patterns of 16S rRNA diversity at a depth of millions of sequences per sample. *Proc. Natl. Acad. Sci.* 108, 4516–4522. doi: 10.1073/pnas.100080107
- Collins, D. A., Akberdin, I. R., and Kalyuzhnaya, M. G. (2017). "Methylobacter" in Bergey's manual of systematics of Archaea and Bacteria (John Wiley & Sons, Ltd), 1–12.
- Collins, M. D., and Green, P. N. (1985). Isolation and characterization of a novel coenzyme Q from some methane-oxidizing bacteria. *Biochem. Biophys. Res. Commun.* 133, 1125–1131. doi: 10.1016/0006-291X(85)91253-7
- Coolen, M. J. L., Talbot, H. M., Abbas, B. A., Ward, C., Schouten, S., Volkman, J. K., et al. (2008). Sources for sedimentary bacteriohopanepolyols as revealed by 16S rDNA stratigraphy. *Environ. Microbiol.* 10, 1783–1803. doi: 10.1111/j.1462-2920.2008.01601.x
- Cordova-Gonzalez, A., Birgel, D., Kappler, A., and Peckmann, J. (2021). Variation of salinity and nitrogen concentration affects the pentacyclic triterpenoid inventory of the haloalkaliphilic aerobic methanotrophic bacterium Methylotulvum microbium alcaliphilum. *Extremophiles* 25, 285–299. doi: 10.1007/s00792-021-01228-x
- Cvejić, J. H., Bodrossy, L., Kovács, K. L., and Rohmer, M. (2000). Bacterial triterpenoids of the hopane series from the methanotrophic bacteria Methylocaldum spp.: phylogenetic implications and first evidence for an unsaturated aminobacteriohopanepolyol. *FEMS Microbiol. Lett.* 182, 361–365. doi: 10.1111/j.1574-6968.2000.tb08922.x
- Doughty, D. M., Coleman, M. L., Hunter, R. C., Sessions, A. L., Summons, R. E., and Newman, D. K. (2011). The RND-family transporter, HpnN, is required for hopanoid localization to the outer membrane of *Rhodospseudomonas palustris* TIE-1. *Proc. Natl. Acad. Sci.* 108, E1045–E1051. doi: 10.1073/pnas.1104209108
- Doughty, D. M., Hunter, R. C., Summons, R. E., and Newman, D. K. (2009). 2-Methylhopanoids are maximally produced in akinetes of *Nostoc punctiforme*: geobiological implications. *Geobiology* 7, 524–532. doi: 10.1111/j.1472-4669.2009.00217.x
- Dunfield, P. F., Yuryev, A., Senin, P., Smirnova, A. V., Stott, M. B., Hou, S., et al. (2007). Methane oxidation by an extremely acidophilic bacterium of the phylum Verrucomicrobia. *Nature* 450, 879–882. doi: 10.1038/nature06411
- Dupont, C. L., Larsson, J., Yooseph, S., Ininbergs, K., Goll, J., Asplund-Samuelsson, J., et al. (2014). Functional tradeoffs underpin salinity-driven divergence in microbial community composition. *PLoS One* 9:e89549. doi: 10.1371/journal.pone.0089549
- Eriksson, E. K., Edwards, K., Grad, P., Gedda, L., and Agmo Hernández, V. (2019). Osmoprotective effect of ubiquinone in lipid vesicles modelling the *E. coli* plasma membrane. *Biochim. Biophys. Acta Biomembr.* 1861, 1388–1396. doi: 10.1016/j.bbamem.2019.04.008
- Ettwig, K. F., Van Alen, T., Van De Pas-Schoonen, K. T., Jetten, M. S. M., and Strous, M. (2009). Enrichment and molecular detection of denitrifying Methanotrophic Bacteria of the NC10 phylum. *Appl. Environ. Microbiol.* 75, 3656–3662. doi: 10.1128/AEM.00067-09
- Flegler, A., Kombeitz, V., and Lipski, A. (2021). Menaquinone-mediated regulation of membrane fluidity is relevant for fitness of *Listeria monocytogenes*. *Arch. Microbiol.* 203, 3353–3360. doi: 10.1007/s00203-021-02322-6
- Franza, T., and Gaudu, P. (2022). Quinones: more than electron shuttles. *Res. Microbiol.* 173:103953. doi: 10.1016/j.resmic.2022.103953
- Guérin, F., and Abril, G. (2007). Significance of pelagic aerobic methane oxidation in the methane and carbon budget of a tropical reservoir. *J. Geophys. Res. Biogeosci.* 112, 1–14. doi: 10.1029/2006JG000393
- Hanson, R. S., and Hanson, T. E. (1996). Methanotrophic bacteria. *Microbiol. Rev.* 60, 439–471. doi: 10.1128/mr.60.2.439-471.1996
- Hiraishi, A. (1999). Isoprenoid quinones as biomarkers of microbial populations in the environment. *J. Biosci. Bioeng.* 88, 449–460. doi: 10.1016/S1389-1723(00)87658-6
- Hiraishi, A., Ueda, Y., and Ishihara, J. (1998). Quinone profiling of bacterial communities in natural and synthetic sewage activated sludge for enhanced phosphate removal. *Appl. Environ. Microbiol.* 64, 992–998. doi: 10.1128/AEM.64.3.992-998.1998
- Hopmans, E. C., Smit, N. T., Schwartz-Narbonne, R., Sinninghe Damsté, J. S., and Rush, D. (2021). Analysis of non-derivatized bacteriohopanepolyols using UHPLC-GRMS reveals great structural diversity in environmental lipid assemblages. *Org. Geochem.* 160:104285. doi: 10.1016/j.orggeochem.2021.104285
- Islam, T., Jensen, S., Reigstad, L. J., Larsen, Ø., and Birkeland, N.-K. (2008). Methane oxidation at 55°C and pH 2 by a thermoacidophilic bacterium belonging to the Verrucomicrobia phylum. *Proc. Natl. Acad. Sci.* 105, 300–304. doi: 10.1073/pnas.0704162105
- Jahnke, L. L., Stan-Lotter, H., Kato, K., and Hochstein, L. I. (1992). Presence of methyl sterol and bacteriohopanepolyol in an outer-membrane preparation from *Methylococcus capsulatus* (Bath). *Microbiology* 138, 1759–1766. doi: 10.1099/00221287-138-8-1759
- Jahnke, L. L., Summons, R. E., Hope, J. M., and Des Marais, D. J. (1999). Carbon isotopic fractionation in lipids from methanotrophic bacteria II: the effects of physiology and environmental parameters on the biosynthesis and isotopic signatures of biomarkers. *Geochim. Cosmochim. Acta* 63, 79–93. doi: 10.1016/S0016-7037(98)00270-1
- Jürgens, U. J., Simonin, P., and Rohmer, M. (1992). Localization and distribution of hopanoids in membrane systems of the cyanobacterium *Synechocystis* PCC 6714. *FEMS Microbiol. Lett.* 92, 285–288. doi: 10.1016/0378-1097(92)90723-2
- Kankaala, P., Huotari, J., Peltomaa, E., Saloranta, T., and Ojala, A. (2006). Methanotrophic activity in relation to methane efflux and total heterotrophic bacterial production in a stratified, humic, boreal lake. *Limnol. Oceanogr.* 51, 1195–1204. doi: 10.4319/lo.2006.51.2.1195
- Kassambara, A., and Mundt, F. (2020). Factoextra: Extract and visualize the results of multivariate data analyses. Available at: <https://cran.r-project.org/web/packages/factoextra/index.html> (Accessed January 29, 2024).
- Kerstens, K., De Vos, P., Gillis, M., Swings, J., Vandamme, P., and Stackebrandt, E. (2006). "Introduction to the Proteobacteria" in *The Prokaryotes: Volume 5: Proteobacteria: Alpha and Beta subclasses*. eds. M. Dworkin, S. Falkow, E. Rosenberg, K.-H. Schleifer and E. Stackebrandt (New York, NY: Springer), 3–37.
- Knief, C. (2015). Diversity and habitat preferences of cultivated and uncultivated aerobic Methanotrophic Bacteria evaluated based on pmoA as molecular marker. *Front. Microbiol.* 6:1346. doi: 10.3389/fmicb.2015.01346
- Krause, S. M. B., Johnson, T., Samadhi Karunaratne, Y., Fu, Y., Beck, D. A. C., Chistoserdova, L., et al. (2017). Lanthanide-dependent cross-feeding of methane-derived carbon is linked by microbial community interactions. *Proc. Natl. Acad. Sci.* 114, 358–363. doi: 10.1073/pnas.1619871114
- Kusch, S., and Rush, D. (2022). Revisiting the precursors of the most abundant natural products on earth: a look back at 30+ years of bacteriohopanepolyol (BHP) research and ahead to new frontiers. *Org. Geochem.* 172:104469. doi: 10.1016/j.orggeochem.2022.104469

- Kusch, S., Shah Walter, S. R., Hemingway, J. D., and Pearson, A. (2018). Improved chromatography reveals multiple new bacteriohopanepolyol isomers in marine sediments. *Org. Geochem.* 124, 12–21. doi: 10.1016/j.orggeochem.2018.07.010
- Lé, S., Josse, J., and Husson, F. (2008). FactoMineR: an R package for multivariate analysis. *J. Stat. Softw.* 25, 1–18. doi: 10.18637/jss.v025.i01
- Martinez-Cruz, K., Sepulveda-Jauregui, A., Walter Anthony, K., and Thalasso, F. (2015). Geographic and seasonal variation of dissolved methane and aerobic methane oxidation in Alaskan lakes. *Biogeosciences* 12, 4595–4606. doi: 10.5194/bg-12-4595-2015
- Moreau, R. A., Powell, M. J., Osman, S. F., Whitaker, B. D., Fett, W. F., Roth, L., et al. (1995). Analysis of intact Hopanoids and other lipids from the bacterium *Zymomonas Mobilis* by high-performance liquid chromatography. *Anal. Biochem.* 224, 293–301. doi: 10.1006/abio.1995.1043
- Mustakhimov, I., Kalyuzhnaya, M. G., Lidstrom, M. E., and Chistoserdova, L. (2013). Insights into denitrification in *Methylobacterium mobilis* from denitrification pathway and methanol metabolism mutants. *J. Bacteriol.* 195, 2207–2211. doi: 10.1128/JB.00069-13
- Neunlist, S., and Rohmer, M. (1985). Novel hopanoids from the methylotrophic bacteria *Methylococcus capsulatus* and *Methylomonas methanica*. (22S)-35-aminobacteriohopane-30, 31, 32, 33, 34-pentol and (22S)-35-amino-3 β -methylbacteriohopane-30, 31, 32, 33, 34-pentol. *Biochem. J.* 231, 635–639. doi: 10.1042/bj2310635
- Newman, D. K., Neubauer, C., Ricci, J. N., Wu, C.-H., and Pearson, A. (2016). Cellular and molecular biological approaches to interpreting ancient biomarkers. *Annu. Rev. Earth Planet. Sci.* 44, 493–522. doi: 10.1146/annurev-earth-050212-123958
- Nowicka, B., and Kruk, J. (2010). Occurrence, biosynthesis and function of isoprenoid quinones. *Biochim. Biophys. Acta Bioenerg.* 1797, 1587–1605. doi: 10.1016/j.bbabi.2010.06.007
- Oksanen, J., Simpson, G. L., Blanchet, F. G., Kindt, R., Legendre, P., Minchin, P. R., et al. (2022). Vegan: Community ecology package. Available at: <https://cran.r-project.org/web/packages/vegan/index.html> (Accessed January 29, 2024).
- Osborne, K. A. (2015). Environmental controls on bacteriohopanepolyol signatures in estuarine sediments. Newcastle, UK: Newcastle University.
- Osborne, K. A., Gray, N. D., Sherry, A., Leary, P., Mejeha, O., Bischoff, J., et al. (2017). Methanotroph-derived bacteriohopanepolyol signatures as a function of temperature related growth, survival, cell death and preservation in the geological record. *Environ. Microbiol. Rep.* 9, 492–500. doi: 10.1111/1758-2229.12570
- Oshkin, I. Y., Belova, S. E., Danilova, O. V., Miroshnikov, K. K., Rijpstra, W. I. C., Sinninghe Damsté, J. S., et al. (2016). *Methylovulum psychrotolerans* sp. nov., a cold-adapted methanotroph from low-temperature terrestrial environments, and emended description of the genus *Methylovulum*. *Int. J. Syst. Evol. Microbiol.* 66, 2417–2423. doi: 10.1099/ijsem.0.001046
- Parada, A. E., Needham, D. M., and Fuhrman, J. A. (2016). Every base matters: assessing small subunit rRNA primers for marine microbiomes with mock communities, time series and global field samples. *Environ. Microbiol.* 18, 1403–1414. doi: 10.1111/1462-2920.13023
- Pol, A., Heijmans, K., Harhangi, H. R., Tedesco, D., Jetten, M. S. M., and Op den Camp, H. J. M. (2007). Methanotrophy below pH 1 by a new Verrucomicrobia species. *Nature* 450, 874–878. doi: 10.1038/nature06222
- Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., et al. (2013). The SILVA ribosomal RNA gene database project: improved data processing and web-based tools. *Nucleic Acids Res.* 41, D590–D596. doi: 10.1093/nar/gks1219
- R Core Team (2023). R: A language and environment for statistical computing. Available at: <https://www.r-project.org/> (Accessed January 29, 2024).
- Raghoebarsing, A. A., Pol, A., Van De Pas-Schoonen, K. T., Smolders, A. J. P., Ettwig, K. F., Rijpstra, W. I. C., et al. (2006). A microbial consortium couples anaerobic methane oxidation to denitrification. *Nature* 440, 918–921. doi: 10.1038/nature04617
- Richter, N., Hopmans, E. C., Mitrović, D., Raposeiro, P. M., Gonçalves, V., Costa, A. C., et al. (2023). Distributions of bacteriohopanepolyols in lakes and coastal lagoons of the Azores archipelago. *Biogeosciences* 20, 2065–2098. doi: 10.5194/bg-20-2065-2023
- Rohmer, M., Bouvier-Nave, P., and Ourisson, G. (1984). (1984). Distribution of Hopanoid triterpenes in prokaryotes. *Microbiology* 130, 1137–1150. doi: 10.1099/00221287-130-5-1137
- Rush, D., Osborne, K. A., Birgel, D., Kappler, A., Hirayama, H., Peckmann, J., et al. (2016). The Bacteriohopanepolyol inventory of novel aerobic methane Oxidising Bacteria reveals new biomarker signatures of aerobic Methanotrophy in marine systems. *PLoS One* 11:e0165635. doi: 10.1371/journal.pone.0165635
- Sáenz, J. P. (2010). Hopanoid enrichment in a detergent resistant membrane fraction of *Crocospaera watsonii*: implications for bacterial lipid raft formation. *Org. Geochem.* 41, 853–856. doi: 10.1016/j.orggeochem.2010.05.005
- Saitou, K., Nagasaki, K., Yamakawa, H., Hu, H.-Y., Fujie, K., and Katayama, A. (1999). Linear relation between the amount of respiratory quinones and the microbial biomass in soil. *Soil Sci. Plant Nutr.* 45, 775–778. doi: 10.1080/00380768.1999.10415843
- Schmerk, C. L., Bernards, M. A., and Valvano, M. A. (2011). Hopanoid production is required for low-pH tolerance, antimicrobial resistance, and motility in *Burkholderia cenocepacia*. *J. Bacteriol.* 193, 6712–6723. doi: 10.1128/JB.05979-11
- Schulenberg-Schell, H., Neuss, B., and Sahn, H. (1989). Quantitative determination of various hopanoids in microorganisms. *Anal. Biochem.* 181, 120–124. doi: 10.1016/0003-2697(89)90403-X
- Seel, W., Flegler, A., Zunabovic-Pichler, M., and Lipski, A. (2018). Increased isoprenoid Quinone concentration modulates membrane fluidity in *Listeria monocytogenes* at low growth temperatures. *J. Bacteriol.* 200:e00148-18. doi: 10.1128/JB.00148-18
- Sévin, D. C., and Sauer, U. (2014). Ubiquinone accumulation improves osmotic-stress tolerance in *Escherichia coli*. *Nat. Chem. Biol.* 10, 266–272. doi: 10.1038/nchembio.1437
- Sherry, A., Osborne, K. A., Sidgwick, F. R., Gray, N. D., and Talbot, H. M. (2016). A temperate river estuary is a sink for methanotrophs adapted to extremes of pH, temperature and salinity. *Environ. Microbiol. Rep.* 8, 122–131. doi: 10.1111/1758-2229.12359
- Søballe, B., and Poole, R. K. (1999). Microbial ubiquinones: multiple roles in respiration, gene regulation and oxidative stress management. *Microbiology* 145, 1817–1830. doi: 10.1099/13500872-145-8-1817
- Talbot, H. M., and Farrimond, P. (2007). Bacterial populations recorded in diverse sedimentary biohopanoid distributions. *Org. Geochem.* 38, 1212–1225. doi: 10.1016/j.orggeochem.2007.04.006
- Talbot, H. M., Handley, L., Spencer-Jones, C. L., Dinga, B. J., Schefuß, E., Mann, P. J., et al. (2014). Variability in aerobic methane oxidation over the past 1.2Myrs recorded in microbial biomarker signatures from Congo fan sediments. *Geochim. Cosmochim. Acta* 133, 387–401. doi: 10.1016/j.gca.2014.02.035
- Talbot, H. M., Sidgwick, F. R., Bischoff, J., Osborne, K. A., Rush, D., Sherry, A., et al. (2016). Analysis of non-derivatised bacteriohopanepolyols by ultrahigh-performance liquid chromatography/tandem mass spectrometry. *Rapid Commun. Mass Spectrom.* 30, 2087–2098. doi: 10.1002/rcm.7696
- Talbot, H. M., Watson, D. F., Murrell, J. C., Carter, J. F., and Farrimond, P. (2001). Analysis of intact bacteriohopanepolyols from methanotrophic bacteria by reversed-phase high-performance liquid chromatography-atmospheric pressure chemical ionisation mass spectrometry. *J. Chromatogr. A* 921, 175–185. doi: 10.1016/S0021-9673(01)00871-8
- Talbot, H. M., Watson, D. F., Pearson, E. J., and Farrimond, P. (2003). Diverse biohopanoid compositions of non-marine sediments. *Org. Geochem.* 34, 1353–1371. doi: 10.1016/S0146-6380(03)00159-1
- Taylor, C. J., Anderson, A. J., and Wilkinson, S. G. (1998). Phenotypic variation of lipid composition in *Burkholderia cepacia*: a response to increased growth temperature is a greater content of 2-hydroxy acids in phosphatidylethanolamine and ornithine amide lipid. *Microbiology* 144, 1737–1745. doi: 10.1099/00221287-144-7-1737
- van Grinsven, S., Sinninghe Damsté, J. S., Harrison, J., and Villanueva, L. (2020). Impact of Electron acceptor availability on methane-influenced microorganisms in an enrichment culture obtained from a stratified Lake. *Front. Microbiol.* 11:715. doi: 10.3389/fmicb.2020.00715
- van Winden, J. F., Talbot, H. M., Kip, N., Reichart, G.-J., Pol, A., McNamara, N. P., et al. (2012). Bacteriohopanepolyol signatures as markers for methanotrophic bacteria in peat moss. *Geochim. Cosmochim. Acta* 77, 52–61. doi: 10.1016/j.gca.2011.10.026
- van Winden, J. F., Talbot, H. M., Reichart, G.-J., McNamara, N. P., Benthien, A., and Sinninghe Damsté, J. S. (2020). Influence of temperature on the $\delta^{13}\text{C}$ values and distribution of methanotroph-related hopanoids in Sphagnum-dominated peat bogs. *Biogeology* 18, 497–507. doi: 10.1111/gbi.12389
- Vences-Guzmán, M. Á., Guan, Z., Ormeño-Orrillo, E., González-Silva, N., López-Lara, I. M., Martínez-Romero, E., et al. (2011). Hydroxylated ornithine lipids increase stress tolerance in *Rhizobium tropici* CIAT899. *Mol. Microbiol.* 79, 1496–1514. doi: 10.1111/j.1365-2958.2011.07535.x
- Welander, P. V., Hunter, R. C., Zhang, L., Sessions, A. L., Summons, R. E., and Newman, D. K. (2009). Hopanoids play a role in membrane integrity and pH homeostasis in *Rhodospseudomonas palustris* TTE-1. *J. Bacteriol.* 191, 6145–6156. doi: 10.1128/JB.00460-09
- Whittenbury, R., Phillips, K. C., and Wilkinson, J. F. (1970). Enrichment, isolation and some properties of methane-utilizing Bacteria. *J. Gen. Microbiol.* 61, 205–218. doi: 10.1099/00221287-61-2-205
- Wickham, H., Chang, W., Henry, L., Pedersen, T. L., Takahashi, K., Wilke, C., et al. (2023). ggplot2: Create elegant data Visualisations using the grammar of graphics. Available at: <https://cloud.r-project.org/web/packages/ggplot2/index.html> (Accessed January 29, 2024).
- Wu, C.-H., Bialecka-Fornal, M., and Newman, D. K. (2015). Methylation at the C-2 position of hopanoids increases rigidity in native bacterial membranes. *eLife* 4:e05663. doi: 10.7554/eLife.05663
- Yilmaz, P., Parfrey, L. W., Yarza, P., Gerken, J., Pruesse, E., Quast, C., et al. (2014). The SILVA and “all-species living tree project (LTP)” taxonomic frameworks. *Nucleic Acids Res.* 42, D643–D648. doi: 10.1093/nar/gkt1209
- Yu, Z., and Chistoserdova, L. (2017). Communal metabolism of methane and the rare earth element switch. *J. Bacteriol.* 199, e00328–e00317. doi: 10.1128/JB.00328-17
- Zhang, J., Kobert, K., Flouri, T., and Stamatakis, A. (2014). PEAR: a fast and accurate Illumina paired-end reAd mergeR. *Bioinformatics* 30, 614–620. doi: 10.1093/bioinformatics/btt593