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Water mass characteristics and hydrodynamics at an inshore versus an offshore mid-Norwegian cold-water coral reef habitat

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Introduction: Cold-water coral reefs form complex benthic habitats, supporting thousands of species. The broadscale environmental tolerances of reef-forming species such as *Lophelia pertusa* are well studied, but small-scale differences between different reef settings have received little attention so far. The controlling factors of thriving cold-water coral reefs and how these habitats differ in terms of framework extent, coral colony morphology, and associated fauna could reveal how these benthic ecosystems form and expand. Information on the natural range of environmental fluctuations could provide a better understanding of the resilience of such ecosystems towards environmental changes. Our study aimed to elaborate small-scale forces on local hydrodynamics and oceanographic parameters at two geographically close but contrasting reef sites in mid-Norway.

Methods: We investigated natural fluctuations and the seasonal variability of environmental conditions of an inshore and an offshore *Lophelia*-dominated reef over an annual cycle by time series monitoring of physical properties by benthic landers and water sampling for biogeochemical variables using CTD casts.

Results and discussion: The flow fields at the extensive reef on the offshore Sula Ridge and a bank reef at Nord-Leksa in a fjord-system differed regarding both short-term and seasonal levels. The inshore flow field was strong and tidally driven, whereas the offshore flow field was slower with large seasonal variability. The local flow regimes and the seasonal atmospheric forcing could explain the observed seasonality of the hydrographic variables and the observed inter-annual variability in biogeochemical variables. Comparison with a flow model showed that the natural short-term and seasonal variability are driven by small-scale forcing that is not represented in model analyses. These results suggest that local hydrodynamics together with sea-floor topography control the reef extent and the morphology of cold-water coral colonies.

KEYWORDS

cold-water corals, *Lophelia pertusa*, *Desmophyllum pertusum*, long-term monitoring, benthic landers, environmental characteristics, biogeochemistry, coral morphology

1 Introduction

Cold-water corals (CWCs) build vast ecosystems, promoting high biodiversity and high species richness in the deeper ocean (Freiwald et al., 2004) equally remarkable as in shallow-water tropical coral reefs (e.g. Freiwald et al., 1999; Henry and Roberts, 2017). The ecological requirements and tolerance thresholds of living reef-forming CWCs have been assessed by field-based analyses and habitat suitability modelling (e.g. Guinotte et al., 2006; Davies et al., 2008; Dullo et al., 2008; Freiwald et al., 2009; Davies and Guinotte, 2011; Brooke et al., 2013; Flögel et al., 2014; Zheng and Cao, 2014; Tong et al., 2023). Reef-forming CWCs are predominantly found in waters with temperatures between 4 and 14°C, a salinity range of 32 to 38 (Freiwald et al., 2004; Taviani et al., 2005; Freiwald et al., 2009), oxygen levels of > 2 mL L⁻¹ (Fink et al., 2012) and aragonite saturation (Ω_{Ar}) levels around or above the aragonite saturation horizon (ASH) of 1 that promote calcification (Davies and Guinotte, 2011). In the Northeast Atlantic, a specific density layer (sigma-theta of 27.35–27.65 kg m⁻³) (Dullo et al., 2008) and low levels of dissolved inorganic carbon (DIC) (< 2,170 $\mu\text{mol kg}^{-1}$) (Flögel et al., 2014) are linked to healthy CWC occurrences. Moreover, CWCs are associated with high surface productivity in combination with strong tidal bottom currents, i.e. regions where fresh labile food particles are transported rapidly from the surface and where an oscillating water flow enhances food supply (e.g., Freiwald et al., 2004; Thiem et al., 2006; Dorschel et al., 2007; Kiriakoulakis et al., 2007; Davies et al., 2009; Soetaert et al., 2016; Juva et al., 2020). Recently, Maier et al. (2023) reviewed the ‘paradox of thriving cold-water coral reefs in the food limited deep sea’ and compared habitat suitability models predicting reef-forming corals with regard to surface productivity and currents. In their global analysis, the authors found that currents/flow speed is a more important driver for CWC reef growth than primary productivity, with the majority of reef-forming corals occurring in areas with above global-average current velocity, while corals face a broad range of primary productivity conditions. Hydrodynamic processes such as internal tidal activity and seasonally driven oscillations cause periodic food pulses to the corals at varying temporal scales rather than constant food supply (Soetaert et al., 2016; Osterloff et al., 2019; de Froe et al., 2022; Maier et al., 2023).

Local hydrodynamics also plays an important role in the morphology of branching corals, i.e. the colony shape depends on the flow speed and direction. Unidirectional flow is linked to ‘bush-like’ growth patterns (Wilson, 1979) with living corals facing the main flow direction (Wagner et al., 2011; Chindapol et al., 2013). In more anisotropic or multi-directional flow fields, coral branches grow in multiple directions creating ‘cauliflower’-shaped colonies (Freiwald et al., 1999). The latter type is more compact and allows for more stability in environments with strong bottom currents, while bush-like colonies are more expanded and have thinner branches that provide better accessibility for capturing food particles (Hunter, 1989). Polyp size is further thought to be related to the size range of utilised food particles (Buhl-Mortensen and Freiwald, 2023) and may thus help to understand the mechanisms of food capture. In Norway, *Lophelia*’s primary food source are crustacean plankton, predominantly copepods

(Järnegren and Kutti, 2014). On the reef scale, the large three-dimensional frameworks created by the corals further influence the flow patterns locally themselves, with their complex structure deflecting near-bed flow and decreasing turbulence levels and flow velocity (Mienis et al., 2019; Bartzke et al., 2021; Corbera et al., 2022), allowing for optimal flow for prey capture and sediment baffling (Hennige et al., 2021; Sanna et al., 2023).

Most of the CWC reefs in the North Atlantic are built by three main species, *Lophelia pertusa* (also referred to as *Desmophyllum pertusum* (Addamo et al., 2016)), *Madrepora oculata*, or *Solenosmilia variabilis* (e.g. Freiwald et al., 2004). Along the continental shelf off Norway *L. pertusa* is the main reef-framework forming species in the Northeast Atlantic (e.g. Buhl-Mortensen et al., 2015a, b). Of all known *L. pertusa* occurrences, 30% are located in Norwegian waters (Järnegren and Kutti, 2014). One of the most pronounced occurrences on the Norwegian margin is the Sula Reef Complex (Figure 1), which stretches about 14 km on the Sula Ridge off the coast of Trøndelag (county with Trondheim being the largest city) in mid-Norway (Freiwald et al., 1999, 2002; Hovland et al., 2005). The Sula Reef Complex consists of ~ 1,000 individual coral mounds (Thorsnes et al., 2016) with Holocene age reaching up to 8,000 years (e.g. Hovland et al., 1998; Freiwald et al., 1999; Hovland et al., 2005). Living *L. pertusa* colonies and the associated fauna thrive at relatively stable environmental conditions with temperatures between 7.6 and 7.8°C and salinities between 35.05 and 35.24 (Freiwald et al., 2002). Cold-water coral reefs off Norway are, however, not restricted to the continental shelf, but extend to the coasts and to narrow fjords from the Oslofjord in the south (Purser et al., 2009) to the Stjernsund in the north (Rüggeberg et al., 2011). Here, CWC occurrences are found on sills, banks and on vertical fjord walls (e.g. Mortensen and Fosså, 2001; Fosså et al., 2002; Hovland et al., 2002; Juva et al., 2021). The Trondheimsfjord in Trøndelag harbours several CWC reefs (Buhl-Mortensen et al., 2015a) including the Nord-Leksa Reef near the fjord entrance (Mortensen et al., 2001; Scheide, 2018) and the shallowest CWC reef known to date at 39–80 m on the Tautra Ridge (Zibrowius, 1980; Fosså et al., 2002).

The environmental conditions in mid-Norway, both on the continental shelf and within the fjords, are driven by large-scale ocean circulation and seasonal forcing. The current system of the Norwegian shelf is dominated by the North Atlantic Current (NAC) and the Norwegian Coastal Current (NCC) (Sætre and Ljøen, 1971; Mork, 1981), which flow northwards along the Norwegian coast (Figure 1C). The NAC brings warm and salty (> 35 g kg⁻¹) Atlantic Water (AW) to the shelf area. North of 63°N, AW is found beneath 100–150 metres (Sætre, 1999). The NAC flows as a slope current with its main core offshore of the shelf break. The NCC flows over the Norwegian shelf, supplied with brackish waters from the Baltic Sea, fjords and rivers along the Norwegian coast, forming fresher (< 35 g kg⁻¹) Norwegian Coastal Water (NCW) (Haugan et al., 1991; Milzer et al., 2013). Its width varies due to season and weather conditions, but on average it stretches 100 km from the coast (Leinebo, 1973). As an effect of the topography, the NCC splits into two northward-flowing branches at about 63°30’N (Ljøen and Nakken, 1969; Eide, 1979; Poulain et al., 1996): the slow meandering branch follows the shelf break on top of the AW, and the fast and more stable branch

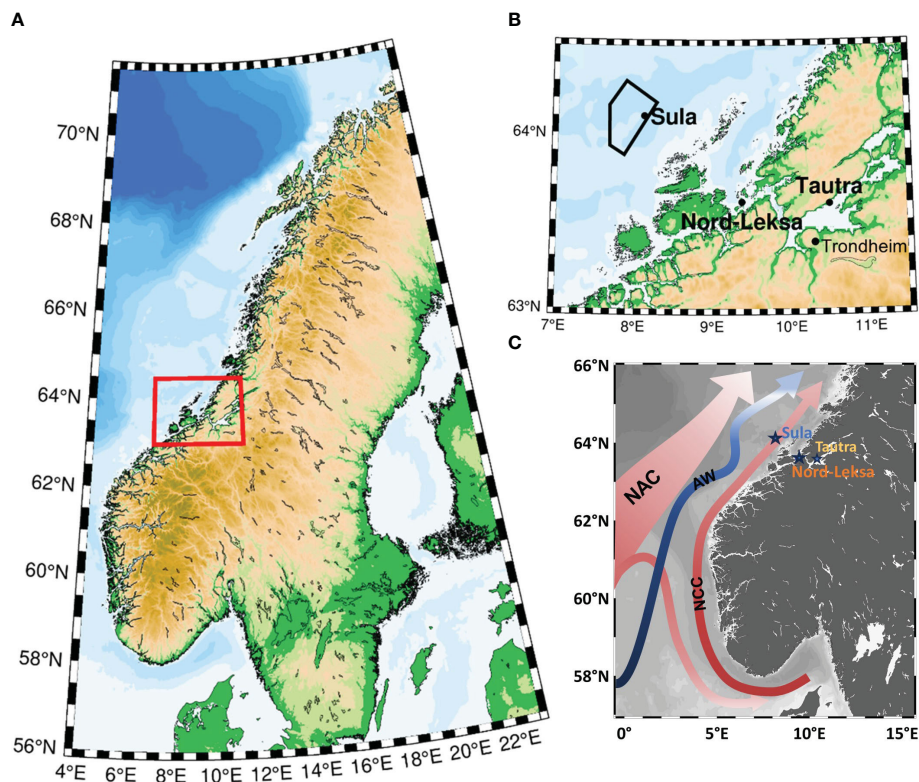


FIGURE 1

(A) The study area (red rectangle) at the coast off mid-Norway. (B) Close-up of the study area with the three study sites: Sula Reef, Nord-Leksa Reef and Tautra Reef. The lander system deployments took place at Sula and Nord-Leksa, while at Tautra, in-reef water sampling was carried out during the first RV *Poseidon* survey in summer 2013. The Sula VME (Vulnerable Marine Ecosystem) protected area is shown with black polygon. (C) Map of south- to mid-Norway showing the prevailing current regime. Surface currents are indicated in red including the North Atlantic Current (NAC) and The Norwegian Coastal Current (NCC). The blue arrow highlights the main northwards moving intermediate water mass Atlantic Water (AW).

flows along the coast (Sætre, 1999). The winter cooling of the upper few tens of metres creates a cool and fresh seasonal water mass – the Winter Mode Water (WMW) – which is less dense but cooler than the NCW. Subsequent warming of the surface waters during spring and summer shifts the WMW into deeper layers of the water column.

Besides the flow regime, the bedrock morphology beneath the CWC reef sets limits to the reef extent (Freiwald et al., 1999). The Sula Reef coral mounds and patches cover areas of several km² with a vertical extent of up to 35 m height and diameters up to 100 m (Thorsnes et al., 2016). The Ridge extends from 230 to 330 m and the largest continuous coral cover is found between 270 and 310 m depth (Freiwald et al., 1999). The highest abundance of CWC mounds in Sula is reported at the steepest part to the southwest of the ridge (Thorsnes et al., 2016). To the northeast, single discrete patch reefs are more common (Mortensen et al., 1995). The Nord-Leksa reef system has two reef tops with the reef area being restricted to a narrow depth range of 138–180 m, but the horizontal extent covering almost 2.3 km² (Scheide, 2018). In comparison, Scheide (2018) estimated the coral cover on Tautra sill to be 0.4–0.8 km² in shallower water depths of 39–80 m. The Tautra CWC site includes several discrete reefs as well (Mortensen and Fosså, 2001).

CWCs within the fjords are likely to experience different seasonal and short-term variability in environmental conditions than CWCs on the shelf, but most studies represent only

snapshots in time of point measurements at specific sites (e.g. Davies et al., 2010; Mienis et al., 2012). Small-scale variability and seasonal or tidal differences are therefore often not considered, but their influence on CWC growth is thought to be critical (Rüggeberg et al., 2011; Findlay et al., 2013; Hennige et al., 2014; Juva et al., 2020). Model approaches comparing two distinct CWC sites in the Northeast Atlantic, one at the Logachev mound province (Southeast Rockall Bank) and one at Condor Seamount (Azores), suggest that changes in small-scale hydrodynamics responding to basin-scale changes in water mass properties and currents through contrasting states of the Atlantic Meridional Overturning Circulation (AMOC), may affect coral occurrences at SE Rockall (with stronger bottom currents and cooler and less saline waters during strong AMOC), whereas waters at Condor Seamount at coral depths remained largely unaffected by AMOC changes (Mohn et al., 2023), demonstrating that some reef sites are periodically exposed to greater variations in environmental conditions, which is likely going to be amplified in the future under ongoing ocean change. In order to resolve small-scale spatial and temporal differences in the environmental conditions of CWC sites, we investigated natural fluctuations and the seasonal variability at two close but contrasting CWC reefs in mid-Norway by long-term observational monitoring. For this, we deployed three benthic

lander systems at thriving offshore and inshore CWC reef sites off Trondheim over an annual cycle. At the time of deployment and recovery of the landers, the water column was additionally characterised by CTD casts, and video footage from the reefs was examined regarding differences of the corals' morphology and the associated fauna. By using a multidisciplinary approach, we investigated the interconnection of biotic and abiotic processes on various scales such as the effect of the reef structure on the local hydrographic and biogeochemical settings, advancing our current understanding of the feedback mechanisms of these important marine ecosystems to the hydrodynamic, biochemical, and geomorphological boundary conditions, which support coral growth in Norwegian reefs.

2 Material and methods

2.1 Studied reef sites

The two studied mid-Norwegian *Lophelia* reef sites represent offshore and inshore environmental conditions of established CWC ecosystems (Figure 1). Offshore, the approximately 14 km long Sula Reef Complex on the Sula Ridge off the coast of Sør-Trøndelag was investigated in the mid-western area of the reef chain at about 300 m water depth (Figures 1B, 2A). Inshore, the CWC bank at Nord-Leksa (Nord-Leksa Reef) just outside the entrance of the Trondheimsfjord and about 40 nm from Trondheim was studied at 170–215 m depth (Figure 1B). For the less studied Nord-Leksa Reef, a bathymetric map (Figure 2B) of the reef was generated based

on single beam echo sounder data recorded through the ship-board navigation system and GPS data of the vessel's route to obtain an idea of the dimensions. The Nord-Leksa Reef rises from about 210 to 145 m water depth and has a dimension of about 1,700 m in west-to-east and about 600 m in north-to-south direction, with a saddle-like depression between two main reef tops (Figure 2B). Similar to the Sula Reef, *Lophelia pertusa* is the most dominant coral species at the inshore reef at Nord-Leksa.

In addition to the two studied reef sites, a side experiment was carried out at the shallow Tautra Reef (Figure 1B) at 39–50 m, where water samples for characterisation of the biogeochemistry were taken by divers very close to the polyps. Water sampling at such close proximity to the polyps is usually not possible by CTD sampling in the deeper reef areas and the information might help to further understand the linkages of organic matter recycling processes within the reef.

The geochemical and physical sample collections and measurements as well as benthic lander deployments and recoveries were carried out during two cruises with RV *Poseidon* (GEOMAR, 2015) in June/July 2013 (POS455) and August 2014 (POS473). Lists of stations and water samples are provided in the cruise reports of POS455 and POS473 by (Form et al., 2014, 2015) and are summarised in Table 1.

2.2 Submersible dives

Dives for positioning the lander systems, coral sampling, and characterisation of the reefs via video and still photos were carried out

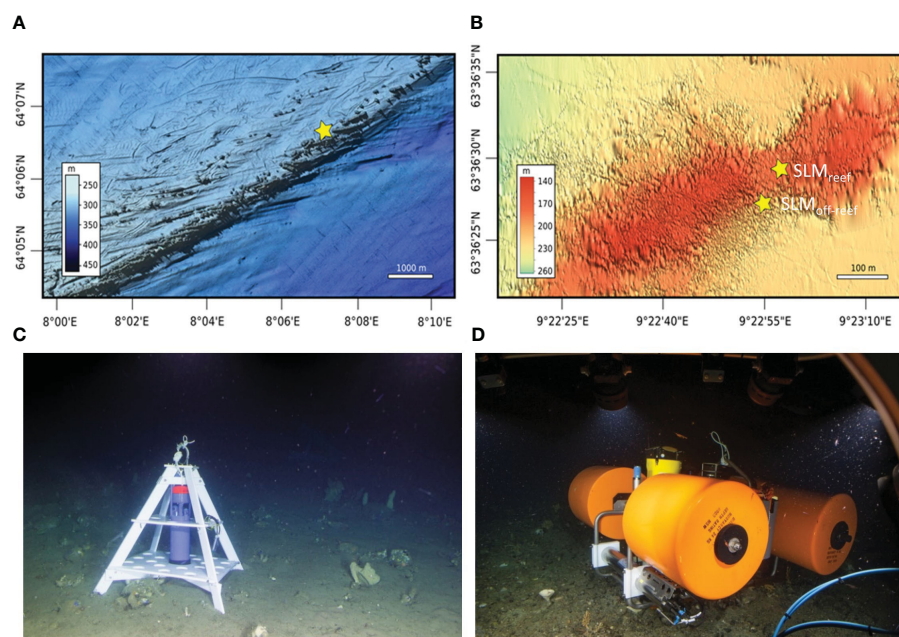


FIGURE 2

Bathymetric maps of (A) the Sula Reef Complex obtained from the 'Marine AREA database for NORwegian waters' (MAREANO) with the water depth profile indicated in a scale bar in metres adapted from Thorsnes et al. (2016), and (B) Nord-Leksa Reef Bank produced based on navigation data during our cruises with RV *Poseidon*. Yellow stars indicate the positions of the SEAGUARD[®] Recording Current Meter (RCM) deployed at Sula (C) and the benthic lander systems (Satellite Lander Module, 'SLM') deployed at Nord-Leksa (SLM_{reef} and SLM_{off-reef}, D) to monitor oceanographic and environmental data such as temperature, conductivity and flow.

TABLE 1 Stations of cruises P455 + P473 with RV *Poseidon* in 2013 and 2014, respectively.

Cruise	Station	Date	Site	Gear	Lat (°N)	Lon (°E)	Depth (m)	Remark
POS455	835/1	29.06.13	Near NLeksa	CTD	63°36.63'	9°19.91'	281	
POS455	837/2	30.06.13	NLeksa	CTD	63°36.57'	9°22.48'	236	WSD: 232, 197, 158, 118, 79, 59, 25
POS455	838/1	30.06.13	NLeksa	JAGO	63°36.43'	9°22.74'	175	
POS455	839/1	01.07.13	NLeksa	JAGO	63°43.31'	9°55.17'	140	
POS455	842/1	02.07.13	NLeksa	SLM _{reef}	63°36.48'	9°22.96'	185	Deployed at 175m
POS455	844/1	02.07.13	NLeksa	SLM _{offreef}	63°36.54'	9°22.93'	216	Deployed at 210m
POS455	846/1	02.07.13	NLeksa	CTD	63°36.40'	9°20.51'	N/A	
POS455	848/1	03.07.13	NLeksa	CTD	63°36.54'	9°22.90'	216	WSD: 211m
POS455	849/1	03.07.13	NLeksa	CTD	63°36.45'	9°22.95'	175	WSD: 172 m
POS455	850/2	04.07.13	Sula	SEAGUARD	64°6.65'	8°7.06'	300	Deployed at 290m
POS455	850/3	04.07.13	Sula	JAGO	64°6.64'	8°7.07'	301	
POS455	851/1	04.07.13	Sula	JAGO	64°6.64'	8°7.13'	299	
POS455	852/1	05.07.13	Sula	CTD	64°6.65'	8°7.23'	300	WSD: 290, 260m
POS455	856/1	07.07.13	NLeksa	CTD	63°36.54'	9°22.96'	214	WSD: 205, 175m
POS455	857/1	07.07.13	NLeksa	CTD	63°36.42'	9°22.93'	217	WSD: 200, 170m
POS455	863/1	11.07.13	Sula	CTD	64°6.70'	8°7.13'	N/A	
POS455	865/1	11.07.13	Sula	CTD	64°6.70'	8°7.02'	302	WSD: 290, 260m
POS473	886-1	18.08.14	NLeksa	CTD	63°36.6'	9° 22.93'	195	WSD: 182,170,20 m
POS473	887-1	18.08.14	NLeksa	JAGO	63°36.5'	9° 23.04'	221	WSD: 180 m
POS473	888-1	18.08.14	NLeksa	CTD	63°36.5'	9° 23.03'	188	
POS473	889-1	19.08.14	NLeksa	CTD	63°36.6'	9° 22.89'	217	WSD: 208,198,20 m
POS473	891-1	19.08.14	NLeksa	CTD	63°36.5'	9° 22.96'	217	
POS473	894-1	20.08.14	NLeksa	SLM _{reef}	63°36.6'	9° 23.02'	207	Recovery
POS473	896-1	21.08.14	NLeksa	SLM _{off-reef}	63°36.5'	9° 22.93'	211	Recovery
POS473	899-1	22.08.14	NLeksa	JAGO	63°36.5'	9° 22.89'	210	WSD: 163 m
POS473	902-1	23.08.14	NLeksa	JAGO	63°36.5'	9° 22.94'	174	
POS473	907-1	25.08.14	NLeksa	JAGO	63°36.5'	9° 22.61'	220	WSD: 219 m
POS473	909-1	26.08.14	Sula	JAGO	64°6.63'	8° 7.07'	304	WSD: 304 m
POS473	910-1	26.08.14	Sula	CTD	64°6.66'	8° 7.12'	301	WSD: 290,280,20 m
POS473	911-1	26.08.14	Sula	JAGO	64°6.63'	8° 7.09'	306	WSD: 275 m
POS473	912-1	27.08.14	Sula	SEAGUARD	64°6.65'	8° 7.14'	300	Recovery
POS473	915-1	28.08.14	Sula	CTD	64°4.91'	8° 1.97'	280	WSD: 270, 260, 20m
POS473	917-1	28.08.14	Sula	CTD	64°4.95'	8° 1.97'	283	

WSD, water sampling depths.

with the manned submersible JAGO (GEOMAR, 2017) at Sula and Nord-Leksa reefs. The highly manoeuvrable research submersible is certified to an operating depth of up to 400 m and accommodates a pilot and an observer. The vehicle is equipped with an underwater navigation and positioning system (USBL), a compass, depth gauges,

vertical and horizontal scanning sonar, underwater acoustic telephone communication, digital video (Full-HD 1080p/50p), still cameras and oceanographic sensors. With JAGO's manipulator arm, sample collection and transportation of small research devices can be carried out such as the positioning of the benthic lander systems.

During the two cruises in 2013 and 2014, JAGO spent 116 hours under water during 40 dives at the two study sites. The JAGO video and still footage was further used to compare general reef characteristics like the reef health status according to Flögel et al. (2014) and the associated fauna between the reef sites. Since the main objectives of the dives were carried out for the purpose of deploying and recovering equipment and collecting samples, the video footage is not suitable for extensive habitat analysis, as distance to the reef changed continuously and no transects were followed. Some marked features can be described and compared between the study sites nonetheless and are described in the results (section 3.1).

2.3 Bottom water monitoring

Three lander systems equipped with oceanographic and environmental data loggers were deployed in the vicinity of the CWC reefs at the two reef locations Sula and Nord-Leksa for over 13 months (Figure 2). The deployment sites were chosen to allow comparison of the hydrodynamics inshore and offshore (Sula vs. Nord-Leksa, Figures 2A, B), as well as the central reef and the surrounding reef margin in Nord-Leksa (Figure 2B). After a detailed survey with JAGO during POS455 in 2013 to find suitable flat bottom topography in the vicinity of live corals for lander deployment, two benthic landers (Satellite Lander Modules, SLM's, GEOMAR, Germany, Figure 2D) were deployed on 2nd July 2013 on the southwestern part of the Nord-Leksa Reef. The first SLM was deployed at 63°36.487'N and 09°22.956'E at a water depth of 175 m within the living reef structure in the valley between the two reef plateaus (referred to as SLM_{reef} hereafter), surrounded by a diverse fauna of many live *L. pertusa*, gorgonians, *Acesta excavata* and *Mycale* sp. within a few metres to the west and dead coral rubble and soft corals to the east. The second SLM was deployed at 63°36.454'N and 09°22.915'E at 217 m depth adjacent to the reef (referred to as SLM_{off-reef} hereafter). The landers were deployed 70 m apart from each other (Figure 2B). Both landers were equipped with the following instruments, located 50 cm above the seafloor: an Acoustic Doppler Current Profiler (SBE 300kHz ADCP, Teledyne RD Instruments) to measure water column flow speed and direction in 10-minute intervals, a conductivity, temperature and pressure measuring instrument (SBE 16 PLUS CTD, Sea-Bird Scientific), and sensors to monitor dissolved oxygen (SBE 43), pH (SBE 27 pH and O.R.P. (Redox) sensor), turbidity and chlorophyll fluorescence (WETLabs ECO-FLNTU(RT)D) measuring in 15-minute intervals.

At Sula, a SEAGUARD[®] Recording Current Meter (TD 262b SEAGUARD[®] RCM, AANDERAA Data Instruments, Bergen, Norway) equipped with a ZPulse Doppler Current Sensor (DCS), a pressure/temperature sensor combination, and a conductivity/temperature sensor combination was mounted in a pyramid-shaped POM frame (Figure 2C) and deployed with the sensors located 75 cm above the seafloor. Oceanographic data comprised tilt-corrected horizontal flow speed and direction, pressure, conductivity, and temperature (both, conductivity and pressure sensors recorded

temperature as well, of which the conductivity sensor-based temperature was used here). The parameters were logged in a 30-minute interval. The SEAGUARD[®] lander was deployed on 4th July 2013 at 64°06.66'N and 08°07.12'E in 305 m water depth 50–75 m from extensive live *L. pertusa* reef framework in the mid-west of the northern Sula Ridge reef area (Figure 2A).

The resolution and accuracy of the measurements of all three landers are shown in Table 2. In August 2014, almost 14 months after deployment, landers were successfully recovered.

In Nord-Leksa, the benthic landers turned out to be too light to cope with the strong bottom currents at the deployment depths. SLM_{off-reef} moved uphill twice during the deployment period, with total vertical movement of 5 m. It first moved between 17th and 19th November 2013 from 216.5 to 214 m, and again between 13th to 14th April 2014 to 211 m depth. SLM_{reef} moved downwards from deployment depth of 175 m to ~ 180 m within the first hour of deployment, where it remained for 5 days. The lander moved several times over the year between depths of 180 and 213 m. At the end of January 2014, the Lander-CTD stopped working at 207 m depth. Both SLMs were stable at the bottom and recording over three periods: 2nd – 7th July 2013, 22nd August – 16th November 2013 and 19th November 2013 – 30th January 2014. The SEAGUARD[®] lander in Sula remained in place and delivered continuous data of all sensors.

TABLE 2 Instrumentation details and metadata for landers and on-board CTDs.

	SEAGUARD [®] lander	SLM landers	CTD RV Poseidon
Flow measurements	Zpulse Doppler Current sensor	RDI Workhorse sentinel 300 kHz	
Velocity acc/res	± 0.15 cm/s/0.01 cm/s	0.5%	
Direction acc/res	± 5°/0.01°		
Echo intensity acc/res		± 1.5 dB	
CTD		SBE 16 plus	SBE 911 plus
T in acc/res [°C]	0.03/0.001	± 0.005/0.0001	± 0.001/0.0002
C in acc/res [mS m ⁻¹]	± 0.18/0.2	± 0.5/0.05	± 0.3/0.04
P in acc/res [% WD ⁻¹]	± 0.02/< 0.0001 (% FSO, 2000 m)	± 0.1/0.002	± 0.015/0.001
Releaser or Deck unit		Video-controlled launcher or K/MT 562	SBE 11 plus deck unit
Additional sensors and instruments		turbidity, dissolved oxygen, pH, chl-a, fluorescence	Fluorescence of chl-a, dissolved oxygen, turbidity, 12×10 rosette

acc, accuracy; res, resolution; T, temperature; C, conductivity; P, pressure; WD, water depth.

From Nord-Leksa, we have reliable time series of ADCP, CTD, oxygen of both landers and turbidity from SLM_{reef}, while the pH sensors of both landers and the turbidity sensor of SLM_{off-reef} malfunctioned. The first five days of deployment were used to compare the environmental conditions within Nord-Leksa and only the measurements of SLM_{off-reef} were used to compare the intra-annual conditions between the inshore and offshore sites. The conductivity sensors had drifts that amounted to -0.3 g kg^{-1} (SLM) and to -1.2 g kg^{-1} (SEAGUARD) over the deployment periods. These drifts have been considered when the hydrographical variables (S_A , Θ and σ_θ) were calculated.

2.4 Water column characteristics and biogeochemistry

To measure water column characteristics at both sites, a CTD system (SBE 911 plus, Sea-Bird Scientific) was used during both cruises. The employed CTD system was built into a rosette housing with 12 10-litre water sampling bottles (Niskin-type). Water samples were taken close to the bottom and in pre-defined depths of the water column (see Table 1). In addition to the CTD system, the rosette frame was equipped with sensors measuring dissolved oxygen, fluorescence of chlorophyll-a, and turbidity. Of the latter two only data from POS455 (2013) are available. Calibrations of the sensors were performed prior to the cruise in the laboratory and all parameters yielded coefficients for a linear fit.

2.4.1 In-reef water sampling at Tautra Reef

At the shallow Tautra Reef in the Trondheimsfjord, water samples were taken by divers with a syringe at close proximity to live coral colonies ($63^\circ36.045'N$, $10^\circ30.466'E$) during POS455 in 2013. The first sample was taken approximately a few tens of metres away from the reef at 50 m water depth on 3rd July 2013. The following day, samples were taken directly from within live coral colonies close to expanded polyps from the lower reef slope at 42 m, the upper reef slope at 40 m and the reef summit at 39 m water depth (Figure 3). Samples were taken using 100 mL syringes (Omnifix[®] Solo, B. Braun Melsungen AG). The water samples remained in the syringes with closed tip until surfacing and were then sterile-filtered (Sartopore[®] sterile capsule, Sartorius Stedim

Biotech GmbH) on-board the diver's zodiac and poisoned with mercuric chloride (HgCl_2) when brought to RV *Poseidon*.

2.4.2 Carbonate chemistry and nutrient analyses

Water samples for measurements of total alkalinity (TA) and dissolved inorganic carbon (DIC) were sterile-filtered (0.2 μm disposable filters, Minisart[®], Sartorius AG), HgCl_2 -poisoned to arrest biological activity, and stored in a cool, dark place until measurement as recommended by Dickson et al. (2007). TA was analysed via potentiometric open-cell titration with an automatic titrator (Titrino 862 Compact Titrosampler, Metrohm). DIC was analysed via infrared detection of CO_2 using an Automated Infra-Red Inorganic Carbon Analyser (AIRICA with LI-COR 7000, Marianda). TA and DIC were calculated accounting for the salinity of the samples and corrected against Certified Reference Materials from A. G. Dickson (Scripps Institution of Oceanography). Additional parameters of the carbonate chemistry were calculated by means of the CO2sys_v.2.1 Excel macro (Pierrot et al., 2011) using the thermodynamic constants of Mehrbach et al. (1973), refitted by Dickson and Millero (1987), on the total scale.

Water samples for analysis of the dissolved inorganic nutrients nitrate (NO_3^-), nitrite (NO_2^-), ammonium (NH_4^+), and phosphate (PO_4^{3-}), were sterile-filtered into 100 mL HDPE vials (Nalgene[®]) and frozen at -20°C until analysis. All nutrients were analysed based on the general methods described by Hansen and Koroleff (1999) using a four-channel Automated Continuous Segmented Flow Analyzer (A3, SEAL Analytical). Detection limits were $0.05 \mu\text{mol L}^{-1}$ for NO_3^- (as combined NO_3^- and NO_2^-), $0.06 \mu\text{g L}^{-1}$ for NO_2^- , $0.02 \mu\text{mol L}^{-1}$ for PO_4^{3-} , and $0.04 \mu\text{mol L}^{-1}$ for NH_4^+ . Nitrate concentrations were determined by subtracting nitrite values from the combined NO_3^- and NO_2^- measurements.

2.5 Data analysis

Further processing of the hydrographic data was performed using the software SBE Data Processing[®] (V7.26.7, Sea-Bird Scientific (2013)) for SLM landers and CTD cast data. For further conversions and visualisation of the oceanographic and environmental data, MATLAB 9.0 (R2016a, The MathWorks, Inc. (1984)) was used. For subsequent data analysis, the raw CTD and ADCP data were converted. CTD data were converted to absolute salinity (S_A), conservative temperature (Θ),

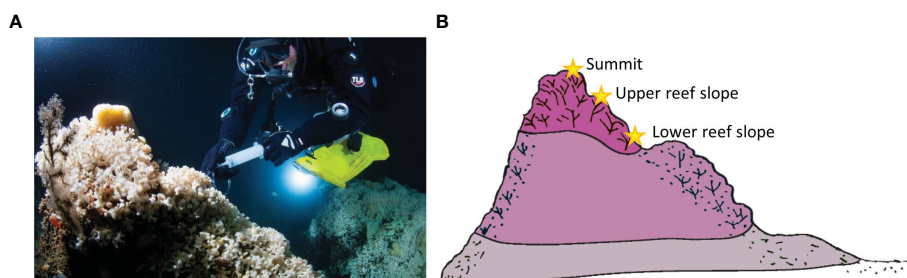


FIGURE 3

Diver taking water samples directly at *Lophelia* reef colonies from the milieu between the skeletal structures at the Tautra Reef (Trondheimsfjord) (A), at the reef summit, the upper and the lower reef slope of live corals (B). Photo credit in (A): Uli Kunz, Schematic (B) modified from Freiwald et al., 2004.

potential density (ρ) and potential density anomaly (σ_θ), i.e. sigma-theta values according to TEOS-10 standard (McDougall and Barker, 2011). The current measurements were corrected for the local magnetic declination based on IGRF-11 model data (Finlay et al., 2010). The site-characteristic dominant tidal frequencies, ω , and their amplitudes, a , were analysed with the harmonic analysis toolbox T_Tide (Pawlowicz et al., 2002). The tidal signals were analysed by using bottom pressure and horizontal velocity fields. Only tidal signals with a signal-to-noise ratio > 2 are considered to be significant.

2.6 Global datasets and oceanographic models

For CWC habitat suitability and conservation analyses, large-scale habitat suitability models are utilised, using global databases and regional ocean models (Davies and Guinotte, 2011; Yesson et al., 2017; Morato et al., 2020). Comparison of field observations to these databases and ocean models is crucial, primarily to ground truth model data and to increase model reliability in general. The carbonate chemistry parameters and nutrients are usually obtained from global databases such as GLODAP (Lauvset et al., 2016) and World Ocean Atlas (Garcia et al., 2013) with low resolution of 1° and down-scaled. Since this resolution is not suitable for inshore sites, we compared the global values from 64.5°N , 8.5°E at 200–250 m depth for Sula and 64.5°N 9.5°E at 125–200 m depth for Nord-Leksa as closest approximation to our studied inshore site. For coastal Norway, the coastal currents are modelled with daily 800-m resolution with the NorKyst-model (Albretsen et al., 2011). This model data was compared with lander deployments from the closest model points for summer 2013 to summer 2014 and with the Tautra Reef area.

3 Results

3.1 Descriptive reef characteristics and biodiversity

Both Nord-Leksa and Sula reef sites belong to category I (healthy reefs and mounds) out of three according to the classification of CWC reef statuses described in Flögel et al. (2014), as both reef habitats are characterised by several 100 m^2 horizontal area and $> 2/3$ vertical extent of living colonies, therewith comprising large biogenic constructions with discrete morphologies (Figure 4). Nevertheless, the colony morphology differs between sites. At Nord-Leksa, the corals form ‘cauliflower’-like colonies with compact polyp arrangement, while at Sula ‘cauliflower’-like colonies are more dispersedly branched with thinner and more extended polyps (Figures 4G, H).

Both sites have rich associated fauna. At Nord-Leksa the living coral reef framework is surrounded by a large coral rubble bed of dead *L. pertusa* fragments from depths of ~ 200 to 160 m, while the reef framework at Sula is surrounded by sandy bottom with sponge-covered boulders or drop stones. In Nord-Leksa, the rubble zone is ‘coral garden’-like, i.e. it is characterised by a benthic community dominated by non-reef-forming soft corals such as the gorgonians

Paragorgia arborea and *Primnoa resaediformes*, *Capnella glomerata*, and *Anthelia borealis*. The rubble zone at Nord-Leksa is abundant in sponges (particularly *Mycale lingua*) and used as a habitat by a range of mobile fauna such as crustaceans (e.g. *Munidopsis* sp.), bivalves (including the large file clam *Acesta excavata*), echinoderms and tunicates. Between about 160 and 152 m, a transition zone follows, which makes up the reef base with first live *L. pertusa* appearing in the white and orange colour morph similar to the Sula Reef, and occasionally *M. oculata* colonies settling on dead coral framework and rubble. Upslope, the reef framework is characterised by an increasing number of live *L. pertusa* colonies accompanied by an increasing number of *M. lingua*, while the abundance of soft corals ceases.

At the offshore study site at Sula, the reef is not surrounded by large rubble fields, but the sandy, rocky seabed (at ~ 302 m) with hard-bottom sponge aggregates transits directly into the base of reef framework of mostly dead *L. pertusa* (~ 302 – 296 m) with a very diverse fauna. Associated species in the Sula Reef encompass the same phyla as in Nord-Leksa but appear to have a larger variety on the species level among poriferans, cnidarians and crustaceans. Numerous large actinians and shrimps (*Pandalus* sp.) are prominent, as well as chaetognaths as abundant zooplankton species. The most remarkable difference between the reef sites (inshore vs. offshore), however, is the abundance and distribution of sponges in the live coral reef zone. The reef top at Nord-Leksa (~ 152 – 140 m) is interspersed with *Mycale* sp. (Figure 4B) and occasionally also gorgonians, cluttered with mobile fauna such as *Munidopsis* sp. and *Gorgonocephalus* sp., whereas the living reef cover at Sula (~ 296 – 283 m) is free of sponges (Figure 4A) and only the dead basis of the reef in between broken-up colonies inhabits a large variety of particularly large sponges (e.g. very commonly *Geodia* sp.) and large gorgonian colonies of *Paragorgia* and *Primnoa*. In comparison to Nord-Leksa, only few *A. excavata* were found at Sula and often only shells. The few individual living *A. excavata* had far brighter, i.e. cleaner shells, whereas at Nord-Leksa shells were overgrown and more strongly eroded by bioeroders.

Similar fish species were spotted at both sites, including saithe (*Pollachius virens*), tusk (*Brosme brosme*), monkfish (*Lophius* sp.), and red fish (*Sebastes* sp.). The latter is the most common species living within the reef framework but was more abundant at Sula. Most common cartilaginous fish included chimaeras at Sula and the velvet belly lantern shark at Nord-Leksa, but also at Sula shark and ray eggs were observed.

3.2 Water column hydrography (CTD casts)

3.2.1 Nord-Leksa

In July 2013, the surface layer was 30 m deep with salinities $< 33.5\text{ g kg}^{-1}$. The WMW was observed as a cool ($\sim 7.5^\circ\text{C}$) and fresh (33.5 – 34.6 g kg^{-1}) layer between 30 and 100 m water depth (Figures 5A, B). It rested above the NCW, which occupied water depths from 100 to 150 m, and the AW in > 150 m water depths (Figure 6). The water column in summer 2013 was therefore composed of surface waters, the WMW, the NCW and the AW. In August 2014, the situation was different, as there was no mid-water temperature minimum (Figure 5A). Below surface waters ($< 33.5\text{ g kg}^{-1}$) the NCW occupied waters

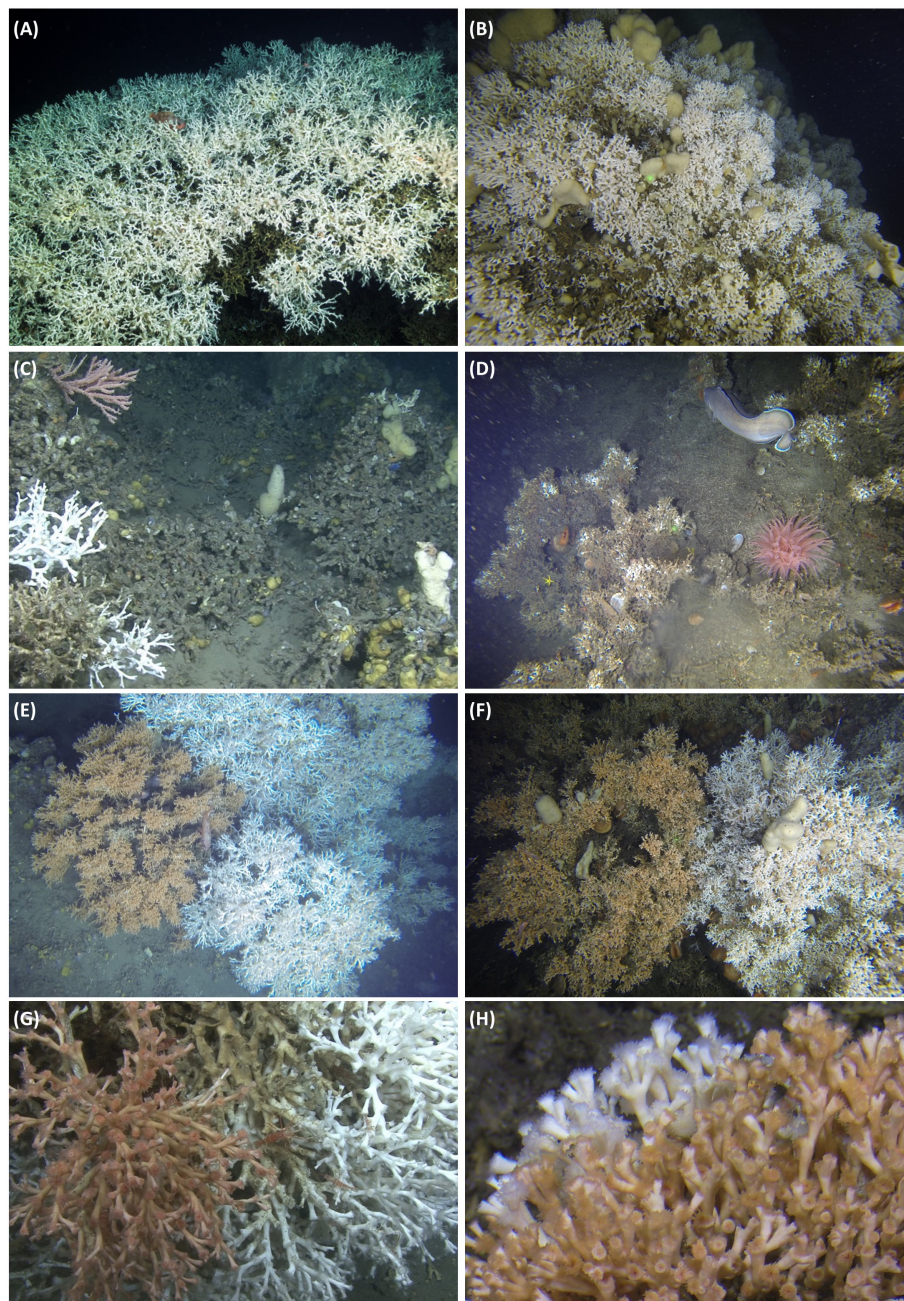


FIGURE 4

Live *L. pertusa* colonies of the study areas at the Sula Reef Complex (left panels) and the Nord-Leksa Reef (right panels) in the vicinity of the deployed lander systems. Pictures (A, B) show the reef top at both study sites, commonly interspersed by *Mycale* sp. at Nord-Leksa Reef (B). (C, D) are pictures from the dead framework zone at the base of the reef, showing the rich and diverse associated fauna. (E, F) show white and orange colonies side by side at both study sites. Finally, (G, H) are close-ups of live colonies demonstrating the different morphologies/growth forms of polyps at Sula and Nord-Leksa. Photographs (A, B) from Max Wisshak, (C-H): JAGO-Team, GEOMAR.

between 10 and 100 metres ($< 8^{\circ}\text{C}$, $< 35 \text{ g kg}^{-1}$) and the AW was found in depths $> 100 \text{ m}$ (Figure 6). The waters were well stratified. In July 2013, the WMW showed lower turbidity values compared to the rest of the water column. Turbidity increased at depths of the CWC reef (150–180 m; Figure 5F). Profiles of oxygen concentration were rather similar between July 2013 and August 2014 with lower values in 2014 (Figure 5D), which could be related to the time-lag of sensor calibration between both years. Oxygen concentrations were supersaturated

in the relatively warm and fresh surface layer of $\sim 20 \text{ m}$. Oxygen saturation decreased with depth to a minimum of $\sim 91\%$ between 60–80 m, below which values increased slightly and remained steadily at values between 92 to 93% up to the seabed.

The environmental parameters at the depths of the CWC reef at Nord-Leksa showed typical values of temperatures of $7.69\text{--}7.76^{\circ}\text{C}$, salinities of $35.01\text{--}35.22 \text{ g kg}^{-1}$, sigma-theta of $27.25\text{--}27.35 \text{ kg m}^{-3}$, oxygen of $5.7\text{--}6.3 \text{ mL L}^{-1}$, fluorescence of $\leq 0.03 \text{ mg m}^{-3}$, and turbidity of $0.11\text{--}0.20 \text{ NTU}$ during both cruises (Figure 5).

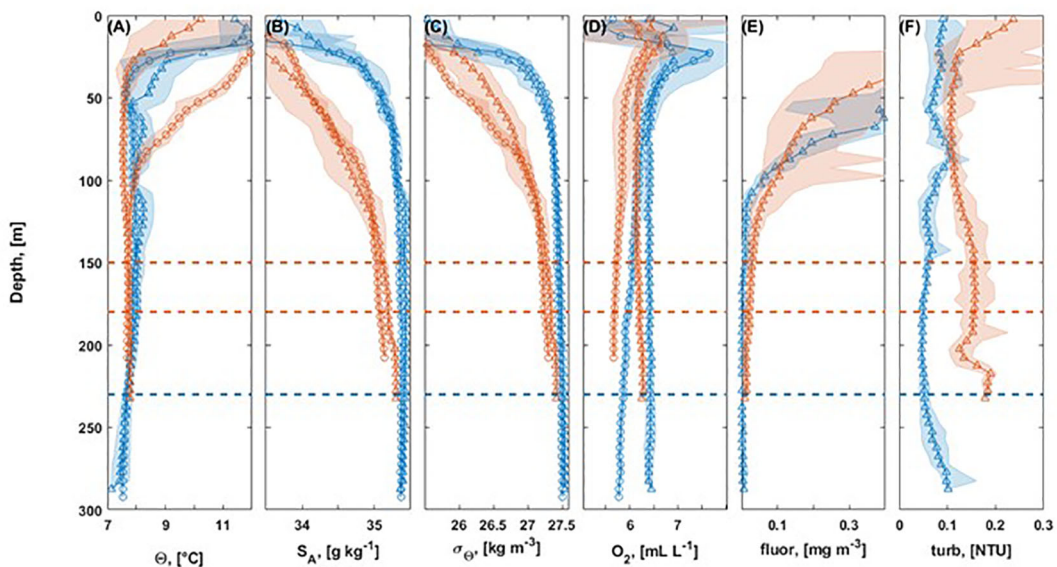


FIGURE 5
Downcast profiles of CTD stations for Sula (blue) and Nord-Leksa (orange) for cruises P455 in July 2013 (triangles) and P473 in July/August 2014 (circles) for conservative temperature (A), absolute salinity (B), sigma-theta (C), oxygen concentration (D), fluorescence (E), and turbidity (F). Dashed lines indicate depths of living cold-water corals at Nord-Leksa (orange) and Sula (blue).

3.2.2 Sula

At Sula, the differences between July 2013 and August 2014 water column characteristics were small compared to Nord-Leksa (Figures 5, 6). In July '13, the water column was saltier and oxygen levels were higher than in August the following year, but in both months, the water column was composed of surface waters (the upper 50 m, $S_A < 35 \text{ g kg}^{-1}$) and the AW below it (Figure 6). In July 2013, the fluorescence peaked at 30 m with 3.14 mg m^{-3} and was $< 0.27 \text{ mg m}^{-3}$ beneath 70 m (Figure 5E). Oxygen profiles at Sula were similar to Nord-Leksa with generally higher oxygen concentrations in both years. Oxygen concentrations peaked at $\sim 23 \text{ m}$ water depth, which coincided with elevated temperatures. Below oxygen concentrations remained steady (2013) or slightly decreased (2014) with depth.

The environmental parameters at depths of the CWC reef at Sula showed typical temperatures of $7.13\text{--}7.86^\circ\text{C}$, salinities of $35.35\text{--}35.44 \text{ g kg}^{-1}$, sigma-theta of $27.50\text{--}27.55 \text{ kg m}^{-3}$, oxygen of $5.7\text{--}6.5 \text{ mL L}^{-1}$, fluorescence of $\leq 0.02 \text{ mg m}^{-3}$, and turbidity of $0.04\text{--}0.16 \text{ NTU}$ during the cruises (Figure 5).

3.3 Short-term environmental variability at Nord-Leksa (Benthic landers)

During the first week after deployment (from 2nd to 7th July 2013), both landers remained stable at their deployment positions (Figure 7). Although SLM_{reef} was located in shallower depths, the water was cooler ($\Theta = 7.78 \pm 0.01^\circ\text{C}$), more saline ($S_A = 35.23 \pm 0.04 \text{ g kg}^{-1}$) and, hence, denser ($\sigma_\Theta = 27.36 \pm 0.03 \text{ kg m}^{-3}$) than at $SLM_{\text{off-reef}}$ ($\Theta = 7.79 \pm 0.01^\circ\text{C}$, $S_A = 35.07 \pm 0.02 \text{ g kg}^{-1}$ and $\sigma_\Theta = 27.23 \pm 0.02 \text{ kg m}^{-3}$) (Figures 7A–F). The variance of the hydrographic variables was higher at SLM_{reef} than at $SLM_{\text{off-reef}}$. Oxygen

concentration levels were similar at both lander positions during the first week of deployment, but fluctuated more at SLM_{reef} as well (Figures 7G, H). The bottom flow was stronger at SLM_{reef} ($U_{\text{mean}} = 13.53 \text{ cm s}^{-1}$, $U_{\text{max}} = 41.59 \text{ cm s}^{-1}$) than at $SLM_{\text{off-reef}}$ ($U_{\text{mean}} = 12.9 \text{ cm s}^{-1}$, $U_{\text{max}} = 37.66 \text{ cm s}^{-1}$) (Figures 7I, J). The flow direction oscillated between SSE and ESE at SLM_{reef} whereas at $SLM_{\text{off-reef}}$ the

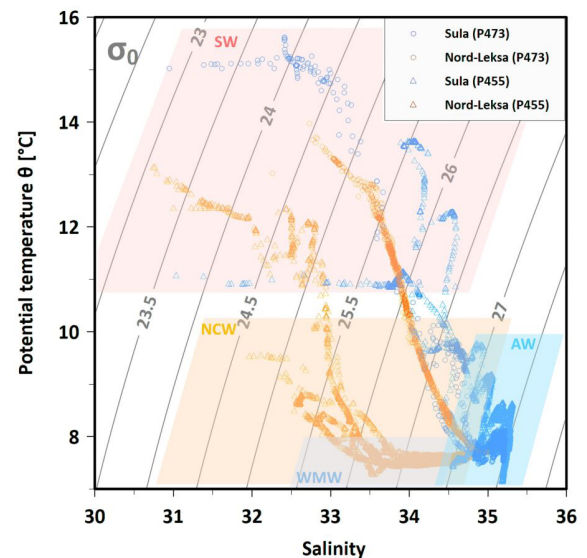


FIGURE 6
Temperature-Salinity plot based on CTD casts from summer 2013 and 2014 inshore (Nord-Leksa, orange) and offshore (Sula, blue), indicating water masses present in mid-Norway in shaded rectangles: Atlantic Water (AW) in blue, Norwegian Coastal Water (NCW) in orange, Winter Mode Water (WMW) in light-blue, Surface Water (SW) in red.

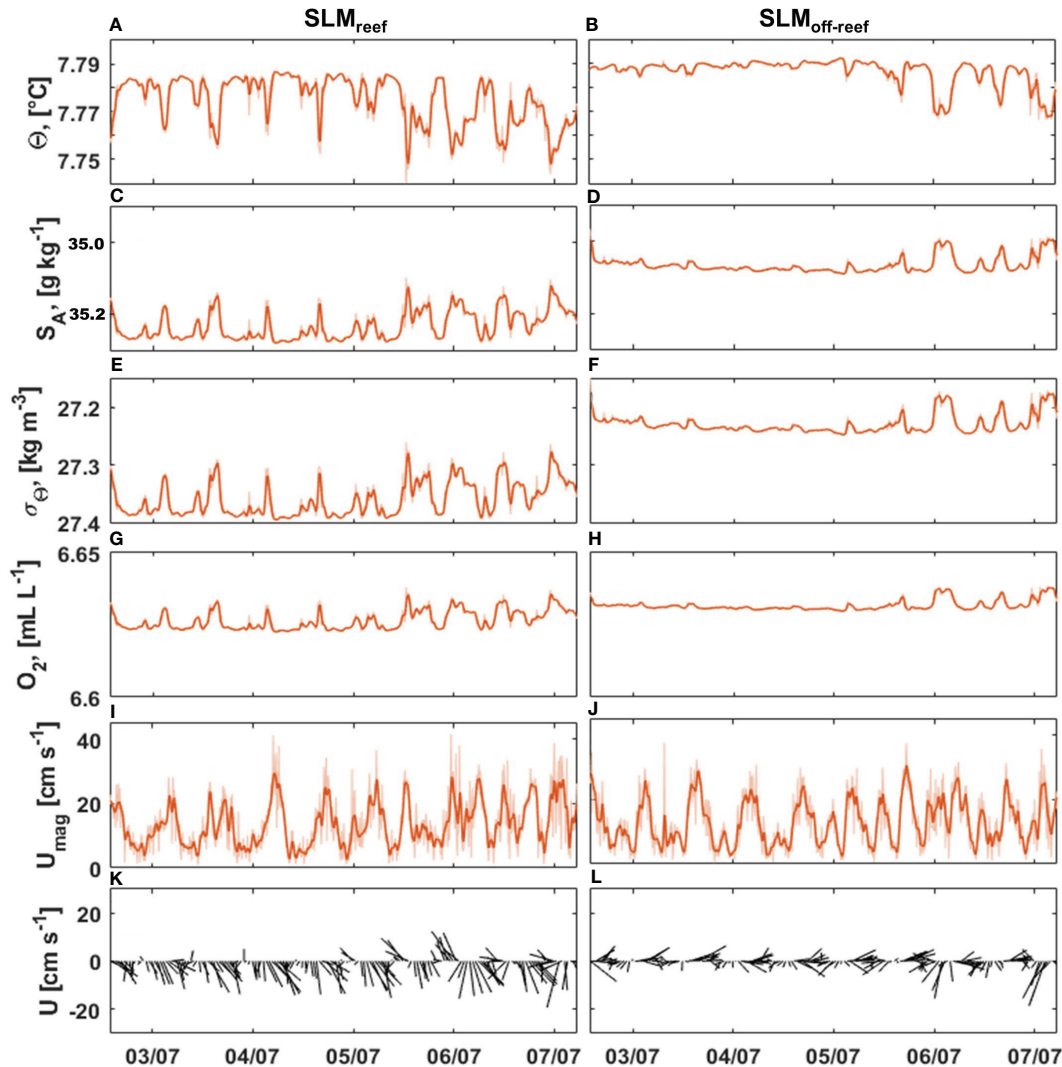


FIGURE 7

Comparison of bottom water characteristics of the first week of deployment for SLM_{reef} (left panels) and $SLM_{off-reef}$ (right panels) for potential temperature (A, B), absolute salinity (C, D), sigma-theta (E, F), oxygen concentration (G, H), flow speed (I, J), and horizontal direction (up - north, down - south) (K, L). The thicker orange line shows the 1-h running mean and the horizontal direction is shown with 1-h resolution. Note the reversed y-axis for absolute salinity (subplots C, D).

flow oscillated between E and WSW (Figures 7K, L). For the other comparable periods (22th August – 16th November 2013 and 19th November 2013 – 30th January 2014), the SLM_{reef} lander measured slightly higher salinities ($\Delta S_A = 0.2 \text{ g kg}^{-1}$) and bottom flow speeds ($\Delta U_{mean} = 1 \text{ cm s}^{-1}$) than $SLM_{off-reef}$, but fluctuations and range of the oceanographic variables were close to each other (not shown).

3.4 Long-term environmental variability at Sula and Nord-Leksa (Benthic landers)

3.4.1 Hydrography

The bottom water temperature, Θ , varied between 7.03°C and 8.91°C in Sula and between 6.83°C and 8.97°C in Nord-Leksa over the year (Figure 8). On average, the bottom waters were warmer in Nord-Leksa ($7.93 \pm 0.34^\circ\text{C}$, Figure 8B) than in

Sula ($7.52 \pm 0.30^\circ\text{C}$, Figure 8A) throughout the year. The bottom water salinity, S_A , varied between 31.12 g kg^{-1} and 36.06 g kg^{-1} in Sula and between 34.11 g kg^{-1} and 35.09 g kg^{-1} in Nord-Leksa. On average, bottom water salinities were higher at Sula ($35.13 \pm 0.50 \text{ g kg}^{-1}$, Figure 8D) than at Nord-Leksa ($34.80 \pm 0.16 \text{ g kg}^{-1}$) (Figure 8E). Sigma-theta, σ_Θ , varied from 24.18 kg m^{-3} to 28.04 kg m^{-3} in Sula (Figure 8G) and from 26.47 kg m^{-3} to 27.25 kg m^{-3} in Nord-Leksa (Figure 8H).

On a seasonal level, bottom waters at Sula were densest between April and September, with relatively high monthly mean salinity ($> 35.2 \text{ g kg}^{-1}$). The bottom water warmed and became fresher during autumn. The warmest values ($> 8.5^\circ\text{C}$) were observed between November and January followed by $> 0.5^\circ\text{C}$ drop in monthly mean temperature between January and February (Figure 9A). Bottom water was fresh ($< 34 \text{ g kg}^{-1}$) between January and February. Lowest average monthly temperatures were observed in July to September ($\sim 7.2^\circ\text{C}$).

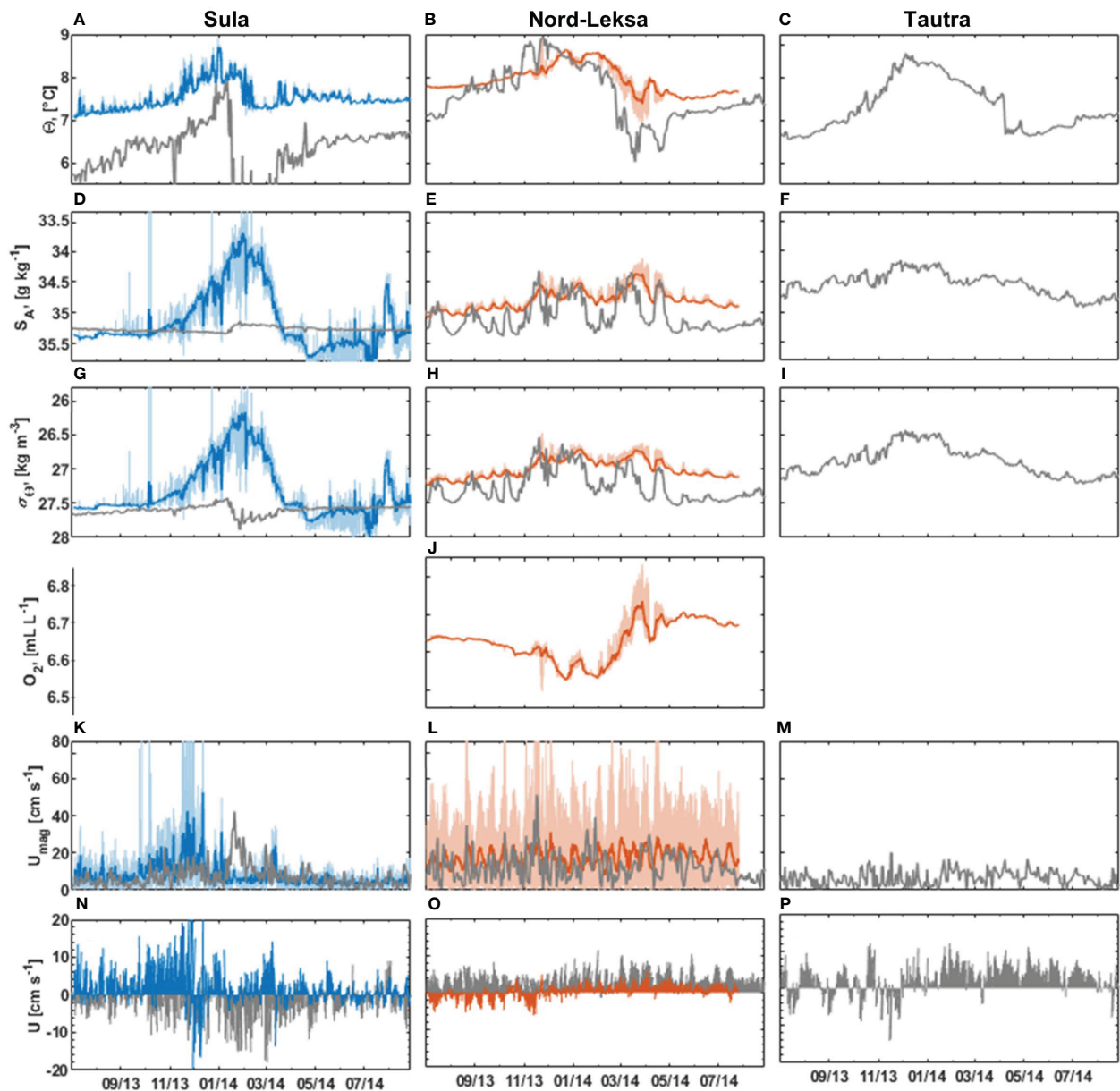


FIGURE 8

Long-term bottom water characteristics for Sula (left panels, blue) and Nord-Leksa (middle panels, orange) compared to Kyst800-model data (grey) for the three sites Sula (left panels), Nordlekxa (middle panels) and Tautra (right panels) showing temperature (A–C), absolute salinity (D–F), sigma-theta (G–I), oxygen concentration (for Nord-Leksa only, J), flow speed (K–M), and horizontal direction (N–P) (up -north, down-south). The thicker blue and orange lines show the 1-d running mean and the horizontal direction is shown with 1-d resolution. Note the reversed y-axes in salinity (D–F) and sigma-theta plots (G–H).

At Nord-Leksa, the bottom water was densest and most saline from July to September. From August to January, bottom water got warmer and fresher. The bottom waters were warmest from December to February with mean temperatures $> 8.3^{\circ}\text{C}$. In March, the temperature dropped to $< 7.5^{\circ}\text{C}$ and salinity to $< 34.5 \text{ g kg}^{-1}$. In March and in April, the temperature fluctuations were large ($> 1^{\circ}\text{C}$) compared to other months. The coolest average monthly temperatures in Nord-Leksa ($\sim 7.5^{\circ}\text{C}$) were observed from May to June (Figure 9B).

Over a semidiurnal (M_2) tidal cycle (12 h), hydrographic parameters varied, on average, $\Delta\Theta = 0.03^{\circ}\text{C}$, $\Delta S_A = 0.08 \text{ g kg}^{-1}$

and $\Delta\sigma_{\theta} = 0.06 \text{ kg m}^{-3}$ in Nord-Leksa, and $\Delta\Theta = 0.05^{\circ}\text{C}$, $\Delta S_A = 0.17 \text{ g kg}^{-1}$ and $\Delta\sigma_{\theta} = 0.14 \text{ kg m}^{-3}$ in Sula. The maximum semidiurnal variability was measured between November and March with the largest change within 12 h of $\Delta\Theta = 0.90^{\circ}\text{C}$, $\Delta S_A = 0.50 \text{ g kg}^{-1}$ and $\Delta\sigma_{\theta} = 0.38 \text{ kg m}^{-3}$ in Nord-Leksa, and $\Delta\Theta = 0.84^{\circ}\text{C}$, $\Delta S_A = 1.64 \text{ g kg}^{-1}$ and $\Delta\sigma_{\theta} = 1.2 \text{ kg m}^{-3}$ in Sula.

The NorKyst-model for hydrographic values (Albretsen et al., 2011) differed from our observations. The model temperatures were lower than the observed temperatures and the variation range in the model was overestimated. Temperature drops in February/March were more rapid in the model than in the observations (Figures 8A,

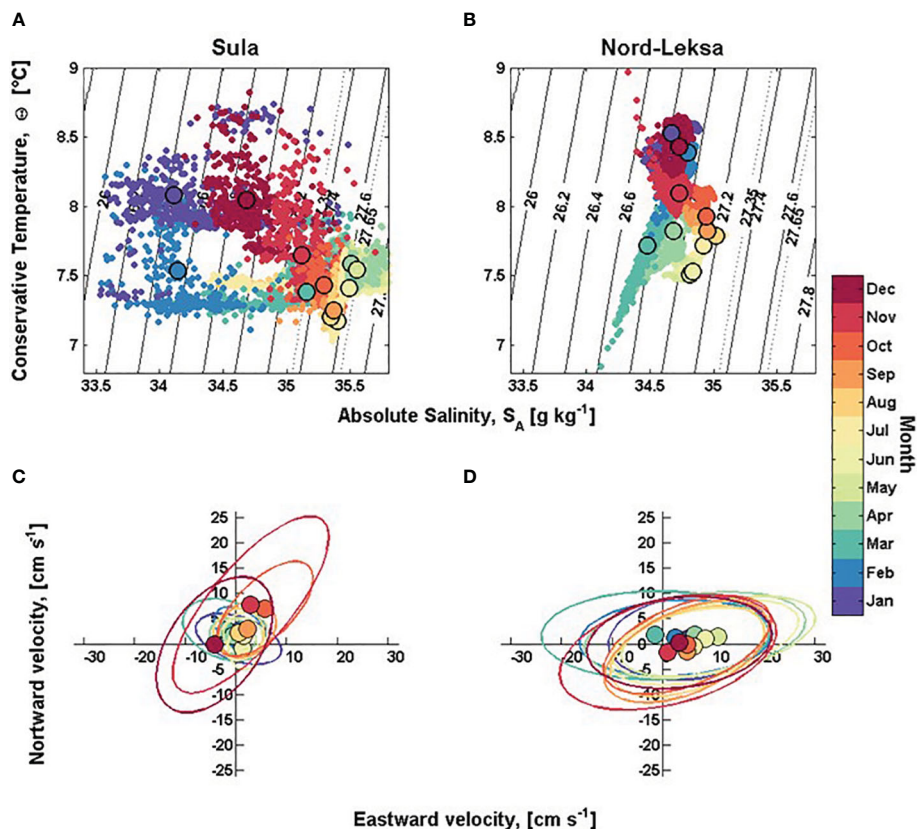


FIGURE 9

(A, B) Conservative Temperature – Absolute Salinity (Θ – S_A) plot of bottom waters from lander deployments from Sula (left) and Nord-Leksa (right). The annual cycle is shown with monthly representative colours and monthly mean values (bigger circles). The sigma-theta range 27.35–27.65 kg m⁻³ from Dullo et al. (2008) is marked with dotted lines. (C, D) Monthly horizontal velocity variance ellipses with mean velocity from Sula (left) and Nord-Leksa SLM_{off-reef} (right) from 0.75 masf and 2 masf, respectively, indicating main flow direction and speed each month. The colour scale shows the respective month of the year. Axis scale is ± 32 cm s⁻¹ in eastward and ± 26 cm s⁻¹ in northward direction.

B). At Tautra, the bottom temperature was estimated to be around $7.31 \pm 0.58^\circ\text{C}$ (Figure 8C). At Sula, the mean bottom salinity values were close to each other ($S_{A,\text{NorKyst}} = 35.27 \pm 0.04$ g kg⁻¹, $S_{A,\text{meas}} = 35.13 \pm 0.5$ g kg⁻¹, Figure 8D). At Nord-Leksa, the model showed more saline waters with overestimated fluctuations of salinity ($S_{A,\text{NorKyst}} = 35.06 \pm 0.26$ g kg⁻¹, $S_{A,\text{meas}} = 34.8 \pm 0.16$ g kg⁻¹, Figure 8E). The Tautra region was estimated to be fresher than sites further offshore by the model with on average 34.54 ± 0.19 g kg⁻¹ (Figures 8D–F). Due to the differences in temperature and salinity, the model showed ~ 0.25 kg m⁻³ denser water than observations with a mean density of 26.89 ± 0.21 kg m⁻³ (Figures 8G–I).

3.4.2 Oxygen

The average oxygen concentration at depth of the CWC reef in Nord-Leksa over the year as recorded by the benthic lander was 6.63 ± 0.05 mL L⁻¹ (210–218 m water depth, Figure 8J). Lowest O₂ concentrations of 6.5–6.6 mL L⁻¹ were found in November to December 2013 and in February 2014 and coincided with peaks in temperature and drops in salinity (Figures 8B, E, J). From February to April 2014, O₂ increased to maximum concentrations

of 6.8 mL L⁻¹, coinciding with decreasing temperatures and increases in salinity.

O₂ concentrations recorded by the lander showed a consistent off-set to the CTD oxygen sensors of ~ 0.3 mL L⁻¹ in 2013 to 0.9 mL L⁻¹ in 2014, which is probably related to different sensors and calibrations. The lower end of values (5.7 mL L⁻¹) recorded in 2014 (Figure 5D) may likely be a result of sensor drift due to lack of calibration between both surveys, as lander values revealed similar concentrations of 6.6 mL L⁻¹ each year at the time of CTD data collection.

3.4.3 Flow characteristics

The time-averaged horizontal bottom flow speed, U_{mean} , was 7.9 cm s⁻¹ at Sula and 17.7 cm s⁻¹ at Nord-Leksa. The peak horizontal flow speed was higher at Sula (157 cm s⁻¹) than at Nord-Leksa (104 cm s⁻¹) (Figures 8K, L). In Sula, the highest mean horizontal speeds were recorded in October and November, with mean direction towards northeast (Figures 8N, 9C). Lowest mean speeds were recorded in January, April and June. The direction of the flow oscillated between NE and SW, except from January to April, when the bottom flow direction oscillated between NW and

SE (Figures 8N, 9C). Between April and September, the flow was slow and isotropic (Figure 9C).

In Nord-Leksa, the highest monthly mean horizontal bottom flow speeds were recorded in May and June. The bottom flow oscillated between east and west with eastward mean flow (Figures 8O, 9D).

At Nord-Leksa, flow-measurements revealed the vertical structure of the water column at water depths between 80 and 208 m (Figure 10). The mean flow speed peaked with 20–27 cm s^{-1} at depths of living CWCs at around 150–180 m. The monthly maximum speed values through the measured water column varied between 50 and 160 cm s^{-1} .

Highest values were recorded from August to November with a peak around CWC depths and lowest maximum values between May and July. The flow oscillated between E and W or between NE and SW. During autumn, the flow was more northward and in winter more eastward. In general, the mean flow direction changed at the depths of the CWCs from east ($90 \pm 20^\circ$) at the lower part of the water column to southwest ($230 \pm 20^\circ$) at the depths above the corals. In December and January, the mean direction was relatively stable through the water column, varying between $76\text{--}109^\circ$ and $46\text{--}97^\circ$, respectively. In November, the change in the mean direction from east to southwest occurred at depths between 100 and 130 m. In March, the mean direction of the lower part of the water column was towards the north.

The NorKyst-model indicates similar flow speeds, but different flow direction as observed (Figures 8K–P). At Sula, the mean flow magnitude was close to observations ($U_{\text{mean, meas}} = 7.9 \text{ cm s}^{-1}$ and $U_{\text{mean, NorKyst}} = 7.32 \text{ cm s}^{-1}$), but the NorKyst mean flow direction was southward, whereas the observed mean flow direction was northward. Peak velocities occurred at different

times in Sula. In observations, the flow peaked in November, whereas in the model, the peak flow was in January. At Nord-Leksa, both the mean and the peak NorKyst flow values were lower than observed. Also, the observed flow in Nord-Leksa was tidal-driven with bi-directional velocity, whereas the model flow magnitude variability was more irregular with northward flow direction. At Tautra, the modelled bottom flow was slower than sites further offshore ($U_{\text{mean, NorKyst}} = 6.39 \text{ cm s}^{-1}$, $U_{\text{mean, NorKyst}} = 20.2 \text{ cm s}^{-1}$) (Figures 8K–P).

3.4.4 Tides

Tidal analysis of the pressure records explains 98.5% and 6.6% of the pressure fluctuations with 37 and 56 significant constituents at Sula and Nord-Leksa, respectively. When excluding the vertical lander movement periods at Nord-Leksa, the pressure fluctuation is > 95% tidal-driven (Table 3). At both sites, the semidiurnal (M_2) signal generates the largest amplitude and it is thus the most significant tidal constituent. It generates an amplitude of 0.70 dbar (Sula) and 0.80 dbar (Nord-Leksa), followed by diurnal K_1 (0.08–0.12 dbar) signal.

The analyses of the tidal constituents from the horizontal velocity records reveal larger differences between the two sites (Table 3). With 5 and 19 significant tidal constituents, tidal analysis explains 6.8% and 56.6% of the horizontal bottom velocity fluctuations at Sula and Nord-Leksa, respectively. The M_2 generates amplitudes of 2 cm s^{-1} at Sula and 20 cm s^{-1} at Nord-Leksa. The direction of tidal flow was towards WNW ($283 \pm 10^\circ$) at Sula and towards ENE ($79 \pm 1^\circ$) at Nord-Leksa. After M_2 , the most significant constituents are other semidiurnal (S_2 and N_2), semi-annual (SSA) and annual (SA) constituents in Sula and

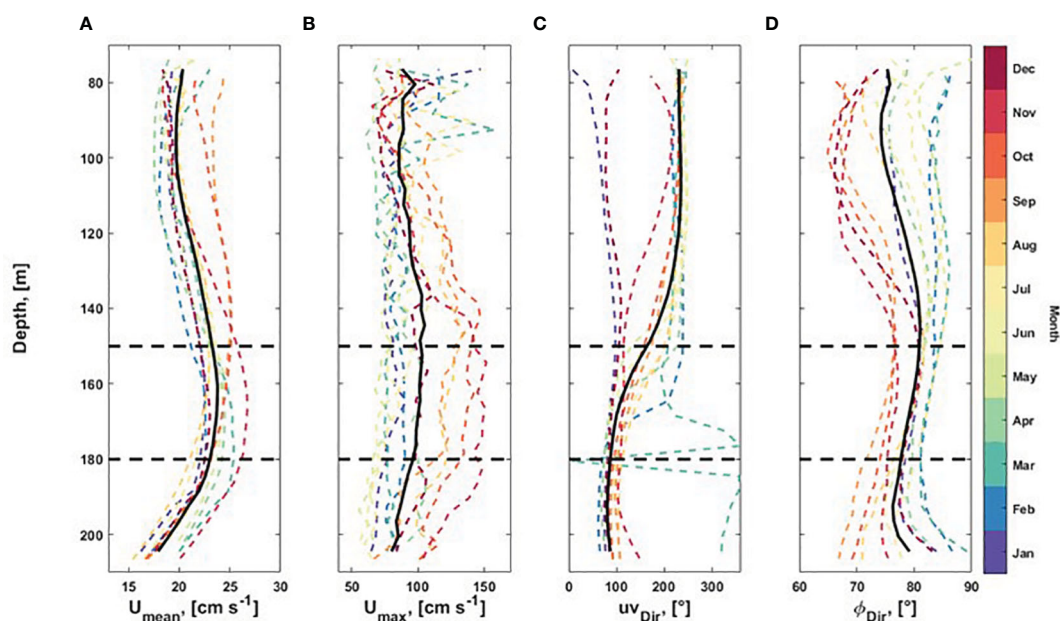


FIGURE 10

The flow characteristics at Nord-Leksa ($SLM_{\text{off-reef}}$) showing (A) the monthly mean and (B) maximum horizontal flow profiles, mean horizontal flow direction (C), and variance ellipse orientation (D). The black line shows the annual mean and the horizontal dashed lines show the CWC living depths.

TABLE 3 Tidal analysis for bottom pressure and flow record based on the harmonic analysis toolbox T_Tide (Pawlowicz et al., 2002).

Site	%	Amplitude (of the most significant constituents)			
		Lower	Diurnal	Semidiurnal	Higher
NL					
p (dbar)	95.8	–	0.07 (K ₁), 0.05 (O ₁)	0.8 (M ₂), 0.27 (S ₂)	0.02 (M ₄)
uv (cm s ⁻¹)	56.6	–	0.8 (K ₁), 0.6 (O ₁)	19.1 (M ₂), 7 (S ₂)	1 (M ₄), 2 (MS ₄)
Sula					
p (dbar)	98.4	–	0.07 (K ₁), 0.05 (O ₁)	0.7 (M ₂), 0.24 (S ₂)	0.01 (M ₄)
uv (cm s ⁻¹)	6.8	2.1 (SA)	–	2.4 (M ₂), 1 (S ₂)	–

For Nord-Leksa, the periods of vertical lander movements are not included in the analysis. Shown are the amplitudes of the two most important constituents in four categories: long, diurnal, semi-diurnal and short period constituents. Explained variance through the tidal model in percent is given next to the parameter. p, pressure; uv, bottom horizontal flow.

semidiurnal (S₂, N₂), diurnal (H₁) and shallow water (MS₄) constituents in Nord-Leksa.

3.5 Carbonate chemistry and dissolved inorganic nutrients

3.5.1 Carbonate system

The carbonate chemistry at the lander positions differed between the times of deployment in 2013 and recovery in 2014 with regard to TA, corresponding to changes in the pH, pCO₂, carbonate ions and the aragonite saturation state (Ω_{Ar}). DIC values were similar (maximum of ± 12 μmol kg_{seawater}⁻¹ in standard deviation among all six means) in both years, whereas TA values were about 95 μmol kg_{seawater}⁻¹ lower in August 2014 compared to July 2013, resulting in lower pH and Ω_{Ar} values, and higher pCO₂ levels and bicarbonate concentrations in 2014 (Table 4). There were no pronounced differences between depths among the two lander positions in Nord-Leksa or the ~ 90 m deeper sampling location in Sula during sampling in 2013. In 2014, both TA and DIC were lower at Sula compared to Nord-Leksa, resulting in lower pCO₂ and higher Ω_{Ar} values at Sula than at Nord-Leksa (Table 4). Both TA (TA_{GLODAP}: 2,294–2,305 μmol kg⁻¹) and DIC (DIC_{GLODAP}: 2,127–2,144 μmol kg⁻¹) concentrations were lower in the GLODAP database than in the measured samples, yielding shifts in all carbonate chemistry variables accordingly.

3.5.2 Dissolved inorganic nutrients

Nitrate (NO₃⁻), nitrite (NO₂⁻), ammonium (NH₄⁺), and phosphate (PO₄³⁻) were similarly indifferent between sites, depths and years (Table 4). Ammonium concentrations were quite high in measurements from both years, but were about twice as high in summer 2014 compared to 2013. Values were, except for ammonium, in a normal open ocean range of 9–13 μmol L⁻¹ in NO₃⁻, < 0.1 μmol L⁻¹

¹ in NO₂⁻ and ~ 0.6–0.7 μmol L⁻¹ in PO₄³⁻. The N:P ratio (calculated from N sources excluding NH₄⁺) was around the typical oceanic ratio (Redfield ratio) at all sites in both years, varying between 14:1 and 22:1. While in 2013, the N:P ratio decreased with depth from 22:1 on-reef in Nord-Leksa to 18:1 off-reef and lowest ratio offshore at Sula (15:1), it was the other way around in the following year with lowest N:P ratios in Nord-Leksa (14:1 on-reef, 17:1 off-reef) and a higher ratio of 19:1 at Sula compared to the fjord and the year before. The GLODAP nitrate and phosphate values were in a similar range as the measured values (NO₃⁻_{GLODAP}: 10.3–12.1 μmol kg⁻¹, PO₄³⁻_{GLODAP}: 0.6–0.8 μmol kg⁻¹).

3.5.3 In-reef sampling at the Tautra Reef

DIC and TA values were slightly lower (ΔDIC ≈ 30 μmol kg⁻¹ and ΔTA ≈ 80 μmol kg⁻¹ on average) at Tautra Reef compared to the deeper Nord-Leksa Reef (Table 5). The pH_{TS} was 0.1 units lower at Tautra than at Nord-Leksa at similar sampling time in 2013 with corresponding lower Ω_{Ar} values (1.8 vs. 2.2).

Within the Tautra Reef area, there is a clear distinction of the samples taken further away from the corals to the samples taken close to live coral colonies and smaller scale changes within the samples taken directly at the colonies. Towards the reef top, pH_{TS} and Ω_{Ar} decreased and the pCO₂ and bicarbonate concentrations increased (Table 5). Off-reef, NO₃⁻ concentration was slightly higher and PO₄³⁻ was lower compared to in-reef samples. NO₂⁻ was noticeably higher at the reef slope and on the top compared to the reef edge and further away from the reef.

4 Discussion

In this study we compare the environmental conditions between an offshore and an inshore thriving CWC reef habitat in mid-Norway. Both sites are characterised by dynamic environmental conditions with large seasonal variability of the physical parameters. The bottom water flow fields are driven by topography-flow-interaction, tidal activity, water column stratification and large-scale atmospheric forcing. These findings could explain the observed differences in associated fauna, coral colony morphology and partly even the reef extent, while biogeochemical water properties between the reef sites were comparable.

4.1 Hydrodynamics and coral colony morphology

The near-reef flow was relatively strong both at Nord-Leksa and Sula but was in the range of flow speeds reported for *L. pertusa* reef distribution in Norwegian waters (Buhl-Mortensen and Freiwald, 2023) and other CWC reefs in the NE Atlantic (White and Dorschel, 2010; Lim et al., 2020), with flow speeds on topographic highs > 20 cm s⁻¹ (Mortensen et al., 2001; Thiem et al., 2006). The mean bottom flow speed was stronger in Nord-Leksa (~ 20 cm s⁻¹) compared to Sula (~ 8 cm s⁻¹). High flow speeds (> 80 cm s⁻¹) occurred more regularly at Nord-Leksa than Sula

TABLE 4 Dissolved inorganic nutrients (nitrate (NO₃⁻), nitrite (NO₂⁻), ammonium (NH₄⁺), and phosphate (PO₄³⁻) and carbonate chemistry and physical seawater parameters analysed from water samples taken at the three lander positions (about 10 m above ground) during POS455 in 2013 and POS473 in 2014 at Nord-Leksa and Sula.

Date	Location	Depth [m]	T [°C]	Sal	DIC [μmol kg ⁻¹]	TA [μmol kg ⁻¹]	pH _{TIS}	pCO ₂ [μatm]	HCO ₃ ⁻ [μmol kg ⁻¹]	CO ₃ ²⁻ [μmol kg ⁻¹]	Ω _{Ar}	NO ₃ ⁻ [μmol L ⁻¹]	NO ₂ ⁻ [μmol L ⁻¹]	NH ₄ ⁺ [μmol L ⁻¹]	PO ₄ ³⁻ [μmol L ⁻¹]
03.07.13	Nord-Leksa _{reef}	175	7.8	35.1	2147.7	2403.9	8.212	265.1	1953.9	181.5	2.74	12.97	0.10	9.52	0.60
04.07.13	Nord-Leksa _{off-reef}	216	7.8	35.1	2142.4	2398.8	8.213	263.8	1948.7	181.4	2.74	10.72	0.00	8.04	0.60
05.07.13	Sula Reef	290	7.2	35.1	2148.7	2408.1	8.217	261.7	1952.8	183.6	2.77	9.17	0.10	5.51	0.60
18.08.14	Nord-Leksa _{reef}	170	7.7	34.9	2147.7	2303.6	8.010	434.8	2010.3	117.1	1.77	9.54	0.00	22.74	0.70
19.08.14	Nord-Leksa _{off-reef}	207	7.7	35.0	2156.4	2307.9	7.998	448.6	2020.7	114.7	1.73	10.67	0.03	14.32	0.64
26.08.14	Sula Reef	290	7.6	35.2	2120.8	2315.0	8.098	346.9	1964.4	140.2	2.12	11.36	0.00	12.72	0.60

Given are analysed values of total alkalinity (TA) and dissolved inorganic carbon (DIC) in μmol kg seawater⁻¹ (means of duplicate samples), in situ temperature (T) in °C and practical salinity (Sal), and calculated values of pH on the total scale (pH_{TIS}), pCO₂ in μatm, bicarbonate (HCO₃⁻) and carbonate (CO₃²⁻) concentrations in μmol kg seawater⁻¹, and the aragonite saturation state (Ω_{Ar}) computed with CO2SYS.

throughout the year. Given the bathymetry, single high flow speed events will likely have a higher impact at the Leksa Reef, where waters are pushed to a narrow strait/fjord. Maximum flow speeds recorded in late autumn were in a similar range at both sites (100–150 cm s⁻¹), which is about five-fold higher than previously reported maximum flow velocities at the Sula Ridge (< 30 cm s⁻¹, Eide, 1978; Roberts et al., 2005). These maximum flow speed events coincided with strong wind speeds (Supplementary Figure S1) and therefore suggests storm events during this time. In November/December, the forcing resulting from this caused the movement of both landers at Nord-Leksa. Similarly high maximum flow speeds of 76 cm s⁻¹ and 114 cm s⁻¹ were recorded at two sites in the upper Porcupine Bank Canyon on the Irish continental shelf, which likewise led to landers being toppled upslope, suggesting even higher current velocities than recorded (Lim et al., 2020). Lim and colleagues suggest that in the Porcupine Bank area a higher live coral abundance is found at sites with lower coral-facing mean flow speeds and that sites that are consistently exposed to higher flow speeds have higher coral rubble percentages due to enhanced physical and biological erosion of the coral framework. In contrast, flow velocity at the inshore site of this study was consistently higher directly at the flourishing reef compared to the off-reef lander location surrounded by rubble where this could be compared, suggesting that the differences in flow direction (De Clippele et al., 2018) and environmental conditions play a greater role in coral distribution at the Nord-Leksa Reef than variations in current speed. In mid-Norway, the reefs are well-developed, large and dense framework structures (Hennige et al., 2014), which could make them less susceptible towards high current velocities than thickets and individual mounds. Nonetheless, current speeds were generally higher at coral sites when compared to non-coral bearing sites in the study by Lim et al. (2020) as well, underlining that corals require high flow speeds to thrive.

Relatively high flow speeds are thought to increase the food encounter rates of the corals (Hunter, 1989; Sebens et al., 1998) and prevent the polyps from clogging with sediments (Brooke et al., 2009; Larsson and Purser, 2011). In contrast, laboratory experiments have demonstrated the efficient prey capture rates of CWCs like *L. pertusa* to peak at relatively low flow speeds < 7 cm s⁻¹ (for zooplankton at flow rates of ~ 2.5 cm s⁻¹ and for phytoplankton ~ 5 cm s⁻¹), as with stronger flow the prey could escape from the polyps (Purser et al., 2010; Orejas et al., 2016). The CWC framework slows down the ambient flow due to friction (Roberts et al., 2009) and dense coral framework creates stability for the colony (Chamberlain and Graus, 1975; Roberts et al., 2009; Hennige et al., 2014). The CWC framework thus acts as a natural sediment trap allowing the corals to capture food and utilise the enhanced particle delivery. The corals are adapted to a very dynamic environment as various hydrodynamic processes on different temporal scales cause food supply to come in periodically rather than constantly (Maier et al., 2023), linked to small-scale processes such as internal tidal activity (Davies et al., 2009; de Froe et al., 2022) as well as larger scale circulation patterns and specific water masses (Schulz et al., 2020; Mohn et al., 2023). On continental margins like the shelf edge along Norway and plateaus such as Rockall Bank, currents continuously interact with the seafloor,

TABLE 5 Carbonate chemistry and physical seawater parameters analysed from water samples taken by divers at different locations at the Tautra Reef in the Trondheimsfjord during POS455 in 2013.

Location	Depth [m]	T [°C]	DIC [$\mu\text{mol kg}^{-1}$]	TA [$\mu\text{mol kg}^{-1}$]	pH _{TS}	pCO ₂ [μatm]	HCO ₃ ⁻ [$\mu\text{mol kg}^{-1}$]	CO ₃ ²⁻ [$\mu\text{mol kg}^{-1}$]	Ω_{Ar}	NO ₃ ⁻ [$\mu\text{mol L}^{-1}$]	NO ₂ ⁻ [$\mu\text{mol L}^{-1}$]	NH ₄ ⁺ [$\mu\text{mol L}^{-1}$]	PO ₄ ³⁻ [$\mu\text{mol L}^{-1}$]
Off- reef	50	6	2105.6	2285.5	8.101	338	1958.8	129.9	1.96	9.17	0.08	10.48	0.13
Reef edge	42	6	2108.8	2279.3	8.080	357	1966.8	124.1	1.87	7.83	0.01	10.58	0.35
Reef slope	40	6	2092.0	2264.6	8.086	348	1949.6	125.0	1.88	8.69	0.48	8.90	0.45
Reef top	39	6	2143.7	2279.6	7.993	447	2017.1	104.2	1.57	8.60	0.35	10.61	0.60

Given are measured values of total alkalinity (TA) and dissolved inorganic carbon (DIC) in $\mu\text{mol kg seawater}^{-1}$ (means of duplicate samples), in situ temperature (T) in °C, and calculated values of pH on the total scale (TS), pCO₂ in μatm , bicarbonate (HCO₃⁻) and carbonate (CO₃²⁻) concentrations in $\mu\text{mol kg seawater}^{-1}$, and the aragonite saturation state (Ω_{Ar}) computed with CO2SYS. Dissolved inorganic nutrient analyses [nitrate (NO₃⁻), nitrite (NO₂⁻), ammonium (NH₄⁺), and phosphate (PO₄³⁻)] from these samples are means of duplicate measurements expressed as $\mu\text{mol per litre}$.

creating good feeding conditions for the corals (Frederiksen et al., 1992; White et al., 2005; Thiem et al., 2006). In the fjords, tidal effects play a more important role for food supply than offshore and higher fluorescence as well as higher turbidity at depths of CWC occurrence at Nord-Leksa may indicate higher local food supply at the time of sampling due to more turbulent conditions compared to Sula.

The colony morphology determines how much the water flow slows down within the framework. Coral colonies with long and thin branches slow down the flow less than more compact colonies with short and thick branches (Kaandorp and Sloom, 2001; Chindapol et al., 2013). The long-branched colony morphology, which is more common at Sula, is associated with low near-reef flow speeds (Kaandorp, 1999; Chindapol et al., 2013). The recorded near-bottom flow speed at Sula was $< 7 \text{ cm s}^{-1}$ for 60% of the time during lander deployment, while at Nord-Leksa it was only 19%, thus most of the time higher flow speeds were recorded during the lander deployment. The difference in the mean near-reef flow speeds likely explains the more extended colony shapes at Sula. High flow speed events ($> 100 \text{ cm s}^{-1}$) in winter may break exposed CWC skeleton in Sula more easily than at Nord-Leksa due to the thinner branches. Considering $< 7 \text{ cm s}^{-1}$ to be the optimal flow rate for prey capture not only in experimental conditions but also in the natural environment suggests that prey capture efficiency is supported better at the Sula Ridge than at Nord-Leksa based on the year-long continuous flow measurements presented here.

However, optimal flow conditions for coral feeding change within a reef system as the flow speed and direction change on a diurnal and seasonal time scale, thus different parts of the reef have different flow conditions for feeding. This affects the growth direction of the corals, which might help them to optimise food capture (Wheeler et al., 2007; Buhl-Mortensen et al., 2010; De Clippele et al., 2017, 2018).

Moreover, local flow dynamics also change with the complexity of the framework, with larger structures modifying their surrounding flow environment and potentially reducing the favourable flow conditions for other colonies in the vicinity, as suggested in a recent modelling study by Hennige et al. (2021) using the 'Goldilocks Principle'. That corals effectively optimise their own

local flow environment through habitat engineering is supported by the finding that surface complexity within and between colonies is similarly variable independent of the differences of the compactness at inshore versus offshore reef sites (i.e., the more compact growth form of corals exposed to strong currents in fjords for example) (Sanna et al., 2023).

The flow regime is controlled by short-term phenomena like tides and storms in Nord-Leksa and by seasonal atmospheric forcing at Sula. The short-term variability in the flow field is controlled by tides and storms. In winter 2013-2014, two storms with winds $> 24.5 \text{ m s}^{-1}$ were recorded over Trøndelag on 16-17th November 2013 (MET-info 2013) and 12th December 2013 (MET-info and Fagerlid, 2014), which coincided with steep increases in flow speeds recorded by the landers (Supplementary Figure S1; Figure 8). The first storm coincided with the vertical lander movements in Nord-Leksa showing that the whole water column in the fjord system is affected when the waters are pushed towards the fjord entrance. At Nord-Leksa, the flow regime is tidally-driven with strong semi-diurnal (M₂) flow ($\sim 20 \text{ cm s}^{-1}$) towards the main current direction. This results in a larger amplitude than the average semi-diurnal tidal current on the central Norwegian shelf (amplitudes of 3-12 cm s^{-1}) (Gjevik and Straume, 1989; Haugan et al., 1991). The tidal flow controls the bi-directional flow at Nord-Leksa supporting the observed 'cauliflower' colony shape. At Sula, the tidal flow is weak ($\sim 2 \text{ cm s}^{-1}$), rotated westward due to an anticyclonic (clockwise) topographic effect of the Haltenbank and differs from the main current direction (northward).

At Sula, the seasonal variations in the flow regime are driven by the quasi-stationary topographical effects of Haltenbank and Frøyabank and the seasonal changes in the NAC-NCC-interaction. The flow is mostly directed by a northward flowing coastal branch of the NCC (Ljøen and Nakken, 1969; Eide, 1979; Poulain et al., 1996; Sætre, 1999). This strong current creates a distinct temperature front between the coastal branch of the NCC and the water masses further offshore (Audunson et al., 1981). The stratification is weakest in late autumn and winter, which yields to observed fresher and warmer surface water at greater depths (Figures 8A, D, 9A). In winter, the anticyclonic eddies are strongest due to changes in the prevailing wind direction (Eide, 1979; Sætre, 1999). This yields intrusions of high-salinity

Atlantic Water (AW) to the shelf (Haugan et al., 1991) and the observed pronounced strong southward flow with increasing salinity in mid-winter (Figures 8D, N, 9A, C).

In the Trondheimsfjord, the seasonal variations in the flow field are controlled by fresh-water input and the dominant wind directions at the coast. Although we do not have flow measurements from Tautra, the flow has previously been reported to be similarly dominated by impacts of fresh-water input and strong tides as in Nord-Leksa (Mortensen and Fosså, 2001). The Trondheimsfjord has a mean fresh water runoff of $725 \text{ m}^3 \text{ s}^{-1}$ with a spring flood maximum in April/May of up to $6,430 \text{ m}^3 \text{ s}^{-1}$ (Sakshaug and Killingtveit, 2000). This partially coincides with the period of northerly winds and the shallow NCC that are believed to explain the inflow of dense AW to the fjord from February to June (Sakshaug and Killingtveit, 2000).

4.2 Environmental conditions

CWCs live in relatively cold ($\Theta = 6.83\text{--}8.97^\circ\text{C}$) and oxygen-rich ($\text{O}_2 = 5.7\text{--}6.6 \text{ mL L}^{-1}$) waters in mid-Norway compared with other known CWC reef sites (Roberts et al., 2006; Dodds et al., 2007; Freiwald et al., 2009; Davies et al., 2010; Davies and Guinotte, 2011; White et al., 2012; Ramos et al., 2017; Hanz et al., 2019). The very low salinity values down to 31.12 g kg^{-1} recorded at Sula are most likely a result of sensor errors (values are peaks in unfiltered data). Disregarding those peaks shows a comparable salinity range ($33.50\text{--}36.06 \text{ g kg}^{-1}$) to other CWC habitats (Freiwald et al., 2004; Davies et al., 2008). However, our results revealed a broader range of salinities than previous studies suggested for the Sula Reef ($35.0\text{--}35.2 \text{ psu}$) but keeping in mind that most comparable studies represent data solely from summer conditions, higher variability throughout the year can be expected. The measured annual range of hydrographic variables was larger than expected (Sula: $> 1.5^\circ\text{C}$ and $> 2 \text{ g kg}^{-1}$ and Nord-Leksa: $> 1.5^\circ\text{C}$ and $> 1 \text{ g kg}^{-1}$), but moderate when compared to the largest measured temperature fluctuations at a CWC site of 9°C within a day in the Cape Lookout area, NW Atlantic (Mienis et al., 2014). Due to coastal fresh-water influence and relatively warm winter bottom water temperatures, the bottom water was less dense over the whole year at Nord-Leksa and less dense from late autumn to winter at Sula than the suggested density range for NE Atlantic CWC sites ($\sigma_\Theta = 27.35\text{--}27.65 \text{ kg m}^{-3}$) (Dullo et al., 2008).

Oxygen concentrations in Nord-Leksa ranged in a well-oxygenated state compared to other CWC reef or mound sites (Hebbeln et al., 2020) and showed relatively little variability throughout the entire year. Despite an off-set ($\sim 0.3\text{--}0.9 \text{ mL L}^{-1}$) between lander and CTD oxygen sensors, both systems revealed values in the same order of magnitude within the range of previously recorded oxygen concentrations in Norwegian and other NE Atlantic CWC habitats (Dullo et al., 2008; Findlay et al., 2014; Hebbeln et al., 2020). Similar to observations in White et al. (2012) maximum values of dissolved oxygen were found in March/April and decreased to lowest values in November/December and is likely due to the supply and decay of organic matter. Lower oxygen concentrations in Nord-Leksa than in Sula may be due to higher O_2 consumption rates inshore as a result of biological activity and

eutrophication, which might be influenced by aquaculture activities as well (Aksnes et al., 2019). Despite being located at the entrance of the fjord and not enclosed in the fjord system as well as experiencing similarly high flow velocities like in Sula, it may also be that the open ocean shelf reef sites have higher renewal rates of dissolved oxygen due to the more direct influence of oceanic AW.

The carbonate chemistry variables measured here were comparable to values previously reported for Norwegian CWCs (Büscher et al., 2019) and indicate saturation of calcium carbonate, promoting calcification at Nord-Leksa and Sula. In the Norwegian Sea, the ASH is $\sim 2,000 \text{ m}$ (Jones et al., 2018). The CWCs at the studied sites thrive in shallower and therewith saturated waters ($\Omega_{\text{Ar}} = 1.57\text{--}2.77$) in a similar range like other CWC occurrences in the NE Atlantic in the Rockall Trough region and the Outer Hebrides (McGrath et al., 2012; Findlay et al., 2014). For CWCs in the Gulf of Mexico in the NW Atlantic, occurring at deeper depths than studied here, between $\sim 400\text{--}600 \text{ m}$, lower aragonite saturation values ($\Omega_{\text{Ar}} \approx 1.15\text{--}1.44$) were reported, as a result of higher mean DIC values (Georgian et al., 2016). Flögel et al. (2014) demonstrated that living CWCs in the NE Atlantic are linked to low DIC concentrations ($< 2,170 \mu\text{mol kg}^{-1}$). This is confirmed for Sula and Nord-Leksa with DIC values ranging from $2,070$ to $2,157 \mu\text{mol kg}^{-1}$, while in waters outside the NE Atlantic, live CWCs occur at higher DIC concentrations up to $2,470 \mu\text{mol kg}^{-1}$ (in the Marmara Sea), attributable to different water masses present in those areas (McCulloch et al., 2012; Findlay et al., 2014). McGrath et al. (2012) have shown that DIC concentrations have increased in the Rockall Trough by nearly $20 \mu\text{mol kg}^{-1}$ in subsurface waters over two decades (1991 – 2010), concomitant with an equivalent decrease in pH and Ω_{Ar} and a shoaling of the ASH. Hence, environmental conditions are changing in the North Atlantic and may shift boundary conditions outside the tolerance limits of benthic organisms such as CWCs.

Nutrient concentrations are scarcely reported from CWC reefs. Findlay et al. (2014) assembled reported data from known CWC reefs sites in the NE Atlantic, revealing a large range for dissolved inorganic nitrate ($4.1\text{--}23.4 \mu\text{mol L}^{-1}$) and a smaller range for phosphate ($0.6\text{--}1.6 \mu\text{mol L}^{-1}$). Analysed nitrate ($7.8\text{--}13.0 \mu\text{mol L}^{-1}$) and phosphate ($0.6\text{--}0.7 \mu\text{mol L}^{-1}$) values were in the lower range of those reported for the Rockall Trough region, but similar to values found in a comparable CWC reef setting (the Mingulay Reef Complex) at the Outer Hebrides.

Phosphate values within the reef structures at Tautra Reef were considerably lower ($< 0.45 \mu\text{mol kg}^{-1}$) than the water above the reef and observed at other NE Atlantic CWC sites. Perhaps, these lower phosphate values close to the polyps point to higher turnover rates in the nutrient cycling within the reef structures. de Froe et al. (2019) recently demonstrated that CWC reef communities and specifically the living CWCs influence benthic nitrogen cycling by circumventing nitrification and NH_4^+ production. High NH_4^+ concentrations close to the coral reefs may thus be indicative of a NH_4^+ release of the corals as a result of nitrogen cycling. However, due to the sensitive nature of nutrient analysis, especially ammonium, and this being only a single point measurement, similar sampling and measurements should be repeated for comparison before jumping to conclusions.

4.3 Associated fauna

Both sites have rich associated fauna using the CWC framework for shelter, reproduction and as feeding grounds (Henry and Roberts, 2007; Baillon et al., 2012; Buhl-Mortensen et al., 2017), but higher diversity of coral-associated invertebrates was previously found in Nord-Leksa and other mid-Norwegian inshore reefs compared to the Sula Reef (Mortensen and Fosså, 2006). The reef framework provides substrate for other benthic filter feeders such as sponges (Orejas and Jiménez, 2017), which can form a substantial component of CWC reef biomass (Hogg et al., 2010). Reef sponges function as nutrient recyclers, substrate stabilizers, bioeroders, and as a food source and habitat for other organisms (Wulff, 2001). At Tisler Reef in the Oslofjord in southern Norway where similar environmental conditions prevail as at Nord-Leksa, both *Geodia* sp. and *M. lingua* were found growing in dead *L. pertusa* framework, but only *M. lingua* was found growing within live *L. pertusa* (De Clippele et al., 2018). Thus, *M. lingua* does not compete with the corals for hard substrate and can co-occur by growing within *Lophelia* colonies as was observed in Nord-Leksa, while *Geodia* sp. presumably competes for the same substrate as the corals and may not be able to outcompete *L. pertusa*'s growth rates in regions where substrate availability is limited (Purser et al., 2013; De Clippele et al., 2018). The coverage of *M. lingua* within the framework of *L. pertusa* at Nord-Leksa, whereas found on dead coral substrate between colonies at Sula, supports the observation that substrate availability is limited at Nord-Leksa and species have to co-exist, despite largely comparable environmental conditions like offshore. The absence of *Geodia* sp. could also be explained by small-scale differences in environmental conditions, e.g. with lower water temperatures by approximately half a degree at Sula on average and considerably lower bottom flow speeds. Temperatures at Tisler Reef, where *Geodia* sp. occur, are however generally higher than temperatures reached at Nord-Leksa at its maximum (compare De Clippele et al., 2018).

5 Conclusion

The data of this study show annual variability of environmental conditions of two thriving cold-water coral reefs in mid-Norway. It sheds new light on the environmental factors controlling habitat features like coral colony morphology and associated fauna on the reef scale. The near-reef flow regime is characterised by high flow velocities both within the fjord and on the shelf, with highest flow speeds peaking in winter (November 2013) and revealing far higher values at times than previously reported ($U_{\max} \sim 150 \text{ cm s}^{-1}$). Inshore, the flow is strongly controlled by tides and storms, whereas on the shelf the seasonal atmospheric forcing determines the major variability in the flow regime. The 'cauliflower' colony morphology is dominating both in the fjord and on the shelf, but with weaker average flow speeds ($U_{\text{mean}} \sim 8 \text{ cm s}^{-1}$ vs. $\sim 20 \text{ cm s}^{-1}$) the coral branches are thinner and longer on the shelf. The databases and oceanographic data used in habitat suitability models have a resolution too coarse to capture variability on a reef-scale. Thus,

taking the natural small-scale temporal and spatial variability into account helps to understand the resilience of CWCs towards environmental fluctuations and would increase the accuracy of the models for future predictions of thriving CWCs.

Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

Author contributions

JB: Conceptualization, Data curation, Formal analysis, Funding acquisition, Investigation, Methodology, Project administration, Resources, Validation, Visualization, Writing – original draft, Writing – review & editing. KJ: Data curation, Formal analysis, Methodology, Software, Visualization, Writing – original draft, Writing – review & editing. SF: Data curation, Methodology, Supervision, Validation, Writing – review & editing. MW: Conceptualization, Data curation, Investigation, Methodology, Supervision, Writing – review & editing. AR: Data curation, Methodology, Validation, Writing – review & editing. UR: Funding acquisition, Resources, Supervision, Writing – review & editing. AF: Conceptualization, Funding acquisition, Investigation, Methodology, Project administration, Resources, Supervision, Writing – review & editing.

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Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2024.1363542/full#supplementary-material>

SUPPLEMENTARY FIGURE 1

Pressure (dbar) (A) as an indication for lander movements and lander flow speed (B) compared with daily wind speeds (metres per second) (C).

References

- Addamo, A. M., Vertino, A., Stolarski, J., García-Jiménez, R., Taviani, M., and Machordom, A. (2016). Merging scleractinian genera: The overwhelming genetic similarity between solitary *Desmophyllum* and colonial *Lophelia*. *BMC Evol. Biol.* 16 (108). doi: 10.1186/s12862-016-0654-8
- Aksnes, D. L., Aure, J., Johansen, P. O., Johnsen, G. H., and Veia Salvanes, A. G. (2019). Multi-decadal warming of Atlantic water and associated decline of dissolved oxygen in a deep fjord. *Estuar. Coast. Shelf Sci.* 228, 106392. doi: 10.1016/j.ecss.2019.106392
- Albretsen, J., Sperrevik, A. K., Staalstrøm, A., Sandvik, A. D., Vikebø, F., and Asplin, L. (2011). NorKyst-800 Report No. 1 User Manual and technical descriptions. *Tech. Rep. Fisker og Havet* 2/2011.
- Audunson, T., Dalen, V., Krogstad, H., Lie, H. N., and Steinbakke, O. (1981). "Some observations of ocean fronts, waves and currents in the surface along the Norwegian coast from satellite images and drifting buoys," in *The Norwegian Coastal Current, Proceedings from Symposium*, eds. R. Saetre and M. Mork. (Norway: University of Bergen), 20–57.
- Baillon, S., Hamel, J. F., Wareham, V. E., and Mercier, A. (2012). Deep cold-water corals as nurseries for fish larvae. *Front. Ecol. Environ.* 10 (7), 351–356. doi: 10.1890/120022
- Bartzke, G., Siemann, L., Büssing, R., Nardone, P., Koll, K., Hebbeln, D., et al. (2021). Investigating the prevailing hydrodynamics around a cold-water coral colony using a physical and a numerical approach. *Front. Mar. Sci.* 8. doi: 10.3389/fmars.2021.663304
- Brooke, S. D., Holmes, M. W., and Young, C. M. (2009). Sediment tolerance of two different morphotypes of the deep-sea coral *Lophelia pertusa* from the Gulf of Mexico. *Mar. Ecol. Prog. Ser.* 390, 137–144. doi: 10.3354/meps08191
- Brooke, S., Ross, S. W., Bane, J. M., Seim, H. E., and Young, C. M. (2013). Temperature tolerance of the deep-sea coral *Lophelia pertusa* from the southeastern United States. *Deep. Res. Part II Top. Stud. Oceanogr.* 92, 240–248. doi: 10.1016/j.dsr.2012.12.001
- Buhl-Mortensen, L., Buhl-Mortensen, P., Dolan, M. F. J., and Holte, B. (2015a). The MAREANO programme – A full coverage mapping of the Norwegian off-shore benthic environment and fauna. *Mar. Biol. Res.* 11, 4–17. doi: 10.1080/17451000.2014.952312
- Buhl-Mortensen, P., Buhl-Mortensen, L., and Purser, A. (2017). "Trophic ecology and habitat provision in cold-water coral ecosystems," in *Marine Animal Forests: The Ecology of Benthic Biodiversity Hotspots*. Eds. S. Rossi, L. Bramanti, A. Gori and C. Orejas (Cham, Springer). doi: 10.1007/978-3-319-21012-4_20
- Buhl-Mortensen, P., and Freiwald, A. (2023). "Norwegian Coral Reefs," in *Coral Reefs of the World*, vol. 19. Eds. E. Cordes and F. Mienis (Springer, Cham). doi: 10.1007/978-3-031-40897-7_5
- Buhl-Mortensen, L., Olafsdottir, S. H., Buhl-Mortensen, P., Burgos, J. M., and Ragnarsson, S. A. (2015b). Distribution of nine cold-water coral species (*Scleractinia* and *Gorgonacea*) in the cold temperate North Atlantic: effects of bathymetry and hydrography. *Hydrobiologia* 759, 39–61. doi: 10.1007/s10750-014-2116-x
- Buhl-Mortensen, L., Vanreusel, A., and Gooday, A. J. (2010). Biological structures as a source of habitat heterogeneity and biodiversity on the deep ocean margins. *Mar. Ecol. Prog. Ser.* 31, 21–50. doi: 10.1111/j.1439-0485.2010.00359.x
- Büscher, J. V., Wisshak, M., Form, A. U., Titschack, J., Nachtigall, K., and Riebesell, U. (2019). *In situ* growth and bioerosion rates of *Lophelia pertusa* in a Norwegian fjord and open shelf cold-water coral habitat. *PeerJ* 7, e7586. doi: 10.7717/peerj.7586
- Chamberlain, J. A., and Graus, R. R. (1975). Water-flow and hydromechanical adaptations of branched reef corals. *Bull. Mar. Sci.*
- Chindapol, N., Kaandorp, J. A., Cronemberger, C., Mass, T., and Genin, A. (2013). Modelling growth and form of the scleractinian coral *Pocillopora verrucosa* and the influence of hydrodynamics. *E.J. PLoS Comput. Biol.* 9, e1002849. doi: 10.1371/journal.pcbi.1002849
- Corbera, G., Lo Iacono, C., Simarro, G., Grinyó, J., Ambroso, S., Huvenne, V. A. I., et al. (2022). Local-scale feedbacks influencing cold-water coral growth and subsequent reef formation. *Sci. Rep.* 12, 20389. doi: 10.1038/s41598-022-24711-7
- Davies, A. J., Duineveld, G. C. A., Lavaley, M. S. S., Bergman, M. J. N., and Van Haren, H. (2009). Downwelling and deep-water bottom currents as food supply mechanisms to the cold-water coral *Lophelia pertusa* (Scleractinia) at the Mingulay Reef complex. *Limnol. Oceanogr.* 54, 620–629. doi: 10.4319/lo.2009.54.2.0620
- Davies, A. J., Duineveld, G. C. A., van Weering, T. C. E., Mienis, F., Quattrini, A. M., Seim, H. E., et al. (2010). Short-term environmental variability in cold-water coral habitat at Viosca Knoll, Gulf of Mexico. *Deep. Res. Part I Oceanogr. Res. Pap.* 57, 199–212. doi: 10.1016/j.dsr.2009.10.012
- Davies, A. J., and Guinotte, J. M. (2011). Global habitat suitability for framework-forming cold-water corals. *PLoS One* 6 (4), e18483. doi: 10.1371/journal.pone.0018483
- Davies, A. J., Wisshak, M., Orr, J. C., and Murray Roberts, J. (2008). Predicting suitable habitat for the cold-water coral *Lophelia pertusa* (Scleractinia). *Deep. Res. Part I Oceanogr. Res. Pap.* 55, 1048–1062. doi: 10.1016/j.dsr.2008.04.010
- De Clippele, L. H., Gafeira, J., Robert, K., Hennige, S., Lavaley, M. S., Duineveld, G. C. A., et al. (2017). Using novel acoustic and visual mapping tools to predict the small-scale spatial distribution of live biogenic reef framework in cold-water coral habitats. *Coral Reefs* 36, 255–268. doi: 10.1007/s00338-016-1519-8
- De Clippele, L. H., Huvenne, V. A. I., Orejas, C., Lundälv, T., Fox, A., Hennige, S. J., et al. (2018). The effect of local hydrodynamics on the spatial extent and morphology of cold-water coral habitats at Tisler Reef, Norway. *Coral Reefs* 37, 253–266. doi: 10.1007/s00338-017-1653-y
- de Froe, E., Maier, S. R., Horn, H. G., Wolff, G. A., Blackbird, S., Mohn, C., et al. (2022). Hydrography and food distribution during a tidal cycle above a cold-water coral mound. *Oceanogr. Res. Papers Deep-Sea Res. I* 189, 103854. doi: 10.1016/j.dsr.2022.103854
- de Froe, E., Rovelli, L., Glud, R. N., Maier, S. R., Duineveld, G., Mienis, F., et al. (2019). Benthic oxygen and nitrogen exchange on a cold-water coral reef in the North-East Atlantic Ocean. *Front. Mar. Sci.* 6. doi: 10.3389/fmars.2019.00665
- Dickson, A. G., and Millero, F. J. (1987). A comparison of the equilibrium constants for the dissociation of carbonic acid in seawater media. *Deep Sea Res. Part A Oceanogr. Res. Pap.* 34, 173–1743. doi: 10.1016/0198-0149(87)90021-5
- Dickson, A. G., Sabine, C. L., and Christian, J. R. (2007). *Guide to best practices for ocean CO₂ measurements*, PICES Spec (British Columbia: North Pacific Marine Science Organization Sydney).
- Dodds, L. A., Roberts, J. M., Taylor, A. C., and Marubini, F. (2007). Metabolic tolerance of the cold-water coral *Lophelia pertusa* (Scleractinia) to temperature and dissolved oxygen change. *J. Exp. Mar. Biol. Ecol.* 349, 205–214. doi: 10.1016/j.jembe.2007.05.013
- Dorschel, B., Hebbeln, D., Foubert, A., White, M., and Wheeler, A. J. (2007). Hydrodynamics and cold-water coral facies distribution related to recent sedimentary processes at Galway Mound west of Ireland. *Mar. Geol.* 244, 184–195. doi: 10.1016/j.margeo.2007.06.010
- Dullo, W. C., Flögel, S., and Rüggeberg, A. (2008). Cold-water coral growth in relation to the hydrography of the Celtic and Nordic European continental margin. *Mar. Ecol. Prog. Ser.* 371, 165–176. doi: 10.3354/meps07623

- Eide, L. I. (1978). *Ocean Currents on the Halten and Malangsgrunnen Banks: Analysis of Observations Carried Out August 1972–September 1976* (Institut for Kontinentalsokkellundersokelser (Continental Shelf Institute).
- Eide, L. I. (1979). Evidence of a topographically trapped vortex on the Norwegian continental shelf. *Deep Sea Res. Part A Oceanogr. Res. Pap.* 26, 601–621. doi: 10.1016/0198-0149(79)90036-0
- Findlay, H. S., Artioli, Y., Moreno Navas, J., Hennige, S. J., Wicks, L. C., Huvenne, V. A. I., et al. (2013). Tidal downwelling and implications for the carbon biogeochemistry of cold-water corals in relation to future ocean acidification and warming. *Glob. Change Biol.* 19, 2708–2719. doi: 10.1111/gcb.12256
- Findlay, H. S., Hennige, S. J., Wicks, L. C., Navas, J. M., Woodward, E. M. S., and Roberts, J. M. (2014). Fine-scale nutrient and carbonate system dynamics around cold-water coral reefs in the northeast Atlantic. *Sci. Rep.* 4, 3671. doi: 10.1038/srep03671
- Fink, H. G., Wienberg, C., Hebbeln, D., McGregor, H. V., Schmiedl, G., Taviani, M., et al. (2012). Oxygen control on Holocene cold-water coral development in the eastern Mediterranean Sea. *Deep Res. Part I Oceanogr. Res. Pap.* 62, 89–96. doi: 10.1016/j.dsr.2011.12.013
- Finlay, C. C., Maus, S., and Beggan, C. D. (2010). International Geomagnetic Reference Field: The eleventh generation. *Geophys. J. Int.* 183, 1216–1230. doi: 10.1111/j.1365-246X.2010.04804.x
- Flögel, S., Dullo, W. C., Pfannkuche, O., Kiriakoulakis, K., and Rüggeberg, A. (2014). Geochemical and physical constraints for the occurrence of living cold-water corals. *Deep Res. Part II Top. Stud. Oceanogr.* 99, 19–26. doi: 10.1016/j.dsr2.2013.06.006
- Form, A. U., Büscher, J., and Hissmann, K. (2014). *RV POSEIDON Cruise Report POS455 LORELEI LOphelia REef Lander Expedition and Investigation, Bremerhaven - (Kristiansund) - Kiel, 24.06. - (12.07.) - 17.07.2013*. (Kiel: GEOMAR Helmholtz-Zentrum für Ozeanforschung), 29. doi: 10.3289/CR_POS_455
- Form, A. U., Büscher, J. V., and Hissmann, K. (2015). *RV POSEIDON Cruise Report POS473 LORELEI II: LOphelia REef Lander Expedition and Investigation II, Tromsø - Bergen - Esbjerg, 15.08. - 31.08. - 04.09.2014*. (Kiel: GEOMAR Helmholtz-Zentrum für Ozeanforschung), 25. doi: 10.3289/CR_POS_473
- Fosså, J. H., Mortensen, P. B., and Furevik, D. M. (2002). The deep-water coral *Lophelia pertusa* in Norwegian waters: Distribution and fishery impacts. *Hydrobiologia*. 471, 1–12. doi: 10.1023/A:1016504430684
- Frederiksen, R., Jensen, A., and Westerberg, H. (1992). The distribution of the scleractinian coral *Lophelia pertusa* around the Faroe Islands and the relation to internal tidal mixing. *Sarsia* 77, 157–171.
- Freiwald, A., Beuck, L., Rüggeberg, A., Taviani, M., and Hebbeln, D. (2009). The white coral community in the central Mediterranean sea revealed by ROV surveys. *Oceanography* 22, 58–74. doi: 10.5670/oceanog.2009.06
- Freiwald, A., Fosså, J. H., Grehan, A., Koslow, T., and Roberts, J. M. (2004). *Cold-water coral reefs*. Cambridge, UK: UNEP-WCMC.
- Freiwald, A., Hühnerbach, V., Lindberg, B., Wilson, J. B., and Campbell, J. (2002). The sula reef complex, Norwegian shelf. *Facies* 47, 179–200. doi: 10.1007/BF02667712
- Freiwald, A., Wilson, J. B., and Henrich, R. (1999). Grounding pleistocene icebergs shape recent deep-water coral reefs. *Sediment. Geol.* 125, 1–8. doi: 10.1016/S0037-0738(98)00142-0
- García, H. E., Locarnini, R. A., Boyer, T. P., Antonov, J. I., Baranova, O. K., Zweng, M. M., et al. (2013). *World Ocean Atlas 2013*, Vol. 4, Dissolved Inorganic Nutrients (phosphate, nitrate, silicate), eds. S. Levitus and A. Mishonov. (Technical Ed. NOAA Atlas NESDIS) 76, 25. doi: 10.7289/V5J67DWD
- GEOMAR Helmholtz-Zentrum für Ozeanforschung (2015). Research vessel POSEIDON. *JLSRF* 1, A36. doi: 10.17815/jlsrf-1-62
- GEOMAR Helmholtz-Zentrum für Ozeanforschung (2017). Manned submersible JAGO. *JLSRF* 3, A110. doi: 10.17815/jlsrf-3-157
- Georgian, S. E., Deleo, D., Durkin, A., Gómez, C. E., Kurman, M., Lunden, J. J., et al. (2016). Oceanographic patterns and carbonate chemistry in the vicinity of cold-water coral reefs in the Gulf of Mexico: implications for resilience in a changing ocean. *Limnol. Oceanogr.* 61, 648–665. doi: 10.1002/lno.10242
- Gjevik, B., and Straume, T. (1989). Model simulations of the M2 and the K1 tide in the Nordic Seas and the Arctic Ocean. *Tellus A*. 41A, 73–96. doi: 10.1111/1016.41A.issue-1
- Guinotte, J. M., Orr, J., Cairns, S., Freiwald, A., Morgan, L., and George, R. (2006). Will human-induced changes in seawater chemistry alter the distribution of deep-sea scleractinian corals? *Front. Ecol. Environ.* 4, 141–146. doi: 10.1890/1540-9295(2006)004[0141:WHCISC]2.0.CO;2
- Hansen, H. P., and Koroleff, F. (1999). “Determination of nutrients,” in *Methods of Seawater Analysis*. Eds. K. Grasshoff, K. Kremling and M. Ehrhardt (Germany: Wiley-VCH GmbH), 159–228.
- Hanz, U., Wienberg, C., and Hebbeln, D. (2019). Environmental factors influencing benthic communities in the oxygen minimum zones on the Angolan and Namibian margins. *Biogeosciences* 16, 4337–4356. doi: 10.5194/bg-16-4337-2019
- Haugan, P. M., Evensen, G., Johannessen, J. A., Johannessen, O. M., and Pettersson, L. H. (1991). Modeled and observed mesoscale circulation and wave-current refraction during the 1988 Norwegian Continental Shelf Experiment. *J. Geophys. Res.* 96, 10487. doi: 10.1029/91JC00299
- Hebbeln, D., Wienberg, C., Dullo, W.-C., Freiwald, A., Mienis, F., Orejas, C., et al. (2020). Cold-water coral reefs thriving under hypoxia. *Coral Reefs* 39, 853–859. doi: 10.1007/s00338-020-01934-6
- Hennige, S. J., Larsson, A. I., Orejas, C., Gori, A., De Clippele, L. H., Lee, Y. C., et al. (2021). Using the Goldilocks Principle to model coral ecosystem engineering. *Proc. R. Soc B*. 288, 20211260. doi: 10.1098/rspb.2021.1260
- Hennige, S. J., Wicks, L. C., Kamenos, N. A., Bakker, D. C. E., Findlay, H. S., Dumousseaud, C., et al. (2014). Short-term metabolic and growth responses of the cold-water coral *Lophelia pertusa* to ocean acidification. *Deep Res. Part II Top. Stud. Oceanogr.* 99, 27–35. doi: 10.1016/j.dsr2.2013.07.005
- Henry, L. A., and Roberts, J. M. (2007). Biodiversity and ecological composition of macrobenthos on cold-water coral mounds and adjacent off-mound habitat in the bathyal Porcupine Seabight, NE Atlantic. *Deep Res. Part I Oceanogr. Res. Pap.* 54, 654–672. doi: 10.1016/j.dsr.2007.01.005
- Henry, L. A., and Roberts, J. M. (2017). “Global biodiversity in cold-water coral reef ecosystems,” in *Marine Animal Forests: The Ecology of Benthic Biodiversity Hotspots* (Cham: Springer International Publishing), 235–256.
- Hogg, M. M., Tendal, O. S., Conway, K. W., Pomponi, S. A., Van Soest, R. W. M., Gutt, J., et al. (2010). “Deep Sea Sponge Grounds: Reservoirs of Biodiversity,” in *UNEP-WCMC Biodiversity Series*. Cambridge, UK: UNEP-WCMC.
- Hovland, M., Mortensen, P. B., Brattegard, T., Strass, P., and Rokoengen, K. (1998). Ahermatypic coral banks off mid-Norway: evidence for a link with seepage of light hydrocarbons. *Palaos*. 13, 189–200. doi: 10.2307/3515489
- Hovland, M., Ottesen, D., Thorsnes, T., Foss, J. H., and Bryn, P. (2005). Occurrence and implications of large *Lophelia*-reefs offshore Mid Norway. *Nor. Pet. Soc. Spec. Publ.* 12, 265–270. doi: 10.1016/S0928-8937(05)80053-0
- Hovland, M., Vasshus, S., Indreeide, A., Austdal, L., and Nilsen, Ø. (2002). Mapping and imaging deep-sea coral reefs off Norway 1982–2000. *Hydrobiologia* 471, 13–17. doi: 10.1023/A:1016576514754
- Hunter, T. (1989). Suspension feeding in oscillating flow: the effect of colony morphology and flow regime on plankton capture by the hydroid *Obelia longissima*. *Biol. Bull.* 176, 41–49. doi: 10.2307/1541887
- Järneggren, J., and Kutti, T. (2014). “*Lophelia pertusa* in Norwegian waters,” in *What have we learned since 2008?* NINA Report 1028. (Norsk institutt for naturforskning), 40.
- Jones, E., Chierici, M., and Skjelvan, I. (2018). Monitoring ocean acidification in Norwegian seas in 2017. *Rapport, Miljødirektoratet, M-XXX|2018*.
- Juva, K., Flögel, S., Karstensen, J., Linke, P., and Dullo, W.-C. (2020). Tidal dynamics control on cold-water coral growth: A high-resolution multivariable study on eastern Atlantic cold-water coral sites. *Front. Mar. Sci.* 7. doi: 10.3389/fmars.2020.00132
- Juva, K., Kutti, T., Chierici, M., Dullo, W.-C., and Flögel, S. (2021). Cold-water coral reefs in the Langenuen Fjord, Southwestern Norway—A window into future environmental change. *Ocean* 2, 583–610. doi: 10.3390/oceans2030033
- Kaandorp, J. A. (1999). Morphological analysis of growth forms of branching marine sessile organisms along environmental gradients. *Mar. Biol.* 134, 295–306. doi: 10.1007/s002270050547
- Kaandorp, J. A., and Sloot, P. M. A. (2001). Morphological models of radiate accretive growth and the influence of hydrodynamics. *J. Theor. Biol.* 209 (3), 257–274. doi: 10.1006/jtbi.2001.2261
- Kiriakoulakis, K., Freiwald, A., Fisher, E., and Wolff, G. A. (2007). Organic matter quality and supply to deep-water coral/mound systems of the NW European Continental Margin. *Int. J. Earth Sci.* 96, 159–170. doi: 10.1007/s00531-006-0078-6
- Larsson, A. I., and Purser, A. (2011). Sedimentation on the cold-water coral *Lophelia pertusa*: Cleaning efficiency from natural sediments and drill cuttings. *Mar. Pollut. Bull.* 62, 1159–1168. doi: 10.1016/j.marpolbul.2011.03.041
- Lauvset, S. K., Key, R. M., and Olsen, A. (2016). A new global interior ocean mapped climatology: The 1° × 1° GLODAP version 2. *Earth Syst. Sci. Data* 8, 325–340. doi: 10.5194/essd-8-325-2016
- Leinebo, R. (1973). Water masses and current in a section across the Norwegian shelf off Stadt. “Meteor. Forsch.-Ergeb 12, 11–23.
- Lim, A., Wheeler, A. J., Price, D. M., O’Reilly, L., Harris, K., and Conti, L. (2020). Influence of benthic currents on cold-water coral habitats: a combined benthic monitoring and 3D photogrammetric investigation. *Sci. Rep.* 10, 19433. doi: 10.1038/s41598-020-76446-y
- Ljøen, R., and Nakken, O. (1969). On the hydrography of the shelf waters off Møre and Helgeland. *FirkDir. Skr. Ser. HavUnders.* 15, 285–294.
- Maier, S. R., Brooke, S., De Clippele, L. H., de Froe, E., van der Kaaden, A. S., Kutti, T., et al. (2023). On the paradox of thriving cold-water coral reefs in the food-limited deep sea. *Biol. Rev.* 98, 1768–1795. doi: 10.1111/brv.12976
- McCulloch, M., Trotter, J., and Montagna, P. (2012). Resilience of cold-water scleractinian corals to ocean acidification: Boron isotopic systematics of pH and saturation state up-regulation. *Geochim. Cosmochim. Acta*. 87, 21–34. doi: 10.1016/j.gca.2012.03.027
- McDougall, T. J., and Barker, P. M. (2011). Getting started with TEOS-10 and the Gibbs seawater (GSW) oceanographic toolbox 28. Available at: <https://www.teos-10.org>.

- McGrath, T., Kivimaa, C., Tanhua, T., Cave, R. R., and McGovern, E. (2012). Inorganic carbon and pH levels in the Rockall Trough 1991–2010. *Deep-Sea Res. I* 68, 79–91. doi: 10.1016/j.dsr.2012.05.011
- Mehrbach, C., Culbertson, C. H., Hawley, J. E., and Pytkowicz, R. M. (1973). Measurement of the apparent dissociation constants of carbonic acid in seawater at atmospheric pressure. *Limnol. Oceanogr.* 18, 897–907. doi: 10.4319/lo.1973.18.6.0897
- MET-info (2013). *Ekstremværrapport - Hending: Hilde, 16.-17.11.2013*.
- MET-info and Fagerlid, G. O. (2014). *Ekstremværrapport Ivar, desember 2013*.
- Mienis, F., Bouma, T. J., Witbaard, R., van Oevelen, D., and Duineveld, G. C. A. (2019). Experimental assessment of the effects of cold-water coral patches on water flow. *Mar. Ecol. Prog. Ser.* 609, 101–117. doi: 10.3354/meps12815
- Mienis, F., Duineveld, G. C. A., and Davies, A. J. (2014). Cold-water coral growth under extreme environmental conditions, the Cape Lookout area, NW Atlantic. *Biogeosciences* 11, 2543–2560. doi: 10.5194/bg-11-2543-2014
- Mienis, F., Duineveld, G. C. A., Davies, A. J., Ross, S. W., Seim, H., Bane, J., et al. (2012). The influence of near-bed hydrodynamic conditions on cold-water corals in the Viosca Knoll area, Gulf of Mexico. *Deep. Res. Part I Oceanogr. Res. Pap.* 60, 32–45. doi: 10.1016/j.dsr.2011.10.007
- Milzer, G., Giraudeau, J., Faust, J., Knies, J., Eynaud, F., and Rühlemann, C. (2013). Spatial distribution of benthic foraminiferal stable isotopes and dinocyst assemblages in surface sediments of the Trondheimsfjord, central Norway. *Biogeosciences* 10, 4433–4448. doi: 10.5194/bg-10-4433-2013
- Mohn, C., Hansen, J. L. S., Carreiro-Silva, M., Cunningham, S. A., de Froe, E., Dominguez-Carrió, C., et al. (2023). Tidal to decadal scale hydrodynamics at two contrasting cold-water coral sites in the Northeast Atlantic. *Progr. Oceanogr.* 214, 103031. doi: 10.1016/j.pocean.2023.103031
- Morato, T., González-Irusta, J. M., and Dominguez-Carrió, C. (2020). Climate-induced changes in the suitable habitat of cold-water corals and commercially important deep-sea fishes in the North Atlantic. *Glob. Change Biol.* 26 (4), 2181–2202. doi: 10.1111/gcb.14996
- Mork, M. (1981). Circulation phenomena and frontal dynamics of the Norwegian coastal current. *Philos. Trans. R. Soc. London. Ser. A Math. Phys. Sci.* 302, 635–647. doi: 10.1098/rsta.1981.0188
- Mortensen, P. B., and Fosså, J. H. (2001). “Coral reefs and other bottom habitats on the ridge of Tautra in Trondheimsfjorden (Norway),” in *TT - Korallrev og Andre Bunnhabitater paa Tautraryggen i Trondheimsfjorden*. Internal Norwegian research report.
- Mortensen, P. B., and Fosså, J. H. (2006). “Species diversity and spatial distribution of invertebrates on deep-water *Lophelia* reefs in Norway,” in *Proceedings of the 10th International Coral Reef Symposium (Okinawa)*. 1849–1868.
- Mortensen, P. B., Hovland, M., Brattgard, T., and Farestveit, R. (1995). Deep water bioherms of the scleractinian coral *Lophelia pertusa* (L.) at 64° n on the Norwegian shelf: Structure and associated megafauna. *Sarsia* 80, 145–158. doi: 10.1080/00364827.1995.10413586
- Mortensen, P. B., Hovland, T., Fosså, J. H., and Furevik, D. M. (2001). Distribution, abundance and size of *Lophelia pertusa* coral reefs in mid-Norway in relation to seabed characteristics. *J. Mar. Biol. Assoc. United Kingdom* 81, 581–597. doi: 10.1017/S002531540100426X
- Orejas, C., Gori, A., and Rad-Menéndez, C. (2016). The effect of flow speed and food size on the capture efficiency and feeding behaviour of the cold-water coral *Lophelia pertusa*. *J. Exp. Mar. Biol. Ecol.* 481, 34–40. doi: 10.1016/j.jembe.2016.04.002
- Orejas, C., and Jiménez, C. (2017). “The builders of the oceans - Part I: Coral architecture from the tropics to the poles, from the shallow to the deep,” in *Marine Animal Forests: The Ecology of Benthic Biodiversity Hotspots*, eds. S. Rossi, L. Bramanti, A. Gori and C. Orejas. (Cham: Marine Animal Forests; Springer). doi: 10.1007/978-3-319-21012-4_10
- Osterloff, J., Nilssen, I., Järnegren, J., Engeland, T. V., Buhl-Mortensen, P., and Nattkemper, T. W. (2019). Computer vision enables short- and long-term analysis of *Lophelia pertusa* polyp behaviour and colour from an underwater observatory. *Sci. Rep.* 9, 6578. doi: 10.1038/s41598-019-41275-1
- Pawlowicz, R., Beardsley, B., and Lentz, S. (2002). Classical tidal harmonic analysis including error estimates in MATLAB using TDE. *Comput. Geosci.* 28, 929–937. doi: 10.1016/S0098-3004(02)00013-4
- Pierrot, D., Lewis, E., and Wallace, D. W. R. (2011). MS Excel program developed for CO₂ system calculations. doi: 10.3334/CDIAC/otg.CO2SYS_XLS_CDIA105a
- Poulain, P. M., Warn-Varnas, A., and Niiler, P. P. (1996). Near-surface circulation of the Nordic seas as measured by Lagrangian drifters. *J. Geophys. Res. C Ocean.* 101, 18237–18258. doi: 10.1029/96JC00506
- Purser, A., Bergmann, M., Lundqvist, T., Ontrup, J., and Nattkemper, T. W. (2009). Use of machine-learning algorithms for the automated detection of cold-water coral habitats: A pilot study. *Mar. Ecol. Prog. Ser.* 397, 241–251. doi: 10.3354/meps08154
- Purser, A., Larsson, A. I., Thomsen, L., and van Oevelen, D. (2010). The influence of flow velocity and food concentration on *Lophelia pertusa* (Scleractinia) zooplankton capture rates. *J. Exp. Mar. Biol. Ecol.* 395, 55–62. doi: 10.1016/j.jembe.2010.08.013
- Purser, A., Orejas, C., Gori, A., Tong, R., Unnithan, V., and Thomsen, L. (2013). Local variation in the distribution of benthic megafauna species associated with cold-water coral reefs on the Norwegian margin. *Cont. Shelf Res.* 54, 37–51. doi: 10.1016/j.csr.2012.12.013
- Ramos, A., Sanz, J. L., Ramil, F., Agudo, L. M., and Presas-Navarro, C. (2017). “The giant cold-water coral mounds barrier off Mauritania,” in *Deep-Sea Ecosystems Off Mauritania*, eds. A. Ramos, F. Ramil and J. Sanz. (Dordrecht: Springer). doi: 10.1007/978-94-024-1023-5
- Roberts, J. M., Peppe, O. C., Dodds, L. A., Mercer, D. J., Thomson, W. T., Gage, J. D., et al. (2005). “Monitoring environmental variability around cold-water coral reefs: the use of a benthic photolander and the potential of seafloor observatories,” in *Cold-Water Corals and Ecosystems. Erlangen Earth Conference Series*, eds. A. Freiwald and J. M. Roberts. (Berlin, Heidelberg: Springer), 483–502. doi: 10.1007/3-540-27673-4_24
- Roberts, J. M., Wheeler, A. J., and Freiwald, A. (2006). Reefs of the deep: The biology and geology of cold-water coral ecosystems. *Sci. (80-)* 312, 543–547. doi: 10.1126/science.1119861
- Roberts, J. M., Wheeler, A. J., Freiwald, A., and Cairns, S. D. (2009). *Cold-water corals: the biology and geology of deep-sea coral habitats*. (Cambridge University Press). doi: 10.1017/CBO9780511581588
- Rüggeberg, A., Flögel, S., Dullo, W. C., Hissmann, K., and Freiwald, A. (2011). Water mass characteristics and sill dynamics in a subpolar cold-water coral reef setting at Stjernsund, northern Norway. *Mar. Geol.* 282, 5–12. doi: 10.1016/j.margeo.2010.05.009
- Sætre, R. (1999). Features of the central Norwegian shelf circulation. *Cont. Shelf Res.* 19, 1809–1831. doi: 10.1016/S0278-4343(99)00041-2
- Sætre, R., and Ljøen, R. (1971). The Norwegian coastal current. *Proc. First Int. Conf. Port Ocean Eng.*
- Sakshaug, E., and Killingtveit, Å. (2000). *Elvene: et stort bidrag*. Eds. E. Sakshaug and J. Sneli (Trondheimsfjorden: Tapir forlag), 65–75.
- Sanna, G., Büscher, J. V., and Freiwald, A. (2023). Cold-water coral framework architecture is selectively shaped by bottom current flow. *Coral Reefs* 42, 483–495. doi: 10.1016/S0278-4343(99)00041-2
- Scheide, M. S. (2018). *Using Deep Learning for Automatic Classification of Marine Habitats in HiSAS Imagery*. (Master's Thesis). Norwegian University of Science and Technology.
- Schulz, K., Soetaert, K., Mohn, C., Korte, L., Mienis, F., Duineveld, G., et al. (2020). Linking large-scale circulation patterns to the distribution of cold water corals along the eastern Rockall Bank (northeast Atlantic). *J. Mar. Syst.* 212, 103456. doi: 10.1016/j.jmarsys.2020.103456
- Sebens, K. P., Grace, S. P., Helmuth, B., Maney, E. J., and Miles, J. S. (1998). Water flow and prey capture by three scleractinian corals, *Madracis mirabilis*, *Montastrea cavernosa* and *Porites porites* in a field enclosure. *Mar. Biol.* 131, 347–360. doi: 10.1007/s002270050328
- Soetaert, K., Mohn, C., Rengstorf, A., Grehan, A., and van Oevelen, D. (2016). Ecosystem engineering creates a direct nutritional link between 600-m deep cold-water coral mounds and surface productivity. *Sci. Rep.* 6, 35057. doi: 10.1038/srep35057
- Taviani, M., Freiwald, A., and Zibrowius, H. (2005). “Deep coral growth in the Mediterranean Sea: an overview,” in *Cold-Water Corals and Ecosystems. Erlangen Earth Conference Series*. eds. A. Freiwald and J. M. Roberts. (Berlin, Heidelberg: Springer), 137–156.
- Thiem, Ø., Ravagnan, E., Fosså, J. H., and Berntsen, J. (2006). Food supply mechanisms for cold-water corals along a continental shelf edge. *J. Mar. Syst.* 60, 207–219. doi: 10.1016/j.jmarsys.2005.12.004
- Thorsnes, T., Bellec, V. K., and Dolan, M. F. J. (2016). Cold-water coral reefs and glacial landforms from Sula Reef, mid-Norwegian shelf. *Geol. Soc. Mem.* 46, 307–308. doi: 10.1144/M46.74
- Tong, R., Davies, A. J., Yesson, C., Yu, J., Luo, Y., Zhang, L., et al. (2023). Environmental drivers and the distribution of cold-water corals in the global ocean. *Front. Mar. Sci.* 10. doi: 10.3389/fmars.2023.1217851
- Wagner, H., Purser, A., Thomsen, L., Jesus, C. C., and Lundälv, T. (2011). Particulate organic matter fluxes and hydrodynamics at the Tisler cold-water coral reef. *J. Mar. Syst.* 85, 19–29. doi: 10.1016/j.jmarsys.2010.11.003
- Wheeler, A. J., Beyer, A., Freiwald, A., de Haas, H., Huvenne, V. A. I., Kozachenko, M., et al. (2007). Morphology and environment of cold-water coral carbonate mounds on the NW European margin. *Int. J. Earth Sci.* 96, 37–56. doi: 10.1007/s00531-006-0130-6
- White, M., and Dorschel, B. (2010). The importance of the permanent thermocline to the cold water coral carbonate mound distribution in the NE Atlantic. *Earth Planet. Sci. Lett.* 296, 395–402. doi: 10.1016/j.epsl.2010.05.025
- White, M., Mohn, C., Stigter, H., and Mottram, G. (2005). “Deep-water coral development as a function of hydrodynamics and surface productivity around the submarine banks of the Rockall Trough, NE Atlantic,” in *Cold-Water Corals and Ecosystems*. Eds. A. Freiwald and J. M. Roberts (Springer-Verlag, Berlin Heidelberg), 503–514.
- White, M., Wolff, G. A., Lundälv, T., Guihen, D., Kiriakoulakis, K., Lavaley, M., et al. (2012). Cold-water coral ecosystem (Tisler Reef, Norwegian shelf) may be a hotspot for carbon cycling. *Mar. Ecol. Prog. Ser.* 465, 11–23. doi: 10.3354/meps09888
- Wilson, J. B. (1979). Patch development of the deep-water coral *lophelia pertusa* (L.) on rockall bank. *J. Mar. Biol. Assoc.* 59 (1), 165–177. doi: 10.1017/S0025315400046257
- Wulff, J. (2001). Assessing and monitoring coral reef sponges: Why and how? *Bull. Mar. Sci.* 69, 831–846.
- Yesson, C., Bedford, F., Rogers, A. D., and Taylor, M. L. (2017). The global distribution of deep-water Antipatharia habitat. *Deep. Res. Part II Top. Stud. Oceanogr.* 145, 79–86. doi: 10.1016/j.dsr2.2015.12.004
- Zheng, M. D., and Cao, L. (2014). Simulation of global ocean acidification and chemical habitats of shallow- and cold-water coral reefs. *Adv. Clim. Change Res.* 5 (4), 189–196. doi: 10.1016/j.accre.2015.05.002
- Zibrowius, H. (1980). The scleractinian corals of the Mediterranean and the northeastern Atlantic. *Mem. l'Institut. Oceanogr.*