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# Comparison of linear and nonlinear modeling approaches to develop an abundance index based on voyage and market data for a data-limited fishery

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**Introduction:** Worldwide coastal fish resources face severe threats from fisheries overexploitation. However, the evaluation of abundance trends in most coastal fisheries is constrained by limited data. This study took blackmouth croaker (*Atrobucca nibe*), a stock depleted by coastal trawl fishery in southwestern Taiwan, as an example to showcase the development of a relative abundance index from data-limited fishery (only landing data were available).

**Methods:** This study employed unique data sourcing from voyage data recorders (VDRs) to estimate fishing effort (in combination with landing data to estimate the catch per unit effort, CPUE) that demonstrated the potential application in global data-limited fisheries and assessed alternative approaches for predictors of fishery-targeting practices to condition effort for producing more accurate metrics of relative abundance. The nominal CPUE was standardized using three statistical models: generalized linear model, generalized additive model (GAM), and vector-autoregressive spatiotemporal models (VASTs) with two treatments of each of the four effects: environmental (sea temperature, salinity, density of mixing layer, seafloor temperature, and chlorophyll), vessel, spatial, and targeting effects. A total of 15 models were designed and compared for these effects, and their explanatory power (EP) was evaluated using cross-validation  $R^2$  and other metrics.

**Results and discussion:** Results indicated that the targeting effect exerted the most significant influence on standardization and was suggested to be addressed through the principal component analysis (PCA) approach. Both vessel and spatial effects demonstrated considerable influence, whereas the environmental effect exhibited a limited impact, possibly due to the small fishing area in this study. Regarding models' EP, given the nonlinear nature of

the PCA algorithm and environmental data, the study highlighted the superiority of the GAM over linear-based models. However, incorporating nonlinear features in VAST (M15) makes it the most effective model in terms of predictive power in this study. Concerning the stock status, despite variations in relative CPUE trends among major models, a general declining trend since 2015 signals the potential decline of the blackmouth stock and urges fishery managers to consider further design of management measures.

#### KEYWORDS

data-limited fisheries, fishing effort estimation, CPUE standardization, VAST, GAM (generalized additive models), spatiotemporal effects, nonlinear feature VAST, five-fold cross-validation

## **1** Introduction

Coastal marine ecosystems, renowned for their high productivity, face severe threats from pervasive human disturbance, with overfishing emerging as a primary concern leading to the depletion of fish stocks (Jackson et al., 2001; Halpern et al., 2008; Nias, 2013). Many data-rich stocks have been assessed with sophisticated approaches, such as stock assessments, and managed accordingly with strengthened measures. Recent analyses suggested a global recovery of the population of well-managed stocks (Duarte et al., 2020; Hilborn et al., 2020), although this is cautioned to be over-optimistic (Duarte et al., 2020). In contrast, many coastal fisheries lack proper management due to insufficient data, creating uncertainty about stock depletion levels and posing challenges in persuading stakeholders to implement necessary measures, such as reducing fishing efforts (King, 2007). The absence of adequate data is attributed to various factors, notably the frequent omission of small-scale coastal fisheries from logbook submission requirements, which hinders the derivation of comprehensive geo-referenced effort data (Chang et al., 2019; Sari et al., 2021).

Various alternative approaches can be explored to address the issue, especially in obtaining abundance indices that provide insights into population abundance changes over time (Chang et al., 2017; Ducharme-Barth et al., 2018). Utilizing market landing data, which includes date information and vessel identity and adheres to specific data quality standards (e.g., minimal discarding due to size-specific high grading, at-sea dumping, and losses during fish processing), can serve as a substitute to represent the catch (Afonso-Dias et al., 2004; Lourenço and Pereira, 2006; Bastardie et al., 2010; FAO, 2023). Fishing effort can be estimated by simple methods such as taking days absent from port or each landing event as a multiplier of fishing day (Greenstreet et al., 2009; Sonderblohm et al., 2014) or, more sophisticatedly, by applying fishery-specific algorithms to auxiliary data such as the data from vessel monitoring system (VMS), coastal surveillance radar system (CSRS), or voyage data recorders (VDRs) (Lee et al., 2010; Gerritsen and Lordan, 2011; Sonderblohm et al., 2014; Chang, 2016; Chang et al., 2017). Catch per unit effort (CPUE) can then be calculated from those data.

CPUE can be influenced by various factors unrelated to population abundance, making the nominal CPUE seldom directly proportional to the abundance over the entire history of exploitation and so needs to be standardized to eliminate the impact of these factors on the abundance index (Maunder and Punt, 2004). The generalized linear model (GLM) is widely employed for this purpose, and the generalized additive model (GAM), incorporating components to capture both linear and nonlinear relationships between CPUE and influencing factors (Sacau et al., 2005), is also commonly used (Winker et al., 2014; Chang et al., 2019). Given that catch data often include a substantial number of zero catches for the studied species, the two-stage delta method [e.g., delta-GLM or delta-GAM, consisting of a positive-catch model (PCM) and a zeroproportion model (ZPM)] is frequently applied to address the impact of high zero-catch records in the data (Lo et al., 1992; Chang et al., 2017), unless the proportion of zero-catch records is exceptionally high (e.g., Wu et al., 2021).

These statistical models offer the flexibility to treat any categorical factor, such as vessel identifier, to allow estimation of the vessel-specific differences in fishing efficiency as either a fixed or random effect. When incorporating random effects, the adaptation of the GLM/GAM to "mixed-effect" models (GLMM/GAMM) becomes essential to address variability associated with random factors (Helser et al., 2004). In addition to these models, an increasing number of studies have embraced vector-autoregressive spatiotemporal models (VASTs) (Thorson, 2019; Thorson and Haltuch, 2019; Hansell et al., 2022), a geostatistical model that employs a delta-GLMM approach and simultaneously integrates spatial and spatiotemporal effects within the statistical R package.

Factors that have a significant influence on fishing efficiency are necessary to be considered in the CPUE standardization models, such as sampling time (year, quarter, or other timespan), location (statistical region, 5° square area, or other resolution), and vessel-related aspects (vessel size or vessel characteristics) (Murray et al., 2013). The targeting effect is particularly pivotal in understanding the intricate dynamics of multispecies fisheries (Maunder and Punt, 2004; Chang et al., 2011; Okamura et al., 2018; Chang et al., 2019), especially in trawl fisheries that frequently capture a diverse array of species (up to 200) simultaneously and change targets seasonally. k-means clustering analysis, commonly used to characterize fish targeting, groups species-specific catch data into clusters showing similar catch patterns (Chang et al., 2011; Wu et al., 2021). One of the inherent challenges in clustering analysis lies in determining the optimal number of groups into which the data should be divided. Alternatively, some studies propose that principal component analysis (PCA) is more effective in selecting species that capture information about the composition of fish market catches (Winker et al., 2013, 2014; Chang et al., 2019). PCA's primary advantage lies in its ability to circumvent the need for categorizing fishing vessels while accounting for the covariance between features in the catch composition data.

The majority of fisheries with a spatial dimension reveal that adjacent locations exhibit similar catch rates, suggesting that the distribution of fish abundance or availability is not random but exhibits a discernible degree of spatial continuity (Booth, 2000; Poulsen and Holm, 2007). Addressing this spatial component is crucial for obtaining accurate estimates of relative abundance (Swartzman et al., 1992). Conversely, the spatial distribution of marine fishes is significantly influenced by environmental factors, encompassing biological elements such as chlorophyll concentration and hydrological factors such as water temperature and salinity (Solanki et al., 2003; Pennino et al., 2020). Therefore, it is imperative to account for the spatial heterogeneity in CPUE data and the spatial effects of environmental and hydrological factors during the standardization process.

Blackmouth croaker, Atrobucca nibe (Jordan and Thompson, 1911), is one of the most important demersal fish species in the southwest coastal waters (including offshore waters hereinafter) of Taiwan, with a high economic value. This species, estimated to have a growth rate of 0.39 year<sup>-1</sup> with  $L_{\infty} = 53.11$  cm, was harvested mainly by small-scale multispecies trawl fishery and gillnet fishery. There has been a significant decline in the domestic catch of Taiwan, dropping from 2,500 tons in 1993 to less than 500 tons after 2011 (Huang et al., 2022; Shao, 2023). However, like the other coastal fisheries around Taiwan, the exemption of submission of catch data from the fishery has hindered informing managers/ industry about stock status derived from scientific evidence. For the stock, only market landing data, no spatiotemporal catch and effort data, were available and were considered a data-limited stock (ICES, 2012). This study uses blackmouth croaker off southwestern Taiwan as an example to demonstrate a set of procedures to develop relative abundance indices that could inform future management measures.

This study conducted a comprehensive analysis involving three levels: (1) nominal CPUE derivation: deriving nominal CPUE from market landing data and VDR data, originally designed for calculating fuel subsidies for small-scale fishing vessels (Chang, 2016). (2) CPUE standardization: standardizing CPUE using three statistical models (delta-GLM, delta-GAM, and VAST). Two treatments were applied for each of the four effects: environmental effect (with and without), vessel effect (fixed and random), spatial effect (with and without), and targeting effect (*k*-means and PCA). This resulted in a total of 15 model runs. (3) Model Explanatory Power (EP) Evaluation: assessing the EP of model runs through cross-validation  $R^2$  (CV- $R^2$ ). The latter, calculated via a cross-validation procedure (Zhang and Yang, 2015; Hsu et al., 2022), determined the overall correlation between observed and predicted values while mitigating overfitting concerns (Chang et al., 2017). The findings of this study offer a scientific foundation for managers to engage with stakeholders regarding stock trends. Additionally, the results provide valuable insights and guidelines for other coastal fisheries in Taiwan or similar regions with similar data-limited situations, aiding in the development of relative abundance indices.

# 2 Materials and methods

## 2.1 The data

Blackmouth croaker is predominantly harvested by coastal trawl and gillnet fisheries in the northern and southwestern regions of Taiwan, with the trawl fishery having a higher catch proportion in recent years (> 60% in general; Fisheries Agency, 2022). Two distinct stocks were identified in the north and southwest (Hwang and Chen, 1984), and this study concentrates on the southwestern stock due to its higher prevalence in the overall catch. The species is captured alongside nearly 200 demersal species throughout the year by the trawl fishery (refer to Section 2.2.1.2), with a moderate higher in catch during the spawning season from January to May (Hsiao et al., 2017). Mitochondrial cytochrome b gene analyses on 59 samples from five locations in the southwestern waters confirm that blackmouth croakers in this region belong to the same stock (unpublished project report by author SK Chang). The majority of blackmouth croaker caught by the trawl fishery in the southwestern waters (> 90%) is landed at Keziliao Fish Market in Zihquan, Kaohsiung City, the primary market for vessels operating in this region, particularly those engaged in coastal trawl fishing (Huang et al., 2022).

Daily vessel-specific landing data, including species-specific weights spanning from 2011 to 2021, were obtained from the market. The data were from all fisheries (no separation of fishing gears) and included a total of 268,176 landing events (annual average = 24,380) for all species and 33,907 landing events (annual average = 3,082) for blackmouth. Interviews with fishers indicated that, given the species' high market value in auction markets for whole fish, at-sea discards or dumping are unlikely. This observation supports the reliability of using individual vessels' landing data as a representative source for their catch information (Chang et al., 2019).

## 2.2 Fishing effort and nominal CPUE

Most small-scale offshore fishing vessels in Taiwan have installed simplified VDR to evaluate fuel subsidies from the active

moving hours of the vessels at sea (Chang, 2016). Akin to the "Black Box" on airplanes, the VDR is an equipment-fitted onboard ships that record the various data on a ship which can be used for reconstruction of the voyage details and vital information during an accident investigation (Bhattacharjee, 2019). It was required to be installed on passenger ships and ships other than passenger ships of 3,000 gross tonnage and upwards constructed on or after 1 July 2002, under International Maritime Organization (IMO) regulations that entered into force on 1 July 2002 (IMO, n.d.). The simplified device provides high-resolution temporal position and speed data like that of VMSs (Figure 1), not in real time but with lower device cost, no data transmission fee, and higher resolution at 3-min intervals. This information incidentally can be used to estimate fishing effort through a general criteria of 2-5 knots for Taiwanese trawl fishery that was developed by a five-step procedure for reviewing fishing patterns and defining speed criteria (Chang, 2016). Constrained by the restriction of the VDR data provision policy of the Fisheries Agency (FA), VDR data of vessels with a 5-year average of blackmouth catch exceeding 1.5 mt were obtained. A total of 29 coastal trawlers (about 25% of the vessels that have landed blackmouth for the past 5 years) were used for this study (the sample vessels), ranging from 20 to 100 gross registered tonnage, whose blackmouth catch during 2011-2021 was 381,743 mt, 62% of the total blackmouth catch, from 20,756 trips. Annual number of sample vessels ranged from 22 to 28.

Based on the work of Chang (2016) and interviews with fishermen, the following rules were established and applied to the VDR data to estimate fishing effort through an automated algorithm: (1) the VDR records that leaving from and later returning to the fishing port were marked as a trip. The number of VDR records between leave and return must be over 40 (about 2h). The VDR records located continuously in port were excluded. (2) The records with a speed of 2–5 knots for at least ten consecutive records (approximately 30 min) were considered as fishing. The short period of slow speed leaving and returning the port was considered as navigating. (3) The location in a resolution of 0.1° square that the vessel stayed longest during the fishing day was defined as the representative fishing location. (4) The records with speed > 5 knots were assumed to be navigating (e.g., transiting to fishing grounds).

The spatial distributions of VDR records identified as fishing according to rule (2) are illustrated in Figure 2. It is important to note that the VDR record distribution may cover a larger area than the CPUE distribution used for standardization analyses, as only one location in a trip was chosen to represent fishing according to rule (3).



#### FIGURE 1

VDR position and vessel speed of a Taiwanese coastal trawler fishing for blackmouth croaker in southwestern Taiwan during January 16–17 of 2011. The upper panel indicates vessel tracks with latitude on the Y axis and longitude on the X axis, and the bottom panel indicates vessel speeds with time and date on the X axis and speed (knots) on the Y axis. The two panels include corresponding color dots to indicate which points on the map (top panel) correspond to which points in the speed time series (bottom panel). The colors rotate, and dots may overlap in the same position in the upper panel. More descriptions on the plots can be found in Supplementary Material.



Following the outlined procedures, a comprehensive dataset of trip records was collected from the sampled vessels. Multiple sets were deployed during each trip, as illustrated in Figure 1. However, the fishing hours per trip exhibited variability based on fishing conditions and weather. Therefore, fishing effort was calculated as fishing hours. Subsequently, the trip-basis effort data (measured in fishing hours) were integrated with the corresponding market landing data of the trip for the sample vessels, to construct a tripbasis logbook-like data with time and location information for the calculation of nominal CPUE (kg per hour). On average, these sample vessels undertook approximately 1,887 trips with 3.5 mt of blackmouth annually, each spanning an average of 5.3 fishing hours per trip (excluding transition time). Since the multispecies feature of the trawl fishery and the relatively low abundance of blackmouth stock, about 60% of the trips contained zero blackmouth catch.

## 2.3 CPUE standardization

## 2.3.1 Key effects considered

## 2.3.1.1 Environmental effect

The spatial distribution of marine fishes is strongly influenced by environmental factors. The studied region encompasses cold water from the China Coastal Current and warm water from the Kuroshio Branch Current in different seasons, so examining the environmental effect should be necessary. Blackmouth is distributed in a depth of 40 m–200 m (Shao, 2023). Hydrological data including sea temperature at depth of 109 m (tem109), salinity at depth of 109 m (sal109), density ocean mixed layer thickness (mlt), chlorophyll (chl), and seafloor potential temperature (sft) were collected from the Copernicus Marine Environment Monitoring Service (CMEMS) (https://resources.marine.copernicus.eu/products), for the region ranging from 117°E to 121°E and 21°N to 24°N (Figure 2). The environmental data were obtained from CMEMS Global Ocean Physics Reanalysis (https://doi.org/10.48670/moi-00021) for 2000– 2019 and from the Operational Mercator global ocean analysis and forecast system (https://doi.org/10.48670/moi-00016) for 2020–2021. These data were cross-checked with similar datasets from the Ocean Data Bank that were sponsored by the National Science and Technology Council, Taiwan, with higher resolution around Taiwan but a shorter time period (2000–2018) (http://www.odb.ntu.edu.tw/odb-services/). While there were differences between the two datasets in absolute values, the trends were basically similar.

Temperature and salinity data exhibit variations at different depths, which can complicate the analysis process. To ensure a standardized analysis, a fixed depth of approximately 109 m was chosen. This depth corresponds to the estimated habitat water depth of the blackmouth croaker (Shao, 2023). Fixing the depth at a specific level allows for consistent comparisons and interpretations of temperature and salinity data across different sampling points, reducing the influence of depth-related variations in the analysis.

Considering the small geographic size of the studied area and small within-month variation of the environmental data observed, these remote-sensing environmental data were computed as monthly averages on a spatial resolution of 0.1° to be consistent with the spatial resolution of the fishery data, using R Statistical Software (v4.2.2; R Core Team, 2021).

## 2.3.1.2 Targeting effect

Many fisheries catch multiple species simultaneously. When fishers specifically target a species during a fishing trip, they may employ fishing operations or techniques that are not optimized for catching other species (Okamura et al., 2018). This fishing strategy (i.e., targeting) may affect the catchability and the representativeness of CPUE as the abundance index and thus requires to be addressed in the standardization process (Maunder and Punt, 2004; Chang et al., 2011, Chang et al., 2019).

Two methods were used to generate targeting covariates. The first one was the k-means clustering method, which is a non-hierarchical clustering technique (Tan et al., 2006). The desired

number of clusters (k) was chosen according to the "elbow method" (Kassambara, 2017) with scree plots (Chang et al., 2019). The species composition of the catch data was partitioned into k groups using this method, and the group ID was subsequently utilized as a categorical variable in the standardization model.

Winker et al. (2013), Winker et al. (2014) suggested a different approach to address the targeting effect in CPUE standardization for multispecies fisheries by employing PCA to calculate continuous principal components (PCs) of catch composition. This study applied the same approach and used these PC scores, which were assigned to each CPUE record, as predictors (continuous variables) in the standardization model to account for changes in fishing tactics over time (i.e., shifts in target species). More information on the approach can be found in the cited works (Winker et al., 2013, 2014; Chang et al., 2019).

Taiwanese offshore trawl fishery in the southwestern region is a multispecies fishery that includes nearly 200 species. Major commercial species such as Loliginidae, Carangidae, Centrolophidae, Priacanthidae, Trichiuridae, and Sciaenidae (including blackmouth croaker and bighead croaker that were targeted simultaneously) were selected and grouped as the primary targets of the fishery, and the remaining fish species (approximately 150 species, including many valuable species to some trawlers) are categorized as the fish group of "others" fishes. Catches of the seven species groups were aggregated by trip and transformed to catch composition in percentage before applying PCA.

#### 2.3.1.3 Spatial effect

Fishery-dependent CPUE data are often characterized by high heterogeneity in the spatial distribution of habitat quality and fishing effort (Hsu et al., 2022; Pourtois et al., 2022). Therefore, considering spatial factors in statistical models can compensate for the limitations of traditional analysis methods. The presence of spatial autocorrelation in fishing catch data can be assessed using the Moran's I test, where a significant result indicates the presence of spatial effects (Cliff and Ord, 1981). Description of the computation of Global Moran's I is provided in Supplementary Section 1.

#### 2.3.2 Statistical models

This study adopts a two-stage procedure (delta procedure), which consists of a PCM and a ZPM to address the high probability of zero catch values in the fishery data (Lo et al., 1992; Chang et al., 2019). For the PCM, ln(CPUE) are modeled assuming a normal distribution, while the ZPM predicts the presence or absence of blackmouth using logistic regression. The standardized index was the product of these model-estimated components.

## 2.3.2.1 Delta-GLM and delta-GAM

GLM and GAM were used for the CPUE standardization. For the PCM, the GLM assumes that the expected value of a transformed response variable is a linear combination of exploratory variables (Guisan et al., 2002) (see Supplementary Section 1). The GLM of this study specifies the response variable to be the natural logarithm of blackmouth CPUE.

The GAM (Hastie and Tibshirani, 2017) (Supplementary Material) is a semi-parametric extension of GLM, with the underlying

assumption that the response variable is related to smooth additive functions of the explanatory variables. GAM was used because of the concern that GLM cannot handle potentially nonlinear relationships between CPUE and PC covariates (Winker et al., 2013).

The covariates considered in the models included: year (2011–2021), quarter (Q1–Q4), targeting factor (k-means or PCs), environmental factor, spatial factor ( $0.1^{\circ} \times 0.1^{\circ}$  grid cell), and vessel factor. Vessel factor, in terms of vessel identification (vessel ID), was considered because each fishing vessel has its own unique characteristics, experience, and fishing capability, in addition to vessel size, which can impact catchability and the representativeness of CPUE as a resource indicator. The factors of year, quarter, vessel size, and target factor of *k*-means were treated as categorical variables, and the rest were treated as continuous variables. The environmental and principal component variables were modeled as smoothers in the GAM.

Spatial factor was treated as a random effect for all standardization models. Vessel factor was treated as a fixed effect in some models and a random effect in others. All remaining factors were treated as fixed effects. Helser et al. (2004) recommended treating the vessel factor as a random effect as it allows the data from various vessels to be combined and a single continuous time series of biomass indices to be developed. When random effects were considered, the mixed-effect models (i.e., GLMM and GAMM) were used for the CPUE standardization (Su et al., 2008; Forrestal et al., 2019; Grüss et al., 2019). A list of all model structures, as specified in the input to R, was provided in the Supplementary Material.

The ZPM predicts the presence or absence of blackmouth using logistic regression in GLM or logistic regression additive model in GAM (Supplementary Material).

The GLM/GAMs (without random effects) and GLMM/GAMMs (with random effects) were constructed using R Statistical Software (v4.2.2; R Core Team, 2021) and the function of the mgcv (Wood and Wood, 2015) and gamm4 package (Wood et al., 2017).

## 2.3.2.2 Spatiotemporal model (VAST)

R package VAST (version 3.10.0) (https://github.com/James-Thorson-NOAA/VAST) developed by Thorson (2019) was applied in this study. By default, VAST is a delta-generalized linear mixed modeling (GLMM) framework that separately estimates the probability of encounter (ZPM) and the mean catch rate of positive catches (PCM) (Supplementary Material).

Both the spatial and spatiotemporal random effects are assumed to be correlated in space; specifically, we smooth the spatial random effect by assuming that residuals at nearby sites are more similar than those at remote sites. The correlation between the spatial and spatiotemporal residuals at two locations (s and s') is assumed to decrease as the distance between them increases. This decrease in correlation is modeled using the Matérn function (Supplementary Material).

#### 2.3.2.3 Nonlinear-featured VAST

By default, VAST assumes a linear effect for each covariate, but it might be enhanced to perform nonlinear features. GAM is suitable for handling potentially nonlinear relationships between CPUE and covariates (e.g., PCs) (Winker et al., 2013). We, therefore, utilize the outcomes from GAM and polynomial basis expansion to capture the nonlinear impact of environmental factors and PCs on CPUE. The model was arbitrarily termed as "VAST (nonlinear)" here, in contrast to "VAST (linear)," and can be expressed as:

$$logit(p_{i}) = \beta_{1}(t_{i}) + L_{\omega_{1}}\omega_{1}(s_{i}) + L_{\varepsilon_{1}}\varepsilon_{1}(s_{i}, t_{i}) + L_{\delta_{1}}\delta_{1}(v_{i}) + \sum_{p=1}^{n_{p}}\sum_{k=1}^{n_{k}}\gamma_{1,k}(p)f_{1,k}(X(s_{i}, t_{i}, p)) + \sum_{j=1}^{n_{j}}\sum_{k=1}^{n_{k}}\lambda_{1,k}(j)g_{1,k}(Q(i, j))$$

and

$$\begin{split} \log(u_i) &= \beta_2(t_i) + L_{\omega_2} \omega_2(s_i) + L_{\varepsilon_2} \varepsilon_2(s_i, t_i) + L_{\delta_2} \delta_2(v_i) \\ &+ \sum_{p=1}^{n_p} \sum_{k=1}^{n_k} \gamma_{2,k}(p) f_{2,k}(X(s_i, t_i, p)) + \sum_{j=1}^{n_j} \sum_{k=1}^{n_k} \lambda_{2,k}(j) g_{2,k}(Q(i, j)) \end{split}$$

where  $\beta(t_i)$  is the intercept for year  $t_i$ ,  $\omega(s_i)$  denotes timeinvariant spatial variations at location  $s_i$ ,  $\varepsilon(s_i, t_i)$  denotes timevarying spatiotemporal variations at location  $s_i$  in year  $t_i$ , and  $\delta(v_i)$  denotes the effect of vessel  $v_i$  on catchability and is normally distributed ( $\delta_l(v_i) \sim N(0,1), l=1,2$ ),  $L_{\omega}$  and  $L_{\varepsilon}$  are the scaling coefficients of the spatial and spatiotemporal random effect distributions. In addition, the  $\gamma(p)$  is the  $p^{th}$  habitat covariate  $X(s_i,$  $t_i,p)$  at location  $s_i$  in year  $t_i$  (i.e., the impact of environmental factors on observed blackmouth CPUE) and the  $\lambda(j)$  is the  $j^{th}$  catchability covariates Q(i,j) for observation *i* (i.e., the impact of target factor (*k*-means or PCs) on observed blackmouth CPUE and the quarter factor) (Thorson, 2019). The nonlinear features are denoted by  $f_{1,p}(\cdot), f_{2,p}(\cdot), g_{1,j}(\cdot)$  and  $g_{2,j}(\cdot)$ , which are the results from a GAMM that uses a polynomial spline for the basis. R-codes of this approach are provided in the Supplementary Material.

Calculations of abundance indices for the statistical models are provided in Supplementary Material.

# 2.4 Models design and explanatory power evaluation

This study constructed a model structure with 15 model designs (M1–M15) to compare the influences of the three statistical models (GLM, GAM, and VAST) and the four key factors: environment factor (with or without), vessel factor (treated as fixed or random effect), spatial factor (with or without), and target factors (addressed by k-means or PCA methods) (Table 1). A flowchart showing the compositions of the models investigated is shown in Figure 3. Each model has a corresponding model to compare with; for example, M1 is to compare with M0 to see the effect of treatment of target factor using k-means versus PCA (the last column of Table 1). The study performed

TABLE 1 The 15 CPUE standardization models designed in this study to compare the influences of three statistical models and four key factors.

| Model code     | Stat. model         | Envir. factors  | Vessel factor | Spatial factor | Target factor | Versus          |  |  |  |
|----------------|---------------------|-----------------|---------------|----------------|---------------|-----------------|--|--|--|
| Set 1 (S1)     |                     |                 |               |                |               |                 |  |  |  |
| M1-glm_fk      | GLM                 |                 | Fixed         |                | k-means       |                 |  |  |  |
| M2-glm_efk     | GLM                 | \$              | Fixed         |                | k-means       | M1              |  |  |  |
| M3-glm_fp      | GLM                 |                 | Fixed         |                | PCA           | M1, M2          |  |  |  |
| M4-glm_efp     | GLM                 | \$              | Fixed         |                | PCA           | M2, M3, M7, M12 |  |  |  |
| M5-gam_fp      | GAM                 |                 | Fixed         |                | PCA*          | M3, M12         |  |  |  |
| Set 2 (S2)     |                     |                 |               |                |               |                 |  |  |  |
| M6-glmm_erk    | GLMM                | \$              | Random        |                | k-means       | M2              |  |  |  |
| M7-glmm_erp    | GLMM                | \$              | Random        |                | РСА           | M4              |  |  |  |
| Set 3 (S3)     |                     |                 |               |                |               |                 |  |  |  |
| M8-VAST_efsk   | VAST (linear)       | \$              | Fixed         | \$             | k-means       | M9              |  |  |  |
| M9-VAST_efsp   | VAST (linear)       | \$              | Fixed         | \$             | PCA           | M8, M11         |  |  |  |
| M10-VAST_ersk  | VAST (linear)       | \$              | Random        | \$             | k-means       | M8              |  |  |  |
| M11-VAST_ersp  | VAST (linear)       | \$              | Random        | \$             | PCA           | M9              |  |  |  |
| Set 4 (S4)     |                     |                 |               |                |               |                 |  |  |  |
| M12-gam_efp    | GAM                 | ×*              | Fixed         |                | PCA*          | M4, M14         |  |  |  |
| M13-glmm_efsp  | GLMM                | \$              | Fixed         | \$             | PCA           | M14             |  |  |  |
| M14-gamm_efsp  | GAMM                | \$ <sup>*</sup> | Fixed         | \$             | PCA*          | M12, M13        |  |  |  |
| M15-VAST_efspn | VAST<br>(Nonlinear) | ¥*              | Fixed         | *              | PCA*          | M9              |  |  |  |

The "x" symbol indicates that the respective factor has been considered in the model. The last column of "Versus" shows the model to be compared with. The "\*" symbol indicates that the respective factors have been smoothed in the model.



four sets of comparisons: the first set (S1) compared M1–M5 for environment factor and treatment of target factor. Including environment factor in the rest models, the second set (S2) compared M6–M7 for vessel factor and target factor, without consideration of spatial factor, and the third set (S3) compared M8–M11 for vessel factor and target factor under the consideration of spatial factor. With fixed vessel factor and PCA-treatment of target factor, the fourth set (S4) compared M12–M15 for statistical models. QQ plots of all the models are provided in Supplementary Material and suggested that they all conformed with the lognormal distribution assumption, affirming the appropriateness of the assumption regarding the error distribution for the CPUE standardization.

The study examined the predictive capability of two categories of methods introduced by Hinton and Maunder (2004) for CPUE model evaluation. The first category applied pseudo- $R^2$  (higher value preferred), a log-likelihood–based measurement representing the improvement in model likelihood over a null model (Faraway, 2016; Chang et al., 2019).

The second category was the cross-validation (CV) method. The study applied a stratified random sampling approach using year and vessel as strata in the fivefold CV procedure. After repeated 20 times, the  $CV-R^2$  that determines the overall correlation between the actual and predicted value (Li et al., 2011; Chang et al., 2017; Zhang et al., 2023) was calculated and averaged.  $CV-R^2$  was used to evaluate model's EP.

Although the Akaike information criterion (AIC), falling into the second category, is also a valuable and widely employed tool for model comparison, it is crucial to recognize that its values are impacted by the sample size, because they are derived from a likelihood function. Therefore, as elucidated in Chang et al. (2019) and the citation,

caution is warranted when comparing one-stage GLMs and twostage delta-GLMs with different sample sizes using AIC. Additionally, specifying the variance parameters of the likelihood function in twostage delta-GLM or delta-GAM cases is a complex undertaking. Therefore, this study used  $CV-R^2$  when evaluating model EP.

# **3** Results

## 3.1 Fishing effort

Figure 1 provides an example of a VDR data track for effort estimation; an explanatory plot modified from Chang (2016) is provided in Supplementary Material. The 2011–2021 data obtained from the FA of Taiwan contains 29 vessels and 45,000 records in the study area. Using the procedures described in Section 2.1, about 43,530 trips with 210 tons of blackmouth catch and 230,809h of fishing effort were constructed from the data. Figure 2 shows the distribution of the fishing efforts, which were used for the following analyses.

# 3.2 CPUE standardizations

## 3.2.1 Targeting effect

Catch composition data were used for the targeting effect analyses, representing potential targeting strategies. For the *k*means method, which minimizes the sum of squares from points to the cluster centers, there is no clear number of clusters suggested by the scree plot from the elbow method, but an *ad hoc* choice of five clusters could be a compromise option between smaller



improvement of within-cluster sum of square and not too large number of clusters. Each of the clusters has different dominant species compositions (Figure 4), indicating five different types of target trips by the sample vessels: the dominant species of Cluster 1 is Sciaenidae, Cluster 2 Carangidae, Loliginidae, and "others" (other fishes), Cluster 3 Centrolophidae and "others," Cluster 4 mainly "others," and Cluster 5 Priacanthidae, Trichiuridae and "others." Cluster 1 has the highest blackmouth catch composition, and the size of this cluster varied by year and was higher in 2011–12 and 2015. The cluster numbers were treated as categorical variables in the standardization model.

Supplementary information on how the PCA results compare to the clustering is provided in the Supplementary Material.

For the PCA method, the scree plot shows that more than 70% of the total variance can be explained by the first four principal components. Figure 5 displays the weights of different species in the first four principal components. PC1 is dominated by "others" fishes, PC2 by Carangidae species, PC3 by Loliginidae species, and PC4 by Trichiuridae and Sciaenidae species. The correlation biplots expressing the loadings of the species composition on the PCs (Figure 6) suggested that, for example, PC2 has a strong positive correlation with both Carangidae and Loliginidae but a weak negative relationship with Sciaenidae. PC2 is the variable with the strongest (negative) effect; higher scores on PC2 indicate a considerable increase in the catch composition of Carangidae and Loliginidae but a relatively lower composition of Sciaenidae. Furthermore, PC4, with a slightly lower effect than PC2, demonstrates a strong negative correlation with the composition of Sciaenidae and a strong positive correlation with Trichiuridae. Higher scores on PC4 signify a substantial increase in the composition of Trichiuridae and a considerable decrease in the composition of Sciaenidae.

## 3.2.2 Spatial effect

The Spatial Autocorrelation (Global Moran's I test) tool was used to assess spatial autocorrelation by considering both the locations and response variables of features simultaneously. It examines whether the observed pattern of the features is clustered, dispersed, or random. The values of Moran's I were statistically significant (I = 0.0047, P < 0.001) in this study, and the spatial factor was included in GLMM and GAMM to test the EPs of spatial and non-spatial models that will be presented in the followings.

In VAST, both the spatial and spatiotemporal random effects for a grid cell are assumed to come from the closest knot in space.



#### FIGURE 5

The weights of contributing species for each principle component (PC) of the PCA on the market landing data from coastal trawlers fishing in the southwestern waters of Taiwan. The size of the points represents the magnitude of the weights, with larger points indicating higher weights.



This assumption simplifies the computational calculations involved in the model. All the spatiotemporal models successfully converged in the study, which was confirmed by the fact that the Hessian matrix was positively definite, and the maximum gradient component was smaller than 0.001. The spatiotemporal distribution of the predicted density of blackmouth showed a stable or slightly increasing pattern before 2016; thereafter, however, the density showed a continuous declining pattern, especially in the coastal areas of Taiwan and some offshore regions, with a rebound in 2018 mainly in the northern region of the coast (Figure 7, based on results of M15).

## 3.3 Model explanatory power evaluation

Two  $R^2$  (pseudo- $R^2$  and CV- $R^2$ ) of the 15 models (M1–M15) are shown in Tables 2 (for model sets S1–S3) and Table 3 (for S4). The environmental effect was evaluated through S1 comparison under the consideration of vessel factor as a fixed effect: the  $R^2$  between M2 (including environment factors) and M1 (without environment factors) are not much different. The same observation was noted for M3/M4 that used PCA for the target factor, compared to M1/M2 that used k-means. Another pair of tests (M5 and M12) using GAM instead of GLM showed that the model with environment factors (M12) has better EP (i.e., higher  $R^2$ ) than without (M5), although the differences were not substantial. The EP of M12 were much better than M2 and M4,

which might suggest that the nonlinear features of environment factors could be better explained by the nonlinear GAM model.

The effect of the method in addressing target factor was explored by comparing M1/M2 with M3/M4 and M8 with M9 when considering vessel factor as a fixed effect and by comparing M6 with M7 and M10 with M11 when considering vessel factor as a random effect (Table 2). All these comparisons suggested that the PCA method performed better than the *k*-means method.

The effect of the treatment of vessel factor was evaluated by comparing of M2 with M6, M4 with M7, M8 with M10, and M9 with M11 (Table 2). These comparisons suggested that models perform slightly better when the vessel factor is considered as a fixed effect instead of a random effect.

The spatial effect and EP of different statistical models were evaluated through comparisons with models of M4 and M12–14 that included environment factors, treated vessel factor as a fixed effect, and addressed target factor by PCA (Table 3). The results suggested that considering spatial random effect (delta-GLMM or delta-GAMM, M13–M14) resulted smaller  $R^2$  than compared to those without (delta-GLM and delta-GAM, M4 and M12), and the M12 using GAM type model (delta-GAM) without consideration of spatial effect has the second-best EP (highest  $R^2$ : pseudo- $R^2$  = 0.604 and CV- $R^2$  = 0.594).

The above comparisons using traditional GLM/GAM type statistical models showed that spatial effect might not be important in the standardization. However, when applying VAST and including spatial and spatiotemporal effects, the EP measures of M9 and M15 were



much higher than those of M13 and M14, which applied GLM/GAM (linear and nonlinear) approaches. It might imply that spatial effect was important but needed a better approach to handle. The nonlinear-featured VAST (M15) demonstrated to have the best EP among all model designs (pseudo- $R^2 = 0.613$  and CV- $R^2 = 0.601$ , Table 3).

model designs (pseudo- $R^2 = 0.613$  and  $CV-R^2 = 0.601$ , Table 3). inclu Table 4 was created for explanatory purposes on the significance of the factors. Table 4 indicates that all the main effects were year,

significantly different from zero at a significance level of  $\alpha = 0.05$  for M4 and M12, except for the chlorophyll factor that was not significant for the PCM sub-model of delta-GLM. Sea temperature and salinity at a depth of 109 m were excluded from PCM but were included and statistically significant in ZPM in the process of model selection. Figure 8 shows the estimates of fixed-effect factors of the year, quarter, and vessel from M12 (GAM), along with the smoothers

| Stat.model | Sub-<br>model | Envir.<br>factors | Target<br>factor | Vessel - Fixed effect |                           |                   | Vessel - Random effect |                                  |                   |
|------------|---------------|-------------------|------------------|-----------------------|---------------------------|-------------------|------------------------|----------------------------------|-------------------|
|            |               |                   |                  | Code                  | Pseudo-<br>R <sup>2</sup> | CV-R <sup>2</sup> | Code                   | Pseudo-<br><i>R</i> <sup>2</sup> | CV-R <sup>2</sup> |
| GLM        | ZPM           | No                | k-means          | M1                    | 0.395                     | 0.394 (0.036)     |                        |                                  |                   |
|            | РСМ           |                   |                  |                       |                           |                   |                        |                                  |                   |
| GLM        | ZPM           | Yes               | k-means          | M2                    | 0.393                     | 0.391 (0.036)     | M6                     | 0.391                            | 0.388 (0.035)     |
|            | РСМ           |                   |                  |                       |                           |                   |                        |                                  |                   |
| GLM        | ZPM           | No                | PCA              | M3                    | 0.566                     | 0.575 (0.001)     |                        |                                  |                   |
|            | PCM           |                   |                  |                       |                           |                   |                        |                                  |                   |
| GLM        | ZPM           | Yes               | PCA              | M4                    | 0.566                     | 0.574 (0.001)     | M7                     | 0.563                            | 0.571 (0.001)     |
|            | РСМ           |                   |                  |                       |                           |                   |                        |                                  |                   |
| GAM        | ZPM           | No                | PCA              | M5                    | 0.599                     | 0.591 (0.002)     |                        |                                  |                   |
|            | PCM           |                   |                  |                       |                           |                   |                        |                                  |                   |
| GAM        | ZPM           | Yes               | PCA              | M12                   | 0.604                     | 0.594 (0.002)     |                        |                                  |                   |
|            | PCM           |                   |                  |                       |                           |                   |                        |                                  |                   |
| VAST       |               | Yes               | k-means          | M8                    | 0.401                     | 0.395 (0.002)     | M10                    | 0.397                            | 0.391 (0.002)     |
|            |               | Yes               | PCA              | M9                    | 0.569                     | 0.557 (0.002)     | M11                    | 0.558                            | 0.556 (0.003)     |

TABLE 2 Summary statistics of CPUE standardization models M1-M12 fitted to blackmouth croaker data off southwestern Taiwan during 2011-2021.

This table evaluates the explanatory power of the models with/without environmental factors, and with different treatments of target factors (k-means or PCA) and vessel effects (fixed or random). Refer to each model code in Table 1 for complete model specifications.

The models with the highest R-squared values when using the GLM/GAM and VAST methods are shown in bold.

TABLE 3 Summary statistics of CPUE standardization models M4, M9, and M12–M15 fitted to blackmouth croaker data off southwestern Taiwan during 2011–2021.

| Stat.model | Sub-model | Without spatial random effect |                       |                   | With spatial random effect |                       |                   |  |
|------------|-----------|-------------------------------|-----------------------|-------------------|----------------------------|-----------------------|-------------------|--|
|            |           | Code                          | Pseudo-R <sup>2</sup> | CV-R <sup>2</sup> | Code                       | Pseudo-R <sup>2</sup> | CV-R <sup>2</sup> |  |
| GLM        | ZPM       | M4                            | 0.566                 | 0.574 (0.001)     | M13                        | 0.426                 | 0.438 (0.001)     |  |
|            | РСМ       |                               |                       |                   |                            |                       |                   |  |
| GAM        | ZPM       | M12                           | <u>0.604</u>          | 0.594 (0.002)     | M14                        | 0.444                 | 0.451 (0.004)     |  |
|            | РСМ       |                               |                       |                   |                            |                       |                   |  |
|            |           |                               |                       |                   | М9                         |                       |                   |  |
| VAST       |           |                               |                       |                   | (GLMM-linear)              | 0.569                 | 0.557 (0.002)     |  |
| VAST       |           |                               |                       |                   | M15<br>(GAMM-nonlinear)    | 0.613                 | 0.601 (0.004)     |  |

The models considered environmental factors, treated vessel factor as a fixed effect, and utilized a PCA-generated target factor. This table evaluates the explanatory power of models with or without spatial random effects and employing different statistical methods. Refer to each model code in Table 1 for complete model specifications.

The models with the highest and second-highest R-squared values when using the GLM/GAM and VAST methods are shown in bold and underlined text. These two models have the best and second-best explanatory power out of the 15 models that were tested.

for the PCA target factor and environment factor. The figure suggested that the highest CPUE occurred in the third and fourth quarters and that the variation among vessels is high. The effect of environmental factors seemed not very strong (Table 4) and without clear patterns (Figure 7). Most of the variation could be explained by the targeting effect, especially the PC1, PC2, and PC4 (Table 4): blackmouth CPUE has general positive (but nonlinear) correlation with PC1 (Figure 7) which was negatively correlated with "others" fishes (Figure 6) (i.e., higher PC1 value has lower "others" fishes composition and higher Sciaenidae composition), and was nonlinearly negative correlated with Sciaenidae (Figure 6).

# 3.4 Trends of standardized indices

The standardized relative CPUE indices are shown for all the 15 models in Figure 9A and a subset of specifically selected eight models in Figure 9B. The general trend shows a decrease to 2013, an increase in 2014 and 2015 that was coincident with a weak to very strong El Niño event (Golden Gate Weather Services, 2023), and a continuous decline since then (although some models, especially the models used VAST, showed a noticeable increase in 2018 and then declined again). In the last year of the analysis in 2021, the indices show an increase to various degrees for all models using GLM- and GAM-type approaches but a continuous decline for models using

TABLE 4 Analysis of variance table for the model M4 (linear delta-GLM) and M12 (nonlinear delta-GAM).

| Delta-GLM (M4) |        |          |         | Delta-GAM (M12) |            |        |          |        |        |
|----------------|--------|----------|---------|-----------------|------------|--------|----------|--------|--------|
|                | ZPM    | ZPM      |         | PCM             |            | ZPM    |          | РСМ    |        |
| Predictors     | Chi.sq | Pr(>Chi) | F       | Pr(>F)          | Predictors | Chi.sq | Pr(>Chi) | F      | Pr(>F) |
| year           | 3433.7 | <0.05    | 113.1   | <0.05           | year       | 845.4  | <0.05    | 59.3   | <0.05  |
| quarter        | 539.0  | <0.05    | 98.5    | <0.05           | quarter    | 177.4  | <0.05    | 76.4   | <0.05  |
| vessel ID      | 614.1  | <0.05    | 22.3    | <0.05           | vessel ID  | 398.5  | <0.05    | 22.9   | <0.05  |
| PC1            | 7416.5 | <0.05    | 1852.7  | <0.05           | s(PC1)     | 5710.4 | <0.05    | 230.4  | <0.05  |
| PC2            | 3113.9 | <0.05    | 4173.9  | <0.05           | s(PC2)     | 3640.4 | <0.05    | 343.9  | <0.05  |
| PC3            | 769.4  | <0.05    | 517.3   | <0.05           | s(PC3)     | 724.9  | <0.05    | 58.4   | <0.05  |
| PC4            | 6633.4 | <0.05    | 10110.4 | <0.05           | s(PC4)     | 4439.9 | <0.05    | 1396.8 | <0.05  |
| tem_109        | 39.0   | <0.05    | -       | -               | s(tem_109) | 82.2   | <0.05    | -      | -      |
| sal_109        | 42.8   | <0.05    | -       | -               | s(sal_109) | 45.9   | <0.05    | -      | -      |
| mlt            | 77.1   | <0.05    | 231.7   | <0.05           | s(mlt)     | 168.3  | <0.05    | 26.1   | <0.05  |
| sft            | 18.3   | <0.05    | 6.9     | <0.05           | s(sft)     | 153.8  | <0.05    | 7.9    | <0.05  |
| chl            | -      | -        | 1.7     | 0.192           | s(chl)     | -      | -        | 12.4   | <0.05  |

The covariates include year, quarter, vessel identification (vessel ID), target factors (four principal components from PCA, PC1–PC4), and environment factors: sea temperature at depth 109 meters (tem\_109), salinity at depth 109 meters (sal\_109), density ocean mixed layer thickness (mlt), seafloor potential temperature (sft), and chlorophyll (chl).



the VAST approach except for models M8 and M10 that used kmeans for targeting effect.

For the specifically selected subset of the CPUE series (Figure 9B): (1) M8 and M9 employ different approaches to handle the target factor, with M8 utilizing k-means and M9 employing PCA, resulting in significant differences in the observed trends. M9 has higher EP than M8. (2) M9 and M11 differ in the way of treating vessel factors, and they are almost identical. (3) M5 and M12 differ in whether the environmental factor is considered, and the trends were basically similar with some years of variations. (4) M12 and M14 differ in whether a spatial factor is considered, and the results were very similar. (5) M4 and M12 differ in the statistical model used (GLM or GAM), and the indices were very different. Finally, (6) M9, M11, and M15 all used VAST but differ in linear or nonlinear treatment of nonlinear factors (environment and PCs). The index of M15 differs

from the M9/M11 only in the beginning period and shows a slightly stronger decline in the last year. All these models exhibit different trends of indices from GLM/GAM type models. M12 and M15 utilized a nonlinear approach with the GAM and VAST methods to address nonlinear environmental factors, and exhibited higher EP compared to M4, M9, and M11, which employed methods featuring linear GLM and VAST.

# 4 Discussion

# **4.1 Nominal CPUE construction**

Logbooks play a crucial role in providing essential catch and effort data for calculating the index of stock abundance. However,



FIGURE 9

Relative CPUEs of all the 15 tested standardization models (A) and of the best two models (M15 and M12) with the nominal CPUE (B) for blackmouth croaker from southwestern Taiwan during 2011–2021. The dash lines are 95% confidence intervals for the M15.

numerous coastal and small-scale fisheries are exempted from logbook submissions. In such cases, particularly for highcommercially valued species, which are more susceptible to intense exploitation, landing data, including date information and vessel identity, serves as a viable alternative to represent the "catch". Additionally, auxiliary data such as VMS, CSRS, VDR, or the automatic identification system (AIS) can be employed to estimate fishing efforts through fishery-specific algorithms (Gerritsen and Lordan, 2011; Chang, 2014, Chang, 2016; Sonderblohm et al., 2014).

Although VMS data have been extensively utilized for scientific purposes, the associated installation and transmission costs are high, making them less feasible for small-scale coastal vessels with limited capital. CSRS has its application limitations, such as the special system requirements to identify fishing vessels from the port and the potential degradation of radar data quality by environmental effects (Chang, 2014).

AIS, introduced by the IMO to enhance maritime safety and prevent ship collisions, offers comparable data to VMS but with improved temporal resolution at a significantly lower cost and can be used to derive fishing efforts and inform fisheries management (Natale et al., 2015; James et al., 2018). This study conducted a preliminary investigation into the utilization of AIS from Taiwanese coastal fishing vessels for scientific purposes, revealing challenges such as the short data period (since the AIS was introduced to fishing vessels late), low data coverage (because installation and operation of AIS are not compulsive for small-scale coastal vessels), and difficulties in identifying fishing vessels (due to the absence of a cross-reference template for AIS databases and fisheries management systems). These challenges are likely prevalent in other coastal countries.

In contrast, the VDR system proves more manageable for the Taiwanese case. The IMO's Maritime Safety Committee mandated passenger ships and large vessels to carry a standard VDR for accident investigations since 2000, with a 2006 amendment introducing a requirement for a simplified VDR (IMO, 2006). Recognizing the potential for fraud in fuel cost subsidies, the FA of Taiwan mandated the installation of simplified VDR in 2010 (Chang, 2016). This cost-reduced VDR, which records date, time, GPS position, speed, and direction of the vessel at 3-min intervals, not only addresses fraudulent activities but also provides data for scientific use at minimal additional cost. The affordability of installation and transmission makes it feasible for sample vessels to carry VDRs for independent research purposes. This study demonstrates the feasibility of using VDR data to construct nominal CPUE and derive an abundance index for declining fish stocks.

## 4.2 Key factors and statistical models

Confounding factors inherent in the nominal CPUE data need to be carefully addressed before the CPUE can represent the abundance. This study designed 15 models to examine the effect of targeting, vessel, and environmental and spatial factors and explore the best model specification for the blackmouth stock off southwestern Taiwan.

## 4.2.1 Targeting effect

Target tactics of the fishers are one of the most important factors influencing the CPUE for a multispecies fishery (Okamura et al., 2018). Of the two commonly used methods to address the target factor (Campbell et al., 2017), this study showed that the PCA approach offered significantly higher EP than the *k*-means approach in terms of  $CV-R^2$  ( $CV-R^2 = 0.391$  vs. 0.575, refer to Section 3.3 and Table 2), regardless of how other factors were treated ( $CV-R^2 =$ 0.391 vs. 0.575, refer to Section 3.3 and Table 2). Moreover, noticeable differences in relative CPUE trends were observed between the two approaches (Figure 9B, M8 vs. M9). This observation aligns with the findings by Chang et al. (2019) during their CPUE standardization research on dolphinfish stock. Although this difference might be driven by the additional flexibility implied by the model structure using PCA, the PCA procedure has been highlighted for its effectiveness in mitigating the impact of targeting on multispecies CPUE (Winker et al., 2014). It achieves this by avoiding the need to determine an optimal number of clusters with rather artificial boundaries, as well as allowing the combinations of different proportions of targeting tactics to be modeled as a continuum of all possible combinations (Winker et al., 2013).

### 4.2.2 Vessel effect

Vessel factor (estimated using individual vessel ID) combined the effects of vessel features and the skill/experiences of the skipper, which have been demonstrated to be a major influential variable in CPUE studies (Bentley et al., 2012; Chang et al., 2019; Hsu et al., 2022). The vessel features may include the size of the vessel, which is usually associated with fishing power and capacity of the vessel and is an important characteristic of trawl vessels. For the Taiwanese case, the first digit of the vessel ID represents the size of the vessel and, thus, vessel size has been considered in the factor. The study showed that treating it as a fixed effect has slightly better EP than as a random effect (Table 2), but the difference is small, and the trends of relative CPUE are similar (Figure 9B, M9 vs. M11). This should be expected as there is more flexibility to fit a covariate to a fixed effect than a random effect, which is (by definition) constrained to be normally distributed with an estimated variance. The lack of difference in the indices may simply point out the fact that the vessel with effects that were shrunk (in the mixed framework) must not have represented a high number of records.

Vessel factor is statistically significant in model M12 (the second-best model in Table 4), although the *F* value is comparatively small. The coefficient of variation (*cv*) of the average blackmouth landings among the vessels was 60.2%, showing a high variation among vessels. Also, removing the vessel factor from M12 produced  $R^2 = 0.576$ , about 3% decline from the original M12, suggesting that the vessel factor was influential in the standardization.

## 4.2.3 Environmental effect

Environmental factors are supposed to have an impact on the CPUE of demersal species (Kempf et al., 2022). The three pairs of

model comparisons (Table 2) suggested that including environment factors performed slightly better than without when using GAM type model (M5 and M12), and on the contrary, performed equally or even slightly worse than without ones when using GLM type model (M1–M4). This suggested that the importance of environmental factors for this study was not high, although two or three of the studied factors showed statistical significance (Table 4). This was speculated owing to the small size, hence less variation, of the studied area (Figure 2). Both the temperature (cv =5.6%) and salinity (cv = 0.3%) factors were not significant just because of their small variation in this area.

## 4.2.4 Spatial effect

Spatial factor, the significance of which was tested using Moran's I, was included as independent and identically distributed random effect in the mixed model analyses (GLMM and GAMM) or directly addressed using the spatiotemporal VAST model as a spatially autocorrelated random effect. The comparison of model EP (Table 3) suggested that including spatial factor did not perform better than those without, and the trends were very similar (Figure 9B, M12 vs. M14). Fish density distributed separately in four regions (Figure 7), and each region featured with different target species (in clockwise): (1) coastal region off Taiwan in the southeastern, targeting Sciaenidae, Priacanthidae, Trichiuridae, Centrolophidae in the southern part of this region, and Sciaenidae, Loliginidae, and "others" in the northern part; (2) South China Sea Slope in the south targeting "others"; (3) Xiapeng Depression in the northwestern targeting Loliginidae; and (4) Chang-Yuen Ridge in the northeastern targeting "others" and some Loliginidae. This suggested that the spatial effect might have largely been explained by the targeting effect in GLM/GAM type models (e.g., high F values for PC1-PC4 in Table 4).

There might be an issue with confounding among spatial, targeting, and environmental effects in the data analysis. It is concerning if the confounding did occur, because the spatial effect, unlike the targeting effect, should not be standardized out of the index, as it reflects true abundance. If the model is unable to distinguish the spatial effect from the targeting effect, it raises questions about the reliability of the standardized index. However, addressing confounding between abundance and catchability covariates is a complex challenge in CPUE standardization.

The spatial area depicted in Figure 2 exhibits some data coverage gaps. Since there is no published study defining the full range of the blackmouth croaker's habitat and the southwestern region is regarded as a single stock (Hwang and Chen, 1984), this research encompassed the vessels' entire distribution area as its study area. The reasons behind these data gaps are unclear, and the representativeness of the predictions in these sparsely covered regions remains uncertain. However, a significant data gap in the central section of Figure 7 could be linked to the presence of the Penghu Archipelago and Taiwan Bank.

When delineating the study area for CPUE analysis, it is important to consider the peripheral regions of the fishery that are rarely fished. It is uncertain whether these areas are unfished due to low CPUE or other reasons. Including such regions in the modeling process will introduce additional uncertainty since estimating abundance in these areas will rely on limited data (Hoyle et al., 2024). One approach to mitigate this concern is to confine the spatial scope to a "core" fishery, comprising the regions where the majority of the catch is obtained (Campbell, 2015; Hoyle et al., 2024).

The significance of spatial factors' interpretation could vary depending on the statistical model employed. When incorporating spatial effects, the EP measures of M9 with VAST surpassed those of M13 (GLMM) and M14 (GAMM) (Table 3), implying that the additional flexibility provided by the Gaussian Markov Random Field in VAST allows for a better representation of spatial effects than a traditional random effect structure. Introducing a new model design (M15) that enables VAST to handle nonlinear features like M12 resulted in even higher EP compared to M12 without spatial effects. This suggests that spatial effects played an important role in standardization, with VAST showcasing their importance more effectively than GLM/GAM models, as indicated above.

Spatial effects are able to smooth the effect of multiple unaccounted drivers and can be modeled using geostatistics (INLA) and 2D smoothing splines of GAM (Paradinas et al., 2023). Modeling spatial effects in continuous space is one of the key appeals of the GAM approach for CPUE standardization (Table 2 of Hoyle et al., 2024). However, as indicated in Section 2.3.2.1, spatial factor was treated as a random effect for all analyses of this study. Exploring models where spatial factors are considered as a continuous 2-D surface may add complexity to the model investigation structure (Figure 3) but holds promise for future investigations.

#### 4.2.5 Statistical models

Three types of statistical models were tested in the study. The models used nonlinear-featured GAM (or GAMM) all have better EP over linear-featured GLM (GLMM) and VAST (GLMM-based) (Tables 2 and 3). This was because the targeting effect is the most influential factor in this study (Table 4), and this factor was better represented by the nonlinear predictors allowed by the GAM.

The advantages of VAST have been successfully demonstrated in many CPUE standardization studies on pelagic and demersal species with spatiotemporal considerations (Grüss et al., 2019; Xu et al., 2019; Hsu et al., 2022; Kawauchi et al., 2023). The advantages, however, did not manifest in this study if using the default linear feature of the model (M9), as the VAST has a limited ability to include nonlinear relationships beyond general quadratic ones compared to GAMs or boosted regression trees (Brodie et al., 2020). This study designed a nonlinear-featured VAST, expanding the approach to include a nonlinear feature (M15) by taking advantage of the inputs from GAM to deal with nonlinear-featured environmental factors and PCs. The result has demonstrated substantial improvement in its predictive EP and, thus, was suggested to be applied to situations with nonlinear covariates.

One key advantage of spatiotemporal models such as VAST is that they allow a straightforward computation of "predict-then-

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aggregate" indices of relative abundance (sensu Hoyle et al., 2024). Since VAST imputes fish density in unsampled locations informed by a distance-based covariance function, the index can be computed from the entire spatial domain at each time step. It uses the area-weighting approach to compute the index of relative abundance. Providing interpolation of fish density distribution across the years (Figure 7) can also facilitate the interpretation of the spatiotemporal change of fish abundance. In comparison, traditional approaches to derive the index based on the year-effect only do not account for spatiotemporal variation in fish density over the entire spatial domain. Since fishing locations usually change over time, the index computed from the traditional year-effect approach may not be representative of the whole population, and, consequently, be proportional to the abundance of the whole population (Maunder et al., 2020).

The main finding from the model comparison exercise revealed that the model configurations with the highest EP, whether related to vessels, targets, or spatial aspects, tended to be the most complex ones. This outcome is not surprising considering the multifaceted nature of the fisheries system, which is influenced by a wide range of factors such as biology, economics, and the environment. As such, it is imperative to consider the nonlinear relationships that arise from these complex drivers when modeling such diverse systems (Glaser et al., 2014).

## 4.3 Abundance index

This study utilized two  $R^2$  as EP metrics. However, it is crucial to emphasize that a higher  $R^2$  do not necessarily equate to a more accurate index of relative abundance. The index of relative abundance, a pivotal input for stock assessment models, directly reflects changes in population abundance over time (Francis, 2011). The primary aim of CPUE standardization is to mitigate the influence of all factors except fish density on CPUE, ensuring that the standardized index of relative abundance approximates population abundance (MaunderTan and Punt, 2004). While a higher  $R^2$  suggests a better fit and more accurate predictions for nominal CPUE data, they do not confirm whether the model accurately identifies and addresses other factors impacting catchability. Therefore, a careful consideration of the history of the fishery and the stock (e.g., in terms of fleet dynamics, possible changes in fishing gear configurations, economic incentives, shifts in species distribution) is required when selecting the best models to use for CPUE standardization.

The abundance indices derived from the best model (M15) and the second-best model (M12) both exhibited a general declining trend for the southwestern blackmouth stock. This marks the first abundance index developed for demersal fish species in the region. Limited academic studies on the stock status of fish resources in the region exist in the literature due to a scarcity of catch and effort data (Liao et al., 2019). Although abundance indices for other demersal fish species are unavailable, their landings indicate a similar declining trend, suggesting a broad degradation of fisheries resources in the region (based on fisheries statistical yearbooks, Fisheries Agency, 2022; Shao, 2023). Indirect methods, such as trophic models (Liu et al., 2009) or fishermen surveys (Liao et al., 2019), along with numerous gray literature sources, caution that overfishing has occurred to Taiwanese coastal blackmouth resources due to the overcapacity of fishing vessels and illegal fishing (Chang et al., 2010; Liao et al., 2019). Comparisons of growth parameters over a decade also imply high fishing pressures on blackmouth, resulting in elevated growth coefficients and smaller asymptotic lengths (Huang et al., 2022).

In addition to overfishing, environmental pollution is asserted to contribute to the degradation of the fishery environment in the region. Repeated pollution events in rivers flowing into the area have been documented for years without complete remediation (Liu et al., 2015; Wang et al., 2021; Liang et al., 2024). However, further studies are required to establish the impact of contaminations on fisheries resources.

# **5** Conclusions

Indices of relative abundance are one of the most essential data for stock assessments and consequent management advice. For datalimited small-scale coastal fisheries that are exempted from logbook submission, the abundance indices can be established through proper standardization on vessel-specific landing data and trip-based fishing effort estimated from geo-referenced position data.

This study utilized unique data sourcing from VDRs to estimate fishing effort, in combination with landing data to estimate the CPUE data for the blackmouth croaker in southwestern Taiwan (Figure 2), which demonstrated the potential application in global data-limited fisheries. The study also assessed alternative approaches for predictors of fishery targeting practices to condition effort for producing more accurate metrics of relative abundance. Fifteen models were designed for CPUE standardization, employing three statistical approaches. Model comparisons were conducted to scrutinize the impacts and treatments of targeting, vessel, environmental, and spatial factors. The results indicated that, in this multispecies fishery, the target factor explained the most variability in the observed catch rates, recommending the utilization of the PCA approach to address the target effect. It was advised to consider the vessel factor in the standardization process, either as a fixed effect with consistent specifications of sample vessels over years, or as a random effect following recommendations from Bentley et al. (2012) and Hsu et al. (2022). The inclusion of environmental factors did not notably enhance model EP, potentially due to the limited fishing area in this study. However, for fisheries with a broader spatial range, this factor should still be thoroughly examined. Because of the nonlinear feature of the PCA approach and environment data, GAM performed better than the linear-based models. However, with the expansion of VAST to include nonlinear features, the model surpassed GLM/GAM-type models in handling spatial random effects.

The best model (Model M15) exhibited a consistent declining trend in CPUE since 2015, with a brief rebound in 2018 (Figure 9B). The second-best model (M12) displayed a similarly sharp decline since 2015 but diverged in 2021. Despite this variation, the overall declining trend across all models signals the potential decline of the blackmouth stock, prompting the consideration of further management measures.

# Data availability statement

The raw data supporting the conclusions of this article contains vessel's identity and their confidential data of fishing locations and catches. It will be made available by the authors to qualified researchers under conditions. Requests to access these datasets should be directed to SC, skchang@faculty.nsysu.edu.tw.

# Author contributions

TY: Formal analysis, Methodology, Software, Visualization, Writing – original draft. HX: Writing – review & editing. BL: Data curation, Software, Writing – review & editing. SC: Conceptualization, Funding acquisition, Resources, Supervision, Writing – original draft, Writing – review & editing.

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# **Conflict of interest**

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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## Supplementary material

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fmars.2024.1344181/ full#supplementary-material

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