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Differential responses of canonical nitrifiers and comammox *Nitrospira* to long-term fertilization in an Alfisol of Northeast China

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The newly identified complete ammonia oxidizer (comammox) that converts ammonia directly into nitrate has redefined the long-held paradigm of two-step nitrification mediated by two distinct groups of nitrifiers. However, exploration of the niche differentiation of canonical nitrifiers and comammox Nitrospira and their ecological importance in agroecosystems is still limited. Here, we adopted quantitative PCR (qPCR) and Illumina MiSeq sequencing to investigate the effects of five long-term fertilization regimes in the variations of ammonia-oxidizing bacteria (AOB), ammoniaoxidizing archaea (AOA), nitrite-oxidizing bacteria (NOB), and comammox Nitrospira abundances and comammox community composition in two soil layers (0-20cm,topsoil; 20–40cm, subsoil) in an Alfisol in Northeast China. The fertilization treatments included no fertilizer (CK); chemical nitrogen (N) fertilizer; chemical N; phosphorus (P) and potassium (K) fertilizers (NPK); recycled organic manure (M) and chemical N, P, K plus recycled manure (MNPK). Compared with CK, manure and/or chemical fertilizer significantly increased the AOB amoA gene abundance. Long-term recycled manure increased soil organic matter (SOM) contents and maintained the soil pH, but decreased the NH₄⁺-N concentrations, which markedly promoted the nxrA and nxrB gene abundances of NOB and the amoA gene abundances of comammox Nitrospira clade A and AOA. Although the comammox Nitrospira clade B abundance tended to decrease after fertilization, the structural equation modeling analysis showed that comammox clade B had direct positive impacts on soil potential ammonia oxidation (PAO; λ =0.59, p<0.001). The long-term fertilization regime altered the community composition of comammox Nitrospira. Additionally, comammox Nitrospira clades A and B had individual response patterns to the soil layer. The relative abundance of clade A was predominant in the topsoil in the N (86.5%) and MNPK (76.4%) treatments, while clade B appeared to be dominant in the subsoil (from 78.7 to 88.1%) with lower ammonium contents, implying niche separation between these clades. Soil pH, NH4⁺-N and SOM content were crucial factors shaping the soil nitrifying microbial abundances and the comammox Nitrospira community. Together, these findings expand the current understanding of the niche specialization and the important role of comammox Nitrospira in terrestrial ecosystems.

KEYWORDS

comammox *Nitrospira*, recycled manure, soil layer, phylogenetic analysis, structural equation modeling analysis, niche differentiation

1. Introduction

In nature, microbial-mediated nitrification processes play a crucial role in the transformation of nitrogen (N), resulting in nitrogen flux from terrestrial ecosystems and water eutrophication resulting from nitrate leaching (Hu et al., 2014; Kuypers et al., 2018). The biological oxidation of ammonia (NH₃) to nitrate (NO₃⁻) *via* nitrite (NO₂⁻), known as classical two-step nitrification, is an important process in the global nitrogen cycle. The first step of nitrification, namely ammonia oxidation, is catalyzed by ammonia-oxidizing bacteria (AOB) and ammonia-oxidizing archaea (AOA; Könneke et al., 2005). The second step, namely, nitrite oxidation, is performed by nitrite-oxidizing bacteria (NOB). In 2015, the complete ammonia oxidation (comammox) process was reported to oxidize at he long-held perspective of the nitrification process (Daims et al., 2015; van Kessel et al., 2015).

To date, all of the comammox *Nitrospira* identified belong to *Nitrospira* sublineage II, harboring an entire set of nitrifying genes including ammonia monooxygenase (AMO), hydroxylamine dehydrogenase (HAO) and nitrite oxidoreductase (NXR; van Kessel et al., 2015; Poghosyan et al., 2019). Metagenomic evidence suggests that the comammox *Nitrospira amoA* gene is phylogenetically different from that of canonical ammonia-oxidizing microorganisms (AOA and AOB). Subsequently, based on the phylogeny of the *amoA* gene, comammox *Nitrospira* has been classified into two separate clades including clade A and clade B (Daims et al., 2015), and the average nucleotide identity values between the genomes of clade A and clade B are in the range of 70 to 72% (Poghosyan et al., 2019). Moreover, clade A can be split into clades A1, A2, and A3 (Xia et al., 2018; Li et al., 2020).

Since the discovery of comammox Nitrospira as a novel ammoniaoxidizing member, the significant importance of different ammonia oxidizers to nitrification has become more complicated and diversified under various soil conditions. Theoretical calculations indicated that comammox Nitrospira prefer to exist in a low-substrate-concentration environment, and kinetic results confirmed that comammox Nitrospira had high affinity for the substrates ammonia and oxygen, demonstrating competitive advantages over canonical nitrifiers in oligotrophic habitats (Kits et al., 2017). However, Li et al. (2019) demonstrated that the addition of (NH₄)₂SO₄ significantly stimulated the amoA gene abundance of Nitrospira clade A in a pasture soil based on a microcosm experiment. Gonzalez et al. (2016) and Palomo et al. (2016) analyzed comammox Nitrospira bacterial genome data and predicted that comammox Nitrospira might have versatile metabolization capability. Numerous metagenomic findings further verified that comammox Nitrospira could competently utilize urea, formate, hydrogen, cyanate, methane, and agmatine as well as degrade organic carbon, highlighting their metabolic flexibility beyond the forms of nitrogen (Palomo et al., 2018; Koch et al., 2019; Vijayan et al., 2020; Yang et al., 2020). In addition to ammonia oxidation, comammox Nitrospira are also capable of oxidizing nitrite, because nitrite oxidoreductase is conserved and highly similar in all Nitrospira genomes (Palomo et al., 2018). It is speculated that the nitrite affinity of Nitrospira inopinata is approximately 50 times higher than that of canonical Nitrospira (Kits et al., 2017). The unique feature of comammox Nitrospira genome seems to be the apparent absence of genes for assimilatory nitrite reduction, resulting in the loss of the potential to use external nitrite nitrogen sources (Palomo et al., 2018). Overall, it is essential to bridge our knowledge gap concerning the ecological characteristics of comammox Nitrospira and re-evaluate the niche specialization of different soil nitrifying groups in terrestrial ecosystems.

Chemical and manure fertilizers have been extensively applied in agricultural soils to maintain soil fertility and improve crop productivity. Long-term experiments with different fertilization regimes are crucial for examining the variations in soil properties and nutrient supply capacity, and they can be used to explore the mechanisms of numerous ecological processes in detail. Many previous studies pointed out that long-term application of different fertilization regimes can significantly impact the abundances and community compositions of nitrogenrelated microorganisms, influencing the contributions and predominance of diverse nitrifier and denitrifier communities involved in nitrogen cycling in soils (Zhang et al., 2012; Gao et al., 2018; Hou et al., 2018). Wang et al. (2019) demonstrated that long-term N and phosphorus (P) fertilizer inputs significantly enhanced AOB amoA gene abundances, while AOA amoA gene abundances were not as sensitive as those of AOB. AOA and AOB exhibited higher amoA gene transcriptional abundances than comammox Nitrospira under longterm chemical nitrogen fertilizer treatment (Wang et al., 2021). In comparison, long-term manure fertilization dramatically increased comammox Nitrospira abundances and influenced their community compositions (Li et al., 2020). Manure fertilizer was also shown to improve the composition and diversity of Nitrospira and Nitrobacter NOB in an acidic red soil (Han et al., 2018). Nevertheless, to our knowledge, although various studies have been conducted to illustrate the changes in nitrifying microbial communities in response to manure application, most of these manure fertilizers were commercial manure or were from external sources, not from within the farmland ecosystem. In fact, manure fertilizers from different sources vary greatly in their nutrient and chemical contents. Importantly, external manure fertilizers in general contain some undesired substances, such as heavy metals or antibiotics, that are detrimental to agroecosystems. Thus, it is encouraged to apply organic manure from local agricultural ecosystems, which is assumed to be a sustainable strategy to maintain agricultural production. However, how local manure fertilizer or the combination of chemical and local manure fertilizer affects the soil nitrifying community is still poorly understood. The soil layer is another factor that shapes the nitrifier community due to the differences in nutrient and oxygen availability (Kaveney et al., 2020). The niches of canonical nitrifiers and the newly discovered comammox Nitrospira under different fertilization regimes and soil layers have not been extensively explored. The relationship between nitrifying microorganism abundances, community compositions and environmental factors remains to be determined.

The objectives of our study were to assess the impacts of a long-term field trial of different fertilization regimes (including chemical fertilizer, recycled manure fertilizer and chemical fertilizer combined with recycled manure fertilizer) and soil layers on the abundance, community composition and diversity of canonical nitrifiers and comammox Nitrospira in an Alfisol. We hypothesized that: (1) comammox Nitrospira are responsive to the fertilization regime and play active roles in fertilized soils due to their extensive metabolic versatility; (2) soil pH, NH4+-N and soil organic matter (SOM) contents are the most influential factors governing the niche specialization of nitrifying guilds and their relative importance to nitrification in soils; and (3) comammox Nitrospira clade A and clade B dominate niches in the 0-20 cm and 20-40 cm fertilized soils, respectively, due to their different ammonia affinities and different nitrogen uptake systems. The results of this study will improve our understanding of the significance of canonical nitrifiers and comammox Nitrospira in agricultural ecosystems and provide a

robust basis for regulation strategies to improve fertilizer use efficiency in agricultural production.

2. Materials and methods

2.1. Site description and experimental design

The experimental site is located at Shenyang experimental station, Chinese Academy of Sciences, Liaoning Province, China (42°32'N, 123°23'E). The site has a temperate semihumid continental climate. The average annual temperature is 7.5°C and the mean precipitation is approximately 680 mm (Ma et al., 2020). The soil in this area is classified as an Alfisol, which is a representative soil type on the slope and piedmont plain of the Liaodong Peninsula. The formation of Alfisols is related to the soil alluvial process. Carbonates and soluble salts are leached from soil, so the leaching effect of the soil is obvious and there are many obvious rust spots in the soil profile. The soil-forming parent material is dominated by moderately acidic bedrock weathering and other non-calcareous sediments, resulting in slightly acidic to neutral soil. Due to the effects of the soil-forming process and soil parent material, the organic matter and nutrient contents in the soil are relatively high. The initial soil pH and the organic carbon, total N, P, and potassium (K) contents were 6.15, 20.9 g·kg⁻¹, 1.13 g·kg⁻¹, 0.44 g·kg⁻¹, and 16.4 g·kg⁻¹, respectively (Ning et al., 2020).

This long-term fertilization experiment consisting of 12 treatments (with three replicated plots) was established in 1990 with completely randomized block design. The area of each field plot was 162 m² (9×18 m). Five treatments were chosen in our study, i.e., no fertilizer (CK), chemical N fertilizer (N), chemical N, P, and K fertilizers (NPK), recycled organic manure (M) and a combined application of chemical N, P, K and recycled manure (MNPK). Maize monoculture has been implemented in each treatment with three replicates since 2012. The input rates of the N, P and K fertilizers for maize were 150 kg·N·ha⁻¹, 25 kg·P·ha⁻¹, and 60 kg·K·ha⁻¹ when respective urea, triple superphosphate and potassium chloride fertilizers were applied. One distinctive trait of this long-term experiment was that the treatments contained recycled organic manure. The recycled organic manure used in the M and MNPK treatments was from the internal nutrient recycling process under the corresponding treatments. Detailed information about the preparation of the recycled manure has been presented in previous studies (Ma et al., 2020; Ning et al., 2020). Because the application amounts of recycled manure were determined by the crop production of the previous year under the corresponding treatment, the application amounts of the recycled manure changed among different years and treatments. The input rates for each treatment in 2018 are listed in Table 1. Detailed information about the historical cultivation system, nutrient internal recycling model and fertilizer application rates is provided in the Supplementary material.

2.2. Soil sampling and analysis of soil properties

Soil samples were collected on October 19, 2018 (maize was seeded on April 1st and harvested on October 18) using stainless steel samplers. It each subplot, two soil depths were sampled corresponding to the topsoil (T, 0–20 cm) and subsoil (S, 20–40 cm). We collected 30 soil samples in total and transferred them to the

laboratory on ice within 2 days. One portion of the soils was sieved using 2-mm mesh and was stored at 4°C before physicochemical property analysis and another portion was stored at-80°C before DNA extraction and subsequent high-throughput sequencing analyses.

Soil pH was determined using a pH meter (1:2.5 soil/deionized water; Mettler-Toledo FiveEasy Plus, United States). The SOM content determined using the K₂CrO₇ oxidation-reduction was spectrophotometric method (Bao, 2000). The soil water content (SWC) was determined based on the weight loss when 2 g of fresh soil was dried to a constant weight at 105°C. Soil total P (TP) was digested with H₂SO₄ and HClO₄ solution using 0.5 g of soil, and soil available P was extracted with 100 ml of 0.5 M NaHCO3 at a 1:20 ratio of soil to solution. A flame atomic absorption spectrophotometer (PerkinElmer 900 T-PinAAcle, MA, United States) was used to analyze the soil available K extracted with ammonium acetate. The available N in soils was determined using the Kjeldahl digestion method. The soil inorganic $\mathrm{NO_3^-}\text{-}\mathrm{N}$ and $\mathrm{NH_4^+}\text{-}\mathrm{N}$ contents were extracted with 2M KCl solution and measured with a continuous flow analyzer (Futura, Alliance Instruments, Pairs, France; Bao, 2000). Soil microbial biomass C and N (SMBC and SMBN) were measured using the chloroform fumigation extraction method with minor modifications as described by Wu et al. (1990). After the soil was grinded and sieved through 100-mesh sieves, the soil total carbon (TC) and total N (TN) contents were measured by dry combustion using a Vario Max Elemental CNS analyzer (Elementar Analysesysteme GmbH, Germany).

2.3. Soil ammonia and nitrite oxidation potentials

Soil potential ammonia oxidation (PAO) was assessed using a modification of the method presented by Kurola et al. (2005). Five grams of equivalent dry mass of fresh soil was incubated with 20 ml of 1 mM $(NH_4)_2SO_4$ and $10 \text{ mg}\cdot\text{L}^{-1}$ potassium chlorate solution (to inhibit nitrite oxidation) dissolved in phosphate buffer solution ($0.2 \text{ g}\cdot\text{L}^{-1} \text{ NaH}_2\text{PO}_4$, $0.2 \text{ g}\cdot\text{L}^{-1}\text{Na}_2\text{HPO}_4$, $8.0 \text{ g}\cdot\text{L}^{-1}$ NaCl, $0.2 \text{ g}\cdot\text{L}^{-1}$ KCl, pH 7.4) for 24 h on a rotating shaker at 170 rpm and 25°C. Then, the suspension was extracted with 2 M KCl and centrifuged at 5,000 rpm for 4 min. The NO₂⁻-N content in the filtered solution was measured with a continuous flow analyzer (Futura, Alliance Instruments, Pairs, France).

The soil potential nitrite oxidation (PNO) was measured using the method modified by Attard et al. (2010). In brief, 5.0 g of equivalent dry mass of fresh soil was agitated in 30 ml of $1.67 \text{ mg}\cdot\text{L}^{-1}$ NaNO₂ solution (each gram of dry soil contained $10\,\mu\text{g}$ of NO₂⁻-N) for 15 h on a rotating shaker at 170 rpm and 25°C. The NO₂⁻-N in the suspension was filtered and measured, and the linear decrease rates were considered PNO.

2.4. Soil DNA extraction and quantitative PCR assay

Soil DNA was extracted from approximately 0.5 g of fresh soil using the FastDNA Spin Kit for soil (MP Biomedicals, CA, United States) following the manufacturer's recommendations. The DNA quality was detected by 1% (w/v) agarose gel electrophoresis and the DNA concentrations were evaluated by a NanoDropTM One UV–Vis spectrophotometer (Nanodrop Technologies, Wilmington, United States). TABLE 1 Experimental design and nutrient inputs of selected treatments in 2018.

Fertilizer regimes	N (kg∙ha ^{−1} ∙year ^{−1})	P (kg∙ha ^{−1} ·year ^{−1})	K (kg∙ha ⁻¹ ·year ⁻¹)
СК	0	0	0
Ν	150	0	0
NPK	150	25	60
Mª	(40.1)	(7.24)	(14.8)
MNPK ^b	199.9 (49.9)	41.4 (16.4)	93.0 (33.0)

^aThe amounts of N, P, and K in M treatment are calculated by their contents in the recycled organic manure, which are listed in the parenthesis.

^bThe amounts of N, P, and K in MNPK treatment are calculated by the sum of their contents in the recycled organic manure and inorganic chemical fertilizers, which are listed outside the parenthesis. The amounts of N, P and K in the parenthesis are the parts from the recycled manure.

The primer sets Arch-amoAF/R (Francis et al., 2005), amoA-1F/2R (Rotthauwe et al., 1997), comaA-244f/659r (a-f), and comaB-244f/659f (a-f; Pjevac et al., 2017) were used to quantify the amoA gene abundances of AOA, AOB, comammox Nitrospira clade A and comammox Nitrospira clade B, respectively. The nxrA gene of Nitrobacter-like NOB and the nxrB gene of Nitrospira-like NOB were amplified using the primers F1norA/R2norA (Poly et al., 2008) and nxrB169F/638R (Pester et al., 2014), respectively. Quantitative PCRs (qPCRs) were conducted using a Bio-Rad CFX384 optical real-time PCR system (Bio-Rad Laboratories Inc., Hercules, CA, United States). The 20 µL PCR mixture contained $10\,\mu L$ of $2\,\times\,SYBR$ Premix Ex Taq^TM (TaKaRa Biotechnology Co., Dalian, China), 0.8μ L of each primer (10 μ M), 2 μ l of DNA template (diluted 10-fold) or 2µL of standard plasmid and 6.4µL of dd H2O. Standard curves were constructed using serially diluted plasmids containing the target genes at the final concentrations of 10^8 to 10^1 gene copies· μ L⁻¹. Negative controls containing water as the template were simultaneously analyzed to eliminate any contamination. Primer set sequences and PCR thermal protocols are listed in Supplementary Table S1. All assays were performed in triplicate and the correlation coefficient (R^2) values above 0.99 were accepted. To evaluate the amplification specificity of qPCR, we conducted melting curve analysis at the end of each run. The efficiencies of qPCR were 94, 101, 96, 102, 95, and 92% for AOB, AOA, comammox Nitrospira clade A, comammox Nitrospira clade B, Nitrospira-like NOB and Nitrobacter-like NOB, respectively.

2.5. Comammox *Nitrospira amoA* gene amplicon sequencing and phylogenetic analysis

For the high-throughput sequencing analysis, the comammox Nitrospira amoA gene was first amplified using the primer set ComaA189Y/ComaC576R (Xia et al., 2018). Then the PCR products were used as templates and amplified using the barcoded primer set ComaA209F/ComaC576R (Xia et al., 2018). The amplification protocols of the two rounds of nested-PCR are listed in Supplementary Table S1. After purification (QIAEX Gel Extraction Kit, Qiagen, Germany), the DNA concentrations were quantified using a Quant-iT[™] PicoGreen[™] dsDNA Assay Kit (Thermo Fisher Scientific, MA, United States). The purified PCR amplicons were incorporated together at equimolar concentrations and sequenced on an Illumina MiSeq platform at Majorbio (Majorbio Biopharm Technology Co., Ltd., Shanghai, China). Raw sequences were processed using Quantitative Insights in Microbial Ecology (QIIME; Caporaso et al., 2010) and Mothur (Schloss et al., 2009). Barcodes were used to split raw reads into different samples. High-quality sequences (quality score higher than 25) were retained. The obtained quality-screened sequences were divided into operational taxonomic units (OTUs) based on 97% sequence similarity for the comammox *Nitrospira amoA* gene using USEARCH (version 1.8.0). One representative sequence was chosen from each OTU for subsequent analysis. The 50 most abundant representative sequences from the remaining high-quality sequences and closely related sequences selected from the NCBI BALST database were aligned with the Clustal X program. Phylogenetic trees were constructed with neighbor-joining (NJ) method using MEGA 7.0 (Tamura et al., 2007), and bootstrap analysis was performed using 1,000 replicates. The acquired comammox *Nitrospira* sequences were deposited in GenBank with accession numbers SRR13319769 to SRR13319798.

2.6. Statistical analysis

The mean values and standard errors for each parameter were calculated from triplicate samples. One-way analysis of variance (ANOVA) was performed with SPSS Statistics 20.0 software (IBM, United States) based on Duncan's test and p < 0.05 was considered statistically significant. Spearman's correlation analyses were conducted to investigate the relationships among the soil physiochemical properties, potential nitrifier activities, and nitrifier functional gene abundances. Sobs, Shannon and phylogenetic diversity (PD) indices were calculated in R 3.6.0 using the 'vegan' package. Canonical correspondence analysis (CCA) was conducted to investigate the influence of the soil physiochemical properties on nitrifier functional gene abundances based on qPCR results and the comammox Nitrospira community based on high-throughput sequencing data. Permutational multivariate analysis of variance (PERMANOVA) was performed to explore the significance of the effects of the fertilization regime and soil layer on the nitrifying microbial community composition. The proportions of variations in the comammox Nitrospira community explained by soil properties were calculated by the "permu.hp" command in the "rdacca.hp" package (Lai et al., 2022). The Mantel test was used to identify potential environmental factors contributing to comammox Nitrospira community dissimilarity. Aggregated boosted tree (ABT) analysis was conducted to explore the relative contribution of soil properties to nitrifier gene abundances using the package "gbmplus" in R (Death, 2007). Structural equation modeling (SEM) was conducted with AMOS 21.0 (Amos, Development Corporation, Meadville, PA, United States) to test the hypothetical causal relationships among the fertilization regime, soil properties, soil microbial biomass (SMBC or SMBN), abundances of nitrifiers and soil nitrification potential (potential ammonia oxidation or potential nitrite oxidization). In the SEM analyses, modeling including all soil variables was conducted first to confirm which factors were significantly related to nitrifier abundances and nitrification potential (PAO or PNO) with the

maximum likelihood estimation method. The adequacy of the SEM was determined by the chi-square/degrees of freedom values (CHI/DF), adjusted goodness of fit index (AGFI), root mean square error of approximation (RMSEA) index and Akaike information criteria (AIC). Favorable model fits adopted were suggested by Duan et al. (2018). Finally, we removed some non-significant variables and chosen the soil organic matter, NH_4^+ -N, NO_3^- -N, total N, and available N contents as soil properties for the best model fit.

3. Results

3.1. Soil physicochemical properties

Most of the measured soil physicochemical properties were significantly affected by the fertilization regime and soil layer (p < 0.05; Table 2). Significant decreases in soil pH occurred in the N-treated soils compared with the CK soils, while the application of recycled manure alone had no obvious effect on the soil pH value. Among the 5 fertilization treatments, the contents of SOM, total N (TN), total P (TP), total K (TK), available N, available P, and available K reached their maximum values under the MNPK treatment, which received the highest rates of organic and chemical fertilizers. The addition of recycled manure alone also increased the SOM, soil TN, TK, TP, and available P contents to a certain degree compared with those of the CK and N treatments. Similarly, the concentrations of NO₃⁻-N and NH₄⁺-N were significantly increased in all fertilized treatments in comparison with those of the CK. The highest concentrations of NO₃⁻-N and NH₄⁺-N were found in the topsoil under the MNPK and N treatments, respectively.

Soil microbial biomass C showed a marked decrease after the application of chemical N, and the lowest content of SMBC in the topsoil was found under the N treatment, with a value of 82.48 mg·kg⁻¹. Recycled manure application markedly increased the SMBC content, and the highest content was found under the M treatment. Although the nutrient inputs of the MNPK treatment were the highest, the SMBC under the MNPK treatment was less than that under the M treatment, which showed the negative effects of chemical N fertilizer on the SMBC. A similar pattern was found for SMBN, and the highest amount of SMBN also appeared under the M treatment. There was no significant difference in the C/N ratio of the topsoil among the different treatments. Unsurprisingly, the SOM, TN, TP, available N, available P, available K, NH_4^+ -N and NO_3^- -N contents of the subsoil were significantly lower than those of the topsoil (Table 2). Conversely, SWC and pH in the subsoil were higher than those in the topsoil.

3.2. Soil potential ammonia oxidation and nitrite oxidation activity

Both soil PAO and PNO varied significantly among soil samples from the different fertilization regimes and soil layers (Figure 1; Supplementary Table S2). Generally, due to the higher nutrient and oxygen contents in the surface soil, the soil PAO and PNO of the topsoil were higher than those of the subsoil under the same fertilization regime except for the soil PAO under the N treatment. In the topsoil, the PAO varied from as low as 16.53 (\pm 0.71) to 324.80 (\pm 79.90) ng·NO₂⁻⁻·g⁻¹·h⁻¹ (Figure 1A) and peaked under the NPK treatment, which was 2.6 times higher than that under the CK. The soil PAO under the CK and MNPK treatments was comparable. Unexpectedly, the lowest PAO of the topsoil and the highest PAO of the subsoil both occurred under the N treatment.

The responses of the soil PNO to the fertilizer regime were different from the changes in the soil PAO (Figure 1B). In the topsoil, the CK treatment had the highest PNO of approximately 468.13 (±51.28) ng·NO₂⁻·g⁻¹·h⁻¹, while the N treatment had the lowest PNO of 178.53 (±47.27) ng·NO₂⁻·g⁻¹·h⁻¹. The soil PNO under the recycled manure treatment was slightly lower than that under the CK treatment but was as much as 2.33 times higher than that under the N treatment. The measured soil PNO of the subsoil ranged from 43.07 (±28.28) to 230.47 (±52.38) ng·NO₂⁻·g⁻¹·h⁻¹, which was significantly lower than that of the topsoil.

3.3. Abundances of canonical ammonia oxidizers, comammox *Nitrospira* and nitrite oxidizers

Quantitative PCR (qPCR) analysis was adopted to quantify the amoA gene abundances of AOB, AOA, and comammox Nitrospira clade A and clade B in response to the fertilization regime and soil layer (Figures 2A-D,G). The results showed that the long-term fertilization regime significantly changed the amoA gene abundances of AOA and AOB in the soil (Supplementary Table S2; Figure 2). In the topsoil, the AOA amoA gene abundance significantly increased under the M treatment, but significantly decreased with the application of chemical N fertilizer. In contrast, the AOB amoA gene abundance demonstrated an increasing tendency accompanied by increasing rates of nutrients, and the highest amoA gene abundance of AOB was found under the MNPK treatment, which was as much as 11.88 times greater than that under the CK. Across all of the treatments, the amoA gene abundances of AOA ranged from 2.37×10^8 to 7.86×10^8 copies g⁻¹ dry soil, which were maximum 48.4 times greater than those of AOB, which ranged from 1.48×10^7 to 4.47×10^8 copies g⁻¹ dry soil (Figures 2A,B). Moreover, the *amoA* gene abundances of AOA and AOB displayed opposite responses to different soil depths.

Similar to AOA, the comammox Nitrospira clade A amoA gene abundances were considerably increased by the application of recycled manure (Figure 2C). The highest abundance of comammox Nitrospira clade A (4.65×10^7 copies per grams of dry soil) occurred in the topsoil under the M treatment. Significant differences were detected in the comammox Nitrospira clade A abundance between the two soil layers (Supplementary Table S2), and the abundances of comammox Nitrospira clade A decreased with increasing soil depth under the CK, NPK, M, and MNPK treatments. The abundance of comammox Nitrospira clade B was much higher than that of comammox Nitrospira clade A in all soil treatments, ranging from 9.68×10^7 to 3.69×10^8 copies g⁻¹ dry soil (Figure 2D). Among all treatments, the highest comammox Nitrospira clade B abundance occurred in the subsoil under the N treatment. There was an increasing trend in the comammox Nitrospira clade B abundance with increasing soil depth except in the CK and NPK treatments. Additionally, the total comammox Nitrospira abundances (the sum of the amoA gene abundances of clade A and clade B) ranged from 1.29×10^8 to 3.88×10^8 copies g⁻¹ dry soil, this range was slightly lower than that of AOA (Figure 2G).

The *Nitrobacter*-like *nxrA* gene and *Nitrospira*-like *nxrB* gene in all soil samples ranged from 1.22×10^7 to 5.90×10^7 and 3.20×10^8 to 1.21×10^9 copies g⁻¹ dry soil, respectively (Figures 2E,F). Manure fertilizer markedly increased the NOB abundance in contrast with the effects of chemical N fertilizer. Significantly higher abundances of the *Nitrobacter nxrA* gene and

Soil layer (cm)	Treatment	SWC (%)**	pH**	SOM (g·kg ⁻¹)**	TN (g⋅kg ⁻¹)**	TP (g⋅kg ⁻¹)**	TK (g⋅kg ⁻¹)	AN (mg⋅kg ⁻¹) **	AP (mg∙kg ^{−1})**	AK (mg∙kg ⁻¹)*	NH₄⁺-N (mg⋅kg⁻¹)**	NO₃ [–] -N (mg⋅kg ^{–1})**	SMBC (mg·kg ⁻¹)**	SMBN (mg·kg ⁻¹)**	C/N
0–20	СК	17.55 (0.96)a	6.92 (0.05)a	16.32 (0.89) cd	0.87 (0.01)c	0.37 (0.02)d	20.10 (0.20)c	96.33 (2.52)b	3.33 (0.51)c	72.33 (3.51)bc	9.76 (0.03)d	6.07 (0.02)c	152.43 (3.32)b	21.28 (1.25)b	10.84 (0.56)a
	N	16.84 (4.02)a	5.16 (0.21)d	15.63 (0.60)d	0.90 (0.02)c	0.37 (0.00)d	20.67 (0.32)b	105.00 (2.65) b	2.13 (0.68)c	60.00 (5.29)c	13.56 (0.35)a	9.50 (0.10)b	82.48 (6.93)c	12.70 (1.73)b	10.05 (0.19)a
	NPK	17.45 (1.72)a	6.24 (0.05)b	17.03 (0.29)c	0.95 (0.02)bc	0.50 (0.01)b	20.83 (0.15)ab	103.33 (2.52) b	19.37 (2.73)b	83.33 (4.04)b	10.72 (0.57)c	6.16 (0.13)c	114.64 (18.69) bc	18.34 (4.56)b	10.36 (0.21)a
	М	18.24 (0.96)a	6.90 (0.08)a	19.24 (0.23)b	1.01 (0.05)b	0.41 (0.01)c	21.17 (0.15)a	98.33 (6.03)b	4.07 (0.23)c	75.67 (1.53)b	7.99 (0.33)e	6.12 (0.03)c	216.33 (51.44)a	30.02 (8.74)a	11.05 (0.74)a
	MNPK	17.58 (0.96)a	5.87 (0.19)c	20.40 (0.29)a	1.12 (0.0.9)a	0.60 (0.05)a	20.57 (0.31)b	123.00 (16.37)a	37.67 (1.15)a	103.00 (14.18)a	11.74 (1.00)b	21.82 (0.05)a	122.40 (11.71) bc	15.68 (2.73)b	10.60 (0.77)a
20-40	СК	21.25 (1.26)a	7.02 (0.14)a	10.98 (0.16)b	0.53 (0.03)c	0.36 (0.02)bc	21.30 (0.26)a	72.33 (6.66)a	2.63 (0.60)c	67.67 (7.57)a	5.94 (0.38)c	1.40 (0.09)d	47.08 (4.97)a	6.31 (0.50)a	12.05 (0.92)a
	N	20.78 (1.96)a	6.50 (0.21)b	10.71 (0.08)b	0.67 (0.01)a	0.34 (0.01)c	20.60 (0.20)b	69.33 (3.21)a	1.57 (0.90)c	66.67 (6.03)a	7.17 (0.06)a	4.47 (0.33)a	48.39 (9.25)a	4.01 (1.09)a	9.25 (0.21)c
	NPK	21.13 (1.56)a	6.89 (0.12)a	11.14 (0.32)b	0.67 (0.03)a	0.38 (0.02)ab	20.23 (0.32)b	72.67 (6.81)a	4.87 (1.00)b	72.67 (4.16)a	6.55 (0.15)b	2.38 (0.09)c	36.40 (8.52)a	4.31 (3.56)a	9.62 (0.55) bc
	М	20.92 (0.32)a	7.01 (0.06)a	10.95 (0.40)b	0.61 (0.02)b	0.36 (0.01)bc	20.27 (0.47)b	67.00 (4.58)a	3.13 (0.71)bc	64.67 (6.03)a	5.66 (0.09)c	1.60 (0.29)d	55.73 (9.40)a	7.38 (2.91)a	10.45 (0.58)b
	MNPK	20.93 (1.45)a	6.85 (0.13)a	12.27 (0.72)a	0.67 (0.05)a	0.40 (0.01)a	20.27 (0.06)b	72.67 (5.77)a	8.70 (1.54)a	74.00 (2.65)a	7.39 (0.31)a	3.80 (0.02)b	49.94 (19.45)a	4.87 (1.63)a	10.62 (0.63)b

TABLE 2 Basic characteristics of soil properties under different fertilization regimes and soil layers.

The values in the parentheses present standard deviations (n=3), and the different letters indicate significant differences among different treatments at P<0.05 level by one-way (ANOVA). Significant difference of soil physicochemical properties between topsoil and subsoil were marked with one asterisk (*) suggesting p<0.05 and two asterisks (**) suggesting P<0.01 based on the Paired-sample *t*-test. SWC, soil water content; SOM, soil organic matter; TN, total N; TP, total N; AP, available N; AP, available N; AP, available K; NH₄⁺-N, ammonium N; NO₃⁻-N, nitrate N; SMBC, soil microbial biomass C; SMBN, soil microbial biomass N; C/N, the ratio of total C to total N.



the *Nitrospira nxrB* gene were found in the topsoil under the M treatment, 2.8 and 3.8 times higher, respectively, than those under the N treatment.

3.4. Soil properties influencing the distribution of nitrifier gene abundances

Canonical correspondence analysis (CCA) was conducted to elucidate variances in the nitrifying guild abundances among different treatments and their responses to the soil properties. The first two CCA dimensions explained 75.3% of the cumulative variations and soils were assembled by layers (Figure 3A). The results showed that AOB and *Nitrobacter*-like NOB were more abundant in the topsoil with higher soil nutrient contents and SMBN, while AOA, *Nitrospira*like NOB and comammox *Nitrospira* clade B tended to be more abundant in the subsoil with a higher pH and soil water content.

Aggregated boosted tree analysis was conducted to elucidate the importance of soil properties in predicting the changes in nitrifier gene abundances (Supplementary Figure S1). The results suggested that different soil properties contributed differently to the variations in the nitrifier gene abundances. The two most crucial soil properties driving the AOA abundances were soil pH (24.3%) and NH₄⁺-N (18.3%). SOM content was the most important contributor to the changes in AOB (21.7%), comammox *Nitrospira* clade A (36.2%), *Nitrobacter*-like NOB (37.9%) and *Nitrospira*-like NOB (16.5%) abundances in all samples. The contribution of the soil available P content (17.2%) to the comammox *Nitrospira* clade B abundances was predominant. Overall, among the determined soil physiochemical properties, the soil pH, NH₄⁺-N, soil organic matter, and available P content were the principal factors that shaped the nitrifier abundances in the fertilized soils.

3.5. Structural equation modeling analysis

Spearman's correlation analysis demonstrated that soil PAO was markedly and positively correlated with the abundances of AOB, comammox Nitrospira clade B and total comammox Nitrospira. Additionally, there were significant and positive relationships between the abundances of AOB, comammox Nitrospira clade A, Nitrobacter-like NOB, Nitrospira-like NOB and soil PNO (Supplementary Table S3). Path analysis was conducted to further assess the direct or indirect relationships among the fertilization regime, soil properties, SMBC or SMBN, the abundances of active nitrifiers (AOA, AOB, comammox Nitrospira clade A, comammox Nitrospira clade B, Nitrobacter-like NOB, and Nitrospira-like NOB) and the soil nitrification potential (PAO and PNO) through structural equation pattern (Figure 4). With respect to soil PAO, the fertilization regime directly affected the soil properties (e.g., soil organic matter, $NO_4{}^+\text{-}N,$ TN, and available N) through a positive path ($\lambda\!=\!0.26,$ p > 0.05) and significantly affected SMBN through a negative path $(\lambda = -0.37, p < 0.01;$ Figure 4A). The comammox *Nitrospira* clade B abundance had a direct positive impact on the soil PAO ($\lambda = 0.59$, p < 0.001) and was affected by the negative effects of SMBN ($\lambda = -0.29$, p > 0.05), followed by the positive effects of the soil properties $(\lambda = 0.23, p > 0.05)$. The AOB abundance also had a direct effect on soil PAO ($\lambda = 0.29$, p > 0.05), but the positive effect of AOB on soil PAO was less than that of comammox Nitrospira clade B. Moreover, soil properties ($\lambda = -0.40$, p > 0.05) and SMBN ($\lambda = 0.27$, p > 0.05) were directly related to soil PAO, although the effects were not significant. Overall, 42% of the variability in soil PAO could be explained by the parameters ($R^2 = 0.42$ in Figure 4A). In addition, the results of the standardized total effects revealed that the abundance of comammox Nitrospira clade B had the strongest positive effects on soil PAO (Supplementary Figure S2).

With respect to soil PNO, the fertilization regime directly affected the soil properties (e.g., soil organic matter and NO₃⁻-N) through a positive path (λ =0.36, p<0.05) and significantly affected SMBC through a negative path (λ =-0.44, p<0.001; Figure 4B). Both the soil properties (λ =0.24, p>0.05) and SMBC (λ =0.51, p<0.05) affected soil PNO directly, and they also affected soil PNO indirectly *via* the influence on the nitrifier abundance. The direct relationship between PNO and the abundances of *Nitrospira*-like NOB was positive but not significant (λ =0.24, p>0.05). Overall, the model explained 63% of the variation in



soil PNO (R^2 =0.63, Figure 4B). Furthermore, the standardized total effects showed that the soil properties and SMBC were important factors, as the soil properties had the strongest indirect effects on soil PNO, and SMBC had the strongest direct effect on the PNO (Supplementary Figure S2).

3.6. Phylogeny of comammox Nitrospira

A total of 2,053,861 high-quality sequences were retrieved from the 30 soil samples. According to the cutoff threshold of 97% nucleotide similarity, 246 unique OTUs were identified. A phylogenetic tree



Canonical correspondence analysis (CCA) plot based on the gene abundances of all nitrifying populations (A) and comammox *Nitrospira* OTU compositions (B). Circles, subsoil samples, S; triangles, topsoil samples, T; black arrows, soil physiochemical properties; dark goldenrod arrows, nitrifiers. Red, blue, green, purple, and orange colors represent CK, M, MNPK, N and NPK treatments, respectively. Abbreviations: SWC, soil water content; TK, total K; AN, available N; AP, available P; AK, available K; NH_4^* -N, ammonium N; NO_3^- -N, nitrate N; SMBN, soil microbial biomass N; C/N, the ratio of total C to total N.

containing the 50 major OTU sequences (the proportion>0.05) accounting for more than 98.9% of the total OTU abundances was constructed using the neighbor-joining method with the p-distance model. Consistent with previous observations, the comammox Nitrospira amoA gene phylogenetic tree showed that the major OTU sequences were clustered into two clades, where 18 out of the 50 sequences were affiliated with comammox Nitrospira clade A within clade A2. The clade A2 cluster was affiliated with the metagenomeassembled genome (MAG) species SG-bin1 (17 OTUs) and SG-bin2 (1 OTU). There were 21 OTUs belonging to clade B (Supplementary Figure S3), affiliated with Nitrospira sp. CG24C and Nitrospira sp. CG24A. Notably, some of the selected-OUT clusters (11 out of 50) were close to those of the AOB amoA sequences due to the nonspecific amplification of the selected comammox Nitrospira amoA gene primer set and partial nested PCR protocol (Supplementary Figure S3). Among the 50 most abundant OTUs, comammox Nitrospira clade B (42%) was the dominant cluster compared with clade A (36%; Supplementary Figure S3).

Cluster analysis showed that the distribution of OTUs belonging to clade A and clade B varied obviously among the different soil treatments (Figure 5A). The topsoil under the CK and M treatments clustered closely with the subsoil. Meanwhile, the relative abundances of clade A and clade B responded differently to the fertilization regime and soil layer (Figure 5B). Clade A was abundant in the topsoil and dominated under the N and MNPK treatments, making up 86.5 and 76.4%,

respectively, of the selected OTUs. In contrast, although clade B seemed to be dominant in the topsoil under the CK (74.3%) and M (56.6%) treatments, the relative abundance of clade B was higher in the low nutrient-replenished subsoil (ranging from 78.7 to 88.1%), indicating niche differentiation between comammox *Nitrospira* clade A and clade B. The highest proportion was found in the subsoil under CK (88.1%), suggesting the preference of clade B in oligotrophic environments.

3.7. Soil properties shaping the community of comammox *Nitrospira*

Canonical correspondence analysis was performed to estimate the main soil properties affecting the community structure of comammox Nitrospira under different treatments, and the first two CCA dimensions explained 47.0% of the cumulative variation in the comammox Nitrospira community compositions (Figure 3B). Notably, pH, SMBN and available N explained 27.1, 17.7 and 14.5%, respectively, of the whole community dissimilarity (p < 0.01), acting as the three most critical factors. Soil samples from the subsoil were closely grouped and obviously differed those from the topsoil. PERMANOVA further demonstrated that the comammox Nitrospira community composition not only significantly diverged with the soil layer ($R^2 = 0.41$, p = 0.001), but also with the fertilization regime ($R^2 = 0.30$, p = 0.001). The compositions of comammox Nitrospira in the topsoil were divided into three groups depending on the fertilization regime. The ranking of the Bray-Curtis distances revealed that there were obvious differences in the comammox Nitrospira communities among the different fertilization treatments in the topsoil (Figure 3B; Supplementary Figure S4). Pearson correlation analysis revealed that the Sobs index of the comammox Nitrospira community (alpha diversity) showed significant positive correlations with pH (r=0.44, p<0.05) and the soil water content (r=0.42, p<0.05), but showed marked negative correlations with available N (r = -0.37, p < 0.05) and NH₄⁺-N (r = -0.37, p < 0.01; Table 3). The Mantel test revealed that the dissimilarity in the beta diversity of comammox Nitrospira composition was significantly correlated with pH and the soil organic matter, TN, available N, NH4+-N and NO3--N contents (Table 3). Considering the influence of soil properties on the abundance and composition of comammox, we confirmed that soil pH and the NH4+-N and soil organic matter contents were the most crucial factors driving the comammox Nitrospira community compositions.

4. Discussion

With the discovery and prevailing occurrence of comammox *Nitrospira* in engineered, coastal and terrestrial environments (Li et al., 2022; Zhu et al., 2022), our previous knowledge about the niche differentiation of different nitrifiers and their relative contribution to nitrification in agricultural ecosystems need to be reconsidered. Distinct fertilization effects on canonical nitrifier abundance and community structure have been reported in many previous studies (He et al., 2007; Wang et al., 2019; Lin et al., 2020; Cai et al., 2021), and our study further confirmed this information. The abundance of AOB was enhanced by the application of chemical N fertilizer, while the changes in the abundance of AOA displayed an opposite pattern, suggesting that the addition of chemical fertilizer created an unfavorable condition for AOA (Figure 2). Compared with chemical fertilizer, recycled manure fertilizer significantly increased the



FIGURE 4

Path diagrams demonstrating the potential causal relationships among the fertilization regime, soil properties (SOM, NO₃⁻⁻N, NO₄⁺⁻N, TN and available N), soil microbial biomass (C and N), the abundances of nitrifiers (AOA, AOB, comammox *Nitrospira* clade A, comammox *Nitrospira* clade A, *Nitrobacter*-like NOB and *Nitrospira* like NOB) and soil PAO (A) or PNO (B). The width of the arrows in proportional to the strength of the potential causal effects adjacent with path coefficients. The red solid and blue dashed lines indicate the positive and negative relationships between the indicators, respectively. The *R*² demotes the percentage of variance explained by the corresponding indicator in the structural equation modeling. Significant levels are indicated as follows: **p* <0.05, ***p* <0.01, and ****p* <0.001. SOM, soil organic matter; TN, total N; AN, available N; NH₄⁺-N, ammonium N; NO₃⁻-N, nitrate N; SMBC, soil microbial biomass C; SMBN, soil microbial N.

abundance of AOA (Figure 2). These distinct responses of AOB and AOA to fertilization regimes can be explained by their different affinities for ammonia substrates (Prosser and Nicol, 2012; Ouyang et al., 2018). The theoretical ammonia concentration in the soil aqueous solution based on the ionization equilibrium ranged from 400 to 14,270 nM (Supplementary Table S4), which facilitated the growth of both AOA and AOB with optimal survival ammonia concentrations ranging from 0.007 to 1,220 nM and 6,730 to 323,000 nM, respectively (He et al., 2012; Hu et al., 2014). Obviously, AOB are better adapted to soils with relatively high ammonia concentrations due to their higher maximum activity and half-saturation constant value for ammonia than AOA (Hu et al., 2014; Li et al., 2022). ABT analysis demonstrated that AOB abundance was positively correlated with the soil TN and NH₄⁺-N contents, while



AOA abundance was significantly and negatively correlated with the NH₄⁺-N content (Supplementary Figure S1). Thus, AOB were more abundant in the 0-20 cm layer of fertilized soils and peaked under the MNPK treatment receiving the highest fertilizer inputs, while AOA preferred oligotrophic environments such as unfertilized soils and subsoils (Figure 2). Notably, long-term manure recycling significantly boosted the AOA abundance. Since manure only contained a small percentage of N in urea or inorganic N form, most of the N was organically derived (Sun et al., 2021), and the negative effect of manure on the AOA abundance was slight. Some AOA species (e.g., Nitrosophaera viennensis and Nitrosopumilus martimus SCM1) are considered to be mixotrophic and possess genes encoding enzymes for assimilating organic compounds (Tourna et al., 2011). Additionally, another study supported the idea that some AOA can use certain organics to act as chemical scavengers to detoxify intracellular H₂O₂ rather than assimilate organics into their membrane lipids (Kim et al., 2016). Therefore, the AOA abundance was significantly stimulated by recycled manure application.

Similarly, *Nitrobacter*-like and *Nitrospira*-like NOB have different lifestyle strategies. *Nitrobacter*-like NOB are known as *r*-strategists, preferring higher N availability and oxygen concentrations (Nowka et al., 2014), whereas *Nitrospira*-like NOB are *K*-strategists, dominating in N-limited environments with higher N substrate affinity (Schramm et al.,

TABLE 3 Correlations between soil properties and comammox *Nitrospira* community diversity in fertilized soils.

Factors	Comammox Nitrospira community								
		α -Diversity	β -Diversity						
	(Pea	rson correla	(Mantel test)						
	Sobs	Shannon	PD	Jaccard	Bray- Curtis				
SWC	0.42*	-0.12	0.56**	0.40**	0.35**				
рН	0.44*	0.31	0.41*	0.41**	0.54**				
SOM	-0.27	0.34	-0.76**	0.38**	0.66**				
TN	-0.34	0.25	-0.78**	0.34**	0.62**				
TP	0.06	0.24	-0.52**	0.17	0.36**				
TK	-0.14	0.07	-0.02	0.01	0.01				
AN	-0.37*	0.15	-0.71**	0.34**	0.63**				
AP	0.13	0.15	-0.48**	0.15	0.29**				
AK	0.15	0.31	-0.39*	0.21	0.33**				
NH4+-N	-0.53**	-0.10	-0.67**	0.46**	0.69**				
NO3 ⁻ -N	-0.12	0.02	-0.57**	0.30*	0.41**				
SMBC	-0.17	0.49**	-0.55**	0.19*	0.49**				
SMBN	-0.22	0.45*	-0.56**	0.20	0.43**				
C/N	0.19	0.28	0.06	0.06	-0.05				

Significant correlations (p < 0.05) are indicated in bold. *p < 0.05, **p < 0.01. Sobs, the observed richness described the community richness; Shannon, the Shannon diversity index described the community diversity; PD, the phylogenetic diversity described the community diversity; Jaccard, transformed data by Jaccard distance; Bray-Curtis, transformed data by Bray-Curtis distance.

1999). Our results supported the idea that the niche differentiation of these two genera of NOB showed that *Nitrobacter*-like NOB were more abundant in the topsoil with high nutrient contents and were more associated with AOB (Figure 3A). The application of recycled manure provided a more stable nutrient supply and increased the soil organic carbon availability and microbial biomass (such as SMBC and SMBN), thereby providing rich and balanced nutrients for both *Nitrobacter*-like and *Nitrospira*-like NOB to grow. In fact, NOB should not be considered obligate chemolithoautotrophs, and some NOB species have been proven to be capable of assimilating simple organic compounds, such as glycerol, formate and pyruvate (Daims et al., 2001; Spieck et al., 2006). Consequently, soil samples under the M treatment significantly increased the *Nitrobacter nxrA* and *Nitrospira nxrB* abundances (Figure 2).

It is well-known that comammox *Nitrospira* are responsive to nitrogen addition in soils because nitrogen as the essential substrate may shape the ecological niche differentiation of comammox *Nitrospira* in agricultural soils (Shi et al., 2018; He et al., 2021). However, knowledge of the effect of different fertilization regimes on comammox *Nitrospira* abundance and community composition is still limited. Our study showed that the abundance of comammox *Nitrospira* (the sum of clade A and clade B) was slightly higher than that of canonical AOB (except in the topsoil under MNPK) and slightly lower than that of AOA (Figure 2), suggesting that comammox *Nitrospira* might be competitively superior in fertilized Alfisols. This means that comammox *Nitrospira* not only have an oligotrophic lifestyle but can also adapt to various environments with abundant substrates (Li et al., 2020; Lu et al., 2020; He et al., 2021), which is consistent with the idea that comammox in terrestrial ecosystems might have evolved ecological features that are more diverse than those in aquatic ecosystems (Hu and He, 2017; Hu et al., 2021).

Both comammox Nitrospira clade A and clade B were observed in the studied soils. Although comammox Nitrospira clade A was proposed to be split into clades A1, A2 and A3 (Xia et al., 2018; Li et al., 2020), only comammox Nitrospira clade A2 was found in the studied soils (Supplementary Figure S3). Unlike the oligotrophic lifestyle of clade A1 (Xia et al., 2018; Sun et al., 2020), clade A2 was dominant in N-rich tidal flat sediment or agricultural soils (Xia et al., 2018; Liu et al., 2020; Xu et al., 2020). Specifically, a recent study demonstrated that there were different ecological preferences in individual subclades (clade A2.1 and clade A2.2) of clade A2 in Ultisols following long-term fertilization (Lin et al., 2022). These findings provide insights that comammox Nitrospira clade A in agricultural soil is not strictly oligotrophic and has adapted to thrive in a wide range of ecological niches. Additionally, comammox Nitrospira clade B, which is oligotrophic, was also found in these fertilized soils. Comammox Nitrospira clade A and clade B abundances responded differently to chemical and/or manure fertilizer addition. Compared with CK and NPK, the application of NPK plus recycled manure (MNPK) or recycled manure alone (M) significantly increased the abundance of comammox Nitrospira clade A. This is consistent with a recent study that provided evidence that comammox Nitrospira clade A might be relatively copiotrophic and thus flourished in NPK plus pig manure treatment with a higher nutrient content (Lin et al., 2022). Generally, comammox Nitrospira clade B is considered oligotrophic and sensitive to the soil NH4+-N content, so clade B did not grow autotrophically when ammonium was added to the soil (Daims et al., 2015; Shi et al., 2018; Wang et al., 2019). Therefore, the application of fertilizer, regardless of chemical fertilizer or recycled manure, decreased the abundance of comammox Nitrospira clade B in all topsoil samples (Figure 2). As shown in Figure 3B, the different long-term fertilization regime also led to marked differences in the community composition of comammox Nitrospira especially in the topsoil. The community structure of comammox under the CK and M treatment was similar and obviously different from that in the treatments containing chemical N, indicating that N forms and contents induced by the fertilization regime were key factors driving the comammox Nitrospira community (Sun et al., 2021; He et al., 2022).

In addition, different types of comammox Nitrospira had unique response patterns to soil layer. The relative abundance of comammox Nitrospira clade A was predominant in the topsoil especially under the chemical fertilizer treatments. In contrast, the relative abundance of clade B appeared to be dominant and increased in the subsoil samples with lower ammonium content (Figures 2, 5). This finding consistently supported the idea that there was niche separation between comammox Nitrospira clade A and clade B owing to their physiological differentiation and genome characteristics. Comammox Nitrospira clade A possesses Rh-type ammonium transporters with a higher uptake capacity and lower affinity for ammonia (homologous Rh-type ammonium transporters discovered in most β -AOB, $K_{m(NH3)} \approx 0.14-0.19 \text{ mM}$) than Amt-type ammonium transporters (Dytczak et al., 2008; Daims et al., 2015; Palomo et al., 2018). Comammox Nitrospira clade B encodes Amt-type ammonium transporters (present in canonical AOA, $K_{m(NH3)} \approx 0.13 \,\mu\text{M}$) with a higher affinity for ammonia (Martens-Habbena et al., 2009; Kits et al., 2017; Koch et al., 2019). In summary, this evidence shows that comammox Nitrospira clade A (mainly clade A2) prefers soil that is relatively rich in ammonium and nutrients, while comammox Nitrospira clade B prevalently occurs in nutrient-limited soils. Different clades of comammox Nitrospira occupied distinct

ecological niches and adaptively responded to long-term fertilization in agricultural soils, which supports our hypothesis.

Comammox Nitrospira can perform both ammonia-oxidation and nitrite-oxidation in the nitrification process, suggesting that they may compete with canonical nitrifiers for the transformation of nitrogen in agroecosystems. Some studies have shown that comammox Nitrospira demonstrated lower activity and less contribution to ammonia oxidation than canonical ammonia oxidizers (Shi et al., 2018; Wang et al., 2019, 2020). Comammox Nitrospira clade B was assumed to contribute more to nitrification in soil without NH₄⁺ addition (Wang et al., 2019). However, as revealed by Spearman's correlation and structural equation modeling analysis, our study revealed that the contributions of comammox Nitrospira clade B and AOB to the soil PAO were positive and significant, and canonical NOB (mainly Nitrospira-like NOB) were the dominant contributors to the soil PNO (Figure 4; Supplementary Table S3). These results confirmed that AOB and comammox Nitrospira might play crucial roles in the nitrification of agricultural soils, as previously reported (Hu et al., 2021). The reason for the important contribution of comammox Nitrospira clade B to soil PAO could be as follows: First, the physiological adaptation of comammox Nitrospira clade B might be more diverse than we previously thought; some clade B specimens found in agricultural soil might adapt well to relatively high ammonia availability in soil (He et al., 2022; Lin et al., 2022), so the high abundance of clade B in the studied soils indicated its importance in soil nitrification. Second, Liu et al. (2020) demonstrated that comammox Nitrospira clade B made greater contributions to the nitrifier abundance on the Qinghai-Tibetan Plateau with a higher pH and lower temperature (9.6 \pm 0.4 °C). The average annual temperature of experimental station was 7.5°C, which might be a favorable temperature for clade B to grow but could generate selective pressure for the other ammonia oxidizers. Third, although canonical ammonia oxidizers and comammox Nitrospira had the same ammonia oxidation function, they might cooperatively coexist rather than be competitively exclusive due to their niche separation (Wang et al., 2019). It is probable that AOB played an important role in the top layer of soil amended with chemical N fertilizer, and comammox Nitrospira clade B preferred to utilize ammonium from mineralized organic N in the subsoil. These two nitrifiers might work cooperatively and occupy different niches to perform ammonia oxidation in longterm chemically and/or manure-fertilized soils. These results support our hypothesis that comammox Nitrospira play an important role in soil ammonia oxidation even with high nutrient inputs. Additionally, it should be noted that the substrate addition method was adopted to measure PAO and PNO in soil, which might not accurately reflect the nitrifier *in situ* rates in soils but only their potential ability (Wang et al., 2019). The active nitrifying guilds responsible for PAO and PNO might be entirely distinct with those under in situ soil conditions. It is thus important to make presumptions for correlating the manipulated PAO and PNO activities with in situ soil microbial diversity and ability by Spearman correlation analysis alone. Moreover, techniques simultaneously combining with selective inhibitors with DNA/RNA-stable isotope probing should be adopted to determine the actual activity and contribution of comammox Nitrospira in agricultural soils.

In our study, the measured soil physicochemical properties were significantly affected by the fertilization regime (p < 0.05; Table 2). It is reported that the application of organic manure can effectively preserve soil carbon and nitrogen pools and buffer soil acidification (Xie et al.,

2021), while repeated application of chemical N fertilizer can significantly decrease the soil pH (Shi et al., 2018). Notably, compared with CK, the decrease of soil pH in the topsoil under chemical NPK treatment was less than that under chemical N treatment. One possible reason for this phenomenon was that the chemical P fertilizer used in NPK treatment was calcium triple superphosphate containing a large amount of calcium cations, which improved the soil ion exchange capacity and decreased the soil pH reduction caused by chemical N fertilizer. Another reason was that the crop yield, root biomass and root exudates under NPK treatment were much higher than those under N treatment (Ning et al., 2020), which promoted the amount of organic matter returned through root stubble under NPK treatment. The higher organic matter content in soil under NPK treatment (Table 2) was conducive to improving soil buffering capacity and alleviating soil acidification.

Several studies have pointed out that environmental discrepancies and biological reciprocal actions are the key factors controlling the variations in the microbial community composition in diverse ecosystems (Bahram et al., 2018; Vijayan et al., 2020). As revealed by CCA and ABT analysis (Figure 3; Supplementary Figure S1), soil pH, NH₄⁺-N and SOM contents were crucial factors shaping nitrifying microbial abundances and comammox Nitrospira communities in soils subjected to long-term fertilization. The vital roles of soil pH and ammonium in controlling the biogeographic distribution, community composition and niche separation of canonical ammonia oxidizers have been thoroughly discussed in many previous studies (He et al., 2012; Prosser and Nicol, 2012; Hu et al., 2013). Zhu et al. (2022) also found that pH was the most important factor affecting comammox Nitrospira abundance on a global scale. Meanwhile, ammonium concentration was identified as the dominant predictor of the variation in comammox Nitrospira abundance and community composition, although the effects of ammonium were inconsistent (Wang et al., 2019; He et al., 2021; Zhu et al., 2022). The reason for the different sensitivities of AOB, AOA and comammox Nitrospira to soil pH could be attributed to their different affinities for NH₃ (Hu and He, 2017; Kits et al., 2017). It is well accepted that the ammonia oxidizers prefer to metabolize NH₃ substrates rather than NH₄⁺ (Hu and He, 2017; Xu et al., 2020). Variations in soil pH under long-term fertilization governed the nitrifier abundances because of the pH-dependent chemical equilibrium between NH₃ and NH₄⁺-N. The results showed that the abundance of comammox Nitrospira was significantly and negatively correlated with soil pH (Supplementary Figure S1), which is consistent with many studies suggesting that comammox Nitrospira exhibits a higher resistance to slightly acidic soils (Shi et al., 2018; Xu et al., 2020; Hu et al., 2021). The recycled manure treatment was the highlight of this long-term field trial, as all of the organic fertilizer applied in this experiment was derived from the ecosystem and then returned to the system itself as a nutrient supply. This procedure was obviously different from employing commercial organic fertilizer. First, the organic fertilizer and nutrients were all from the system itself and depended on the biological mass product. Second, the amount of organic fertilizer varied from year to year. Ultimately, although the input rates of N, P and K were not as high as those under the NPK treatment, recycled manure significantly increased the soil SOM and TN contents and elevated the levels of SMBN (Table 2), but decreased the soil NH4+-N content, similar to the findings in many previous investigations (Xu et al., 2012; Ma et al., 2020). The main reason for the accumulation of SOM contents under the M and MNPK treatments was that the manure contained a large proportion of recalcitrant organic carbon, which could facilitate the formation of macroaggregates and protect the organic matter from being decomposed (Tao et al., 2017; Li et al., 2020; Lin et al., 2022). High SOM contents provided plentiful nutrients to increase the total abundance of soil microorganisms, which indirectly caused the increase in comammox Nitrospira (Wang et al., 2019). Additionally, the application of manure provided available carbon as well as rich and stable nutrients for nitrifier growth, which stimulated comammox Nitrospira clade A abundance and promoted the nitrification process, leading to a considerable decrease in the NH4⁺-N contents (Han et al., 2018; Li et al., 2020; Lin et al., 2022). Based on comparative genome analysis, some functional genes associated with gluconeogenesis, glycolysis reactions, oxidative TCA cycle and the pentose phosphate pathway were discovered in the core genome of certain comammox Nitrospira (Lücker et al., 2010). The *phaZ* gene that encodes polyhydroxybutyrate (PHB) depolymerase was exclusively discovered in comammox Nitrospira genomes (Palomo et al., 2018). Overall, this evidence suggests that comammox Nitrospira has the potential to grow mixotrophically similar to the other Nitrospira (Daims et al., 2001; Koch et al., 2019), so it was not surprising that comammox Nitrospira could be shaped by organic manure application. Furthermore, the relative abundances of comammox Nitrospira clade A and clade B responded differently to the application of manure, implying that SOM content is a crucial property shaping the niche differentiation of comammox Nitrospira in agricultural soil. The application of manure might provide an unfavorable environment for comammox Nitrospira clade B grow (He et al., 2022). Notably, other soil properties, such as the available phosphorus content (Liu et al., 2019; Lin et al., 2022), have been found to be vital factors determining the comammox Nitrospira community (Figure 3; Supplementary Figure S1). Comammox Nitrospira might be more competitive than canonical ammonia oxidizers under phosphorus limited conditions due to possessing the alkaline phosphatase enzyme (Palomo et al., 2016). However, because no isolated comammox strain was obtained from the soil, more efforts determine whether comammox Nitrospira can utilize organic substrates directly for growth or which genomic characteristics determine the niche specialization of comammox Nitrospira in soil are needed in the future studies.

5. Conclusion

Our results showed that diverse long-term fertilization regimes changed the soil properties and resulted in changes in AOB, AOA, *Nitrospira*-like NOB, *Nitrobacter*-like NOB, and comammox *Nitrospira* abundances in Alfisols. Niche separation existed among canonical nitrifying microorganisms and comammox *Nitrospira*. AOB abundances were significantly stimulated by the nitrogen input rates, while the abundances of AOA were increased by recycled manure application. Comammox *Nitrospira* clade A and clade B responded differently to chemical and/or manure fertilizer, and the relative importance of clade B to soil PAO was confirmed by SEM analysis. Variations in the relative abundances of comammox *Nitrospira* clade A and clade B in different soil layers showed that niche differentiation certainly existed between them. Soil pH, NH₄⁺-N, and SOM contents were crucial factors driving the comammox *Nitrospira* community in Alfisols under long-term fertilization regimes. However, an extension of our knowledge about the contribution to nitrification and the physiological and genomic characteristics of new comammox *Nitrospira* bacteria is needed, and studies of pure comammox *Nitrospira* cultures should be integrated to comprehensively understand the nitrogen cycle under agricultural management.

Data availability statement

The datasets presented in this study can be found in online repositories. The names of the repository/repositories and accession number(s) can be found at: https://www.ncbi.nlm.nih.gov/, SRR13319769 to SRR13319798.

Author contributions

YW, XZ, QM, WY, and ZZ performed the field experiment and analyzed the abundances of canonical nitrifiers and comammox *Nitrospira* in soils. YZ, NZ, and LX analyzed and interpreted the effects of different fertilization regimes on soil properties and comammox *Nitrospira* community. YW was the major contributor in writing the manuscript. XZ reviewed and edited the manuscript. All authors read and approved the final manuscript.

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Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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Supplementary material

The Supplementary material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fmicb.2023.1095937/full# supplementary-material

References

Attard, E., Poly, F., Commeaux, C., Laurent, F., Terada, A., and Smets, B. F. (2010). Shifts between *Nitrospira*-and Nitrobacter-like nitrite oxidizers underlie the response of soil potential nitrite oxidation to changes in tillage practices. *Environ. Microbiol.* 12, 315–326. doi: 10.1111/j.1462-2920.2009.02070.x

Bahram, M., Hildebrand, F., Forslund, S. K., Anderson, J. L., Soudzilovskaia, N. A., Bodegom, P. M., et al. (2018). Structure and function of the global topsoil microbiome. *Nature* 560, 233–237. doi: 10.1038/s41586-018-0386-6

Bao, S. D. (Ed.) (2000). Chemical analysis for agricultural soil. Beijing: China Agricultural Press.

Cai, F. F., Luo, P. Y., Yang, J. F., Irfan, M., Zhang, S. Y., An, N., et al. (2021). Effect of longterm fertilization on ammonia-oxidizing microorganisms and nitrification in brown soil of Northeast China. *Front. Microbiol.* 11:622454. doi: 10.3389/fmicb.2020.622454

Caporaso, J. G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F. D., Costello, E. K., et al. (2010). QIIME allowes analysis of high-throughput community sequencing data. *Nat. Methods* 7, 335–336. doi: 10.1038/nmeth.f.303

Daims, H., Lebedeva, E. V., Pjevac, P., Han, P., Herbold, C., Albertsen, M., et al. (2015). Complete nitrification by *Nitrospira* bacteria. *Nature* 528, 504–509. doi: 10.1038/ nature16461

Daims, H., Nielsen, J. L., Nielsen, P. H., Schleifer, K. H., and Wagner, M. (2001). In situ characterization of *Nitrospira* like nitrite-oxidizing bacteria active in wastewater treatment plants. *Appl. Environ. Microbiol.* 67, 5273–5284. doi: 10.1128/AEM.67.11.5273-5284.2001

Death, H. (2007). Boosted trees for ecological modeling and prediction. *Ecology* 88, 243–251. doi: 10.1890/0012-9658(2007)88[243:BTFEMA]2.0.CO;2

Duan, P. P., Wu, Z., Zhang, Q. Q., Fan, C. H., and Xiong, Z. Q. (2018). Thermodynamic responses of ammonia-oxidizing archaea and bacteria explain N_2O production from greenhouse vegetable soils. *Soil Biol. Biochem.* 120, 37–47. doi: 10.1016/j.soilbio.2018.01.027

Dytczak, M. A., Londry, K. L., and Oleszkiewicz, J. A. (2008). Nitrifying genera in activated sludge may influence nitrification rates. *Water Environ. Res.* 80, 388–396. doi: 10.2175/106143007X221373

Francis, C. A., Roberts, K. J., Beman, J. M., Santoro, A. E., and Oakley, B. B. (2005). Ubiquity and diversity of ammonia-oxidizing archaea in water columns and sediments of the ocean. *Proc. Natl. Acad. Sci. U. S. A.* 102, 14683–14688. doi: 10.1073/pnas.0506625102

Gao, S. J., Chang, D. N., Zou, C. Q., Cao, W. D., Gao, J. S., Huang, J., et al. (2018). Archaea are the predominant and responsive ammonia oxidizing prokaryotes in a red paddy soil receiving green manures. *Eur. J. Soil Biol.* 88, 27–35. doi: 10.1016/j.ejsobi.2018.05.008

Gonzalez, M. A., Rodriguez, S. A., van Loosdrecht, M. C., Gonzalez, L. J., and Vahala, R. (2016). Detection of comammox bacteria in full-scale wastewater treatment bioreactors using Tag-454-pyrosequencing. *Environ. Sci. Pollut. Res.* 23, 25501–25511. doi: 10.1007/s11356-016-7914-4

Han, S., Zeng, L. Y., Luo, X. S., Xiong, X., Wen, S. L., Wang, B. R., et al. (2018). Shifts in Nitrobacter-and *Nitrospira*-like nitrite-oxidizing bacterial communities under long-term fertilization practices. *Soil Biol. Biochem.* 124, 118–125. doi: 10.1016/j.soilbio.2018.05.033

He, J. Z., Hu, H. W., and Zhang, L. M. (2012). Current insights into the autotrophic thaumarchaeal ammonia oxidation in acidic soils. *Soil Biol. Biochem.* 55, 146–154. doi: 10.1016/j.soilbio.2012.06.006

He, S. S., Li, Y. W., Mu, H. B., Zhao, Z. R., Wang, J. W., Liu, S. F., et al. (2021). Ammonium concentration determines differential growth of comammox and canonical ammonia-oxidizing prokaryotes in soil microcosms. *Appl. Soil Ecol.* 157:103776. doi: 10.1016/j. apsoil.2020.103776

He, J. Z., Shen, J. P., Zhang, L. M., Zhu, Y. G., Zheng, Y. M., Xu, M. G., et al. (2007). Quantitative analyses of the abundance and composition ammonia-oxidizing bacteria and ammonia-oxidizing archaea of a Chinese upland red soil under long-term fertilization practices. *Environ. Microbiol.* 9, 2364–2374. doi: 10.1111/j.1462-2920.2007.01358.x

He, Z. Y., Sun, A. Q., Jiao, X. Y., Ge, A. H., Hu, H. W., Jin, S. S., et al. (2022). Fertilization has a greater effect than rhizosphere on community structures of comammox *Nitrospira* in an alkaline agricultural soil. *Appl. Soil Ecol.* 175:104456. doi: 10.1016/j.apsoil.2022.104456

Hou, S. P., Ai, C., Zhou, W., Liang, G. Q., and He, P. (2018). Structure and assembly cues for rhizospheric nirK-and nirS-type denitrifier communities in long-term fertilized soils. *Soil Biol. Biochem.* 119, 32–40. doi: 10.1016/j.soilbio.2018.01.007

Hu, H. W., and He, J. Z. (2017). Comammox-a newly discovered nitrification process in the terrestrial nitrogen cycle. J. Soils Sediments 17, 2709–2717. doi: 10.1007/s11368-017-1851-9

Hu, H. W., Xu, Z. H., and He, J. Z. (2014). Ammonia-oxidizing archaea play a predominant role in acid soil nitrification. *Adv. Agron.* 125, 261–302. doi: 10.1016/ B978-0-12-800137-0.00006-6

Hu, H. W., Zhang, L. M., Dai, Y., Di, H. J., and He, J. Z. (2013). pH-dependent distribution of soil ammonia oxidizers across a large geographical scale as revealed by high-throughput pyrosequencing. *J. Soils Sediments* 13, 1439–1449. doi: 10.1007/s11368-013-0726-y

Hu, J. J., Zhao, Y. X., Yao, X. W., Wang, J. Q., Zheng, P., Xi, C. W., et al. (2021). Dominance of comammox *Nitrospira* in soil nitrification. *Sci. Total Environ.* 780:146558. doi: 10.1016/j. scitotenv.2021.146558

Kaveney, B., Condon, J., Doran, G., Galea, F., and Rigg, J. (2020). The distribution of soil nitrifiers and efficacy of 3,4-dimethylpyrazole phosphate changes with soil depth and calcium carbonate application. *Soil Biol. Biochem.* 150:108009. doi: 10.1016/j. soilbio.2020.108009

Kim, J. G., Park, S. J., Sinninghe, D. J. S., Schouten, S., Rijpstra, W. I., Jung, M. Y., et al. (2016). Hydrogen peroxide detoxification is a key mechanism for growth of ammoniaoxidizing archaea. *Pro. Natl. Acad. Sci. U.S.A.* 113, 7888–7893. doi: 10.1073/ pnas.1605501113

Kits, K. D., Sedlacek, C. J., Lebedeva, E. V., Han, P., Bulaev, A., Pjevac, P., et al. (2017). Kinetic analysis of a complete nitrifiers reveals an oligotrophic lifestyle. *Nature* 549, 269–272. doi: 10.1038/nature23679

Koch, H., Van Kessel, M. A., and Lücker, S. (2019). Complete nitrification: insights into the ecophysiology of comammox *Nitrospira*. *Appl. Microbiol. Biotech.* 103, 177–189. doi: 10.1007/s00253-018-9486-3

Könneke, M., Bernhard, A. E., de la Torre, J. R., Walker, C. B., Waterbury, J. B., and Stahl, D. A. (2005). Isolation of an autotrophic ammonia-oxidizing marine archaeon. *Nature* 437, 543–546. doi: 10.1038/nature03911

Kurola, J., Salkinoja-Salonen, M., Aarnio, T., Hultman, J., and Romantschuk, M. (2005). Activity, diversity and population size of ammonia-oxidising bacteria in oilcontaminated landfarming soil. *FEMS Microbiol. Lett.* 250, 33–38. doi: 10.1016/j. femsle.2005.06.057

Kuypers, M. M. M., Marchant, H. K., and Kartal, B. (2018). The microbial nitrogencycling network. *Nat. Rev. Microbiol.* 16, 263–276. doi: 10.1038/nrmicro.2018.9

Lai, J. S., Zou, Y., Zhang, J. L., and Peres-Neto, P. (2022). Generalizing hierarchical andvariation partitioning in multiple regression and canonical analyses using therdacca Hp R package. *Methods Ecol. Evol.* 13, 782–788. doi: 10.1111/2041-210X.13800

Li, X., Han, S., Wan, W. J., Zheng, L. X., Chen, W. L., and Huang, Q. Y. (2020). Manure fertilizers alter the nitrite oxidizer and comammox community composition and increase nitrification rates. *Soil Till. Res.* 204:104701. doi: 10.1016/j.still.2020.104701

Li, C. Y., He, Z. Y., Hu, H. W., and He, J. Z. (2022). Niche specialization of comammox Nitrosipra in terrestrial ecosystems: oligotrophic or copiotrophic? *Crit. Rev. Env. Sci. Tech.* 53, 161–176. doi: 10.1080/10643389.2022.2049578

Li, C. Y., Hu, H. W., Chen, Q. L., Chen, D. L., and He, J. Z. (2019). Comammox *Nitrospira* play an active role in nitrification of agricultural soils amended with nitrogen fertilizers. *Soil Biol. Biochem.* 138:107609. doi: 10.1016/j.soilbio.2019.107609

Li, C. Y., Hu, H. W., Chen, Q. L., Chen, D., and He, J. Z. (2020). Growth of comammox *Nitrospira* is inhibited by nitrification inhibitors in agricultural soils. *J. Soils Sediments* 20, 621–628. doi: 10.1007/s11368-019-02442-z

Lin, Y. X., Fan, J. B., Hu, H. W., Duan, C. J., Ye, G. P., Wan, S., et al. (2022). Differentiation of individual clusters of comammox *Nitrospira* in an acidic Ultisol following long-term fertilization. *Appl. Soil Ecol.* 170:104267. doi: 10.1016/j.apsoil.2021.104267

Lin, Y. X., Ye, G. P., Ding, W. X., Hu, H. W., Zheng, Y., Fan, J. B., et al. (2020). Niche differentiation of comammox *Nitrospira* and canonical ammonia oxidizers in soil aggregate fractions following 27-year fertilizations. *Agric. Ecosys. Environ.* 304:107147. doi: 10.1016/j. agee.2020.107147

Liu, S. F., Wang, H. Y., Chen, L. M., Wang, J. W., Zheng, M. S., Liu, S. T., et al. (2020). Comammox *Nitrospira* within the Yangtze River continuum: community, biogeography, and ecological drivers. *ISME J.* 14, 2488–2504. doi: 10.1038/s41396-020-070-8

Liu, T. L., Wang, Z. H., Wang, S. L., Zhao, Y. P., Wright, A. L., and Jiang, X. J. (2019). Responses of ammonia-oxidizers and comammox to different long-term fertilization regimes in a subtropical paddy soil. *Euro. J. Soil Biol.* 93:103087. doi: 10.1016/j. ejsobi.2019.103087

Liu, Z. B., Zhang, C. J., Wei, Q. Y., Zhang, S. Y., Quan, Z. X., and Li, M. (2020). Temperature and salinity drive comammox community composition in mangrove ecosystem across southeastern China. *Sci. Total Environ.* 742:140456. doi: 10.1016/j. scitotenv.2020.140456

Lu, X. D., Taylor, A. E., Myrold, D. D., and Neufeld, J. D. (2020). Expanding perspectives of soil nitrification to include ammonia-oxidizing archaea and comammox bacteria. *Soil Sci. Soc. Am. J.* 84, 287–302. doi: 10.1002/saj2.20029

Lücker, S., Wagner, M., Maixner, F., Pelletier, E., Koch, H., Vacherie, B., et al. (2010). A *Nitrospira* metagenome illuminates the physiology and evolution of globally important nitrite-oxidizing bacteria. *Proc. Natl. Acad. Sci. U. S. A.* 107, 13479–13484. doi: 10.1073/pnas.1003860107

Ma, Q., Li, S. L., Xu, Z. Q., Fan, S. B., Xia, Z. Q., Zhou, C. R., et al. (2020). Changes in supply pathways under different long-term fertilization regimes in Northeast China. *Soil Till. Res.* 201:104609. doi: 10.1016/j.still.2020.104609

Martens-Habbena, W., Berube, P. M., Urakawa, H., de la Torre, J. R., and Stahl, D. A. (2009). Ammonia oxidation kinetics determine niche separation of nitrifying archaea and bacteria. *Nature* 461, 976–979. doi: 10.1038/nature08465

Ning, C. C., Ma, Q., Yuan, H. Y., Zhou, C. R., Xia, Z. Q., and Yu, W. T. (2020). Mineral fertilizers with recycled manure boost crop yield and P balance in a long-term field trial. *Nutr. Cycl. Agroeosyst.* 116, 271–283. doi: 10.1007/s10705-019-10043-y

Nowka, B., Daims, H., and Spieck, E. (2014). Comparison of oxidation kinetics of nitriteoxidizing bacteria: nitrite availability as a key factor in niche differentiation. *Appl. Environ. Microbiol.* 81, 745–753. doi: 10.1128/aem.02734-14

Ouyang, Y., Evans, S. E., Friesen, M. L., and Tiemann, L. K. (2018). Effect of nitrogen fertilization on the abundance of nitrogen cycling genes in agricultural soils: a metaanalysis of field studies. *Soil Biol. Biochem.* 127, 71–78. doi: 10.1016/j.soilbio.2018.08.024 Palomo, A., Flower, S. J., Gulay, A., Rasmussen, S., Sicheritze-Pontén, T., and Smets, B. F. (2016). Metagenomic analysis of rapid gravity sand filter microbial communities suggests novel physiology of *Nitrospira* spp. *ISEM J.* 10, 2569–2581. doi: 10.1038/ismej.2016.63

Palomo, A., Pedersen, A. G., Flower, S. J., Dechesne, A., Sicheritze-Pontén, T., and Smets, B. F. (2018). Comparative genomics sheds light on niche differentiation and the evolutionary history of comammox *Nitrospira*. *ISEM J.* 12, 1779–1793. doi: 10.1038/s41396-018-0083-3

Pester, M., Maixner, F., Berry, D., Rattei, T., Koch, H., and Lücker, S. (2014). NxrB encoding the beta subunit of nitrite oxidoreductase as functional and phylogenetic marker for nitrite-oxidizing *Nitrospira*. *Environ*. *Microbiol*. *Rep.* 16, 3055–3071. doi: 10.1111/1462-2920.12300

Pjevac, P., Schauberger, C., Poghosyan, L., Herbold, C. W., Van Kessel, M. A., Daebeler, A., et al. (2017). AmoA-targeted polymerase chain reaction primers for the specific detection and quantification of comammox *Nitrospira* in the environment. *Front. Microbiol.* 8:1508. doi: 10.3389/fmicb.2017.01508

Poghosyan, L., Koch, H., Lavy, A., Frank, J., Van Kessel, M. A., Jetten, M. S., et al. (2019). Metagenomic recovery of two distinct comammox *Nitrospira* from the terrestrial subsurface. *Environ. Microbiol.* 21, 3627–3637. doi: 10.1111/1462-2920.14691

Poly, F., Wertz, S., Brothier, E., and Degrange, V. (2008). First exploration of Nitrobacter diversity in soils by a PCR cloning-sequencing approach targeting functional gene nxrA. *FEMS Microbiol. Ecol.* 63, 132–140. doi: 10.1111/j.1574-6941.2007.00404.x

Prosser, J. I., and Nicol, G. W. (2012). Archaeal and bacterial ammonia-oxidisers in soil: the quest for niche specialization and differentiation. *Trends Microbiol.* 20, 523–531. doi: 10.1016/j.tim.2012.08.001

Rotthauwe, J. H., Witzel, K. P., and Liesack, W. (1997). The ammonia monooxygenase structural gene amoA as a functional marker: molecular fine-scale analysis of natural ammonia-oxidizing populations. *Appl. Environ. Microbiol.* 63, 4704–4712. doi: 10.1128/aem.63.12.4704-4712.1997

Schloss, P. D., Westcott, S. L., Ryabin, T., Hall, J. R., Hartmann, M., Hollister, E. B., et al. (2009). Introducing mothur: open-source, platform-independent, community-supported software for describing and comparing microbial communities. *Appl. Environ. Microbiol.* 75, 7537–7541. doi: 10.1128/AEM.01541-09

Schramm, A., De Beer, D., Den Heuvel, J. C., Ottengraf, S. P., and Amann, R. (1999). Microscale distribution of populations and activities of *Nitrosospira* and *Nitrospira* spp. along a macroscale gradient in a nitrifying bioreactor: quantification by in situ hybridization and the use of microsensors. *Appl. Environ. Microbiol.* 65, 3690–3696. doi: 10.1128/AEM.65.8.3690-3696.1999

Shi, X. Z., Hu, H. W., Wang, J. Q., He, J. Z., Zheng, C. Y., Wan, X. H., et al. (2018). Niche separation of comammox *Nitrospira* and canonical ammonia oxidizers in an acidic subtropical forest soil under long-term nitrogen deposition. *Soil Biol. Biochem.* 126, 114–122. doi: 10.1016/j.soilbio.2018.09.004

Spieck, E., Hartwig, C., McCormack, I., Maixner, F., Wagner, M., Lipski, A., et al. (2006). Selective enrichment and molecular characterization of a previously uncultured *Nitrospira*like bacterium from activated sludge. *Environ. Microbiol.* 8, 405–415. doi: 10.1111/j.1462-2920.2005.00905.x

Sun, D. Y., Tang, X. F., Zhao, M. Y., Zhang, Z. X., Hou, L. J., Liu, M., et al. (2020). Distribution and diversity of comammox *Nitrospira* in coastal wetlands of China. *Front. Microbiol.* 11:589268. doi: 10.3389/fmicb.2020.589268

Sun, P., Zhang, S. X., Wu, Q. H., Zhu, P., Ruan, Y. Z., and Wang, Q. (2021). pH and ammonium concentration are dominant predictors of the abundance and community composition of comammox bacteria in long-term fertilized Mollisol. *Appl. Soil Ecol.* 168:104139. doi: 10.1016/j.apsoil.2021.104139

Tamura, K., Dudley, J., Nei, M., and Kumar, S. (2007). MEGA4: molecular evolutionary genetics analysis (MEGA) software version 4.0. *Mol. Biol. Evol.* 24, 1596–1599. doi: 10.1093/molbev/msm092

Tao, R., Wakelin, S. A., Liang, T. C., and Chu, G. X. (2017). Response of ammoniaoxidizing archaea and bacteria in calcareous soil to mineral and organic fertilizer application and their relative contribution to nitrification. *Soil Biol. Biochem.* 114, 20–30. doi: 10.1016/j.soilbio.2017.06.027

Tourna, M., Stieglmeier, M., Spang, A., Könneke, M., Schintlmeister, A., Urich, T., et al. (2011). Nitrososphaera viennensis, an ammonia oxidizing archaeon from soil. *Proc. Natl. Acad. Sci. U. S. A.* 108, 8420–8425. doi: 10.1073/pnas.1013488108

van Kessel, M. A., Speth, D. R., Albertsen, M., Nielsen, P., Op den Camp, H. Z., Kartal, B., et al. (2015). Complete nitrification by a single microorganism. *Nature* 528, 555–559. doi: 10.1038/nature16459

Vijayan, A., Jayadradhan, R. K., Pillai, D., Geetha, P. P., Joseph, V., and Sarojini, B. S. (2020). *Nitrospira* as versatile nitrifiers: taxonomy, ecophysiology, genome characteristics, growth, and metabolic diversity. *J. Basic Microbiol.* 61, 88–109. doi: 10.1002/jobm.202000485

Wang, Z. H., Cao, Y. Q., Barker, X. Z., Nicol, G. W., Wright, A. L., Jia, Z. J., et al. (2019). Comammox *Nitrospira* clade B contributes to nitrification in soil. *Soil Biol. Biochem.* 135, 392–395. doi: 10.1016/j.soilbio.2019.06.004

Wang, F., Liang, X. L., Ma, S. H., Liu, L. Z., and Wang, J. K. (2021). Ammonia-oxidizing archaea are dominate over comammox in soil nitrification under long-term nitrogen fertilization. *J. Soils Sediments* 21, 1800–1814. doi: 10.1007/s11368-021-02897-z

Wang, X. M., Wang, S. Y., Jiang, Y. Y., Zhou, J. M., Han, C., and Zhu, J. B. (2020). Comammox bacterial abundance, activity, and contribution in agricultural rhizosphere soils. *Sci. Total Environ.* 727:138563. doi: 10.1016/j.scitotenv.2020.138563

Wang, J. C., Wang, J. L., Rhodes, G., He, J. Z., and Ge, Y. (2019). Adaptive responses of comammox *Nitrospira* and canonical ammonia oxidizers to long-term fertilizations: implications for the relative contributions of different ammonia oxidizers to soil nitrogen cycling. *Sci. Total Environ.* 668, 224–233. doi: 10.1016/j.scitotenv.2019.02.427

Wu, J. S., Joergensen, R. G., Pommerening, B., Chaussod, R., and Brooked, P. C. (1990). Measurement of soil microbial biomass C by fumigation-extraction-an automated procedure. *Soil Biol. Biochem.* 22, 1167–1169. doi: 10.1016/0038-0717(90)90046-3

Xia, F., Wang, J. G., Zhu, T., Zou, B., Rhee, S. K., and Quan, Z. X. (2018). Ubiquity and diversity of complete ammonia oxidizer (Comammox). *Appl. Environ. Microbiol.* 84, e01390–e01318. doi: 10.1128/AEM.01390-18

Xie, S. W., Yang, F., Feng, H. X., Yu, Z. Z., Liu, C. S., Wei, C. Y., et al. (2021). Organic fertilizer reduced carbon and nitrogen in runoff and buffered soil acidification in tea plantations: evidence in nutrient contents and isotope fractionations. *Sci. Total Environ.* 762:143059. doi: 10.1016/j.scitotenv.2020.143059

Xu, S. Y., Wang, B. Z., Li, Y., Jiang, D. Q., Zhou, Y. T., Ding, A. Q., et al. (2020). Ubiquity, diversity and activity of comammox *Nitrospira* in agricultural soils. *Sci. Total Environ.* 706:135684. doi: 10.1016/j.scitotenv.2019.135684

Xu, Y. G., Yu, W. T., Ma, Q., and Zhou, H. (2012). Responses of bacterial and archaeal ammonia oxidisers of an acidic luvisols soil to different nitrogen fertilization rates after 9 years. *Biol. Fert. Soil.* 48, 827–837. doi: 10.1007/s00374-012-0677-2

Yang, Y., Daims, H., Liu, Y., Herbold, C. W., Pjevac, P., Lin, J. G., et al. (2020). Activity and metabolic versatility of complete ammonia oxidizer in full-scale wastewater treatment systems. *MBio* 11, e03175–e03119. doi: 10.1128/mBio.03175-19

Zhang, L. M., Hu, H. W., Shen, J. P., and He, J. Z. (2012). Ammonia-oxidizing archaea have more important role than ammonia-oxidizing bacteria in ammonia oxidation of strongly acidic soils. *ISME J.* 6, 1032–1045. doi: 10.1038/ismej.2011.168

Zhu, G. B., Wang, X. M., Wang, S. Y., Yu, L. B., Armanbek, G., Yu, J., et al. (2022). Towards a more labor-saving way in microbial ammonium oxidation: a review on complete ammonia oxidization (comammox). *Sci. Total Environ.* 829:154590. doi: 10.1016/j. scitotenv.2022.154590