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# Phytoplankton carbon to chlorophyll *a* model development: a review

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The cellular carbon content and chlorophyll *a* (Chl *a*) concentration are two of the most significant indices for assessing phytoplankton biomass. Recording and monitoring these biomasses are essential tasks in phytoplankton research, and the carbon-to-chlorophyll *a* (C:Chl *a*) ratio serves as a crucial conversion tool between them. Although the C:Chl *a* ratio varies widely, it is influenced by external environmental factors, making modeling studies of C:Chl *a* particularly important. This paper provides an overview of the historical development of the C:Chl *a* model, beginning with early empirical models and progressing to the development of mechanistic models. This discussion is followed by an examination of existing gaps and future challenges in current C:Chl *a* modeling, particularly the potential underestimation of carbon biomass in existing C:Chl *a* models for dinoflagellates exhibiting multiple growth strategies. Finally, it is suggested that future C:Chl *a* models should strive to achieve a balance between reliability and applicability.

## KEYWORDS

phytoplankton, carbon to chlorophyll *a* ratio, empirical model, mechanistic model, environmental regulation

## 1 Introduction

Unicellular phytoplankton are significant contributors to ocean productivity, accounting for approximately half of global primary production (Worden et al., 2015; Levine and Leles, 2021). They play a crucial role in global biogeochemical cycles and climate change (Falkowski, 1994; Boyce et al., 2010; Litchman et al., 2015; Behrenfeld et al., 2016). Phytoplankton metabolic processes significantly influence global marine carbon and oxygen production. In addition to regulating food web dynamics, phytoplankton are responsible for cycling essential biological elements such as carbon and nitrogen (Litchman et al., 2015; Behrenfeld et al., 2016).

Phytoplankton in the surface ocean convert dissolved carbon and nutrients into organic matter, which is then transported downward through sinking particles. This process creates a biological pump that reduces atmospheric CO<sub>2</sub> concentrations (DeVries et al., 2012;

Agustí et al., 2015; Tréguer et al., 2018). Understanding the elemental composition of marine photoautotrophic phytoplankton is crucial for studying global ecosystems and Earth's climate (Falkowski, 2012; Schoo et al., 2013; Kwiatkowski et al., 2018). Several studies on phytoplankton conducted in recent years have demonstrated that community structure, composition, and diversity significantly influence phytoplankton stoichiometry (Tréguer et al., 2018; McCain et al., 2021). This stoichiometry, in turn, reflects the quality of the food supply and the pathways of energy transfer in tropical ecosystems (Sardans et al., 2021).

Because of its widespread presence in all phytoplankton taxa, including both prokaryotes and eukaryotes, chlorophyll *a* (Chl *a*) concentration is one of the most commonly used indicators of phytoplankton biomass. The measurement of chlorophyll concentration is conducted using various methods, such as spectrophotometry (Ergun et al., 2004), fluorescence (Kruskopf and Flynn, 2006), and high-performance liquid chromatography (Moorhouse et al., 2018). Chl *a* measurements, in addition to being convenient and continuously operable, can be conducted remotely using satellite remote sensing methods. Remote sensing of ocean water color is an effective monitoring tool for the simultaneous observation of phytoplankton distribution at the sea surface on a large scale and for the detection of phytoplankton blooms (Dierssen, 2010; Kaymaz and Ates, 2018; He et al., 2020). However, there were differences in Chl *a* concentrations among taxonomic groups. Letelier et al. (1993) summarized seven distinct taxonomic groups of Chl *a* algorithms based on diagnostic pigments specific to each algal group. The Chl *a* concentration of Bacillariophytes was primarily determined by fucoxanthin, while the Chl *a* concentration of Dinoflagellates exhibited a linear correlation with peridinin. The Chl *a* concentration of Cyanobacteria was influenced by zeaxanthin and chlorophyll *b* (Chl *b*). Additionally, the Chl *a* concentration of Chrysophytes was linearly correlated with 19'-butanoyloxyfucoxanthin, and the Chl *a* concentration of Prymnesiophytes was linearly correlated with 19'-hexanoyloxyfucoxanthin. The Chl *a* concentration of Prasinophytes was linearly correlated with prasinoxanthin, and the Chl *a* concentration of *Prochlorococcus* spp. was determined by both Chl *b* and prasinoxanthin. It is important to note that the phytoplankton community counts also included both mixotrophic and heterotrophic dinoflagellates that do not contain Chl *a*. In addition, because changes in Chl *a* concentration may not be influenced by phytoplankton carbon biomass, Chl *a* concentrations do not fully represent phytoplankton biomass (Jackson et al., 2017; Sathyendranath et al., 2020).

To address the limitations of Chl *a* as an indicator of biomass, scientists have employed alternative biological indicators, with biological carbon being the most widely used. Carbon is typically found in high concentrations relative to other elements, and importantly, biomass measured using biocarbon markers can be directly linked to the carbon cycle (Graff et al., 2012). However, unlike Chl *a*, directly measuring phytoplankton carbon content is quite challenging. Currently, our techniques do not permit direct monitoring of phytoplankton carbon biomass over large areas of the ocean. Traditionally, phytoplankton carbon biomass has been estimated from cell volume through microscopic counts, which

are then converted to carbon per cell. This information is subsequently used to determine the distribution and variation of carbon biomass throughout the community (Menden-Deuer and Lessard, 2000). However, the methods of microscopic counting and estimating cell volume for biomass still have limitations. Some species, such as diatoms, possess a large cell volume that is primarily occupied by a vacuole, which may lead to an overestimation of their cell carbon content (Hansen and Visser, 2019).

To estimate the carbon content of phytoplankton based on its Chl *a* concentration, a conversion factor known as the phytoplankton carbon to Chl *a* ratio (C: Chl *a*), has been established (Strickland, 1960). This conversion requires quantifying the C:Chl *a* ratios, as they are essential for relating Chl *a* concentration to carbon biomass. However, the C:Chl *a* ratio in phytoplankton is not constant; it varies not only between species but is also significantly influenced by external environmental factors such as light, nutrients, and temperature (Geider et al., 1997; Behrenfeld et al., 2002; Wang et al., 2009; Li et al., 2010). Therefore, the accuracy with which ecosystem models can predict the C:Chl *a* ratio in natural ecosystems is one of the fundamental challenges in contemporary ocean carbon cycle research (Armstrong, 2006). Phytoplankton cells under nutrient-deficient or senescent conditions exhibited a higher C: Chl *a* ratio compared to healthy phytoplankton cells, and this ratio was significantly influenced by environmental factors. Lorenzen (1968) found that the C:Chl *a* ratio of phytoplankton was positively correlated with light intensity. Eppely determined the C: Chl *a* of phytoplankton in two areas within the La Jolla Sea Area and found that the C: Chl *a* in the saturated nutrient area was about 22-28 g C g Chl *a*<sup>-1</sup>, whereas it was higher in the oligotrophic area, about 91-105 g C g Chl *a*<sup>-1</sup> (Eppely, 1968). Like higher plants, phytoplankton convert solar energy into chemical energy and inorganic matter into storable organic matter through photosynthesis. Factors such as light and temperature can influence this photosynthetic process, thereby altering the carbon to chlorophyll *a* ratio in phytoplankton. Additionally, nutrients, trace elements, and other factors can affect the composition of materials within phytoplankton cells.

In early phytoplankton modeling studies, the relationship between phytoplankton growth rates and various environmental factors was often analyzed separately. Light is the environmental factor most frequently utilized in modeling, and several models have been developed to describe the light dependence of growth and photosynthetic rates. These include modifications of the Monod equation, hyperbolic tangent functions, and Poisson functions (Jassby and Platt, 1976; Aiba, 1982; Yun and Park, 2003; García-Malea et al., 2006; Costache et al., 2013). For nutrients, the Monod equation is commonly used to describe the external nutrient dependence of phytoplankton growth rates, while the Droop equation illustrates the relationship between growth rates and cellular quotas of limiting nutrients (Morel, 1987; Sunda et al., 2009; Mei et al., 2011; Edwards et al., 2012; Thomas et al., 2017). Under equilibrium growth conditions, the Monod and Droop equations align with each other and with Michaelis-Menten nutrient uptake kinetics (Geider et al., 1998). Additionally, various equations describe the relationship between growth rate and temperature (Kremer et al., 2017); however, this relationship is

typically represented using an exponential function or the Arrhenius equation (Ahlgren, 1987; Blackford et al., 2004; Taucher and Oschlies, 2011; Sherman et al., 2016; Barton et al., 2020).

A substantial body of laboratory experiments has demonstrated the reliability of these early models in describing the dependence of phytoplankton growth and photosynthesis on environmental factors (Burmester, 1979; Megard et al., 1984; Sukenik et al., 1987, 1991; Grima et al., 1996; Xin et al., 2010). However, these models are only applicable under the constraints of balanced growth, which necessitates that nutrient uptake, light capture, and carbon fixation be strictly coupled (Eppley, 1980). Furthermore, when phytoplankton are limited by a combination of environmental factors, early modeling efforts lacked consensus on how to integrate these factors concerning growth rates and develop new descriptive equations (Rodhe, 1978; Droop, 1983). Some authors have stated that phytoplankton growth processes are limited by a single factor (Rodhe, 1978), while other researchers have demonstrated that there is a multiplicative interaction between light and nutrient limitations (Droop, 1983). Phytoplankton dynamic models have been extensively utilized in marine biological and global marine ecological models (Moore et al., 2001; Fennel and Boss, 2003; Faugeras et al., 2004; Li et al., 2010; Schourup-Kristensen et al., 2014; Álvarez et al., 2018). However, research has demonstrated that the outputs of these models are not entirely reliable, and that time-series changes in Chl *a*, phytoplankton carbon biomass, and C:Chl *a* ratios do not align consistently with observations (Lefevre et al., 2003; Doney et al., 2009). Therefore, in addition to model simulation studies, there is also a need for evaluation studies to assess the actual performance of the models.

This paper introduces the C:Chl *a* model and reviews the history of the development of this type of model and highlights the characteristics of each of these models. There is limited understanding of how light, nutrients and temperature interact together to regulate large-scale changes in phytoplankton C:Chl *a* in the ocean. The purpose of this study is to describe the research ideas of C:Chl *a* models from empirical to mechanistic models and to explore their gaps by summarizing the existing C:Chl *a* models. Then, the gaps and future challenges facing C:Chl *a* models are discussed in terms of particular limiting circumstances and species differences. Finally, the balance between reliability and applicability of the model is emphasized.

## 2 C:Chl *a* model development: from empirical to mechanistic models

Similar to most biological models, early C:Chl *a* models were based on a substantial amount of laboratory experimental data, with empirical models derived through linear or nonlinear fitting to this data. These empirical models typically describe C:Chl *a* as a function of one or more environmental factors. They are characterized by their simplicity and convenience, allowing for rapid transformation to obtain C:Chl *a* when the parameters of the culture environment are specified.

## 2.1 Empirical models

### 2.1.1 Geider empirical model (1987)

In early models of phytoplankton physiology, C:Chl *a* ( $\theta$ ) was an important variable (Kiefer and Mitchell, 1983; Geider et al., 1986), but its fundamental role was implicit in the model (Steele, 1962; Eppley and Sloan, 1966; Shuter, 1979). Geider (1987) provided an empirical model for the variation of C:Chl *a* with light and temperature for microalgae and cyanobacteria under nutrient-sufficient conditions based on from laboratory experiments data. The results show that under nutrient-sufficient conditions, it is possible to describe most of the variation in  $\theta$  using functions of light intensity, temperature and four coefficients.

By summarizing experiments with 13 phytoplankton (Table 1), Geider found that the light dependence of C:Chl *a* can be expressed as a linear function when a single phytoplankton species is at constant temperature and under

TABLE 1 Microalgae species and culture conditions summarized by Geider (1987).

Species	$T$ (°C)	$I$ ( $\mu\text{mol photons m}^{-2} \text{s}^{-1}$ )
<i>Skeletonema costatum</i>	0	4-46
	5	2-92
	10	5-80
	16	5-94
	22	19-119
<i>S. costatum</i>	20	15-650
<i>Leptocylindrus danicus</i>	5	6-72
	10	6-79
	15	7-132
	20	6-127
<i>Thalassiosira weissflogii</i>	18	30-600
<i>T. weissflogii</i>	20	2-105
<i>Thalassiosira pseudonana</i>	18	14-512
<i>Phaeodactylum tricornutum</i>	23	1-230
	25	12-230
<i>P. tricornutum</i>	25	52-277
<i>Fragillaria crotonensis</i>	20	13-154
<i>Scenedesmus</i> sp.	20	15-82
<i>Nannochloris atomus</i>	23	1-200
<i>Euglena gracilis</i>	25	9-483
<i>Chlorella pyrenoidosa</i>	26	11-80
<i>Dunaliella tertiolecta</i>	34	\
<i>Oscillatoria redekii</i>	15	15-250
<i>Microcystis aeruginosa</i>	29	20-565

nutrient-sufficient conditions:

$$\theta = \theta_0 + \varepsilon \cdot I \quad (1.1 - 1)$$

where  $\theta$  is the C: Chl *a* ratio (g C g Chl  $a^{-1}$ ),  $I$  is the photon flux density ( $\mu\text{mol photons m}^{-2} \text{s}^{-1}$ ).  $\theta_0$  is the value of  $\theta$  at  $I = 0$ , and  $\varepsilon$  is the empirical regression coefficient in g C g Chl  $a^{-1} \text{m}^2 \text{s} \mu\text{mol photons}^{-1}$ . The intercept  $\theta_0$  varied from 6 to 40 g C g Chl  $a^{-1}$ . The slope  $\varepsilon$  varied over a wide range from 0.04 to 1.9 g C g Chl  $a^{-1} \text{m}^2 \text{s} \mu\text{mol photons}^{-1}$ .

Geider found a linear decrease in  $\theta_0$  as the temperature increased from 15°C to 30°C. In addition, the regression coefficient  $\varepsilon$  has a negative exponential relationship with temperature ( $T$ ):

$$\varepsilon = 1.85 * \exp(-0.126T) \quad (1.1 - 2)$$

Combining Equations 1.1-1 and 1.1-2, the change in  $\theta$  can be described in terms of light and temperature:

$$\theta = (a - bT) + c \exp(-dT) \quad (1.1 - 3)$$

where the 4 regression coefficients  $a = 43.4$ ,  $b = 1.14$ ,  $c = 1.85$ , and  $d = 0.126$ , respectively. The variation of  $\theta$  with light at three temperature conditions (5°C, 15°C and 25°C) was predicted using Equation 1.1-3. The results showed that the tendency of  $\theta$  becoming larger with increasing light was more obvious at low temperatures. In addition,  $\theta$  was larger at low temperatures, which is consistent with the results of several studies (Eppley, 1971; Li, 1980; William and Morris, 1982). This phenomenon is widespread in nature (Berry and Bjorkman, 1980; Lapointe et al., 1984), and it is exacerbated by increased light (Liu et al., 2021). This is due to the fact that at low temperatures, phytoplankton can suffer from low-temperature chlorosis due to chlorophyll loss (Eppley, 1971).

Geider empirical model revealed a relationship between C: Chl *a* ( $\theta$ ), light intensity and temperature in microalgae. It was possible to describe changes in  $\theta$  using a single function of temperature and light intensity, independent of species. However, because the study included only a few species and limited observations at low temperatures, this empirical model has limited applicability.

### 2.1.2 Cloern model (1995)

In the previous section, Geider proposed a model for the light-temperature dependence of C: Chl *a* under nutrient-sufficient conditions, but much of the ocean is oligotrophic, so the model is not applicable in these regions. Cloern proposed a new empirical equation that describes much of the variation in Chl: C ratios expressed by phytoplankton of laboratory experiments. The modeling approach differs from that of Geider (1987), which incorporates an explicit link between nutrient-limited growth rates and Chl: C. The Cloern model is based on a theory describing the link between biochemical components and phytoplankton growth proposed by previous modeling studies (Laws and Bannister, 1980; Kiefer and Mitchell, 1983; Laws and Chalup, 1990).

Cloern summarized data from 12 published studies, most of which were conducted under different light-limited or nutrient-

limited conditions and reported temperature  $T$  (°C), irradiance  $I$  (mol quanta  $\text{m}^{-2} \text{d}^{-1}$ ), photoperiod, nutrient-limited growth rate  $\mu'$  ( $\text{d}^{-1}$ ) and Chl: C, to obtain an empirical equation applicable to temperate coastal waters:

$$\text{Chl : C} = 0.003 + 0.0154 \cdot \exp(0.050T) \cdot \exp(-0.059I) \cdot \mu' \quad (1.2 - 1)$$

Cloern model possessed three features: Firstly, Chl: C possesses a minimum value of about 0.003 mg Chl *a* (mg C) $^{-1}$ . Secondly, Chl: C is linearly related to  $\mu'$  when light and temperature are stable (Laws and Bannister, 1980; Sakshaug et al., 1989; Chalup and Laws, 1990). Third, the linear relationship between Chl:C and  $\mu'$  changes with light. For mixed waters,  $I$  is related to the depth of the mixing layer,  $H$  (Cullen, 1990):

$$I = (1/H) \cdot \int_0^H I_\phi \exp(-kz) dz \quad (1.2 - 2)$$

$$= (I_\phi/kH) \cdot [1 - \exp(-kH)]$$

$I_\phi$  is the daily irradiance below the surface. The Monod equation was used to describe the limitation of growth rate  $\mu'$  by nutrients:

$$\mu' = N/(K_N + N) \quad (1.2 - 3)$$

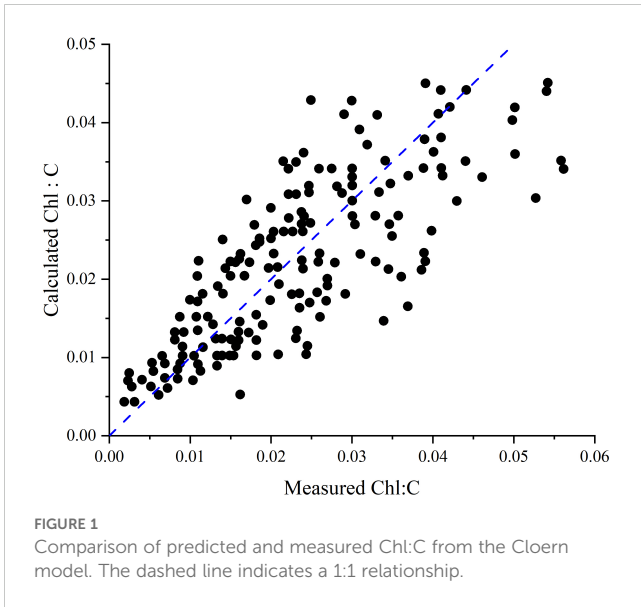
$N$  is the limiting nutrient concentration, and  $K_N$  is the half-saturation constant, i.e., the nutrient concentration at which the growth rate reaches 1/2 of its maximum, which also represents  $\mu'$  sensitivity to changes in nutrient concentration. Combining the three equations gives an equation that is directly described by light, temperature and nutrient concentration together:

$$\text{Chl : C} = 0.003 + 0.0154 \cdot \exp(0.050T) \cdot \exp\{-0.059(I_\phi/kH) \cdot [1 - \exp(-kH)]\} \cdot [N/(K_N + N)] \quad (1.2 - 4)$$

Although the Cloern empirical model is sufficient to describe most of the variability in phytoplankton Chl:C in the steady state of laboratory experiments (Figure 1), a considerable number of questions remain unanswered. Examples include the  $P$ - $I$  parameters (Harding et al., 1981), the dynamics of photoadaptation (Lewis et al., 1984), and differential adaptations between different phytoplankton taxa (Chan, 1980).

### 2.1.3 Shortcomings of empirical models

Although the empirical model initially fit the data well, it was limited to laboratory experiments that overlooked the biological processes of phytoplankton, such as photosynthesis, chlorophyll synthesis, and respiration. This limitation hindered the understanding of how environmental factors specifically affect the physiological state of phytoplankton. In more complex growth environments, the empirical model's fitting results become unsatisfactory, indicating that it is not suitable for estimating the C:Chl *a* ratio of phytoplankton communities in natural waters. Consequently, more comprehensive mechanistic modeling approaches are necessary to address the challenges associated with modeling C:Chl *a*.



## 2.2 Mechanism models

The natural environment is often in a dynamic state of flux, and the cellular chemistry of phytoplankton changes in response to changes in the environment, an adjustment known as acclimation. Typically, environmental acclimation that regulate phytoplankton growth and physiology include principally light, nutrients and temperature. There have been several studies on the individual modeling of these factors on phytoplankton growth or photosynthesis (Jassby and Platt, 1976; McCarthy, 1981; Li, 1980; Droop, 1983; Morel, 1987). In natural environment, phytoplankton growth is often co-regulated by multiple factors, and models of single-factor dependence are not fully applicable.

In both procaryotic and eukaryotic phytoplankton, photoacclimation involves the down-regulation of pigment synthesis at high irradiance (Falkowski and LaRoche, 1991; Alderkamp et al., 2011; McKew et al., 2013). Two of the most widely used indicators of the state of phytoplankton photoacclimation are Chl *a*: C ( $\eta$ ) and Chl *a*-specific light-saturated photosynthetic rate (Geider, 1993). Geider et al. (1996) described a dynamic model of phytoplankton growth and photoacclimation under nutrient-sufficient conditions. The main features of the model are that it explicitly describes the role of the ratio of photosynthesis to light harvesting in regulating the biosynthesis of light-harvesting pigments, shows how light acclimation arises from the dynamic partitioning between intracellular carbon pools, and accounts for equilibrium growth and transient responses to changes in irradiance. Subsequently, Geider et al. (1997) proposed an improvement, with a new model providing an analysis of the response of  $\eta$  to irradiance in equilibrium growth, and the limiting effects of nutrient limitation and temperature on the rate of light-saturated photosynthesis were incorporated into the model.

### 2.2.1 Geider initial model (1997)

C:Chl *a* has a wide range (Taylor et al., 1997; Wang et al., 2009; Jakobsen and Markager, 2016) but is highly regulated by irradiance, nutrient effectiveness, and temperature (Goldman, 1980; Geider, 1987, 1993; Cloern et al., 1995). Geider initially built a dynamic model that C:Chl *a* under different conditions of irradiance, nutrient-limitation and temperature. Changes in Chl *a* depend on the ratio of energy supplied by light absorption and photosynthetic energy conversion to the energy required for growth.

The light dependence of photosynthesis (*P-I* curve) in traditional studies is usually described as a saturation function (Jassby and Platt, 1976). Geider built on this to establish a light (*I*) dependence function for the carbon-specific photosynthetic rate ( $P^C$ ):

$$P^C = P_m^C \left[ 1 - \exp\left(\frac{-\alpha^{\text{Chl}} I \eta}{P_m^C}\right) \right] \quad (2.1 - 1)$$

where  $\eta$  is Chl *a*:C,  $P^C$  is the carbon-specific photosynthetic rate,  $\alpha^{\text{Chl}}$  is the Chl *a*-specific initial slope of the *P-I* curve, and  $P_m^C$  is the carbon-specific light saturated photosynthetic rate. Changes in  $\eta$  were assumed to depend on the relative changes in the rates of net Chl *a* synthesis and net carbon fixation. The net rate of carbon fixation ( $dC/dt$ ) is derived from the difference between photosynthetic and respiration rates (Equation 2.1-2). The net rate of Chl *a* accumulation ( $d\text{Chl}/dt$ ) is derived from the difference between synthesis and degradation rates (Equation 2.1-3):

$$\frac{dC}{dt} = P^C C - R^C C \quad (2.1 - 2)$$

$$\frac{d\text{Chl}}{dt} = \rho_{\text{Chl}} P^C C - R^{\text{Chl}} \text{Chl} \quad (2.1 - 3)$$

$R^C$  and  $R^{\text{Chl}}$  are the degradation rate constants for carbon and Chl *a*, respectively, and  $\rho_{\text{Chl}}$  is the ratio of Chl *a* synthesis to carbon fixation, which can be expressed as:

$$\rho_{\text{Chl}} = \eta_m \left( \frac{P^C}{\alpha^{\text{Chl}} I \eta} \right) \quad (2.1 - 4)$$

$\eta_m$  is the Chl *a*:C maximum. Combining Equations 2.1-1 to 2.1-4 yields a light-dependent function of  $\eta$ :

$$\eta = \frac{\eta_m}{1 + \left( \frac{\alpha^{\text{Chl}} I \eta_m}{2P_m^C} \right)} \quad (2.1 - 5)$$

For temperature and nutrient limitation, Geider assumed that both factors affect only the light-saturated photosynthetic rate, the temperature dependence of  $P_m^C$  was described using the Arrhenius equation (Li, 1980), and the nutrient limitation of  $P_m^C$  was described as the Monod equation (Morel, 1987), with the combined effect of the two factors regarded as a multiplicative function of temperature and nutrient effectiveness:

$$P_m^C(T, N) = P_{\text{ref}}^C \left( \frac{N}{N + K_N} \right) \times \exp\left[ -\frac{E_a}{R} \left( \frac{1}{T} - \frac{1}{T_{\text{ref}}} \right) \right] \quad (2.1 - 6)$$

$P_m^C(T, N)$  is the light-saturated photosynthetic rate at a given temperature and nutrient concentration.  $T$  is the temperature,  $T_{\text{ref}}$  is

the reference temperature of 293 K,  $P_{ref}^C$  is the reference value of  $P_m^C$  under nutrient-sufficient conditions at the temperature of  $T_{ref}$ .  $N$  denotes the nutrient concentration,  $K_N$  is the half-saturation constant of growth, and  $E_a/R$  is the slope of the curve of the Arrhenius equation.

Geider initial model was a more simplified model, and although based on intracellular sources and losses of carbon and chlorophyll, there were still some parts that were not included (e.g., excretion of DOC, biosynthesis consumption of C) in order to make the model easy to use.

### 2.2.2 Geider model (1998)

Based on previous work (Geider et al., 1996; Geider et al., 1997), Geider developed a more complete model of the C:Chl mechanism, which will be referred to simply as the Geider model (Geider et al., 1998) (Table 2). The Geider model was based on the adaptation of phytoplankton growth and physiology to light, nutrient concentration, and temperature, and treats nutrient uptake and photosynthesis rates as a function of environmental factors and cell chemical composition (Chl:C and N:C). Multiple environmental variables together determine the instantaneous rates of light utilization, carbon assimilation, Chl synthesis and nutrient assimilation (Table 3).

The Geider model is based on three features of the phytoplankton adaptation process. First, it includes the down-regulation of pigment content at high irradiance and/or when growth rates are limited by nutrient availability or temperature (Falkowski and LaRoche, 1991; Geider et al., 1996). Second, it includes the accumulation of energy-storing polymers when growth rates are light-saturated and/or nutrient-limited, and the

TABLE 3 The variables of Geider Model (1998).

Variable	Definition	Units
$C$	Phytoplankton carbon	$gC\ m^{-3}$
$N$	Phytoplankton nitrogen	$gN\ m^{-3}$
Chl	Chl $a$	$gChl\ a\ m^{-3}$
$P_{phot}^C$	C-specific rate of photosynthesis	$d^{-1}$
$P_{max}^C$	Maximum value of $P_{phot}^C$ at temperature $T$	$d^{-1}$
$P_{ref}^C$	Value of $P_{max}^C$ at temperature $T_{ref}$	$d^{-1}$
$V_N^C$	Phytoplankton carbon-specific nitrate uptake rate	$gN\ (gC)^{-1}\ d^{-1}$
$V_{max}^C$	Maximum value of $V_N^C$ at temperature $T$	$gN\ (gC)^{-1}\ d^{-1}$
$V_{ref}^C$	Value of $V_{max}^C$ at temperature $T_{ref}$	$gN\ (gC)^{-1}\ d^{-1}$
$\alpha^{Chl}$	Chl $a$ -specific initial slope of the photosynthesis-light curve	$g\ C\ m^2\ (\mu mol\ photons\ g\ Chl\ a)^{-1}$
$Q$	The cell quota of nitrogen (N: C)	$gN\ (gC)^{-1}$
$Q_{min}$	Minimum value of $Q$	$gN\ (gC)^{-1}$
$Q_{max}$	Maximum value of $Q$	$gN\ (gC)^{-1}$
$E_0$	Incident scalar irradiance	$\mu mol\ photons\ m^{-2}\ s^{-1}$
$N_i$	Inorganic nitrogen concentration	$\mu M$
$R^C$	Maintenance respiration rate constant	$d^{-1}$
$R^N$	N remineralization rate constant	$d^{-1}$
$R^{Chl}$	Chl $a$ degradation rate constant	$d^{-1}$
$R_{ref}$	Degradation rate constant at the reference temperature	$d^{-1}$
$T$	Temperature	K
$T_{function}$	Temperature-response function	Dimensionless
$T_{ref}$	Reference temperature	K
$\eta^C$	Chl $a$ to C ratio	$gChl\ a\ (gC)^{-1}$
$\eta^N$	Chl $a$ to N ratio	$gChl\ a\ (gN)^{-1}$
$\eta_{max}^N$	Maximum value of $\eta^N$	$gChl\ a\ (gN)^{-1}$
$\phi_{max}$	Maximum photon efficiency of photosynthesis	$mol\ C\ (mol\ photons)^{-1}$
$\rho_{chl}$	Chl $a$ synthesis regulation term	Dimensionless
$\zeta$	Cost of biosynthesis	$gC\ (gN)^{-1}$
$A_E$	Slope of the linear region of the Arrhenius plot	Dimensionless
$n$	Shape-factor describing dependence of $V_{max}^C$ on $Q$	Dimensionless
$K_{nit}$	Half-saturation constant for nitrate uptake	$\mu M$
$a^{Chl}$	Chl-specific light absorption coefficient	$m^2\ (g\ Chl\ a)^{-1}$

TABLE 2 The equations of Geider Model (1998).

$\frac{1}{C} \frac{dC}{dt} = C_{phot} - R^C - \zeta V_N^C$	(2.2-1)
$\frac{1}{N} \frac{dN}{dt} = \frac{V_N^C}{Q} - R^N$	(2.2-2)
$\frac{1}{Chl} \frac{dChl}{dt} = \frac{\rho_{Chl} V_N^C}{\eta^C} - R^{Chl}$	(2.2-3)
$P_{phot}^C = P_{max}^C \left[ 1 - \exp\left(\frac{-\alpha^{Chl} \eta^C E_0}{P_{max}^C}\right) \right]$	(2.2-4)
$P_{max}^C = P_{ref}^C \left( \frac{Q - Q_{min}}{Q_{max} - Q_{min}} \right) T_{function}$	(2.2-5)
$V_N^C = V_{max}^C \left( \frac{N_i}{N_i + K_{nit}} \right)$	(2.2-6)
$V_{max}^C = V_{ref}^C \left( \frac{Q_{max} - Q}{Q_{max} - Q_{min}} \right)^n T_{function}$	(2.2-7)
$\rho_{Chl} = \eta_{max}^N \frac{C}{\alpha^{Chl} \eta^C E_0}$	(2.2-8)
$R^C = R^N = R^{Chl} = R_{ref} T_{function}$	(2.2-9)
$T_{function} = \exp \left[ A_E \left( \frac{1}{T} - \frac{1}{T_{ref}} \right) \right]$	(2.2-10)

subsequent release of these polymers when light is limited or nutrients are replenished (Foy and Smith, 1980). Third, it includes feedbacks between nitrogen and carbon metabolism. This model differs from previous equilibrium growth steady-state models (Bannister, 1979; Shuter, 1979; Kiefer and Mitchell, 1983; Laws et al., 1983; Geider et al., 1997) in that it explicitly takes into account the time-dependence of biomass and pigment accumulation under non-equilibrium growth steady-state conditions and incorporates both nitrogen limitation and N:C ratio changes are included.

The Geider model is based on phytoplankton mass balances of carbon (C), nitrogen (N) and chlorophyll *a* (Chl) (Equations 2.2-1, 2.2-2, 2.2-3). The light dependence of photosynthesis is considered as a Poisson function (Equation 2.2-4), nutrient uptake rates are described using the Michaelis-Menten function (Equation 2.2-6), and temperature affects light-saturated photosynthesis, maximum nutrient uptake rate, and maintenance respiration rate through the Arrhenius relationship (Equation 2.2-10).

In the Geider model, the constants that maintain C respiration, N remineralization and Chl degradation are assumed to be equal. In the Geider model, the carbon-specific light-saturated photosynthetic rate depends on the nitrogen status inside the cell (Equation 2.2-5), and the carbon-specific light-limited photosynthetic rate depends on the Chl: C ratio (Equation 2.2-4). The synthesis of Chl *a* requires nitrogen assimilation (Equation 2.2-3), and the maximum rate of nitrogen assimilation is regulated by the nitrogen status inside the cell (Equation 2.2-7). Chl *a* synthesis is down-regulated when the rate of light absorption exceeds the photon utilization rate of carbon fixation (Equation 2.2-3), and the extent of the down-regulation depends on the imbalance between the rates of light absorption and photosynthesis (Equation 2.2-8). Respiration rate is linked to the rate of nitrogen assimilation through the cost of biosynthesis (Equation 2.2-1).

The Geider model provides a more complete dynamic model of phytoplankton physiology based on the sources and destinations of elements within the phytoplankton than previous models. The introduction of dynamic regulation of nitrogen in addition to carbon and Chl *a* has linked pigment synthesis to nitrogen assimilation in the model. The Geider model is the first representative model in C:Chl modeling research, and it provides a foundation idea for future C:Chl modeling research.

### 2.2.3 Pahlow model (2013)

In the Geider model, the nutrient dependence of nutrient uptake rates is described using the Michaelis-Menten equation (MM equation), but the MM equation is still flawed. For two limiting nutrients, nitrogen (Rhee, 1974) and phosphorus (Droop, 1974; Rhee, 1974), the MM equation is unable to describe the steady-state uptake response over a wide range of concentrations with constant half-saturation constants,  $K_s$ , and nutrient uptake saturation rates,  $V_{max}$ . Therefore, scientists have proposed a new hypothesis - optimal uptake (OU) kinetics. Optimal uptake (OU) kinetics assumes that there is a physiological trade-off between the efficiency of nutrient contact at the cell surface ( $A_s$ ) and the maximum rate of nutrient assimilation ( $V_{max,s}$ ). The main idea is that phytoplankton change the

number of their surface uptake sites (or ion channels), which determine the time scale of nutrient encounters, and the number of internal enzymes, which assimilate the nutrients after they have been encountered. Nitrogen is recognized as a key resource for partitioning between these two uses, as both uptake sites and enzymes are predominantly composed of proteins and therefore require large amounts of nitrogen (Pahlow, 2005).

Based on this assumption, Pahlow proposed an optimality-based model for phytoplankton (hereafter referred to as the Pahlow model). In the Pahlow model, phytoplankton cells are assumed to transiently optimize the Chl *a* and N content of their cells for maximum net growth, i.e., the net accumulation of carbon fixation minus the energy cost of photosynthesis and nutrient uptake. The N of the Pahlow model is divided into three components: nutrient acquisition, light-trapping organs and structural proteins (Figure 2).

The Pahlow model has three optimization levels without considering P limitation. The first is to optimize the Chl:C ratio ( $\hat{\eta}_0$ ) within the chloroplasts such that net photosynthesis is maximised.  $\hat{\eta}_0$  can be obtained by solving the gradient function of net photosynthesis of a chloroplast with respect to  $\hat{\eta}_0$ :

$$\hat{\eta}_0 = \frac{1}{\zeta^{chl}} + \frac{V_0^C}{\alpha^{chl} I} \left\{ 1 - W_0 \left[ \left( 1 + \frac{R_M^{chl}}{L_d V_0^C} \right) e^{\left( 1 + \frac{\alpha^{chl} I}{\zeta^{chl} V_0^C} \right)} \right] \right\} \quad (2.3 - 1)$$

if  $I > I_0$

$$\hat{\eta}_0 = \eta_{min} \quad \text{if } I > I_0$$

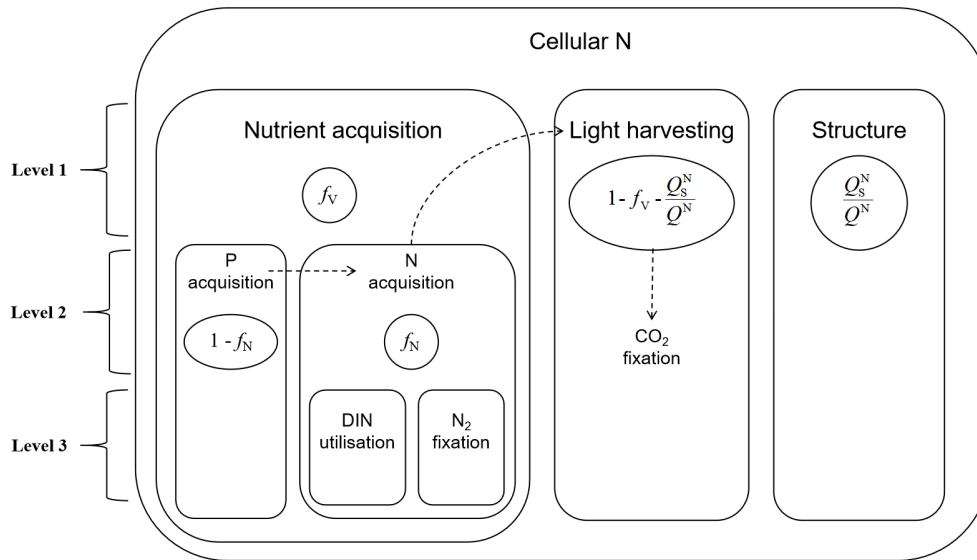
where  $I$  is irradiance ( $W m^{-2}$ ),  $\alpha^{chl}$  is the chlorophyll-specific initial slope of the photosynthesis-irradiance ( $P-I$ ) curve, and  $\zeta^{chl}$  is the cost coefficient of photosynthesis.  $R_M^{chl}$  is the maintenance cost of Chl.  $L_d$  is the length of the day.  $W_0$  is the zero branch of the Lambert-W function.  $\eta_{min}$  is the minimum chlorophyll-to-carbon ratio.  $I_0$  is the threshold light level for chlorophyll synthesis.

$$I_0 = \frac{\zeta^{chl} R_M^{chl}}{L_d \alpha^{chl}} \quad (2.3 - 2)$$

The second level of optimization is to balance the energy allocation between photosynthesis and nitrogen uptake. The net relative growth of a cell is equal to the net photosynthesis of the whole cell minus the respiratory cost of nitrogen uptake. The optimal cellular nitrogen quota,  $Q^N$ , can be calculated by solving the gradient function of net relative growth against the nitrogen uptake allocation factor ( $f_V$ ):

$$Q^N = Q_S^N \left[ 1 + \sqrt{1 + \frac{1}{Q_S^N \left( \frac{\hat{\mu}^I}{V^N} + \zeta^N \right)}} \right] \quad (2.3 - 3)$$

where  $Q_S^N$  is the nitrogen quota for structural proteins, which is half of the minimum nitrogen quota ( $Q_0^N$ ).  $\hat{V}^N$  is the potential nutrient uptake rate.  $\hat{\mu}^I$  is the light-dependent growth rate after accounting for photosynthesis and chlorophyll maintenance.  $\zeta^N$  stands for the cost of nitrogen assimilation.  $f_V$  can be calculated as:



**FIGURE 2** Three levels of nitrogen partitioning in the Pahlow model. Level 1 includes nutrient acquisition ( $f_v$ ), structural proteins ( $Q_s^N/Q^N$ ) and light harvesting ( $1 - f_v - Q_s^N/Q^N$ ). Level 2 divides nutrient acquisition into P acquisition ( $1 - f_N$ ) and N acquisition ( $f_N$ ). Level 3 divides dissolved inorganic nitrogen (DIN) uptake and N<sub>2</sub> fixation.

$$f_v = \frac{Q_s^N}{Q^N} - \zeta^N (Q^N - 2Q_s^N) \quad (2.3 - 4)$$

Chl:C ( $\eta$ ) for the whole cell can be calculated as.

$$\eta = \hat{\eta}_0 \left( 1 - \frac{Q_s^N}{Q^N} - f_v \right) \quad (2.3 - 5)$$

The third dimension of optimization concerns the trade-off between nutrient uptake at the cell surface and nutrient assimilation within the cell (Pahlow, 2005; Smith et al., 2009). The potential nutrient uptake rate ( $\hat{V}^N$ ) is calculated as.

$$\hat{V}^N = \frac{V_0^N N}{\frac{V_0^N}{A_0^N} + 2\sqrt{\frac{V_0^N N}{A_0^N}} + N} \quad (2.3 - 6)$$

where  $V_0^N$  is the maximum potential nitrogen uptake rate.  $A_0^N$  is the maximum potential affinity.  $N$  is the nitrogen concentration in the environment.

The success of the Pahlow model is demonstrated by the fact that it provided the first theoretical derivation of the well-known Droop quota model (Pahlow and Oschlies, 2013) and has been extensively validated (Fernández-Castro et al., 2016; Smith et al., 2016) with laboratory datasets (Pahlow et al., 2013) and ocean observations (Arteaga et al., 2014). However, the Pahlow model is still flawed, and under low light conditions, the prediction of Chl a:C by the Pahlow model may be lower than the actual value (Chen and Smith, 2018).

### 2.3 Applications of several models

Several C:Chl *a* models are also used in a wide range of applications, from indoor incubation to parts of the ocean, and in

global ocean (Table 4). For the empirical models, most of them have been applied in various sea areas, especially the Cloern model (Pondaven et al., 1999; Kettle, 2009; Omta et al., 2009; Itoh et al., 2015; Kooi et al., 2017), due to the advantage of ease of use of empirical models, which can be directly used for estimation. For mechanism models, mostly global ocean and biogeochemical models are applicable, due to the fact that mechanistic models cover phytoplankton biological processes and are closely linked to the ocean carbon cycle. Because the Pahlow model describes the nutrient uptake of phytoplankton cells more rationally, the Pahlow model has begun to gradually replace the Geider model in recent studies (Arteaga et al., 2016; Masuda et al., 2021; Sasai et al., 2022; Kerimoglu et al., 2023).

## 3 Future development challenges

### 3.1 Light

Light is one of the most crucial external factors in phytoplankton modeling, particularly in carbon-to-chlorophyll (C:Chl) modeling. This is because light directly influences carbon fixation and plays a key role in regulating chlorophyll a synthesis. However, there are still aspects of light research that require further exploration.

Most of the data used to validate the models come from laboratory experiments, which are primarily conducted under constant irradiance using artificial light sources. In contrast, the only light source for phytoplankton growth in the natural environment is sunlight, whose intensity varies continuously throughout the day and with depth in the water column. Since the C:Chl model is intended for application in natural



TABLE 4 Application of several types of models.

Models	Application area	Sources
Geider empirical model	Manukau Harbour	Gallegos and Vant, 1996
Cloern model	the Antarctic Circumpolar Current and the North Pacific Ocean	Pondaven et al., 1999
	Mozambique Channel	Omta et al., 2009
	Atlantic	Kettle, 2009
	Kuroshio–Oyashio Extension region	Itoh et al., 2015
	Modeling of marine microplastics	Kooi et al., 2017
Geider initial model	California coastal	Li et al., 2010
	Global ocean	Jackson et al., 2017
	the northern Gulf of Alaska	Coyle et al., 2019
Geider model	Indoor incubation	Armstrong, 2006
		Ross and Geider, 2009
	TOPAZ model	Henson et al., 2009
	ReCOM2 model	Schourup-Kristensen et al., 2014
	Global ocean	Álvarez et al., 2018
	the Southern Ocean	Losa et al., 2019
Pahlow Model	Global ocean	Arteaga et al., 2016
		Masuda et al., 2021
	the North Pacific Ocean	Sasai et al., 2022
	FABM (Framework of Aquatic Biogeochemical Model)	Kerimoglu et al., 2023

environments, this limitation reduces the applicability of laboratory experiment data for the model.

In addition, photoinhibition is a phenomenon that cannot be overlooked. Microalgal cells undergo three processes depending on light intensity (Figure 3). When irradiance exceeds saturating light levels, the photosynthetic system becomes impaired, leading to a decrease in the photosynthetic rates phenomenon known as photoinhibition (Marshall et al., 2000). Due to light attenuation, most of the seawater is in a photo-restricted state, with only a small portion experiencing photoinhibition. However, as mentioned in the introduction, the application of the C:Chl *a* model necessitates the integration of satellite remote sensing data, which measures Chl *a* from the ocean's surface layer, the region most significantly affected by photoinhibition. Álvarez et al. (2018) developed a preliminary C:Chl *a* model that incorporates photoinhibition by including the photoinhibition term function proposed by Platt et al. (1980). Nevertheless, the photoinhibition resulting from the transient inactivation of PSII or the reduction of its absorption cross-section remains inadequately explained.

## 3.2 Nutrients

Currently, nutrient uptake is often described using either the Monod equation or the Pahlow model. Chen and Smith (2018) compared these two models and found that the assumption of a trade-off between photosynthesis and nutrient uptake in the Pahlow model was useful, suggesting that optimal uptake dynamics should be considered in marine biogeochemical models. In the Monod model, only one nutrient is assumed to be limiting for other nutrients such as phosphorus, silicon, and iron. However, Browning and Moore (2023) conducted a global analysis of nutrient limitation in marine phytoplankton and discovered that surface seawater is often close to a state of nutrient co-limitation. Although the Pahlow model incorporates phosphorus uptake, its computational steps are quite cumbersome, making it challenging to obtain a reliable and user-friendly nutrient uptake component for the C:Chl *a* model.

## 3.3 Temperature

In most cases, the role of temperature is incorporated into the model in a multiplicative form. The temperature dependence of phytoplankton growth has been demonstrated by Bernard and Rémond (2012), showing that the growth rate of phytoplankton increases gradually with rising temperatures up to an optimum level, after which it decreases sharply when temperatures exceed this optimum (Figure 4). This sharp decline is attributed to the inactivation of enzymes or denaturation of proteins at elevated temperatures, leading to metabolic disorders and potentially algal cell death (Renaud et al., 2002; Ras et al., 2013; Serra-Maia et al., 2016). The Arrhenius equation, while a useful model for representing growth rates at low temperatures, has a limited temperature range in which it is applicable and fails to account for the declining portion of the thermal growth curve. In the Geider model, temperature is assumed to influence only the light-saturated photosynthetic rate (see Section 2.1). Additionally, some ecosystem models treat the respiration rate as a constant, typically set at  $0.01 \text{ d}^{-1}$  (Schourup-Kristensen et al., 2014). Notably, the impact of temperature on respiration is more significant compared to its effect on photosynthesis (Barton et al., 2020). Furthermore, temperature also influences enzyme activity, indicating that reactive processes such as nutrient uptake may be affected by temperature fluctuations. However, nutrient limitation can inhibit the temperature dependence of phytoplankton metabolism (Marañón et al., 2018).

In addition to its effects on phytoplankton cells at the microscopic level, temperature also influences the stratification of the water column, resulting in changes to the vertical distribution of phytoplankton (Berger et al., 2010; Hordoir and Meier, 2012). Over time, seasonal variations in temperature significantly impact phytoplankton communities. In previous studies, the influence of temperature has been regarded as secondary to that of light and nutrients. However, in light of rising global and ocean temperatures, the significance of temperature must be emphasized.

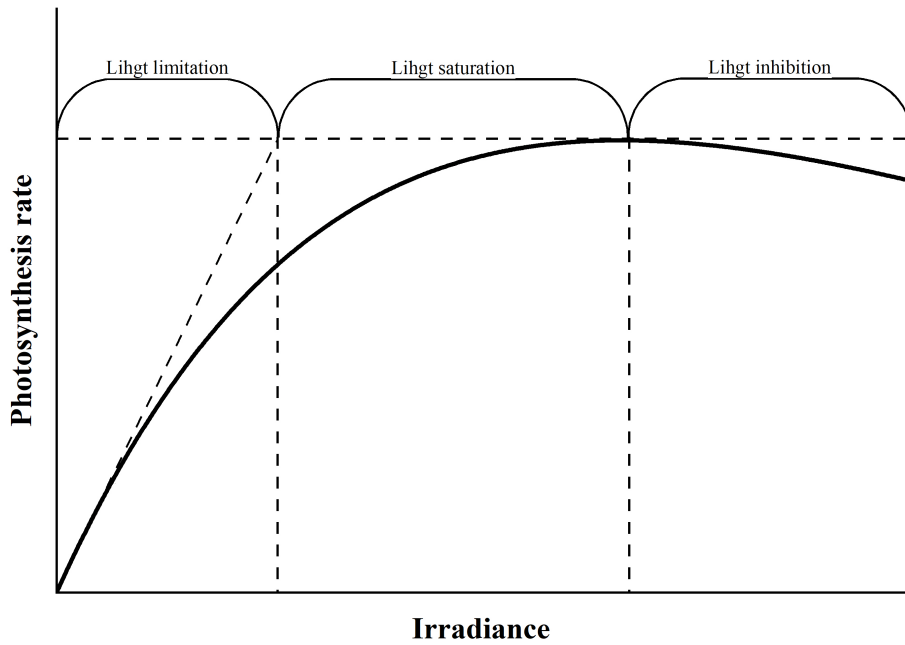


FIGURE 3 Schematic diagram of the *P-I* curve. Phytoplankton cells undergo 3 stages of increasing light: (1) light limitation, (2) light saturation, and (3) light photoinhibition.

### 3.4 Communities

The impact of phytoplankton species has largely been overlooked in current C:Chl modeling studies, as these models typically estimate the carbon biomass of the entire phytoplankton community across various marine environments. However, certain specific cases warrant consideration, including mixotrophic, methanotrophs, and heterotrophs. Most dinoflagellates are mixotrophic and play a dominant role in several marine ecosystems (Burkholder et al., 2008; Taylor et al., 2008; Zhang et al., 2013; Stoecker et al., 2017; Jeong et al., 2021). Cohen et al. (2021) demonstrated that

dinoflagellates are a significant component of the microbial community in the Pacific Ocean, exhibiting high relative abundance across a trophic gradient. Unlike other phytoplankton, dinoflagellates exhibit a diverse range of trophic strategies, including autotrophy, heterotrophy, and mixotrophy (Cohen et al., 2021). This diversity suggests that the C:Chl model may underestimate carbon biomass by not adequately accounting for dinoflagellate mixotrophy when assessing the phytoplankton community dominated by these organisms. It is unclear how much the carbon biomass acquired by dinoflagellates through heterotrophy impacts the total carbon biomass of the entire phytoplankton community.

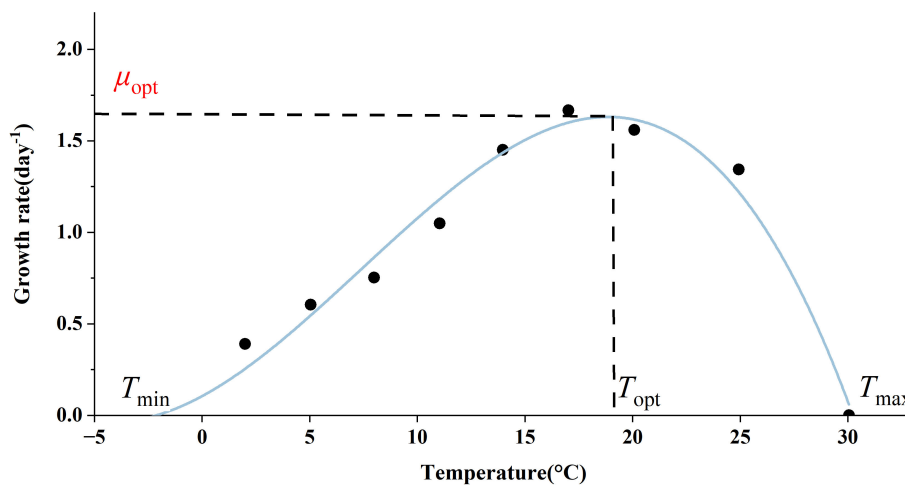


FIGURE 4 Thermal growth curves of microalgae as exemplified by *Astrionella formosa* (from Bernard and Rémond, 2012).

### 3.5 Improved methodology

Currently, mechanistic models, such as the Geider model and the Pahlow model, are widely utilized in biogeochemical modeling studies due to the cycling processes within phytoplankton that involve elements such as carbon (C) and nitrogen (N). However, there are still aspects of the Geider model that can be enhanced. Firstly, for photosynthetic processes (Equations 2.1-1, 2.2-4), a photoinhibition term can be incorporated (Platt et al., 1980).

$$P^C = P_m^C \cdot \left[ 1 - \exp\left(\frac{-\alpha^{Chl} I \eta}{P_m^C}\right) \right] \cdot \exp\left(\frac{-\beta^{Chl} I \eta}{P_m^C}\right) \quad (3.5 - 1)$$

where  $\beta^{Chl}$  represents the empirical photoinhibition parameter. The inclusion of this term enables the Geider model to provide a more accurate estimation of phytoplankton carbon biomass in the ocean's surface waters. Additionally, for nutrient uptake, optimal uptake kinetics are more appropriate than the Monod or Droop equations (Pahlow, 2005):

$$V_{OU} = \frac{1}{(f_A A_{0,S} S)^{-1} + [(1 - f_A) V_{0,S}]^{-1}} \quad (3.5 - 2)$$

$f_A$  is the proportion of the total nitrogen pool allocated to the cell surface, and  $(1 - f_A)$  is the proportion assigned to intracellular enzymes.  $A_{0,S}$  and  $V_{0,S}$  are the potential affinity maximum value and maximum uptake rate for nutrient  $S$ , respectively. Thirdly, irreversible damage caused by high temperatures can be described using the modified Arrhenius equation (Grimaud et al., 2017):

$$T_{function} = k_r \exp\left(\frac{T_A}{T_{ref}} - \frac{T_A}{T}\right) \quad (3.5 - 3)$$

where  $T_{ref}$  is the reference temperature,  $T_A$  is the Arrhenius temperature (i.e., the slope of the straight line in the Arrhenius plot), and  $k_r$  is the reaction rate at  $T_{ref}$ . Currently, there is no superior method to exclude the effects of mixotrophic or heterotrophic dinoflagellates in natural phytoplankton taxa. Additionally, more culture experiments, such as mixed

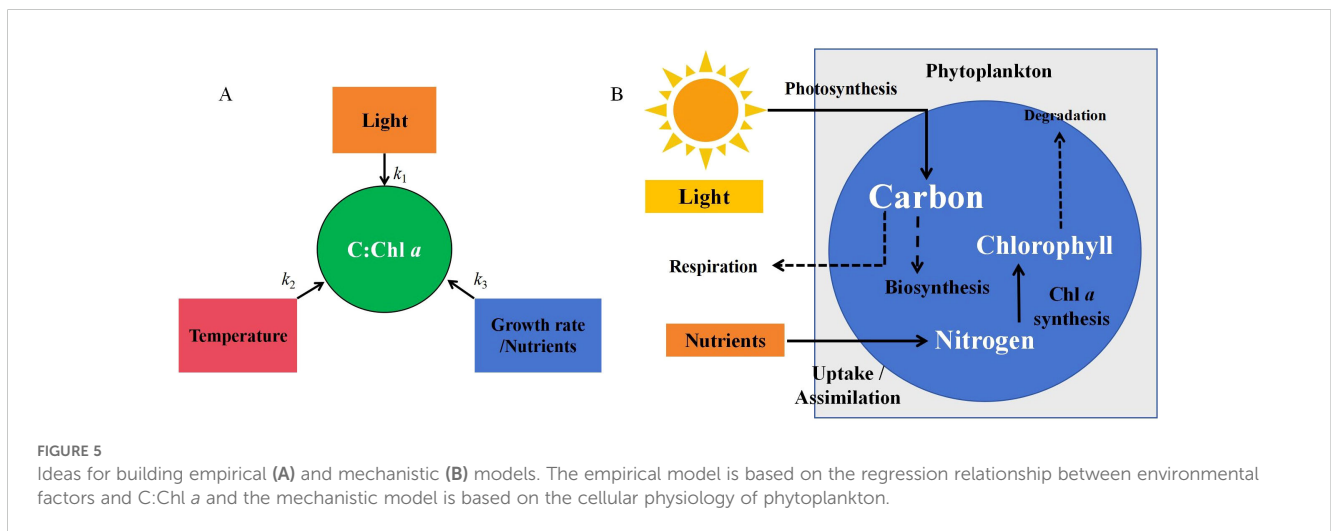
incubations of diatoms and dinoflagellates, are necessary to summarize potential empirical relationships.

### 3.6 Special events

There are several episodic and special events in the oceans that can alter the phytoplankton community and, consequently, its biomass. Phytoplankton blooms frequently occur in both lakes and oceans, some of which are harmful algal blooms (HABs) driven by coastal upwelling or significant anthropogenic nutrient enrichment (Pitcher et al., 2010; Richlen et al., 2010; Andersen et al., 2017; López-Cortés et al., 2019). Research indicates that the frequency and distribution of these blooms are expected to increase with future climate change (Dai et al., 2023). Therefore, it is possible to develop a specialized C:Chl  $a$  model for algal blooms and to analyze later whether it is feasible to predict the likelihood of blooms based on C:Chl  $a$  ratios. The Southern Ocean is commonly characterized as a high-nutrient, low-chlorophyll (HNLC) region, a phenomenon attributed to the limitation of trace element iron (Broecker, 1992). Current projections of carbon biomass for the global ocean primarily rely on Monod dynamics or optimal uptake dynamics that describe nitrogen limitation. However, for the HNLC region, C:Chl  $a$  modeling that accounts for iron limitation is essential.

## 4 Conclusion

The key to the transition of the C:Chl  $a$  model from a simple empirical framework to a dynamically regulated mechanistic model lies in a fundamental shift in modeling concepts. Initially, empirical models were primarily characterized by values of C:Chl  $a$ . While there was some consideration of the individual or combined effects of environmental factors (Figure 5A), the relationship between these factors and C:Chl  $a$  was direct, with only specific regression



coefficients influencing C:Chl *a*. Consequently, the primary purpose of the empirical models was to investigate the mathematical relationships derived from the regression results. Thus, the aim of these models was to explore the mathematical connections within the regression findings. There has been significant advancement in the theory of mechanistic models, including the Geider model, which is based on cellular elemental sources and destinations, and the Pahlow model, which focuses on optimal uptake kinetics. Both models are grounded in the cellular physiological regulation of phytoplankton (Figure 5B). The regulatory role of environmental factors is more clearly defined in these models.

This paper reviews the C:Chl *a* models that have been commonly applied to date, which primarily consider the combined effects of light, temperature, and nutrients on phytoplankton biological processes. However, these considerations are not comprehensive. C:Chl *a* models that incorporate photodamage still fail to explain the photoinhibition caused by the transient inactivation of PSII or the reduction of its absorption cross-section. While optimality absorption kinetics may offer a more plausible explanation, its computational steps are significantly more complex. The impact of temperature is likely underestimated and should be more thoroughly considered in the context of global warming. Additionally, since dinoflagellates exhibit multiple trophic strategies, new models may need to be incorporated when dinoflagellates are dominant to minimize the underestimation of carbon biomass.

In addition to the necessity of being as comprehensive as possible to encompass all types of scenarios and enhance the reliability of the output results, the modeling process should not be overly complex. For instance, while the Pahlow model's optimality-based assumption makes it superior to the Monod equation in describing the nutrient uptake of phytoplankton cells, it is rarely employed in marine biogeochemical modeling due to its complicated computational requirements. Consequently, the model's structure should remain straightforward and user-friendly.

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JS: Funding acquisition, Writing – review & editing. JG: Writing – original draft.

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## Conflict of interest

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