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# Spatial distribution of the demersal chondrichthyan community from the western Mediterranean trawl bycatch

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**Introduction:** Chondrichthyans (sharks, batoids and chimaeras) play key roles in the regulation of marine food webs dynamics. However, more than half of the assessed species in the Mediterranean are threatened, primarily by fishing pressure and compounded by habitat degradation and climate change. Nevertheless, there is an important knowledge gap in identifying the underlying drivers of their community structure and spatial distribution.

**Methods and Results:** We provide insights into the current bycatch rates of chondrichthyans in the western Mediterranean commercial bottom trawling fishery by accurately depicting the unaltered practices of the local fleet. A total of 17 species were recorded in the studied fishing grounds (ranging from 50 to 800 m deep), including 7 sharks, 9 batoids, and 1 chimaera, although the total catch was dominated by few species. Furthermore, we tested the effect of environmental and fishing-related factors on multiple community descriptors by using analysis of community structure (multidimensional scaling and analysis of similitude) and generalized linear mixed models to further understand the drivers of the chondrichthyan community distribution and structure. This study revealed the importance of combining environmental and anthropogenic drivers to further understand the spatial distribution patterns of chondrichthyan communities. Our results indicate that depth was the main driver of the community structure, with deeper areas within the fishing grounds hosting a higher diversity. Sea bottom temperature and substrate type also influenced the distribution of the community, with substrate effects changing depending on the intensity of fishing pressure. In all cases, density, biomass, and diversity of chondrichthyans were negatively impacted by increasing fishing effort.

**Discussion:** Understanding the drivers of the structure and distribution of the chondrichthyan community is crucial to understand the potential impacts that increased fishing pressure, habitat loss and global change may entail. The ongoing challenges that the Mediterranean chondrichthyans and their ecosystems are facing highlights the need for continued monitoring and improved chondrichthyan-focused fisheries management.

#### KEYWORDS

environmental drivers, fishing pressure, commercial bottom trawling, diversity, elasmobranchs, sharks, rays, modelling

## Introduction

Chondrichthyans (sharks, batoids and chimaeras) play a key role in the regulation of marine ecosystems by controlling the size and dynamics of their prey populations and connecting trophic webs across habitats (Stevens et al., 2000; Sheaves, 2009; Barriá et al., 2015a). Their life history traits (slow growth, late maturity, low fecundity, and longevity) make them particularly vulnerable to the anthropogenic pressure which is primarily propitiated by fisheries (for both target and non-target species), followed by habitat loss and degradation and to a lesser extent climate change and pollution (Cortés, 2000; Dulvy et al., 2021). Given their important ecological role, the depletion of chondrichthyans induces changes in the structure and functioning of marine ecosystems (Stevens et al., 2000), ultimately leading to a loss of biodiversity and productivity (Heithaus et al., 2008). Considered as keystone species, some chondrichthyans can be used as indicators to monitor overfishing and ecosystem stability (Stevens et al., 2000; Heithaus et al., 2008).

The global situation for chondrichthyans is highly concerning, constituting the most threatened group of marine animals, with more than one third of the species estimated to be threatened (Dulvy et al., 2021). This trend is even exacerbated in the Mediterranean, with 53% out of the 73 assessed species classified as threatened (Dulvy et al., 2016). Fishing activities are the main cause for the decline of the Mediterranean chondrichthyan community (Follesa et al., 2019; Serena et al., 2020). Chondrichthyans may constitute an important proportion of the catch for multi-specific fisheries, but only a minor part is reported on landings as they are frequently captured as non-target species (bycatch) and discarded at sea (Ferretti et al., 2008; Peristeraki et al., 2020). Despite efforts by the General Fisheries Commission for the Mediterranean (GFCM) to reduce the capture of sharks by prohibiting onboard finning and the sale of the chondrichthyan species listed in the Annex II of the Barcelona Convention (FAO, 2020), the populations of many species are still poorly managed (Walls and Dulvy, 2021; Koehler et al., 2022). Moreover, reliable data on the stock status of chondrichthyans is scarce in many regions.

Monitoring programs on the demersal Mediterranean marine resources started at the beginning of the 20th century describing the

regional chondrichthyan community (De Buen, 1935; Maurin, 1968; Aldebert, 1997). In the early 1990s international programs based in fisheries-independent data increased importantly their capacity to monitor the Mediterranean demersal resources. Part of the published research covered large areas of the Mediterranean (e.g. Bertrand et al., 2000; Follesa et al., 2019; Serena et al., 2020), but most occurred at regional scales along the European coasts (e.g. Guijarro et al., 2012; Peristeraki et al., 2020; Geraci et al., 2021), while remaining particularly scarce in the African region (e.g. Ordines et al., 2011; Colloca et al., 2017; Serena et al., 2020). These studies described the regional chondrichthyan communities and their spatial distribution (e.g. Tserpes et al., 2013; Lauria et al., 2015; Porcu et al., 2020). They also elucidated temporal trends for some species (e.g. Damalas et al., 2010; Damalas and Vassilopoulou, 2011; Maravelias et al., 2012), documenting the drastic community composition changes suffered since the early fishing exploitations (e.g. Jukic-Peladic et al., 2001; Ferretti et al., 2013; Barausse et al., 2014). Starting from the 1950s, the north-western Mediterranean trawling grounds were populated with some of the currently most threatened species (Maurin, 1968; Aldebert, 1997), and although after the 2000s certain stability was reached in terms of diversity (Ramírez-Amaro et al., 2020) the abundance of an important part of the species continues to decline (Ligas et al., 2013; Marongiu et al., 2017; Ramírez-Amaro et al., 2020; Walls and Dulvy, 2020; Walls and Dulvy, 2021).

Recently, with the use of spatially explicit modelling approaches, the research efforts have been focused on the effects of the changing environments on the structure and dynamics of these communities (e.g. Follesa et al., 2019; Maioli et al., 2022), and only rarely they have integrated the anthropogenic pressures in the modelling approaches as predictors in shaping the composition, abundance, and distribution of the demersal chondrichthyans in the Mediterranean (Navarro et al., 2016; Peristeraki et al., 2020). From an ecological perspective, it is important to understand how species and the community respond to certain factors. However, from a management standpoint, it may be more useful to quantify how multiple factors interact and combine in space to shape distribution patterns (Melo-Merino et al., 2020). Assessing the combined effects of both environmental and fishing-related drivers on the chondrichthyan community might enhance not only the capacity

to determine important areas and conditions for conservation, but also the ability to make predictions where data is scarce or absent (Ovaskainen et al., 2017). This is especially important to build capacity in poorly researched areas, determine thresholds for sustainable practices, and understand the potential effects of increasing fishing pressure, global change and habitat loss (Melo-Merino et al., 2020).

In this study, we updated information on the status of the demersal chondrichthyan community in the Mediterranean trawling grounds from the Northern Spain GFCM geographic subarea (GSA06). First, we collected data from commercial bottom trawling surveys to provide accurate chondrichthyan bycatch rates during regular commercial practices. Then, we assessed the influences of environmental and fishing-related drivers on the structure and distribution of the chondrichthyan community, aiming to improve our understanding of the status of the western Mediterranean chondrichthyan community for a more efficient management.

## Materials and methods

### Study site and fishery

This work covered diverse bottom trawling grounds across the Mediterranean waters of the eastern Iberian Peninsula, i.e. the “Northern Spain” geographical subarea (GSA06) established by the GFCM (Figure 1), occurring in the soft bottoms of the

continental margin between 50 and 800 m deep. The Spanish trawl fleet operating in the GSA06 is censused at 32 home ports ranging from Cartagena to Roses. As in December 2022, the region counts with a fleet of 421 licensed bottom trawlers, although the fleet continues to decrease progressively during the last decades. Surveys were conducted onboard of 20 of these commercial fishing vessels (F/V) from 14 home ports ranging from Cartagena to Roses to cover most of the GSA06 (Figure 1). The surveys were conducted through onboard observation campaigns performed without altering the conventional fishing practices of the F/V nor the location of the conducted hauls, therefore, being these randomly selected. The surveyed hauls were tracked geographically using GPS and bathymetrically using the SCANMAR system (Catch Control Systems, Scanmar AS, Åsgårdstrand, Norway). The sampling protocol accounted with a proportional number of hauls analyzed for the three different métiers (i.e. group of fishing operations targeting a specific assemblage of species) that occur in the region: (1) the “hake métier” targets *Merluccius merluccius*, *Mullus* spp., and *Parapenaeus longirostris*; (2) the “Norway lobster métier” targets *Nephrops norvegicus* with main accessory species being *Micromesistius poutassou*, *Phycis blennoides* and *P. longirostris*; and (3) the “blue and red shrimp métier” targets *Aristeus antennatus*, with the main accessory species being *Geryon longipes* and *Plesionika* spp. Fishing métiers slightly overlap geographically and bathymetrically but the hake métier mainly occurs in the continental shelf (approximately 50–250 m deep), the Norway lobster métier fishes mainly in the upper continental slope

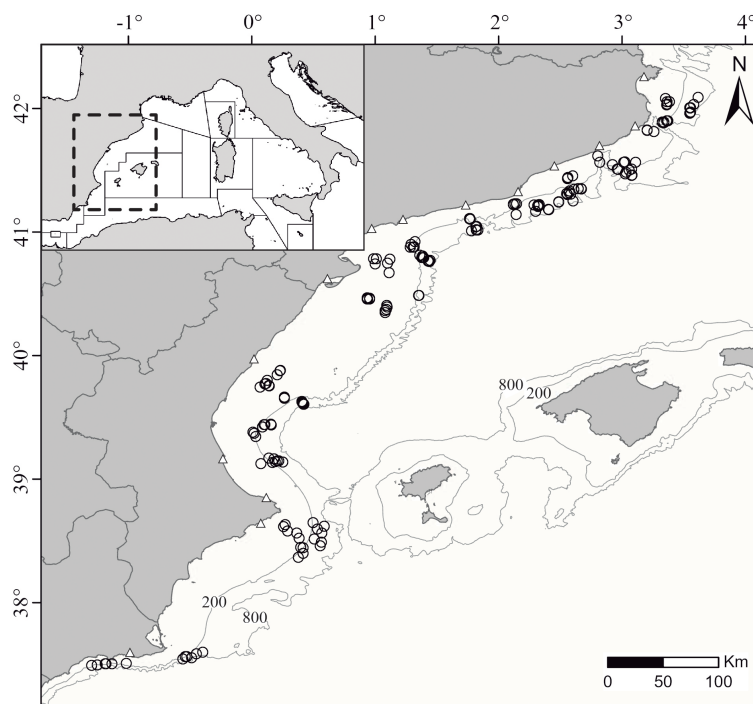


FIGURE 1

Geographical position of the study area within the western Mediterranean including the respective existent Geographical Sub-Areas (GSAs). The points indicate the position of the 189 fishing hauls analyzed in this study occurring across the GSA06 (Northern Spain) at depths ranging from 50 to 800 m. The triangles denote the location of the home ports included in the study and the isobaths denote 200 and 800 m deep.

(approximately 200-600 m deep) and the blue and red shrimp métier mostly operates in the lower continental slope (approximately 400-800 m deep). The conducted sampling aimed to accurately represent the normal commercial practices of the local fleet while using a standardized protocol.

## Experimental procedures

The data set consisted of 189 hauls conducted by 20 different commercial bottom trawlers (F/V) between June 2020 and January 2022. A consistent stratified-random sampling design was used, as haul locations were randomly selected but a proportional division among the number of analyzed hauls was considered for each port, season, depth range and métier. Data was collected on a seasonal basis at each of the 14 studied ports, accounting with hauls from each of the three previously described métiers and occurring proportionally across the continental margin (continental shelf, upper slope, and lower slope) with minimum and maximum depths ranging from 50 to 737 m. The number of hauls per season and depth range is provided in Table 1. These surveys used F/V of similar capability and fishing gear per each métier to minimize any effects of gear selectivity, although some variation occurred. All fishing gears consisted of a bottom otter trawl design, with the horizontal opening of the net averaging  $24.1 \pm 7.0$  m and fitted with a squared 40 mm cod-end mesh, except in Palamós, where the co-management measures for the blue and red shrimp métier require a 50 mm squared mesh. Each haul was conducted without altering the conventional commercial practices of the F/V. The mean distance covered in each haul averaged  $18.0 \pm 8.7$  km, the effective haul duration was  $3.3 \pm 1.4$  h, and the average towing speed was  $2.9 \pm 0.4$  knots. The depth range covered in each haul varied in average  $77.5 \pm 59.7$  m, but the variation was larger in the upper and lower slope ( $111.3 \pm 68.7$  m,  $107.1 \pm 52.4$  m, respectively) than in the shelf ( $35.2 \pm 25.7$  m). The raising and lowering of net from and to the sea bottom, as well as the horizontal and vertical openings of the net were estimated using the SCANMAR system. For each haul every chondrichthyan captured was identified and classified and all of them or a representative random sample was counted and measured to the nearest millimeter using total length (TL; tip of the snout to posterior end of caudal fin) for sharks and batoids and pre-supracaudal length (PSCL; tip of the snout to the posterior edge of the supracaudal fin) for chimaeras.

## Community descriptors

We calculated five descriptive metrics of the chondrichthyan community: (1) density, (2) biomass, (3) diversity, (4) evenness, and (5) richness. Capture per unit of effort (CPUE) was calculated as both density and biomass measures, defined as the number of individuals and kilograms of chondrichthyans, respectively, caught per square kilometer pooling all species together (i.e. inds/km<sup>2</sup> and kg/km<sup>2</sup>, respectively). We estimated the individual weight by using species-specific length-weight relationships presented in previous studies from the study area (Barría et al., 2015b). Species diversity, richness and evenness were calculated using the Shannon index ( $H'$ ), Margalef's index ( $d$ ), and Pielou's index ( $J$ ), respectively, accounting not only for the number of species but also with their relative proportion in the catch. Community assemblages were further described using multidimensional scaling (MDS) on the standardized density dataset. Principal Coordinate Analysis (PCoA) was performed on a Bray-Curtis dissimilarity matrix to summarize and visualize patterns in the species assemblages based on their relative density. To meet the assumptions of PCoA, we tested for multivariate normality and applied logarithmic transformations to the data. Analyses were conducted using the package *vegan* (Oksanen et al., 2022).

## Predictor variables

The community descriptor metrics and the principal coordinates (PCoA) axis that accounted for the largest variance in the data were related to different explanatory variables *a priori* selected including both environmental and fishing-related predictors. Environmental predictors included depth, slope angle, seabed substrate type, and sea bottom temperature (Figure 2). The predictors selected have previously been associated with species diversity and community structure (e.g. Lucifora et al., 2012; Pennino et al., 2013; Lauria et al., 2015). The mean depth of each haul was calculated using the SCANMAR system incorporated in the F/V and was then supported using EMODnet Bathymetry, resulting in a continuous variable ranging from 54 to 685 m. The slope angle was calculated using the raster R package (Hijmans, 2022) from the EMODnet Bathymetry. Substrate data was extracted from the European Marine Observation Data Network (EMODnet) Seabed Habitats project

TABLE 1 Number of hauls analyzed onboard of commercial bottom trawlers per season and depth strata (shelf, 50-200 m deep; upper slope, 200-500 m deep; and lower slope, 500-800 m deep).

Depth strata	Autumn	Winter	Spring	Summer	Total
Shelf	20 (70)	18 (89)	22 (91)	20 (75)	80 (81)
Upper slope	13 (100)	12 (100)	17 (100)	14 (100)	56 (100)
Lower slope	13 (100)	11 (100)	14 (100)	15 (93)	53 (98)
Total	46 (87)	41 (95)	53 (96)	49 (88)	189 (92)

Data shows in brackets the percentage (%) of hauls in which chondrichthyans were present.

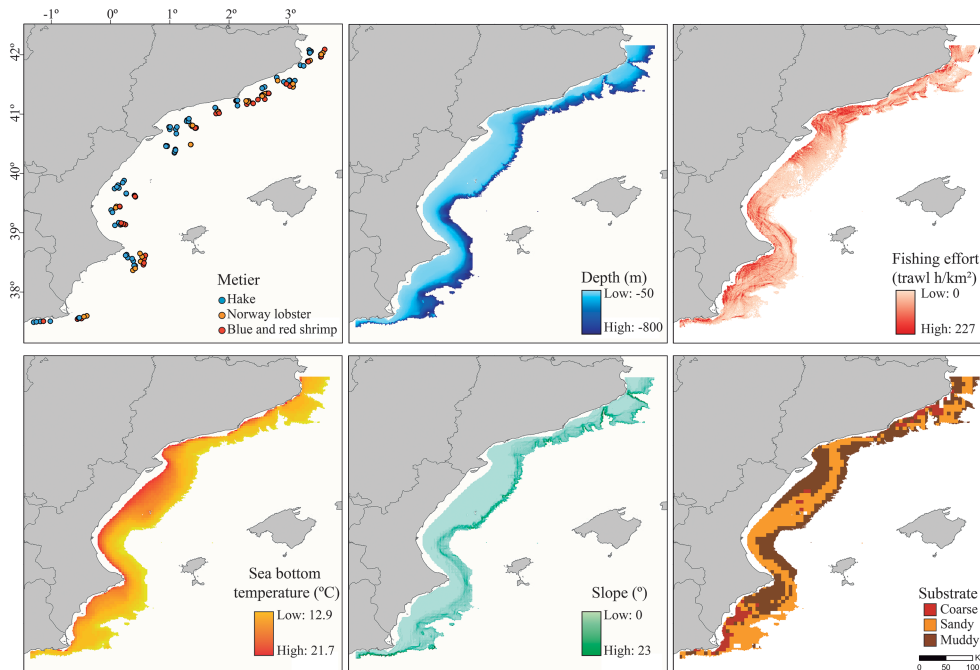


FIGURE 2

Environmental and fishing-related drivers used as predictors to model the distribution and structure of the chondrichthyan community, including: (A) métier, (B) depth, (C) fishing effort, (D) sea bottom temperature, (E) slope, and (F) seabed substrate. The values for each variable are spanned within the study region (GSA06) from 50 to 800 m deep.

(emodnet.ec.europa.eu/en/seabed-habitats) and categorized as: (1) coarse, (2) sandy, and (3) muddy. Sea bottom temperature (SBT) data was obtained from the Copernicus Marine Environment Monitoring Service (CMEMS) dataset of Mediterranean Sea Physics Reanalysis maps on a  $1/24^\circ$  resolution grid spanning the complete Mediterranean. We extracted monthly sea bottom temperature estimates for the study area from 50 to 800 m deep and we matched the values with each trawl haul, yielding values ranging from 12.9 to 16.6 °C.

The fishing-related variables included métier and fishing effort (Figure 2). The métier of each haul was classified according to the three categories previously described. Apparent fishing effort was extracted from Global Fishing Watch (globalfishingwatch.org), which uses automatic identification system (AIS) data to identify and track fishing vessels. The apparent effort was calculated as the number of hours trawled within each squared kilometer of the study area for the 2019–2022 period. The time period selected was intended to represent not only the effort conducted during the survey period but the pressure that has been exerted during recent years. This allowed us to highlight areas of high fishing effort, with a maximum effort of 226.8 h/km<sup>2</sup> and calculate the average fishing effort in the location where each haul was conducted, ranging from 5.5 to 183.8 h/km<sup>2</sup>. Fishing effort was higher across the continental shelf (< 200 m deep) and, especially, in areas located near large ports and cities, likely influenced by the market demands, whereas the deepwater fishing grounds (> 200 m deep) were concentrated in certain regions leaving others unfished. Multiple overlapping ring buffers of 5 km<sup>2</sup> were used to extract mean values of continuous variables matching to haul locations.

## Statistical analyses

We used generalized linear mixed effect models (GLMMs) to analyze the effect of environmental and fishing-related variables on each community descriptor metric and the first principal coordinate (PCoA) axis describing the composition of the community assemblages. The fixed effects included all the previously described predictors (depth, slope, substrate type, sea bottom temperature, fishing effort, and métier). We considered a quadratic form of depth to account for possible nonlinear effects of this predictor variable. We included an interaction term between depth and sea bottom temperature, as the later changes progressively with depth and may affect the response variable differently at different depths. We also included an interaction term between fishing effort and substrate to account for indirect effects of trawling in relation to different substrates (Collie et al., 2017). In addition, we considered F/V (20 levels) as independent random effects to account for the potential heterogeneity between F/V (e.g. skipper behavior and experience).

We created a full model for each response variable including all fixed and random effects (Bolker et al., 2009). We fitted the full models using the package lme4 (Bates et al., 2015) and a Gaussian error distribution and estimated the parameters using restricted maximum likelihood (REML). We standardized the continuous explanatory variables (i.e. transforming them to yield a mean of zero and standard deviation of one) to reduce collinearity effects among variables. We checked for multicollinearity using the variance inflation factor (VIF) and all predictor variables were used for model selection as they were uncorrelated (i.e. VIF <

3.5). Homoscedasticity was examined through the visual evaluation of residual plots across categories and continuous predictors. We tested for normality by visually inspecting residual histograms, normal Q-Q plots, and the Shapiro–Wilk test. Where normality and homoscedasticity assumptions were violated, logarithmic transformations were applied. We first estimated the optimal structure of the random component using the package *lmerTest* (Kuznetsova et al., 2017). When the random component was not significant, generalized linear models were fitted instead. Both backwards and forward stepwise approaches based on the Akaike Information Criterion (AIC) were used to select the best model fixed structure (Zuur et al., 2009). The linear term was forced into the model each time the quadratic term was selected in the model (Zurell et al., 2009). Once the optimal model was achieved, we rechecked the assumptions of homoscedasticity and normality. In addition, we examined the presence of spatial autocorrelation (SAC; Dormann, 2007) in the residuals based on Moran's I test; no SAC was detected in any of the models.

The center of gravity (CoG) was calculated to better understand the influence of bathymetry on the chondrichthyan community structure by determining the average depth at which each species was most commonly found (Stefanescu et al., 1993). In addition, the average density and biomass of each species was calculated for each feature of the continental margin, defined as the continental shelf (50–200 m), upper slope (200–500 m), and lower slope (500–800 m). Finally, analyses of similarities (ANOSIM) and similarity

percentages (SIMPER) were conducted on the density Bray-Curtis matrix to determine whether the species assemblages were significantly different among each of the features of the continental margin and to further describe each of them.

## Results

### Collected data

A total of 189 hauls were examined, covering an area of 53.1 km<sup>2</sup> proportionally distributed across the examined geographic and bathymetric range in the GSA06. Chondrichthyan specimens were found in 91.5% of these hauls (173 out of 189), with a total of 11,997 specimens captured. No chondrichthyan specimens were recorded on the remaining 16 hauls (Table 1). Chondrichthyan specimens were present all along the geographical and bathymetrical range.

### Community description

A total of 17 chondrichthyan species were caught in the studied area during the surveyed period, including 7 sharks (belonging to 3 orders and 7 families), 9 batoids (belonging to 3 orders and 4 families) and 1 chimaera species (Table 2). The most frequently caught species across the study range were *Scyliorhinus canicula*,

TABLE 2 List of the chondrichthyan species reported in the surveyed hauls, including their IUCN red list status categories in the Mediterranean according to IUCN (2022), preferred depth ranges and overall bycatch rates.

Order	Family	Species	IUCN status	D (m)	COG (m)	F (%)	Density (inds/km <sup>2</sup> )	Biomass (kg/km <sup>2</sup> )
Carcharhiniformes	Scyliorhinidae	<i>Scyliorhinus canicula</i>	LC	62–649	241 ± 150	86.7	239.7 ± 480.9	39.9 ± 57.2
	Pentanchidae	<i>Galeus melastomus</i>	LC	130–685	535 ± 114	50.3	87.6 ± 194.3	13.3 ± 22.6
Hexanchiformes	Hexanchidae	<i>Hexanchus griseus</i>	LC	534–537	535 ± 2	1.2	0.02 ± 0.2	1.5 ± 20.4
Squaliformes	Centrophoridae	<i>Centrophorus uyato</i>	VU	432–588	446 ± 110	1.2	0.01 ± 1.3	0.1 ± 1.3
	Dalatiidae	<i>Dalatias licha</i>	VU	475–673	591 ± 51	2.9	0.1 ± 0.7	0.02 ± 0.2
	Etmopteridae	<i>Etmopterus spinax</i>	LC	207–685	557 ± 105	26.6	7.8 ± 19.5	0.6 ± 1.1
	Oxynotidae	<i>Oxynotus centrina</i>	CR	95–214	113 ± 84	1.2	0.07 ± 0.7	0.3 ± 3.5
Myliobatiformes	Myliobatidae	<i>Aetomylaeus bovinus</i>	CR	73	73	0.6	0.02 ± 0.3	0.2 ± 2.8
		<i>Myliobatis aquila</i>	VU	73	73	0.6	0.01 ± 0.2	0.03 ± 0.4
	Dasyatidae	<i>Dasyatis pastinaca</i>	VU	133	133	0.6	0.01 ± 0.09	0.07 ± 0.9
		<i>Pteroplatytrygon violacea</i>	LC	62–78	68 ± 11	1.2	0.06 ± 0.5	0.4 ± 3.9
Rajiformes	Rajidae	<i>Dipturus oxyrinchus</i>	NT	329	329	0.6	0.02 ± 0.3	0.004 ± 0.05
		<i>Raja asterias</i>	NT	57–124	66 ± 48	1.2	0.06 ± 0.6	0.02 ± 0.1
		<i>Raja clavata</i>	NT	88–329	179 ± 69	4.1	3.2 ± 37.0	1.8 ± 20.0
		<i>Raja polystigma</i>	LC	63–181	117 ± 37	4.6	0.3 ± 1.9	0.1 ± 0.6
Torpediniformes	Torpedinidae	<i>Torpedo marmorata</i>	LC	62–257	104 ± 48	17.3	3.3 ± 11.7	0.7 ± 2.5
Chimaeriformes	Chimaeridae	<i>Chimaera monstrosa</i>	NT	410–595	42 ± 69	1.7	0.1 ± 1.1	0.003 ± 0.03

Centre of gravity (COG), density and biomass values are given as mean ± SD and abbreviations stand for: D, depth range of appearance and F, frequency of occurrence.

*Galeus melastomus*, *Etmopterus spinax* and *Torpedo marmorata*, which also had the highest levels of abundance. Threatened species were also present in the area with low frequency and abundance, including batoids such as *Aetomylaeus bovinus*, *Myliobatis aquila* and *Dasyatis pastinaca*, found in the continental shelf, and sharks such as *Oxynotus centrina*, *Centrophorus uyato* and *Dalatis licha* mostly occurring in the continental slope. Table 2 summarizes the depth range and center of gravity at which each species was caught, as well as their bycatch rates in terms of frequency of capture and CPUE for the overall studied range, measured as mean density and mean biomass. Additional information on bycatch rates in terms of frequency, density, and biomass of each species at each section of the continental margin (shelf, upper slope and lower slope) is provided in the supplementary material (Supplementary Table S1).

## Environmental and fishing-related drivers of the structure and distribution of the community

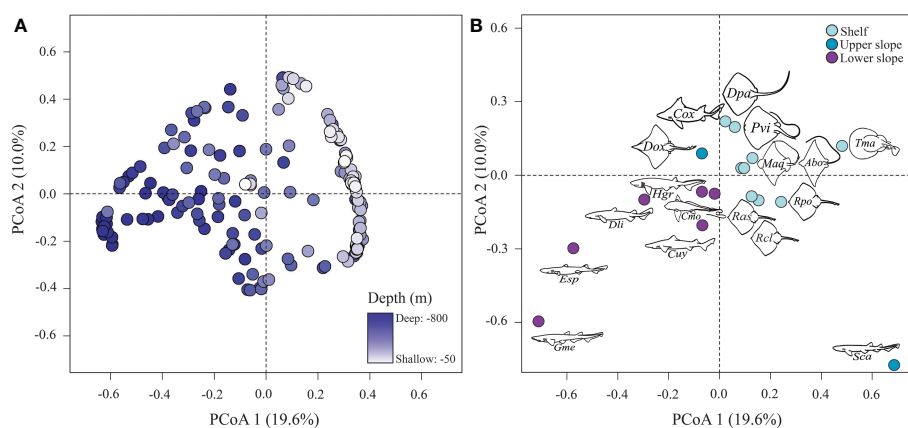
The principal coordinates analysis (PCoA) plot based on the relative density of each species revealed dissimilarities in the composition (Figure 3). The first unconstrained index of compositional dissimilarity (PCoA axis 1) explained almost 20% of the total variation in the relative density of the species, grouping them in relation to depth (Table 3; Figure 4). The species assemblage occurring in each of the regions of the continental margin was significantly different (ANOSIM,  $p < 0.001$ ;  $R = 0.45$ ). According to the SIMPER analysis, *G. melastomus* and *S. canicula* were the main species contributing to this difference, followed by *E. spinax*, *T. marmorata*, and *D. licha* (Table 4). The continental shelf was characterized by high densities of *S. canicula*, followed by *T. marmorata* and the skates *Raja polystigma* and *Raja clavata*, occurring in >90%, >40, >10 and >5% of the hauls, respectively (Table 4; Supplementary Table S1). Despite their low overall abundance, batoids made up the largest part of the species

diversity in the continental shelf fishing grounds, with 8 out of 9 species preferring this bathymetric range. Sharks, on the other hand, constitute the most diverse group in the continental slope. The upper slope is characterized by high abundance levels of *S. canicula*, followed by *G. melastomus* and *E. spinax*, with a frequency of appearance of >90%, >60%, and >20%, respectively. This bathymetric range was also characterized by the occurrence of low densities of *C. uyato*, *Dipturus oxyrinchus*, and *Chimaera monstrosa*, which had their centres of gravity in this range (Table 4; Supplementary Table S1). The lower slope was dominated by *G. melastomus* with a frequency of appearance of >90%, followed by *S. canicula* and *E. spinax*, both appearing in >60% of the hauls conducted in this range. *Hexanchus griseus* and *D. licha* also occurred predominantly in this range in low abundance levels (Table 4; Supplementary Table S1).

The overall density of chondrichthyan specimens, including all trawls regardless of whether they captured any chondrichthyans, was  $252 \pm 298$  inds/km<sup>2</sup>. Density values ranged from 0 to 1498 inds/km<sup>2</sup>. The selected model for chondrichthyan density included the interaction between fishing effort and substrate. The effect of fishing effort on chondrichthyan density varied depending on the substrate, with a stronger negative effect of fishing pressure on sandy and coarse sediments compared to muddy sediments. Larger densities at low fishing effort occurred in sandy and coarse sediments but higher overall mean density occurred in muddy bottoms as density decreased more rapidly in sandy and coarse sediments with increasing fishing effort (Table 3; Figure 4).

The biomass of chondrichthyans averaged  $44 \pm 52$  kg/km<sup>2</sup> and ranged from 0 to 323 kg/km<sup>2</sup>. The selected model for chondrichthyan biomass included the depth and the interaction between fishing effort and substrate. The impact of fishing pressure on chondrichthyan biomass varied depending on the substrate, analogously as for density. In addition, biomass increased significantly with depth (Table 3; Figure 4).

Mean diversity using Shannon index was  $0.37 \pm 0.32$ , ranging from 0 to 1.1. The selected model for chondrichthyan diversity



**FIGURE 3**  
**(A)** Principal coordinate analysis, where the locations of points reflect compositional differences among hauls based on species-specific density data. The color of the points represents the mean depth of the haul. **(B)** The same ordination showing scores for each species driving differences in the composition of the community. Each point represent an species and is accompanied by its illustration and initials (the first letter for genus and the two other for species). Point colors represents the preferred depth range of the particular species according to COG analysis (Table 2).

TABLE 3 Results of the generalized mixed effect models (GLMM) of the five community descriptor metrics and the first principal coordinate (PCoA) axis.

Parameter	Density <sup>a</sup>	Biomass <sup>a</sup>	Diversity <sup>a</sup>	Richness	Evenness	PCoA 1
Intercept	3.81 (0.39)***	2.64 (0.38)***	0.17 (0.08)*	0.74 (0.05)***	0.62 (0.07)***	0.37 (0.04)***
depth	–	0.24 (0.11)*	0.05 (0.06)	0.07 (0.04)*	–	-0.001 (<0.001)***
depth <sup>2</sup>	–	–	–	–	–	–
slope	–	–	–	–	–	–
SBT	–	–	0.17 (0.06)**	–	0.06 (0.03)*	–
depth*SBT	–	–	0.11 (0.05)*	–	–	–
Norway lobster métier	–	–	0.13 (0.08)	–	0.23 (0.07)**	–
blue and red shrimp métier	–	–	0.25 (0.12)*	–	0.25 (0.07)***	–
fishing effort	-1.74 (0.46)***	-1.82 (0.48)***	-0.31 (0.08)***	–	0.08 (0.03)**	–
sandy	0.89 (0.32)**	0.49 (0.35)	0.07 (0.07)	–	-0.22 (0.09)*	–
muddy	0.92 (0.32)**	0.41 (0.36)	0.03 (0.07)	–	-0.22 (0.09)*	–
fishing effort*sandy	0.86 (0.43)*	1.04 (0.46)*	0.29 (0.09)**	–	–	–
fishing effort*muddy	1.64 (0.45)***	1.66 (0.47)***	0.25 (0.09)**	–	–	–
Pearson coefficient	0.78	0.72	0.49	0.60	0.41	0.86

<sup>a</sup>Response variables that were log-transformed.

Significance codes: \*\*\*<0.001, \*\*<0.01, \*<0.05.

Estimates of fixed effects (S.E.) and significance are presented. The results correspond to the optimal models, with variable selection conducted using AIC. ‘–’ indicates variables not selected using AIC criteria.

included the métier and the interactions between fishing effort and substrate, as well as between sea bottom temperature and depth. The effect of fishing pressure on chondrichthyan diversity varied depending on the substrate, with a stronger negative effect on coarse sediments, while its effect was milder for both sandy and muddy substrates. In all the cases diversity decreased with increasing fishing effort. The effect of sea bottom temperature on diversity also varied depending on depth, with major changes occurring at deeper areas where higher temperatures were associated with higher diversity. Temperature fluctuated more in the continental shelf, ranging from 12.9 to 16.6 °C, whereas in the slope it was more stable, ranging from 12.6 to 13.8 °C. In addition, chondrichthyan diversity varied depending on the métier. Higher diversity of chondrichthyans observed in the catches of the blue and red shrimp métier, followed by the Norway lobster métier, both métiers occurring in the continental slope. The diversity captured in the hake métier was lower, occurring mostly in the continental shelf (Table 3; Figure 4).

Richness (Margalef's index) values for chondrichthyans in this study ranged from 0 to 1.95, averaging  $0.51 \pm 0.44$ , with a maximum of 5 different species found in a single haul. The selected model for chondrichthyan richness only included the predictor depth, having a positive effect (Table 3; Figure 4).

Evenness (Pielou's index) averaged at  $0.59 \pm 0.28$  and ranged between 0.04 and 1. The optimal model for chondrichthyan evenness included the predictor variables of fishing effort, sea bottom temperature, métier, and substrate. Evenness increased with higher levels of fishing effort and sea bottom temperature. It

was lower for the hake métier and in sandy and muddy substrates (Table 3; Figure 4).

## Discussion

Our study provides updated information on the status and composition of the chondrichthyan community in the Mediterranean trawling grounds of the eastern Iberian Peninsula (GSA06) and contributes to a deeper understanding of the drivers leading its structure and spatial distribution. A total of 17 species were recorded, including 7 sharks, 9 batoids, and 1 chimaera. The chondrichthyan catch was dominated by few species including *S. canicula*, *G. melastomus*, *E. spinax* and *T. marmorata*. In this study we provide bycatch rates for all the reported species by accurately depicting the unaltered practices of the local fleet. The bycatch rates provided for each species may aid to understand their current level of interaction with commercial bottom trawling. This information is essential to inform management plans, considering that most of the current research uses fishing-independent methods and most of the chondrichthyan biomass captured is either not landed or labelled incorrectly in fisheries statistics (Barriá and Colmenero, 2019; Cashion et al., 2019). The species composition, their relative abundance and the values of the analyzed community descriptors are consistent with what has been described in recent years using fisheries-independent research in the same region (Ramirez-Amaro et al., 2015; Follesa et al., 2019; Ramirez-Amaro et al., 2020). Although, chondrichthyan diversity has been historically considered greatest in the western basin, the demersal community



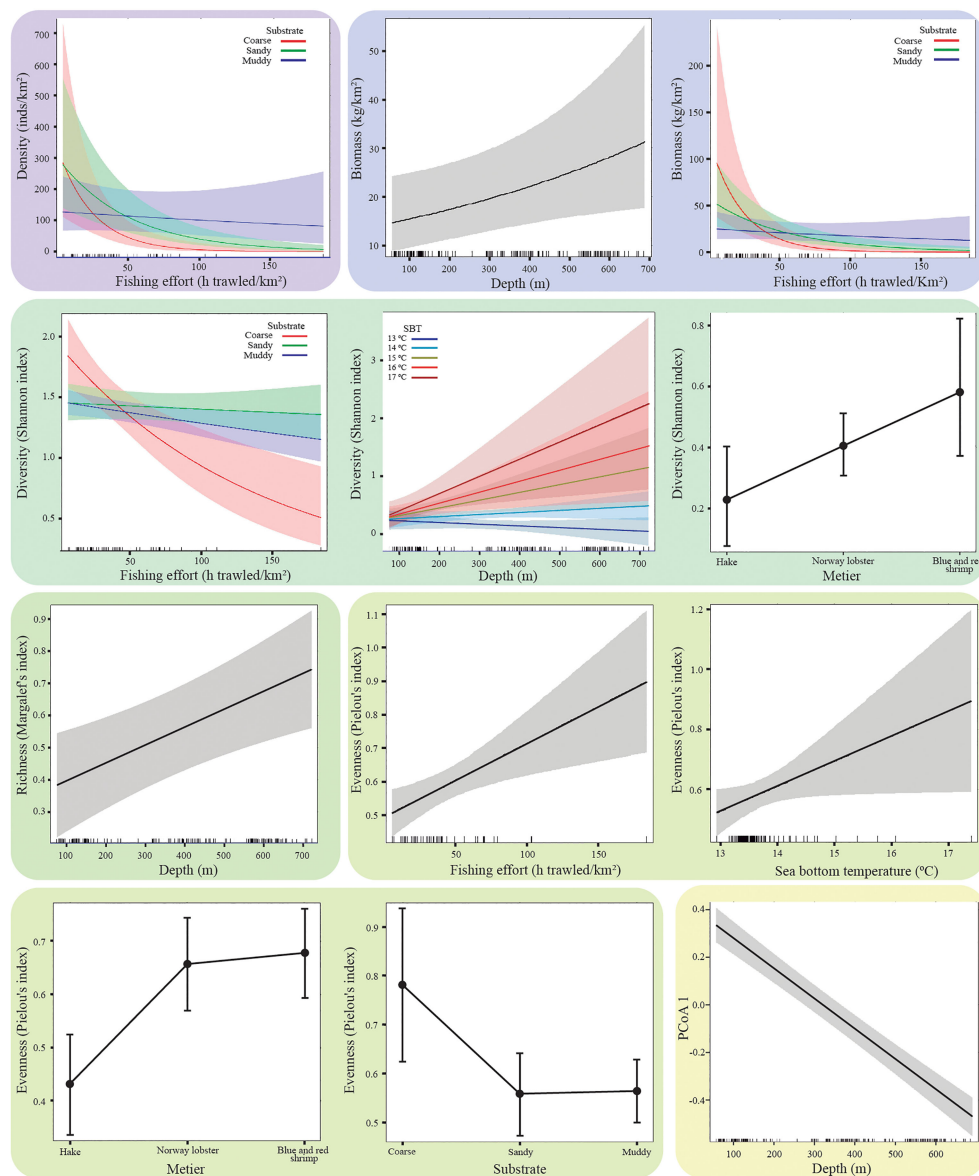


FIGURE 4

Partial effects of predictor variables included in the selected models for chondrichthyan density (inds/km<sup>2</sup>), biomass (kg/km<sup>2</sup>), diversity (Shannon index), richness (Margalef's index), evenness (Pielou's index) and the principal coordinate (PCoA) axis. The box of each plot is coloured referring to its respective response variable.

inhabiting in the bottom trawling grounds has been suggested to be higher in the Strait of Sicily and the Aegean Sea, located in the central and eastern Mediterranean, respectively, while little is known about the African Mediterranean (Dulvy et al., 2016; Follesa et al., 2019; Serena et al., 2020). We found that the distribution and structure of the chondrichthyan community is influenced by environmental drivers such as depth, temperature, and seabed substrate, but also by the fishing pressure exerted in the area.

## Environmental drivers

Depth was found to be the most important factor in shaping chondrichthyan community composition. Most of the batoids (*A.*

*bovinus*, *M. aquila*, *Raja asterias*, *R. clavata*, *R. polystigma*, *T. marmorata*, *D. pastinaca*, and *Pteroplatytrygon violacea*) were found in the continental shelf (50–200 m depth). In contrast, most sharks (*G. melastomus*, *E. spinax*, *D. licha*, and *H. griseus*), preferred the deepest part of the continental slope (>500 m depth). The skate *D. oxyrinchus*, the shark *C. uyato*, and the chimaera *C. monstrosa* were primarily found in the upper continental slope (200–500 m depth). *S. canicula* was present all along the studied bathymetrical range although its abundance was higher in the continental shelf and upper slope, with center of gravity at about 250 m depth. Chondrichthyan biomass increased with depth although density did not exhibit any significant trend, indicating that the specimens in deeper areas had a larger body biomass. Diversity and richness indices also increased with depth. Similar

TABLE 4 Similarity percentages (SIMPER) analysis results for the comparison of the assemblages (continental shelf, upper slope, and lower slope) identified from principal coordinates analysis.

	Density (individuals/km <sup>2</sup> )					
	$\bar{X}$	$\bar{Y}$	$\hat{S}_i$	$\hat{S}_i/SD$	% $\hat{S}_i$	$\Sigma\% \hat{S}_i$
Shelf vs. upper slope (average dissimilarity = 51.27%)						
<i>G. melastomus</i>	0.85	118.71	0.14	1.17	31.94	31.94
<i>S. canicula</i>	340.58	302.36	0.15	1.09	31.36	63.30
<i>T. marmorata</i>	8.22	0.31	0.09	0.80	14.06	77.36
<i>E. spinax</i>	0.00	5.21	0.08	0.44	7.10	84.46
Upper slope vs. lower slope (average dissimilarity = 51.48%)						
<i>S. canicula</i>	302.36	46.86	0.19	1.27	37.73	37.28
<i>G. melastomus</i>	118.71	160.42	0.18	1.36	34.35	71.63
<i>E. spinax</i>	5.21	20.23	0.11	1.24	22.4	94.03
<i>D. licha</i>	0.00	0.31	0.01	0.29	1.37	95.40
Shelf vs. lower slope (average dissimilarity = 75.33%)						
<i>G. melastomus</i>	0.85	160.42	0.29	2.51	37.34	37.87
<i>S. canicula</i>	340.58	46.86	0.21	1.28	27.59	65.46
<i>E. spinax</i>	0.00	20.23	0.12	1.26	16.34	81.80
<i>T. marmorata</i>	8.22	0.00	0.07	0.78	8.94	90.74

$\bar{X}$  mean density at group 1;  $\bar{Y}$  mean density at group 2;  $\hat{S}_i$  average contribution to overall dissimilarity; %  $\hat{S}_i$  percentage contribution to dissimilarity.

patterns have been reported by previous research using fisheries-independent trawling surveys in Northern Spain, finding higher chondrichthyan diversity in the slope, as well as in the Aegean Sea and in the Strait of Sicily, where the diversity in the chondrichthyan community peaked at 600 m deep (Lauria et al., 2015; Ramírez-Amaro et al., 2015; Peristeraki et al., 2020). However, in other subregions the peak in chondrichthyan diversity has been detected in shallower areas as in the Balearic Sea, where it occurs in the continental shelf, condition that has been attributed to a lower fishing pressure occurring in the continental shelf of the Balearic Sea (Gouraguine et al., 2011; Ramírez-Amaro et al., 2015). The changes in the species composing the community in relation to depth is related to the niches available to each species (including habitat and inter and intra-specific competition) depending on its adaptations to the different parameters that vary with depth (e.g. light, temperature, currents, pressure, food, oxygen, substrate) (Gage and Tyler, 1991). Therefore, the higher diversity levels in the slope cannot be disregarded as a result of the more extensive depth range covered by each haul in this region. However, lower diversity levels in the continental shelf may be related to the fishing pressure exerted in the area as suggested in prior research (Aldebert, 1997; Ramírez-Amaro et al., 2015). Fishing pressure in Northern Spain decreases with depth, being higher in the continental shelf, especially in areas located near large ports and cities, likely influenced by the market demands, as in other areas of the Mediterranean (Peristeraki et al., 2020). Additionally, fisheries beyond 1,000 m depth are banned since 2006 as decided by the GFCM. The generally lower fishing pressure in the continental slope may provide deepwater chondrichthyan with certain refuge

from anthropogenic pressures, maintaining higher diversity levels (García et al., 2008). However, a shift of the fleet towards deepwaters has occurred in the western Mediterranean during the last decades consequence of technological improvement of fishing activities, the overexploitation status of the continental shelf stocks and the higher economic value of the targets *N. norvegicus* and *A. antennatus* (Moranta et al., 2008; Fernandes et al., 2017). It has been suggested that this shift has promoted certain stability for the chondrichthyan inhabiting in the Mediterranean continental shelf (Marongiu et al., 2017; Ramírez-Amaro et al., 2020), but continued elevated fishing pressure may drive to declining population trends for the species dwelling in the slope, as it has been observed for *E. spinax* and *D. oxyrinchus* in Northern Spain (Ramírez-Amaro et al., 2020). Special caution must be taken with deepwater species considering their longer turnover times (i.e. slower growth, later age at maturity and higher longevity; Pardo and Dulvy, 2022).

Sea bottom temperature has been previously suggested as a relevant driver of chondrichthyan habitat selection (e.g. Follesa et al., 2019; Maioli et al., 2022). In the Mediterranean, increased shark densities have been previously documented at higher bottom temperatures using fisheries-independent trawling surveys (Follesa et al., 2019), although species-specific patterns and ontogenetic shifts also occur (Damalas et al., 2010; Maioli et al., 2022). As chondrichthyan are ectothermic, changes in the environmental temperature affect their physiological processes (Brown et al., 2004; Savage et al., 2004). Higher temperatures speed chemical and metabolic reactions as hypothesized by metabolic theory, which may lead to a redistribution of chondrichthyan to thermoregulate (Osgood et al., 2021). Temperature decreases with depth and so

does light, oxygen, and primary production, and consequently, food availability for chondrichthyans (Gage and Tyler, 1991; Jahnke, 1996; Rex et al., 2006). Thus, the metabolic capacity of chondrichthyans also generally decreases with depth (Seibel and Drazen, 2007; Pardo and Dulvy, 2022), limiting their energetic capabilities to forage in an environment where food availability is generally low. Our study indicates that sea bottom temperature is a significant predictor for chondrichthyan diversity and evenness. Particularly, the effect of bottom temperature upon diversity was remarkable at major depths where diversity is higher in the slightly warmer areas. Although temperature in the slope was fairly stable, it ranged from 12.6 to 13.8 °C. Slightly warmer waters in deeper areas may enable for a higher activity in chondrichthyans, attracting a larger diversity of species (Grady et al., 2019). Additionally, temperature may also affect the availability of prey which in turn affects chondrichthyan distribution (Cotton et al., 2005; Lauria et al., 2015). Our study supports previous insights highlighting the importance of sea bottom temperature on structuring the chondrichthyan community, however, continued research and monitoring on the effects of this driver over extended time periods is essential to understand potential effects of global change. As most species are narrowly thermally adapted, it must be considered that increasing temperatures due to ocean warming may influence species distribution and abundance, potentially shifting their habitat ranges to select deeper and cooler waters in order to enhance physiological processes, potentially entailing important changes in the community dynamics (Sunday et al., 2012; Osgood et al., 2021; Vilmar and Di Santo, 2022).

Habitat preferences of demersal chondrichthyan species have been previously found to be influenced by the type of seabed substrate (e.g. Pennino et al., 2013; Lauria et al., 2015). Additionally, these preferences may change as the species undergoes ontogeny, potentially following the distribution of their preferred prey (Martin et al., 2012). In the Strait of Sicily, higher richness has been documented in coarse sediments using trawling surveys, although certain species including *R. clavata*, *Mustelus mustelus* and *Squalus blainvillei* preferred fine sediments (Lauria et al., 2015). Similarly, in the Adriatic *R. clavata*, *T. marmorata* and *Scyliorhinus stellaris* have been associated to muddy environments, while *S. canicula* and *R. miraletus* to sandy substrates (Maioli et al., 2022). In the southern region of the GSA06 the sharks *S. canicula*, *G. melastomus* and *E. spinax* have been documented to prefer sandy substrates rather than muddy bottoms using commercial trawling data (Pennino et al., 2013). Our results indicate that sandy and coarse seabeds in Northern Spain may host a higher abundance and diversity of chondrichthyans at low fishing pressure levels. However, as fishing effort increased in sandy and coarse sediment areas, these levels drop rapidly in comparison to muddy environments, where declining occurred more progressively. This pattern is more marked for abundance than for diversity levels where sandy bottoms exhibited a similar trend to muddy bottoms, but in both cases resulted in larger overall mean values in muddy sediments. As areas with hard sediments are usually not trawlable, they may constitute important *de facto* refugia for demersal chondrichthyans as it has been previously suggested (Shephard et al., 2012). However, chondrichthyans in these areas may be more

abruptly affected by increased fishing effort than muddy bottoms. The use of remotely operated vehicles (ROVs) constitutes a useful tool to acquire a better understanding of the importance of hard substrate areas for chondrichthyans. Certainly, prior research using this technology indicates that hard substrate areas may provide shelter for chondrichthyans and act as nurseries for some demersal species (e.g. Treude et al., 2011; Henry et al., 2013; Micarelli and Reinero, 2021). However, many of these areas remain poorly researched.

The differential effect of muddy sediments with increasing fishing effort may be a consequence of the indirect effects of trawling, which are particularly important in these environments (Collie et al., 2017). Sediment resuspension rates are higher in muddy bottoms, usually found in areas of low natural disturbance where the time to sediment again is generally larger (Duplisea et al., 2001; Kaiser et al., 2006). The physical disturbance caused during trawling in these areas may increase food availability for mesopredators in the short-term, displacing burial biota, increasing carrion, and resuspending organic matter (Link and Almeida, 2002; Collie et al., 2017). Although, the long-term consequences are not well-known and may depend on the species' feeding habits and the environmental conditions (Collie et al., 2017), chronically trawled sediments in the north-western Mediterranean have experimented significant decreases in organic matter (Pusceddu et al., 2014). Thus, although high fishing effort in muddy environments may lead to depletion, low to medium fishing efforts may retain mesopredators if they are attracted to the short-term pulses of food (Hiddink et al., 2006; Hinz et al., 2009; Collie et al., 2017). A correlation between the abundance of *G. melastomus* and turbid areas caused by high levels of natural vorticity, has been described in the north-western Mediterranean (Amores et al., 2014). Certainly, chronic trawling along the continental slope of the Northern Spain was reported to cause a similar effect on resuspension and turbidity as natural high levels of vorticity (Paradis et al., 2017; Arjona-Camas et al., 2021). Therefore, we hypothesise that the resuspension effect may mask the direct impact of fishing in muddy areas. However, in all cases higher levels of fishing effort led to changes in the chondrichthyan community structure with declining abundance and diversity and increasing evenness. Nevertheless, it must be considered that trawling surveys are mostly limited to soft substrate, being harder substrate areas not trawlable. Thus, the observed distribution patterns may be partially influenced by limitations in the sampling procedure when using trawling data.

## Fishing effects

Our results indicate significant declines in density, biomass, diversity and evenness of chondrichthyans as fishing effort increases. This is in line with previous studies, supporting the idea that fishing effort contributes importantly into shaping the community of chondrichthyans (Navarro et al., 2016; Peristeraki et al., 2020). Fishing effort played an important role in the modification of the community during the early exploitation period (1960-1990). In the north-western Mediterranean, some of

the currently most threatened species were commonly caught during the 1950s, including from more to less common: *Squalus* spp., *S. stellaris*, *Mustelus* spp., *Leucoraja circularis*, *Galeorhinus galeus*, *O. centrina*, *Leucoraja melitensis*, *Rhinobatos* spp., and *Squatina* spp. (Maurin, 1968; Aldebert, 1997; Jukic-Peladic et al., 2001). These species are now rare catch in bottom trawling or locally extinct species as their life-history parameters and distribution patterns made them especially susceptible to overexploitation (Walls and Dulvy, 2020; Walls and Dulvy, 2021). After the 2000s, it has been suggested that the GSA06 reached certain stability in terms of diversity (Ramírez-Amaro et al., 2020). Nevertheless, the current species-specific patterns still suggest a trend towards community simplification, with declining population trends generally occurring for species with lower intrinsic rates of population growth (Ligas et al., 2013; Marongiu et al., 2017; Ramírez-Amaro et al., 2020; Walls and Dulvy, 2020; Walls and Dulvy, 2021). Even though most of these chondrichthyan species are discarded due to their low economical value or protection status, their survival rates are likely to be low (Ellis et al., 2017). Further research on post-capture mortality and handling methods to enhance survival are key to inform management plans.

On the other hand, some species such as *S. canicula*, *G. melastomus*, *T. marmorata* and *R. clavata* exhibit generally increasing population trends in the north-western Mediterranean (Ligas et al., 2013; Marongiu et al., 2017; Ramírez-Amaro et al., 2020). These species can withstand the effects of fishing effort, potentially taking the released niches by accounting with higher rates of intrinsic population growth and post-capture survival (Revell et al., 2005; Ellis et al., 2017; Walls and Dulvy, 2020; Walls and Dulvy, 2021). Additionally, some of them may take advantage of fisheries discards (Olaso et al., 2002). This is in line with the elevated abundance levels that were detected for these species and especially, for *S. canicula* and *G. melastomus*, in the fishing grounds studied here. In general, the Mediterranean chondrichthyan communities are shifting towards the predominance of species with lower age at maturity and higher reproductive rates as result of the selection force occasioned by the elevated fishing pressure (Ramírez-Amaro et al., 2020; Maioli et al., 2022). Furthermore, there is evidence of decreasing size at maturity in *S. canicula* and *G. melastomus* in the Spanish Mediterranean during recent decades, potentially as adaptative response to overfishing (Ramírez-Amaro et al., 2020). These tendencies indicate that fishing pressure have had an important role in shaping the current structure of the chondrichthyan community, acting as a selective pressure in the Mediterranean and, particularly, in our study region. The results from this study are in line with these insights, revealing that the abundance and diversity of chondrichthyans decreases importantly in areas with increased fishing pressure. Considering that the GSA06 is one of the most exploited sub-areas in the Mediterranean (Armelloni et al., 2021), continued research in this region is essential to monitor changes and inform management measures. Particular attention must be given to species exhibiting decreasing population trends, but also to species that are generally commercialised, such as *S. canicula*, *G. melastomus*, and the species of the genus *Raja* spp. as the overexploitation of many fish stocks

may lead to an increase in the catch and retain of species with lower economical interest, including these (Davidson et al., 2016).

Future research may consider both species and ontogenetic-stage specific distribution modelling approaches with additional detail in both environmental and human factors to further understand the drivers of their distribution and the potential effects of increased fishing pressure, global change, and habitat loss. Furthermore, it is crucial to improve chondrichthyan-focused fisheries management in the Mediterranean to halt the declines of these species and prevent regional species extinctions (Colloca et al., 2013; Fernandes et al., 2017; Walls and Dulvy, 2021). Although, there have been diverse suggestions for addressing the overfishing problems in the Mediterranean fisheries, only few have been effectively implemented to protect sharks, rays, and chimaeras (Colloca et al., 2013; Walls and Dulvy, 2021).

## Conclusions

This study revealed the importance of combining environmental and fishing-related drivers to understand the spatial distribution patterns of chondrichthyan communities. Our results indicate that depth is the main driver of the composition of the community in the studied bottom trawling fishing grounds (ranging from 50 to 800 m deep). Deeper areas host a higher diversity of chondrichthyans species, potentially related to the higher fishing effort exerted in the continental shelf. The distribution of the chondrichthyan community is also influenced by sea bottom temperature with slight variances in temperature in deeper areas entailing significant changes in diversity, and by substrate type, with different sediment types exhibiting different responses to increased fishing effort. However, in all cases, density, biomass, and diversity of chondrichthyans is importantly affected by the intensity of fishing pressure, with higher pressure leading to declining values. These insights are crucial to further understand the effects of increased fishing effort on the chondrichthyan community as well as the potential effects of global change and habitat loss. Our approach provide capacity to make more accurate predictions when determining thresholds for sustainable practices and estimating the situation for chondrichthyans where data is scarce or absent. This may aid to build management capacity in poorly researched areas and prevent future threats. Additionally, our work highlights the need for continued research and monitoring in this area, given the ongoing challenges and uncertainties that the Mediterranean chondrichthyans and their ecosystems are facing. It is crucial to improve chondrichthyan-focused fisheries management in the Mediterranean to halt the declines of these species and prevent regional species extinctions.

## Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

## Ethics statement

Ethical review and approval was not required for the animal study because the chondrichthyan specimens analyzed in the present work were obtained from commercial fisheries. The activity was conducted with the observation of the Regulation of the European Parliament and the Council for fishing in the General Fisheries Commission for the Mediterranean (GFCM) Agreement area and amending Council Regulation (EC) No. 1967/2006. This Regulation is de facto the unique authorization needed to conduct this type of activity.

## Author contributions

CB, JR, and DR-G conceived and designed the study. JR, DR-G, and CB acquired funding and administrated the project. DR-G, AC, JC, LR, and CB contributed to data collection. DR-G, DM, and FQ performed the statistical analysis. DR-G wrote the draft of the manuscript. All authors contributed to the article and approved the submitted version.

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## References

- Albeert, Y. (1997). Demersal resources of the gulf of lions (NW mediterranean). impact of exploitation on fish diversity. *Vie Milieu*. 47, 275–284.
- Amores, A., Rueda, L., Monserrat, S., Guijarro, B., Pasqual, C., and Massuti, E. (2014). Influence of the hydrodynamic conditions on the accessibility of *Aristeus antennatus* and other demersal species to the deep water trawl fishery off the Balearic islands (western Mediterranean). *J. Mar. Syst.* 138, 203–210. doi: 10.1016/j.jmarsys.2013.11.014
- Arjona-Camas, M., Puig, P., Palanques, A., Durán, R., White, M., Paradis, S., et al. (2021). Natural vs. trawling-induced water turbidity and suspended sediment transport variability within the palamós canyon (NW Mediterranean). *Mar. Geophys. Res.* 42, 1–22. doi: 10.1007/s11001-021-09457-7
- Armelloni, E. N., Tassetti, A. N., Ferrà, C., Galdelli, A., Scanu, M., Mancini, A., et al. (2021). AIS data, a mine of information on trawling fleet mobility in the Mediterranean Sea. *Mar. Policy* 129, 104571. doi: 10.1016/j.marpol.2021.104571
- Barausse, A., Correale, V., Curkovic, A., Finotto, L., Riginella, E., Visentin, E., et al. (2014). The role of fisheries and the environment in driving the decline of elasmobranchs in the northern Adriatic Sea. *ICES J. Mar. Sci.* 71, 1593–1603. doi: 10.1093/icesjms/fst222
- Barría, C., Coll, M., and Navarro, J. (2015a). Unravelling the ecological role and trophic relationships of uncommon and threatened elasmobranchs in the western Mediterranean Sea. *Mar. Ecol. Prog. Ser.* 539, 225–240. doi: 10.3354/meps11494
- Barría, C., and Colmenero, A. I. (2019). La pesca de tiburones y rayas en el noroeste del mediterráneo, una situación compleja. *da*. 10, 105–111. doi: 10.5565/rev/da.468
- Barría, C., Navarro, J., Coll, M., Fernandez-Arcaya, U., and Sáez-Liante, R. (2015b). Morphological parameters of abundant and threatened chondrichthyans of the northwestern Mediterranean Sea. *J. Appl. Ichthyol.* 31, 114–119. doi: 10.1111/jai.12499
- Bates, D., Mächler, M., Bolker, B., and Walker, S. (2015). Fitting linear mixed-effects models using lme4. *J. Stat. Software* 67, 1–48. doi: 10.18637/jss.v067.i01
- Bertrand, J., Gil-de-Sola-Simarro, L., Papakostantinou, C., Relini, G., and Souplet, A. (2000). Contribution on the distribution of elasmobranchs in the Mediterranean (from MEDITS surveys). *Biol. Mar. Medit.* 7, 1–15.
- Bolker, B. M., Brooks, M. E., Clark, C. J., Geange, S. W., Poulsen, J. R., Stevens, M. H. H., et al. (2009). Generalized linear mixed models: a practical guide for ecology and evolution. *Trends Ecol. Evol.* 24, 127–135. doi: 10.1016/j.tree.2008.10.008
- Brown, J. H., Gillooly, J. F., Allen, A. P., Savage, V. M., and West, G. B. (2004). Toward a metabolic theory of ecology. *Ecology* 85, 1771–1789. doi: 10.1890/03-9000

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## Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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## Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2023.1145176/full#supplementary-material>

- Cashion, M. S., Bailly, N., and Pauly, D. (2019). Official catch data underrepresent shark and ray taxa caught in Mediterranean and black Sea fisheries. *Mar. Policy* 105, 1–9. doi: 10.1016/j.marpol.2019.02.041
- Collie, J., Hiddink, J. G., van Kooten, T., Rijnsdorp, A. D., Kaiser, M. J., Jennings, S., et al. (2017). Indirect effects of bottom fishing on the productivity of marine fish. *Fish Fish.* 18, 619–637. doi: 10.1111/faf.12193
- Colloca, F., Cardinale, M., Maynou, F., Giannoulaki, M., Scarcella, G., Jenko, K., et al. (2013). Rebuilding Mediterranean fisheries: A new paradigm for ecological sustainability. *Fish Fish.* 14, 89–109. doi: 10.1111/j.1467-2979.2011.00453.x
- Colloca, F., Scarcella, G., and Libralato, S. (2017). Recent trends and impacts of fisheries exploitation on Mediterranean stocks and ecosystems. *Front. Mar. Sci.* 4. doi: 10.3389/fmars.2017.00
- Cortés, E. (2000). Life history patterns and correlations in sharks. *Rev. Fish. Sci.* 8, 299–344. doi: 10.1080/10408340308951115
- Cotton, P. A., Sims, D. W., Fanshawe, S., and Chadwick, M. (2005). The effects of climate variability on zooplankton and basking shark (*Cetorhinus maximus*) relative abundance off southwest Britain. *Fish. Oceanogr.* 14, 151–155. doi: 10.1111/j.1365-2419.2005.00331.x
- Damalas, D., Maravelias, C. D., Katsanevakis, S., Karageorgis, A. P., and Papaconstantinou, C. (2010). Seasonal abundance of non-commercial demersal fish in the eastern Mediterranean Sea in relation to hydrographic and sediment characteristics. *Estuar. Coast. Shelf Sci.* 89, 107–118. doi: 10.1016/j.ecss.2010.06.002
- Damalas, D., and Vassilopoulou, V. (2011). Chondrichthyan by-catch and discards in the demersal trawl fishery of the central Aegean Sea (Eastern Mediterranean). *Fish. Res.* 108, 142–152. doi: 10.1016/j.fishres.2010.12.012
- Davidson, L. N., Krawchuk, M. A., and Dulvy, N. K. (2016). Why have global shark and ray landings declined: improved management or overfishing? *Fish Fish.* 17, 438–458. doi: 10.1111/faf.12119
- De Buen, F. (1935). “Fauna ictiológica,” in *Catálogo de los peces ibéricos: de la planicie continental, aguas dulces, pelágicos y de los abismos próximos* (Madrid: Instituto Español de Oceanografía: Primera Parte).
- Dormann, C. F. (2007). Effects of incorporating spatial autocorrelation into the analysis of species distribution data. *Global Ecol. Biogeogr.* 16, 129–138. doi: 10.1111/j.1466-8238.2006.00279.x
- Dulvy, N. K., Allen, D. J., Ralph, G. M., and Walls, R. (2016). *The conservation status of sharks, rays and chimaeras in the Mediterranean Sea [Brochure]* (Malaga, Spain: IUCN).
- Dulvy, N. K., Pacoureau, N., Rigby, C. L., Pollom, R. A., Jabado, R. W., Ebert, D. A., et al. (2021). Overfishing drives over one-third of all sharks and rays toward a global extinction crisis. *Curr. Biol.* 31, 4773–4787.e8. doi: 10.1016/j.cub.2021.08.062
- Duplisa, D. E., Jennings, S., Malcolm, S. J., Parker, R., and Sivyer, D. B. (2001). Modelling potential impacts of bottom trawl fisheries on soft sediment biogeochemistry in the north Sea. *Geochem. Trans.* 14, 1–6. doi: 10.1186/1467-4866-2-112
- Ellis, J. R., McCully Phillips, S. R., and Poisson, F. (2017). A review of capture and post-release mortality of elasmobranchs. *J. Fish Biol.* 90, 653–722. doi: 10.1111/jfb.13197
- FAO (2020). “The state of Mediterranean and black Sea fisheries 2020,” in *General fisheries commission for the Mediterranean*. (Rome: Food and Agriculture Organization of the United Nations). doi: 10.4060/cb2429en
- Fernandes, P. G., Ralph, G. M., Nieto, A., García Criado, M., Vasilakopoulos, P., Maravelias, C. D., et al. (2017). Coherent assessments of Europe’s marine fishes show regional divergence and megafauna loss. *Nat. Ecol. Evol.* 1, 1–9. doi: 10.1038/s41559-017-0170
- Ferretti, F., Myers, R. A., Serena, F., and Lotze, H. K. (2008). Loss of large predatory sharks from the Mediterranean Sea. *Conserv. Biol.* 22, 952–964. doi: 10.1111/j.1523-1739.2008.00938.x
- Ferretti, F., Osio, G. C., Jenkins, C. J., Rosenberg, A. A., and Lotze, H. K. (2013). Long-term change in a meso-predator community in response to prolonged and heterogeneous human impact. *Sci. Rep.* 3, 1–11. doi: 10.1038/srep01057
- Follesa, M. C., Marongiu, M. F., Zupa, W., Bellodi, A., Cau, A., Cannas, R., et al. (2019). Spatial variability of Chondrichthyes in the northern Mediterranean. *Sci. Mar.* 83, 81–100. doi: 10.3989/scimar.04998.23A
- Gage, J. D., and Tyler, P. A. (1991). *Deep-Sea biology: A natural history of organisms at the deep-sea floor* (UK: Cambridge University Press).
- García, V. B., Lucifora, L. O., and Myers, R. A. (2008). The importance of habitat and life history to extinction risk in sharks, skates, rays and chimaeras. *Proc. R. Soc. B.* 275, 83–89. doi: 10.1098/rspb.2007.1295
- Geraci, M. L., Ragonese, S., Scannella, D., Falsone, F., Gancitano, V., et al. (2021). Batoid abundances, spatial distribution, and life history traits in the strait of Sicily (central Mediterranean sea): Bridging a knowledge gap through three decades of survey. *Animals* 11, 2189. doi: 10.3390/ani11082189
- Gouraguine, A., Hidalgo, M., Moranta, J., Bailey, D. M., Ordines, F., Guisjarro, B., et al. (2011). Elasmobranch spatial segregation in the western Mediterranean. *Sci. Mar.* 75, 653–664. doi: 10.3989/scimar.2011.75n4653
- Grady, J. M., Maitner, B. S., Winter, A. S., Kaschner, K., Tittensor, D. P., Record, S., et al. (2019). Metabolic asymmetry and the global diversity of marine predators. *Science* 363, eaat4220. doi: 10.1126/science.aat4220
- Guisjarro, B., Quetglas, A., Moranta, J., Ordines, F., Valls, M., González, N., et al. (2012). Inter- and intra-annual trends and status indicators of nektonic elasmobranchs off the Balearic islands (northwestern Mediterranean). *Sci. Mar.* 76, 87–96. doi: 10.3989/scimar.03432.22A
- Heithaus, M. R., Frid, A., Wirsing, A. J., and Worm, B. (2008). Predicting ecological consequences of marine top predator declines. *Trends Ecol. Evol.* 23, 202–210. doi: 10.1016/j.tree.2008.01.003
- Henry, L. A., Navas, J. M., Hennige, S. J., Wicks, L. C., Vad, J., and Roberts, J. M. (2013). Cold-water coral reef habitats benefit recreationally valuable sharks. *Biol. Conserv.* 161, 67–70. doi: 10.1016/j.biocon.2013.03.002
- Hiddink, J. G., Jennings, S., Kaiser, M. J., Queirós, A. M., Duplisa, D. E., and Piet, G. J. (2006). Cumulative impacts of seabed trawl disturbance on benthic biomass, production, and species richness in different habitats. *Can. J. Fish. Aquat. Sci.* 63, 721–736. doi: 10.1139/f05-266
- Hijmans, R. (2022). *Raster: Geographic data analysis and modeling*. R Package version 3, 6–11. Available at: <https://cran.r-project.org/web/packages/raster/raster.pdf>.
- Hinz, H., Prieto, V., and Kaiser, M. J. (2009). Trawl disturbance on benthic communities: chronic effects and experimental predictions. *Ecol. Appl.* 19, 761–773. doi: 10.1890/08-0351.1
- IUCN (2022) *The IUCN red list of threatened species* (Version 2022-2). Available at: <https://www.iucnredlist.org> (Accessed January 15, 2023).
- Jahnke, R. A. (1996). The global ocean flux of particulate organic carbon: Areal distribution and magnitude. *Global Biogeochem. Cycles* 10, 71–88. doi: 10.1029/95GB03525
- Jukic-Peladic, S., Vrgoc, N., Krstulovic-Sifner, S., Piccinetti, C., Piccinetti-Manfrin, G., Marano, G., et al. (2001). Long-term changes in demersal resources of the Adriatic Sea: a comparison between trawl surveys carried out in 1948 and 1998. *Fish. Res.* 53, 95–104. doi: 10.1016/S0165-7836(00)00232-0
- Kaiser, M. J., Clarke, K. R., Hinz, H., Austen, M. C., Somerfield, P. J., and Karakassis, I. (2006). Global analysis of response and recovery of benthic biota to fishing. *Mar. Ecol. Prog. Ser.* 311, 1–14. doi: 10.3354/meps311001
- Koehler, L., Giovos, I., and Lowther, J. (2022). The application of precaution in elasmobranch conservation and management in the Mediterranean Sea. *Mar. Policy* 135, 104830. doi: 10.1016/j.marpol.2021.104830
- Kuznetsova, A., Brockhoff, P. B., and Christensen, R. H. B. (2017). lmerTest package: Tests in linear mixed effects models. *J. Stat. Software* 82, 1–26. doi: 10.18637/jss.v082.i13
- Lauria, V., Gristina, M., Attrill, M., and Garofalo, G. (2015). Predictive habitat suitability models to aid conservation of elasmobranch diversity in the central Mediterranean Sea. *Sci. Rep.* 5, 13245. doi: 10.1038/srep13245
- Ligas, A., Osio, G. C., Sartor, P., Sbrana, M., and De Ranieri, S. (2013). Long-term trajectory of some elasmobranch species off the Tuscany coasts (NW Mediterranean) from 50 years of catch data. *Sci. Mar.* 77, 119–127. doi: 10.3989/scimar.03654.21C
- Link, J. S., and Almeida, F. P. (2002). Opportunistic feeding of longhorn sculpin (*Myoxocephalus octodecemspinosus*): Are scallop fishery discards an important food subsidy for scavengers on georges bank? *Fish. Bull.* 100, 381–385.
- Lucifora, L. O., García, V. B., Menni, R. C., and Worm, B. (2012). Spatial patterns in the diversity of sharks, rays, and chimaeras (Chondrichthyes) in the southwest Atlantic. *Biodivers. Conserv.* 21, 407–419. doi: 10.1007/s10531-011-0189-7
- Maioli, F., Weigel, B., Chiarabelli, E., Manfredi, C., Anibaldi, A., Isailović, I., et al. (2022). Influence of ecological traits on spatio-temporal dynamics of an elasmobranch community in a heavily exploited basin (Accessed January 15, 2023).
- Maravelias, C. D., Tserpes, G., Pantazi, M., and Peristeraki, P. (2012). Habitat selection and temporal abundance fluctuations of demersal cartilaginous species in the Aegean Sea (eastern Mediterranean). *PLoS One* 7, e35474. doi: 10.1371/journal.pone.0035474
- Marongiu, M. F., Porcu, C., Bellodi, A., Cannas, R., Cau, A., Cuccu, D., et al. (2017). Temporal dynamics of demersal chondrichthyan species in the central western Mediterranean Sea: The case study in Sardinia island. *Fish. Res.* 193, 81–94. doi: 10.1016/j.fishres.2017.04.001
- Martin, C. S., Vaz, S., Ellis, J. R., Lauria, V., Coppin, F., and Carpentier, A. (2012). Modelled distributions of ten demersal elasmobranchs of the eastern English channel in relation to the environment. *J. Exp. Mar. Biol. Ecol.* 418, 91–103. doi: 10.1016/j.jembe.2012.03.010
- Maurin, C. (1968). Ecologie ichthyologique des fonds chaulutables atlantiques (de la baie ibéro-marocaine à la mauritanie) et de la méditerranée occidentale. *Rev. Trav. Inst. Pêches Marit.* 32, 95–124.
- Melo-Merino, S. M., Reyes-Bonilla, H., and Lira-Noriega, A. (2020). Ecological niche models and species distribution models in marine environments: A literature review and spatial analysis of evidence. *Ecol. Modell.* 415, 108837. doi: 10.1016/j.ecolmodel.2019.108837
- Micarelli, P., and Reinerio, F. R. (2021). Preliminary data about nursery survey of Nursehound, *Scyliorhinus stellaris* (Linnaeus, 1758), around the peninsula of Monte Argentario (Tuscany, Italy) using a newtool: Poseidon R.O.V. (Remote Operative Vehicle). *Accademia Lett.* Article 2013. doi: 10.20935/AL2013
- Moranta, J., Quetglas, A., Massuti, E., Guisjarro, B., Hidalgo, M., and Diaz, P. (2008). Spatio-temporal variations in deep-sea demersal communities of the Balearic islands (western Mediterranean). *J. Mar. Syst.* 71, 346–366. doi: 10.1016/j.jmarsys.2007.02.029

- Navarro, J., Cardador, L., Fernández, Á.M., Bellido, J. M., and Coll, M. (2016). Differences in the relative roles of environment, prey availability and human activity in the spatial distribution of two marine mesopredators living in highly exploited ecosystems. *J. Biogeogr.* 43, 440–450. doi: 10.1111/jbi.12648
- Oksanen, J., Simpson, G., Blanchet, F., Kindt, R., Legendre, P., Minchin, P., et al. (2022). *Vegan: Community ecology package* (R package version 3.0.3). Available at: <https://CRAN.R-project.org/package=vegan>.
- Olaso, I., Sánchez, F., Rodríguez-Cabello, C., and Velasco, F. (2002). The feeding behaviour of some demersal fish species in response to artificial discarding. *Sci. Mar.* 66, 301–311. doi: 10.3989/scimar.2002.66n3301
- Ordines, F., Massutí, E., Moranta, J., Quetglas, A., Guijarro, B., and Fliti, K. (2011). Balearic Islands vs Algeria: Two nearby western Mediterranean elasmobranch assemblages with different oceanographic scenarios and fishing histories. *Sci. Mar.* 75, 707–717. doi: 10.3989/scimar.2011.75n4707
- Osgood, G. J., White, E. R., and Baum, J. K. (2021). Effects of climate-change-driven gradual and acute temperature changes on shark and ray species. *J. Anim. Ecol.* 90, 2547–2559. doi: 10.1111/1365-2656.13560
- Ovaskainen, O., Tikhonov, G., Dunson, D., Grøtan, V., Engen, S., Sæther, B., et al. (2017). How are species interactions structured in species-rich communities? a new method for analysing time-series data. *Proc. R. Soc. B* 284, 20170768. doi: 10.1098/rspb.2017.0768
- Paradis, S., Puig, P., Masqué, P., Juan-Díaz, X., Martín, J., and Palanques, A. (2017). Bottom-trawling along submarine canyons impacts deep sedimentary regimes. *Sci. Rep.* 7, 1–12. doi: 10.1038/srep43332
- Pardo, S. A., and Dulvy, N. K. (2022). Body mass, temperature, and depth shape the maximum intrinsic rate of population increase in sharks and rays. *Ecol. Evol.* 12, e9441. doi: 10.1002/ece3.9441
- Pennino, M. G., Muñoz, F., Conesa, D., López-Quílez, A., and Bellido, J. M. (2013). Modeling sensitive elasmobranch habitats. *J. Sea Res.* 83, 209–218. doi: 10.1016/j.seares.2013.03.005
- Peristeraki, P., Tserpes, G., Kavadas, S., Kallianiotis, A., and Stergiou, K. I. (2020). The effect of bottom trawl fishery on biomass variations of demersal chondrichthyes in the eastern Mediterranean. *Fish. Res.* 221, 105367. doi: 10.1016/j.fishres.2019.105367
- Porcu, C., Marongiu, M. F., Olita, A., Bellodi, A., Cannas, R., Carbonara, P., et al. (2020). The demersal bathyal fish assemblage of the central-Western Mediterranean: Depth distribution, sexual maturation and reproduction. *Deep Sea Res. part I oceanogr. Res. Pap.* 166, 103394. doi: 10.1016/j.dsr.2020.103394
- Pusceddu, A., Bianchelli, S., Martín, J., Puig, P., Palanques, A., Masqué, P., et al. (2014). Chronic and intensive bottom trawling impairs deep-sea biodiversity and ecosystem functioning. *PNAS* 111, 8861–8866. doi: 10.1073/pnas.1405451111
- Ramírez-Amaro, S., Ordines, F., Esteban, A., García, C., Guijarro, B., Salmerón, F., et al. (2020). The diversity of recent trends for chondrichthyes in the Mediterranean reflects fishing exploitation and a potential evolutionary pressure towards early maturation. *Sci. Rep.* 10, 1–18. doi: 10.1038/s41598-019-56818-9
- Ramírez-Amaro, S., Ordines, F., Terrasa, B., Esteban, A., García, C., Guijarro, B., et al. (2015). Demersal chondrichthyes in the western Mediterranean: Assemblages and biological parameters of their main species. *Mar. Freshw. Res.* 67, 636–652. doi: 10.1071/MF15093
- Revill, A. S., Dulvy, N. K., and Holst, R. (2005). The survival of discarded lesser-spotted dogfish (*Scyliorhinus canicula*) in the Western English channel beam trawl fishery. *Fish. Res.* 71, 121–124. doi: 10.1016/j.fishres.2004.07.006
- Rex, M. A., Etter, R. J., Morris, J. S., Crouse, J., McClain, C. R., Johnson, N. A., et al. (2006). Global bathymetric patterns of standing stock and body size in the deep-sea benthos. *Mar. Ecol. Prog. Ser.* 317, 1–8. doi: 10.3354/meps317001
- Savage, V. M., Gillooly, J. F., Brown, J. H., West, G. B., and Charnov, E. L. (2004). Effects of body size and temperature on population growth. *Am. Nat.* 163, 429–441. doi: 10.1086/381872
- Seibel, B. A., and Drazen, J. C. (2007). The rate of metabolism in marine animals: environmental constraints, ecological demands and energetic opportunities. *Phil. Trans. R. Soc. B* 362, 2061–2078. doi: 10.1098/rstb.2007.2101
- Serena, F., Abella, A. J., Bargnesi, F., Barone, M., Colloca, F., Ferretti, F., et al. (2020). Species diversity, taxonomy and distribution of Chondrichthyes in the Mediterranean and black Sea. *Eur. Zool. J.* 87, 497–536. doi: 10.1080/24750263.2020.1805518
- Sheaves, M. (2009). Consequences of ecological connectivity: The coastal ecosystem mosaic. *Mar. Ecol. Prog. Ser.* 391, 107–115. doi: 10.3354/meps08121
- Shephard, S., Gerritsen, H., Kaiser, M. J., and Reid, D. G. (2012). Spatial heterogeneity in fishing creates de facto refugia for endangered celtic sea elasmobranchs. *PLoS One* 7, e49307. doi: 10.1371/journal.pone.0049307
- Stefanescu, C., Lloris, D., and Rucabado, J. (1993). Deep-sea fish assemblages in the Catalan Sea (western Mediterranean) below a depth of 1000 m. *Deep-Sea Res. Part I* 40, 695–707. doi: 10.1016/0967-0637(93)90066-C
- Stevens, J. D., Bonfil, R., Dulvy, N. K., and Walker, P. A. (2000). The effects of fishing on sharks, rays, and chimaeras (chondrichthyans), and the implications for marine ecosystems. *ICES J. Mar. Sci.* 57, 476–494. doi: 10.1006/jmsc.2000.0724
- Sunday, J. M., Bates, A. E., and Dulvy, N. K. (2012). Thermal tolerance and the global redistribution of animals. *Nat. Clim. Change* 2, 686–690. doi: 10.1038/nclimate1539
- Treude, T., Kiel, S., Linke, P., Peckmann, J., and Goedert, J. L. (2011). Elasmobranch egg capsules associated with modern and ancient cold seeps: A nursery for marine deep-water predators. *Mar. Ecol. Prog. Ser.* 437, 175–181. doi: 10.3354/meps09305
- Tserpes, G., Maravelias, C. D., Pantazi, M., and Peristeraki, P. (2013). Distribution of relatively rare demersal elasmobranchs in the eastern Mediterranean. *Estuar. Coast. Shelf Sci.* 117, 48–53. doi: 10.1016/j.ecss.2012.09.020
- Vilmar, M., and Di Santo, V. (2022). Swimming performance of sharks and rays under climate change. *Rev. Fish. Biol. Fish.* 32, 765–781. doi: 10.1007/s11160-022-09706-x
- Walls, R. H., and Dulvy, N. K. (2020). Eliminating the dark matter of data deficiency by predicting the conservation status of northeast Atlantic and Mediterranean Sea sharks and rays. *Biol. Conserv.* 246, 108459. doi: 10.1016/j.biocon.2020.108459
- Walls, R. H. L., and Dulvy, N. K. (2021). Tracking the rising extinction risk of sharks and rays in the northeast Atlantic ocean and Mediterranean Sea. *Sci. Rep.* 11, 15397. doi: 10.1038/s41598-021-94632-4
- Zurell, D., Jeltsch, F., Dormann, C. F., and Schröder, B. (2009). Static species distribution models in dynamically changing systems: How good can predictions really be? *Ecography* 32, 733–744. doi: 10.1111/j.1600-0587.2009.05810.x
- Zuur, A. F., Ieno, E. N., Walker, N. J., Saveliev, A. A., and Smith, G. M. (2009). *Mixed effects models and extensions in ecology with R*. (USA: New York: Springer).