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Biogenic silica cycling in the Skagerrak

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Dissolved silicate (H_4SiO_4) is essential for the formation of the opaline skeletal structures of diatoms and other siliceous plankton. A fraction of particulate biogenic silica (bSi) formed in surface waters sinks to the seabed, where it either dissolves and returns to the water column or is permanently buried. Global silica budgets are still poorly constrained since data on benthic bSi cycling are lacking, especially on continental margins. This study describes benthic bSi cycling in the Skagerrak, a sedimentary depocenter for particles from the North Sea. Biogenic silica burial fluxes, benthic H_4SiO_4 fluxes to the water column and bSi burial efficiencies are reported for nine stations by evaluating data from in-situ benthic landers and sediment cores with a diagenetic reaction-transport model. The model simulates bSi contents and H_4SiO_4 concentrations at all sites using a novel power law to describe bSi dissolution kinetics with a small number of adjustable parameters. Our results show that, on average, $1100 \text{ mmol m}^{-2} \text{ yr}^{-1}$ of bSi rains down to the Skagerrak basin seafloor, of which 50% is released back to overlying waters, with the remainder being buried. Biogenic silica cycling in the Skagerrak is generally consistent with previously reported global trends, showing higher Si fluxes and burial efficiencies than deep-sea sites and similar values compared to other continental margins. A significant finding of this work is a molar bSi-to-organic carbon burial ratio of 0.22 in Skagerrak sediments, which is distinctively lower compared to other continental margins. We suggest that the continuous dissolution of bSi in suspended sediments transported over long distances from the North Sea leads to the apparent decoupling between bSi and organic carbon in Skagerrak sediments.

KEYWORDS

biogenic silica, dissolved silicate, *in situ* measurement, burial, marine cycling, diagenetic model, sedimentation rate, Skagerrak

1 Introduction

Global marine biogeochemical cycling of dissolved silicate (H_4SiO_4) and biogenic silica (bSi) is mainly controlled by riverine inputs and by outputs by burial at the seafloor (Wollast, 1974; DeMaster, 1981; Nelson et al., 1995; Tréguer et al., 1995; Ragueneau et al., 2000; DeMaster, 2002; Tréguer and de la Rocha, 2013; Rahman et al., 2017; DeMaster, 2019). Silicon (Si) is an essential nutrient in the ocean for a number of siliceous organisms such as diatoms, silicoflagellates, siliceous sponges and radiolaria (DeMaster, 2002). The bSi and carbon (C) cycles are coupled *via* biomineralization in the surface ocean and subsequent dissolution of detritus in the water column and burial in sediments. The tightness of this coupling is thus important for the sequestration of atmospheric CO_2 (Dugdale et al., 1995; Ragueneau et al., 2006; Tréguer and de la Rocha, 2013; Tréguer et al., 2018; Tréguer et al., 2021).

The global ocean is undersaturated with respect to bSi (Hurd, 1973; Nelson et al., 1995; Ragueneau et al., 2000; Tréguer and de la Rocha, 2013). Biogenic siliceous material dissolves as it sinks through the water column and only a fraction of the bSi exported from the surface ocean reaches the seafloor (Nelson and Gordon, 1982; Wollast and Mackenzie, 1981; Nelson et al., 1991; Nelson et al., 1995; Tréguer et al., 1995). Upon sedimentation, bSi is subject to further dissolution, although the regenerated dissolved H_4SiO_4 can reprecipitate into authigenic minerals (Mackenzie et al., 1981; Mackenzie and Kump, 1995; Michalopoulos and Aller, 1995; Loucaides et al., 2010; Ehlert et al., 2016) or adsorb to mineral surfaces (Huang, 1975; Sigg and Stumm, 1981; Swedlund, 1999; Davis et al., 2001; Davis et al., 2002). The competition between these processes determines the proportion of bSi that is reintroduced to the water column as a benthic flux of H_4SiO_4 and the proportion that is permanently buried. Tréguer and de la Rocha (2013) underlined the importance of the benthic bSi cycle by showing that most of the marine bSi dissolves at the sediment-water interface rather than during the sinking of silica particles.

Based on molar bSi:C ratios and particulate organic carbon (POC) burial rates, DeMaster (2002) updated global estimates of bSi burial on continental margins and demonstrated that the contribution of these regions to the global Si budget is more important than previously thought. Since then, more studies have focused on the margins, e.g. in the North Atlantic (Schlüter and Sauter, 2000; Schmidt et al., 2001; Oehler et al., 2015b). Yet, current global Si budgets still lack sufficient data, especially with regard to benthic fluxes and burial efficiencies, to accurately constrain bSi burial and dissolution on continental margins.

To our knowledge, no comprehensive study on bSi cycling has been published for the Skagerrak; a continental margin representing the largest depocenter for sediments transported from the North Sea (Eisma and Kalf, 1987; Van Weering et al., 1987; Van Weering et al., 1993). Here, we present a dataset on bSi cycling for the Skagerrak and quantify benthic fluxes of H_4SiO_4 , bSi rain rates to the seafloor, burial fluxes and burial efficiencies using a combination of geochemical observations and modelling. We propose a novel model for bSi cycling in Skagerrak sediments with potential broader applicability. We compare our results with other regions to highlight bSi cycling in the Skagerrak in a global context.

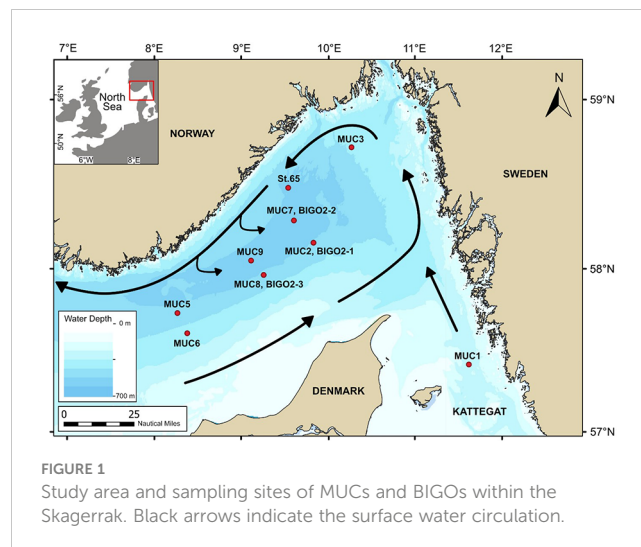


FIGURE 1
Study area and sampling sites of MUCs and BIGOs within the Skagerrak. Black arrows indicate the surface water circulation.

2 Study area

The Skagerrak strait is located between Denmark, Norway and Sweden and links the North Sea and the Kattégat with maximum water depths of approximately 700 m (Figure 1). Surface waters in the Skagerrak circulate anticlockwise. Water from the North Sea enters the Skagerrak through the Jutland Current from the south, which, together with the Baltic Current, results in the outflowing Norwegian Coastal Current leaving the Skagerrak to the north (Van Weering et al., 1987; Otto et al., 1990). Annual total sediment deposition in the Skagerrak is $46 \cdot 10^6$ tons yr^{-1} (De Haas and Van Weering, 1997). Sediment composition varies between sand (<40% clay), mainly along the Danish coast, and fine-grained silt and clay sediments in the deeper parts (Stevens et al., 1996). Substantial animal burrows are present in the upper 10 - 20 cm in the fine-grained areas (Canfield et al., 1993a; Kristensen et al., 2018). Skagerrak sediments are characterized by a large lateral input of mostly lithogenic material from the North Sea (Van Weering et al., 1993; De Haas and Van Weering, 1997). Nutrient supply from external waters and local rivers support an annual primary productivity of $130 \text{ g C}_{\text{org}} \text{ m}^{-2} \text{ yr}^{-1}$ (Beckmann and Liebezeit, 1988). Both siliceous (diatoms) and carbonaceous (haptophytes) frustule-building phytoplankton are present in the Skagerrak, with diatoms being dominant during the early bloom in February-March and a second bloom in May-June (Gran-Stadniczeňko et al., 2019).

3 Materials and methods

Data are presented from nine stations (65 to 677 m water depth) visited over two sampling campaigns, AL557 and AL561, with R/V Alkor in June and August 2021, respectively (Schmidt, 2021; Thomas et al., 2022). Only data from St. 65 are from June 2021 (Figure 1; Table 1). At each station, a short sediment core was recovered (<50cm) using a multiple-corer (MUC). Sediments at these sites were dominated by fine-grained material except MUC6, which was taken from a more sandy area (Table 1). At three stations during AL561, autonomous benthic landers (Biogeochemical

TABLE 1 Summary of sampling sites and general information on multicorer (MUC) and biogeochemical observatory (BIGO) stations.

Station	Latitude N	Longitude E	Water depth (m)	Porosity ^a	Grain sizes ^b D ₅₀ (μm)	bSi ^a (%)	TOC ^a (%)	CaCO ₃ ^a (%)	Lithogenic ^c material (%)
MUC1	57° 27.030'	11° 30.480'	65	0.75	11	3.2	1.7	12.0	80
MUC6	57° 38.086'	08° 23.998'	185	0.50	94	0.6	0.6	6.3	91
MUC3	58° 44.876'	10° 13.437'	215	0.77	11	2.5	2.0	14.1	78
MUC5	57° 45.191'	08° 17.173'	434	0.77	9	2.6	2.1	13.8	78
MUC8	57° 59.286'	09° 14.305'	490	0.77	14	2.0	2.5	14.0	77
MUC2	58° 10.884'	09° 47.624'	500	0.78	9	2.5	2.5	13.4	77
St. 65	58° 30.068'	09° 29.887'	530	0.79		2.1	2.2	12.5	79
MUC9	58° 04.352'	09° 05.736'	604	0.80	8	2.2	2.1	11.4	80
MUC7	58° 18.785'	09° 34.335'	677	0.82	7	2.3	2.1	11.4	81
BIGO-2-1	58° 10.969'	09° 47.423'	502						
BIGO-2-2	58° 18.778'	09° 34.362'	678						
BIGO-2-3	57° 59.220'	09° 14.300'	490						

^aPorosity, and contents of bSi (as wt% SiO₂), total organic carbon (TOC) and CaCO₃ represent mean values for the whole core.

^bGrain sizes are given as D₅₀ (50% of particles are smaller than this value), provided as average of all samples from the core.

^cLithogenic material was calculated as: Lithogenic material = 100 - (bSi + CaCO₃ + 2.8 · TOC), where the factor 2.8 converts the mass of TOC to total particulate organic matter (Sayles et al., 2001).

Observatories, BIGO) were deployed to determine benthic H₄SiO₄ fluxes.

3.1 Sampling

Sediment samples for porosity and solid phase analyses from the MUC were taken every centimetre and subsequently stored refrigerated at 4°C. An additional core from each station was subsampled at the same resolution for ²¹⁰Pb radionuclide and grain size analyses. From a third core, sediment for porewater analysis was sampled in an argon-filled glove bag in a refrigerated laboratory adjusted to bottom water temperatures (ca. 7°C). The samples were centrifuged at 4000 rpm at 8°C for 20 min to separate porewater from the solid phase. Subsequently, the supernatant water was filtered through a 0.2 μm cellulose-acetate syringe filter inside the glove bag. Porewater from St. 65 (AL557) was sampled with syringes and rhizones (0.15 μm pore size).

3.2 In-situ flux measurements

For the measurement of total solute fluxes, the GEOMAR BIGO type lander (Biogeochemical Observatory) was deployed as described in detail by (Sommer et al., 2009). BIGO-2 contained two circular flux chambers (internal diameter 28.8 cm area 651.4 cm²). A TV-guided launching system allowed smooth placement of the observatory on the sea floor. Approximately 2 hours (BIGO-2-1) or 4 hours (BIGO-2-2 & -3) after the observatories were placed on the seabed, the chambers were slowly driven into the sediment (~30 cm h⁻¹). During this initial period, where the bottom of the chambers was not closed by the sediment, the water inside the flux chamber was

periodically replaced with ambient bottom water. The water body inside the chamber was replaced once more with ambient bottom water after the chamber had been driven into the sediment to flush out solutes that might have been released from the sediment during chamber insertion. To determine fluxes of H₄SiO₄, eight sequential water samples were removed with glass syringes (volume ~ 47 ml) using syringe water samplers. The syringes were connected to the chamber using 1 m long Vygon tubes with a dead volume of 5.2 ml. Prior to deployment, these tubes were filled with distilled water.

3.3 Analytical methods

Porewater H₄SiO₄ analysis was done onboard immediately after filtering using standard methods (Grasshoff et al., 1999) with a Hitachi U-2001 spectrophotometer. The analytical precision of the analysis was better than 2 μmol L⁻¹. The porewater samples from St. 65 (AL557) were stored frozen until they were analyzed for H₄SiO₄ six months after sampling with a TECAN infinite 200 plate reader at 810 nm according to the method by Ringuet et al. (2011). Porewater sub-samples for total dissolved iron (Fe), potassium (K) and lithium (Li) were acidified with 10 μL of suprapure concentrated HNO₃ per mL of sample and stored refrigerated at 4°C in the dark for later analyses at GEOMAR by inductively coupled plasma optical emission spectroscopy (ICP-OES, Varian ICP 720-ES). A seawater standard (IAPSO) was used for quality control. The analytical accuracy was better than 2%.

Porosity was determined from the loss of water after freeze-drying. Biogenic silica contents were analyzed on freeze-dried and homogenized sediment by a wet-leaching procedure based on the method of Müller and Schneider (1993) using 1 M NaOH as the leaching solution. An internal sediment standard was used for quality control and the analytical accuracy was better than 7%.

The presented bSi contents are given as wt% SiO₂. This method may underestimate bSi content if particles are coated with metal oxides that protect bSi from dissolution or if bSi is incorporated into authigenic clay minerals (Michalopoulos and Aller, 2004; Rahman et al., 2017). Biogenic silica may also be overestimated if coexisting non-biogenic compounds are extracted by the alkaline leach (DeMaster, 1991; Kamatani and Oku, 2000). In particular, for sediments with high contents of lithogenic silica, alkaline leaching of lithogenic silica can interfere with the measured bSi content (Ragueneau and Tréguer, 1994; Zhu et al., 2023). Total carbon (TC) and total organic carbon (TOC) were analyzed on freeze-dried and homogenized sediment using a EuroEA 3000 element analyzer. For TOC analysis, sediments were first decarbonized with 0.25 N HCl on a hotplate at ~70°C. Total inorganic carbon (TIC) was calculated by subtracting TOC from TC. TIC values are reported as weight percent CaCO₃ assuming that all TIC occurs as calcium carbonate. The analytical accuracy was better than 5% for TOC and TC based on repeated analyses of certified reference material BSTD1 (soil standard).

Analysis of ²¹⁰Pb was carried out at stations MUC1, MUC3, MUC5, MUC9 and St.65. Freeze-dried and homogenized sediment was embedded into containers sealed with a two-component epoxy resin. Samples were left standing for two weeks to ensure steady state equilibration between ²²⁶Ra and ²¹⁴Bi. Activities of total ²¹⁰Pb were measured *via* its peak intensity at 46.5 keV by gamma spectrometry on n-type planar or coaxial Canberra Ge-detectors at GEOMAR (MUC5), Göttingen (St.65) and IAF Dresden (MUC1, MUC3, MUC9). Total ²¹⁰Pb activities were corrected for the natural background decay of ²²⁶Ra at 295 keV in marine sediments to obtain excess ²¹⁰Pb (²¹⁰Pb_{ex}) values.

3.4 Grain size determination

Bulk grain size distributions were determined for each station except St. 65. Samples were treated with 10% - 30% hydrogen peroxide and 60% acetic acid to dissolve organic and carbonate compounds. Subsequently, samples were dispersed in water using tetra-sodium diphosphate decahydrate. Grain size distributions were determined at the CEN, University of Hamburg, with a laser-diffraction particle-sizer (Sympatec HELOS/KF Magic; range 0.5/18 to 3500 μm). Accuracy of measurements and absence of a long-term instrumental drift was ensured by regular analysis of an in-house standard (standard deviation for mean grain size and D₅₀ over the analysis period was < 1.1 μm). Statistical evaluation of the grain size distribution was based on the graphical method (Folk and Ward, 1957), calculated using GRADISTAT software (Blott and Pye, 2001).

3.5 Steady-state calculations

Biogenic silica burial fluxes (bSi_{bur}) were calculated as follows:

$$bSi_{bur} = ds(1 - \phi) SR bSi \quad (1)$$

Where *ds* is the density of dry solids, ϕ is the mean porosity of the whole sediment core, *bSi* is the mean bSi content of each sediment core in wt% SiO₂ and *SR* is the sediment accumulation or sedimentation rate in compacted sediments. Mean bSi contents were used due to scattering of bSi and an absence of clear trends with sediment depth. *SR* was either determined from ²¹⁰Pb_{ex} data (Eq. 7) or taken from literature values close to the station (Table 2).

At BIGO stations, benthic fluxes of H₄SiO₄ (*J_{Si}*) were calculated from least-square linear regression fits of concentration versus time plots of the data obtained by benthic chamber in-situ measurements and the height of overlying water in the benthic chamber (Dale et al., 2021). The H₄SiO₄ benthic fluxes represent the mean flux of both benthic chambers at each station. At non-BIGO stations, porewater gradients of H₄SiO₄ were used to calculate H₄SiO₄ benthic fluxes by applying Fick's First Law. This approach only considers diffusive transport and does not account for additional non-local transport by bioirrigation, in contrast to in-situ BIGO measurements that provide the total flux. The enhancement of solute transport by bioturbation and bioirrigation was obtained by calculating a correction factor, α (Stahl et al., 2004). The factor (mean $\alpha = 2.2$) was calculated as the ratio between benthic fluxes determined from in-situ measurements and porewater gradients at the three BIGO stations. The correction factor was then applied to the diffusive flux at non-BIGO stations to obtain total H₄SiO₄ benthic fluxes:

$$J_{Si} = -\phi D_{sed} \frac{d[C]}{dx} \alpha \quad (2)$$

Where *J_{Si}* is the H₄SiO₄ benthic flux, ϕ is porosity, $d[C]/dx$ is the concentration gradient between porewater (taken at 0.5cm) and bottom water (taken from overlying water of the MUCs), and *D_{sed}* is the molecular diffusion coefficient of H₄SiO₄ corrected for tortuosity ($0.58 \cdot 10^{-5} \text{ cm}^2 \text{ s}^{-1}$) and bottom water temperatures (7°C) taken from Rebreanu et al. (2008). H₄SiO₄ fluxes were always directed out of the sediment. In this study, they are reported as positive numbers.

Rain rates of bSi (bSi_{rr}) were then calculated as the sum of bSi burial fluxes and H₄SiO₄ benthic fluxes:

$$bSi_{rr} = bSi_{bur} + J_{Si} \quad (3)$$

The percent bSi burial efficiency (bSi_{be}) was then calculated as follows:

$$bSi_{be} = \frac{bSi_{bur}}{bSi_{rr}} 100\% \quad (4)$$

3.6 Model description

To simulate bSi turnover, we used a steady-state numerical transport-reaction model for dissolved H₄SiO₄ and particulate bSi and ²¹⁰Pb_{ex}. The following one-dimensional partial differential equations (Berner, 1980) were applied to solve for the concentration profiles of H₄SiO₄ and bSi:

TABLE 2 Measured and modelled results for bSi cycling.

Measured and modelled data	MUC1 65m	MUC6 185m	MUC3 215m	MUC5 434m	MUC8 490m	MUC2 500m	St.65 530m	MUC9 604m	MUC7 677m
H ₄ SiO ₄ bottom water (μmol L ⁻¹)	7.5	0.0	7.1	15.0	12.5	10.5	22.6 ^a	20.2	22.6
bSi content (wt%) ^b	3.2	0.6	2.5	2.6	2	2.5	2.1	2.2	2.3
Sedimentation rate, SR (cm yr ⁻¹)	0.14	0.12 ^c	0.14	0.32	0.21 ^c	0.38 ^c	0.26	0.19	0.24 ^c
bSi burial flux (mmol m ⁻² yr ⁻¹)	460	150	340	790	390	850	490	360	410
H ₄ SiO ₄ benthic flux (mmol m ⁻² yr ⁻¹)	670 ^d	30 ^d	600 ^d	500 ^d	660	420	510 ^d	450 ^d	590
Bioirrigation (%) ^e	43	33	43	42	45	43	41	49	42
bSi rain rate, bSi _{rr} (mmol m ⁻² yr ⁻¹) ^f	1130	180	940	1290	1060	1270	1070	1000	1000
bSi burial efficiency, bSi _{be} (%) ^f	41	83	36	61	37	67	49	44	41
Integrated bSi dissolution (mmol m ⁻² yr ⁻¹) ^f	670	30	600	500	660	420	510	450	590
Kinetic constant, B ₀ (mmol cm ⁻³ B ₂ yr ⁻¹)	0.40	0.06	0.28	0.26	0.35	0.26	0.32	0.27	0.42
Kinetic constant, B ₁ (cm)	2.3	2.7	2.3	2.3	2.3	2.3	2.4	2.5	2.4
Kinetic constant, B ₂ (-)	-2.6	-1.5	-2.4	-2.5	-2.5	-2.3	-2.4	-2.5	-2.6
Bioturbation coefficient, D _{B0} (cm ² yr ⁻¹)	50	26	52	6	20	25	5	13	15
Bioturbation coefficient, x _B (cm)	20.0	4.0	19.5	3.0	2.5	2.3	1.5	3.0	3.0
Bioirrigation coefficient, D ₁₀ (yr ⁻¹)	39	24	33	22	33	19	17	32	23
Bioirrigation coefficient, x ₁ (cm)	4.1	1.5	4.3	6.0	5.7	6.3	6.9	6.9	8.0
Porosity at sediment-water interface, φ ₀ (-)	0.85	0.50	0.86	0.87	0.89	0.89	0.90	0.88	0.89
Porosity in compacted sediment, φ _c (-)	0.73	0.49	0.76	0.74	0.75	0.77	0.76	0.78	0.80
Attenuation coefficient, px (cm)	0.12	0.10	0.11	0.10	0.14	0.17	0.15	0.13	0.11

^aBottom water H₄SiO₄ concentration at St. 65 was taken from the adjacent station MUC7.

^bBiogenic silica contents are given as mean values for the whole sediment core.

^cSedimentation rates were taken from reported values measured in cores close to our stations: MUC2: Van Weering et al. (1993); MUC6: Van Weering et al. (1987); MUC7: Paetzel et al. (1994); MUC8: Erlenkeuser and Pederstad (1984).

^dH₄SiO₄ benthic fluxes were calculated by the diffusive fluxes and the correction factor α (Eq. 2).

^eBioirrigation contribution to total H₄SiO₄ benthic fluxes.

^fIn the model, bSi rain rates and burial efficiencies were prescribed using field data (Eq. 14, 15). Consequently, depth-integrated net bSi dissolution rates were also predefined and equal to the H₄SiO₄ benthic flux.

Stations are arranged with increasing water depth. Si fluxes in the table are rounded to two significant figures.

$$ds(1 - \phi) \frac{\delta bSi}{\delta t} = \frac{\delta}{\delta x} \left(ds(1 - \phi) D_B \frac{\delta BSi}{\delta x} - ds(1 - \phi) us BSi \right) - ds(1 - \phi) RbSi \tag{5}$$

$$\phi \frac{\delta H_4SiO_4}{\delta t} = \frac{\delta}{\delta x} \left(\phi D_s \frac{\delta H_4SiO_4}{\delta x} - \phi upw H_4SiO_4 \right) + \phi D_I (BW - H_4SiO_4) + \phi RbSi \tag{6}$$

where x is sediment depth, t is time, ds is the density of dry solids (= 2.5 g cm⁻³), φ is porosity, D_s is the tortuosity-corrected molecular diffusion coefficient, D_B is the bioturbation coefficient, D_I is the bioirrigation coefficient, BW denotes the concentration of H₄SiO₄ in ambient bottom water, us and upw are burial velocities for solids and porewater, respectively, and RbSi is the rate of bSi dissolution.

The following equation was used to simulate particle-bound ²¹⁰Pb_{ex} and derive burial velocities and bioturbation rates:

$$ds(1 - \phi) \frac{\delta Pb_{ex}}{\delta t} = \frac{\delta}{\delta x} \left(ds(1 - \phi) D_B \frac{\delta Pb_{ex}}{\delta x} - ds(1 - \phi) us Pb_{ex} \right) - ds(1 - \phi) \lambda Pb_{ex} \tag{7}$$

where Pb_{ex} is the excess activity of ²¹⁰Pb in sediments and λ is the ²¹⁰Pb decay constant (0.031 yr⁻¹).

Sediment compaction was considered in the model by fitting the following function to the porosity data at each station:

$$\phi = \phi_c + (\phi_0 - \phi_c) e^{-\frac{x}{px}} \tag{8}$$

where φ₀ is the porosity at the sediment surface, φ_c is the porosity in compacted sediment and px is the attenuation coefficient. Burial velocities of solids and solutes were then described as:

$$us = \frac{SR(1 - \phi_c)}{(1 - \phi)} \tag{9}$$

$$upw = \frac{SR\phi_c}{\phi} \tag{10}$$

The distribution of $^{210}\text{Pb}_{\text{ex}}$ in surface sediments was used to constrain bioturbation rates:

$$D_B = D_{B0} \exp\left(-\frac{x^2}{2x_B^2}\right) \tag{11}$$

where D_{B0} is the bioturbation coefficient at the sediment-water interface and x_B controls the bioturbation mixing depth.

Flushing of animal burrows by bioirrigation was mathematically described as:

$$D_I = D_{I0} \exp\left(-\frac{x^2}{2x_I^2}\right) \tag{12}$$

where D_{I0} is the bioirrigation coefficient at the sediment-water interface and x_I controls the bioirrigation depth. Bioirrigation rates were constrained from porewater H_4SiO_4 concentrations.

Most models that simulate bSi dissolution use a rate that decreases rapidly with sediment depth to simulate porewater H_4SiO_4 concentrations (e.g. McManus et al., 1995; Rabouille et al., 1997; Khalil et al., 2007). Three different explanations have been put forward to explain the decrease in bSi dissolution; (i) saturation control by high porewater H_4SiO_4 levels (McManus et al., 1995; Rabouille et al., 1997), (ii) different phases of bSi undergoing dissolution, each with its own reactivity (Boudreau, 1990; Archer et al., 1993; McManus et al., 1995), and (iii) retardance of dissolution due to progressive coating of particle surfaces by Al-containing minerals (Kamatani et al., 1988; McManus et al., 1995). Since the underlying mechanisms are unknown in Skagerrak sediments, we made no attempt to resolve the different factors controlling the rate decline with depth. Instead, we applied an empirical power law model to simulate the combined effect of these processes, which implicitly includes dissolution, reprecipitation and adsorption, and which we refer to as the depth-dependent net bSi dissolution rate (RbSi). Thus, true bSi dissolution rates may be higher if large amounts of H_4SiO_4 are removed from porewaters by adsorption or reprecipitation.

Based on previous modelling work (Middelburg, 1989; Boudreau and Ruddick, 1991; Boudreau et al., 2008), Stolpovsky et al. (2015) proposed a depth-dependent power function to simulate POC degradation in bioturbated sediments:

$$\text{RPOC} = B_0 (x + B_1)^{B_2} \tag{13}$$

where RPOC is the rate of POC degradation with sediment depth ($\text{mmol cm}^{-3} \text{yr}^{-1}$) and B_0 ($\text{mmol cm}^{-3-B_2} \text{yr}^{-1}$), B_1 (cm) and B_2 (dimensionless) are parameters defining the decrease in POC degradation rate. Given that the rates of both bSi dissolution and POC degradation decrease with sediment depth, we assumed that the power law from Stolpovsky et al. (2015) is applicable to describe bSi dissolution kinetics:

$$\text{RbSi} = B_0 (x + B_1)^{B_2} \tag{14}$$

where RbSi is the depth-dependent net rate of bSi dissolution ($\text{mmol cm}^{-3} \text{yr}^{-1}$). The values of B_0 , B_1 and B_2 for bSi dissolution are

expected to be different to those for POC degradation that were constrained from a global database of benthic oxygen and nitrate fluxes (Stolpovsky et al., 2015). Following Stolpovsky et al. (2015), we assume that the rain rate of bSi to the seafloor provides an upper limit of the total amount of bSi available for dissolution and that its burial efficiency describes the limit of bSi preservation. Consequently, bSi rain rates and burial efficiencies are model input parameters. If bSi_{rr} and bSi_{be} are known, the value of one of the parameters B_0 , B_1 or B_2 can be determined from the other two, i.e. for B_0 :

$$B_0 = \frac{((1 + B_2)(-100 + \text{bSi}_{\text{be}}) \text{bSi}_{\text{rr}})}{(100(B_2^{1+B_2} - B_1(B_1 + L)^{B_2} - L(B_1 + L)^{B_2}))} \tag{15}$$

where L is the depth at which bSi no longer dissolves. Since the depth profiles of bSi and H_4SiO_4 indicate that bSi dissolution stops near the bottom of the sediment core, we assumed L to be 50 cm.

Two adjustable parameters, B_1 and B_2 , are required to simulate net bSi dissolution. These were constrained from the measured bSi and H_4SiO_4 data. In total, 10 parameters (B_1 , B_2 , SR, ϕ_0 , ϕ_c , p_x , D_{I0} , x_I , D_{B0} , x_B) were adjusted to fit the porosity, $^{210}\text{Pb}_{\text{ex}}$, H_4SiO_4 and bSi data. The parameters SR, D_{B0} and x_B were evaluated by a Monte Carlo procedure (2000 runs) after Dale et al. (2021). For each model run, the parameters were randomly varied from 0.05 to 4 cm yr^{-1} for SR, 1 to 30 $\text{cm}^2 \text{yr}^{-1}$ for D_{B0} and 1 to 3 cm for x_B . These ranges were chosen based on literature values and visual identification of the bioturbation layer in the $^{210}\text{Pb}_{\text{ex}}$ data. At stations MUC1 and MUC3, no clear bioturbation layer was observed. Here, ranges of D_{B0} and x_B for the Monte Carlo runs were extended to 1 - 150 $\text{cm}^2 \text{yr}^{-1}$ and 1 - 30 cm, respectively.

Upper boundary conditions were set as constant concentrations for H_4SiO_4 and constant particulate fluxes to the seafloor for bSi. For $^{210}\text{Pb}_{\text{ex}}$, a constant rate of supply was set at the upper boundary from the measured integrated $^{210}\text{Pb}_{\text{ex}}$ activity in the sediment and the decay constant λ :

$$\text{FPb}_{\text{ex}} = \lambda ds (1 - \phi_0) \int_0^\infty \text{Pb}_{\text{ex}} dx \tag{16}$$

where FPb_{ex} is the flux of $^{210}\text{Pb}_{\text{ex}}$ to the seafloor. At the lower boundary at 50 cm, a zero gradient condition was imposed for all model variables.

Model input data for each station are summarized in Table 2. The model was run to steady-state using the partial differential equation solver implemented in Mathematica 12.2. Mass conservation > 99% was achieved in all model runs. Mean deviations between the model and measured H_4SiO_4 , bSi and $^{210}\text{Pb}_{\text{ex}}$ data, expressed as root mean square errors, were 27 $\mu\text{mol L}^{-1}$, 0.4 wt% and 0.6 dpm g^{-1} , respectively.

4 Results

4.1 $^{210}\text{Pb}_{\text{ex}}$ activities, sedimentation rates and mixing

At stations MUC5, MUC9 and St.65, $^{210}\text{Pb}_{\text{ex}}$ activities decreased exponentially below the upper centimetres where $^{210}\text{Pb}_{\text{ex}}$ activities tended to be more constant (Figure 2). In contrast, stations MUC1 and MUC3 showed an almost linear $^{210}\text{Pb}_{\text{ex}}$ decrease throughout

the whole core. Sedimentation rates determined by the model ranged from 0.14 to 0.32 cm yr⁻¹ (Table 2) and varied with water depth, but no clear trend was observed.

Sediment mixing by bioturbation was limited to the upper 5 - 10 cm at most sites (Figure 3C, Eq. 11) with bioturbation coefficients (D_{B0}) and mixing coefficients (x_B) of 5 - 26 cm² yr⁻¹ and 1.5 - 4.0 cm, respectively (Table 2). In contrast, bioturbation was predicted to proceed throughout the whole sediment core at stations MUC1 and MUC3. At these sites, extreme D_{B0} and x_B values in the range of 50 cm² yr⁻¹ and 20 cm, respectively, were required to simulate the measured ²¹⁰Pb_{ex} data. Bioirrigation was most intense in the upper layers and extended up to 5 - 20 cm depth, with deeper bioirrigation observed with increasing water depth (Figure 3C).

4.2 Dissolved and solid phase geochemistry

Sediments were generally fine-grained with mean D_{50} values ranging from 7 to 14 μm (Table 1). MUC6 contained a higher proportion of sand with a D_{50} of 94 μm. Solid phase bSi contents varied between 1 and 5 wt%, except for the sandy station MUC6, where lower values of 0.4 - 0.9 wt% were observed (Figure 3A). Biogenic silica showed no clear trend with sediment depth. CaCO₃ varied from ca. 6 - 13% and lithogenic material from 77 - 91% (Table 1).

H₄SiO₄ concentrations in the Skagerrak generally increased with sediment depth up to asymptotic concentrations of 300 - 660 μmol L⁻¹ (Figure 3B). Although asymptotic values were not reached at every station, the data indicate that the maximum asymptotic concentrations decreased with water depth. Furthermore, concentration plateaus, or local minima, were observed between 2 - 15 cm and were most pronounced at the deeper stations (>400 m). Total dissolved Fe concentrations were below the detection limit in the upper 5 - 10 cm, followed by an increase up to a maximum concentration of 40 - 135 μmol L⁻¹ (Figure 4A). Thereafter, dissolved Fe concentrations tended to decrease. The zone of increasing dissolved Fe concentrations coincided with the concentration plateaus of H₄SiO₄ at the deeper sites. Dissolved K and Li concentrations showed near-constant concentrations with sediment depth, ranging between 10 and 12

mmol L⁻¹ and 24 and 36 μmol L⁻¹, respectively, across the different sites (Figure 4B).

4.3 Si fluxes and burial efficiencies

Burial fluxes varied between 150 and 850 mmol m⁻² yr⁻¹. (Table 2). H₄SiO₄ fluxes at the three BIGO stations ranged between 420 and 660 mmol m⁻² yr⁻¹, which is similar to reported in-situ flux measurements of 360 - 609 mmol m⁻² yr⁻¹ in the Skagerrak basin between 411 - 682 m water depth (Hall et al., 1996). Benthic fluxes at stations that were calculated based on porewater gradients and the correction factor α for non-diffusive transport (Eq. 2) were between 450 - 670 mmol m⁻² yr⁻¹, except for the sandy station MUC6 that showed a lower rate of 30 mmol m⁻² yr⁻¹. Bioirrigation accounted for 33 - 49% of the total benthic H₄SiO₄ flux. Biogenic silica rain rates varied between 180 and 1290 mmol m⁻² yr⁻¹ with burial efficiencies of 36 - 83%. Fluxes and burial efficiencies varied between the different stations, but no clear trend was observed with water depth (Figure 5).

4.4 Modelled bSi dissolution

The model predicted that most of the net bSi dissolution takes place within the top sediment layer (Figure 3C). The adjustable parameters B_1 and B_2 that describe bSi dissolution varied between 2.3 and 2.5 cm, and -2.3 and -2.6 (dimensionless), respectively, excluding sandy station MUC6 (Table 2). At this station, B_1 and B_2 were 2.7 and -1.5, respectively. There were no clear trends in B_0 , B_1 and B_2 across different sites.

5 Discussion

5.1 Biogenic silica cycling

The major fluxes of bSi in Skagerrak sediments are investigated in this study through numerical interpretation of the geochemical data. In our approach, the down-core decrease in net bSi dissolution, bioirrigation and bioturbation can be quantified with our novel kinetic bSi dissolution model with reasonable confidence. The two adjustable parameters needed to describe bSi dissolution, B_1 and B_2 , show consistent values across the stations with fine-grained material, demonstrating the potential to apply the model elsewhere in comparable environments. Our kinetic description of bSi dissolution represents an alternative to previous approaches. It is especially applicable for regions where bSi dissolution rates and underlying mechanisms that control dissolution rates are unknown but where data for bSi rain rate and burial are available.

Biogenic silica burial fluxes were calculated based on ²¹⁰Pb derived sedimentation rates that are in the range of reported values from nearby locations (Erlenkeuser and Pederstad, 1984; Van Weering et al., 1993; Deng et al., 2020). A mean bSi burial efficiency of 50% in fine-grained sediments in the Skagerrak basin indicates that about half of the annual bSi raining onto the seafloor,

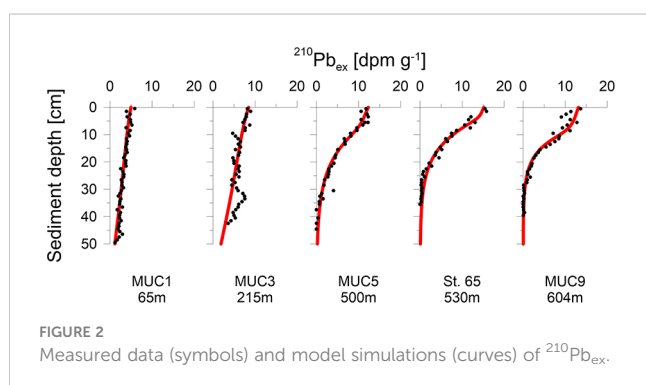
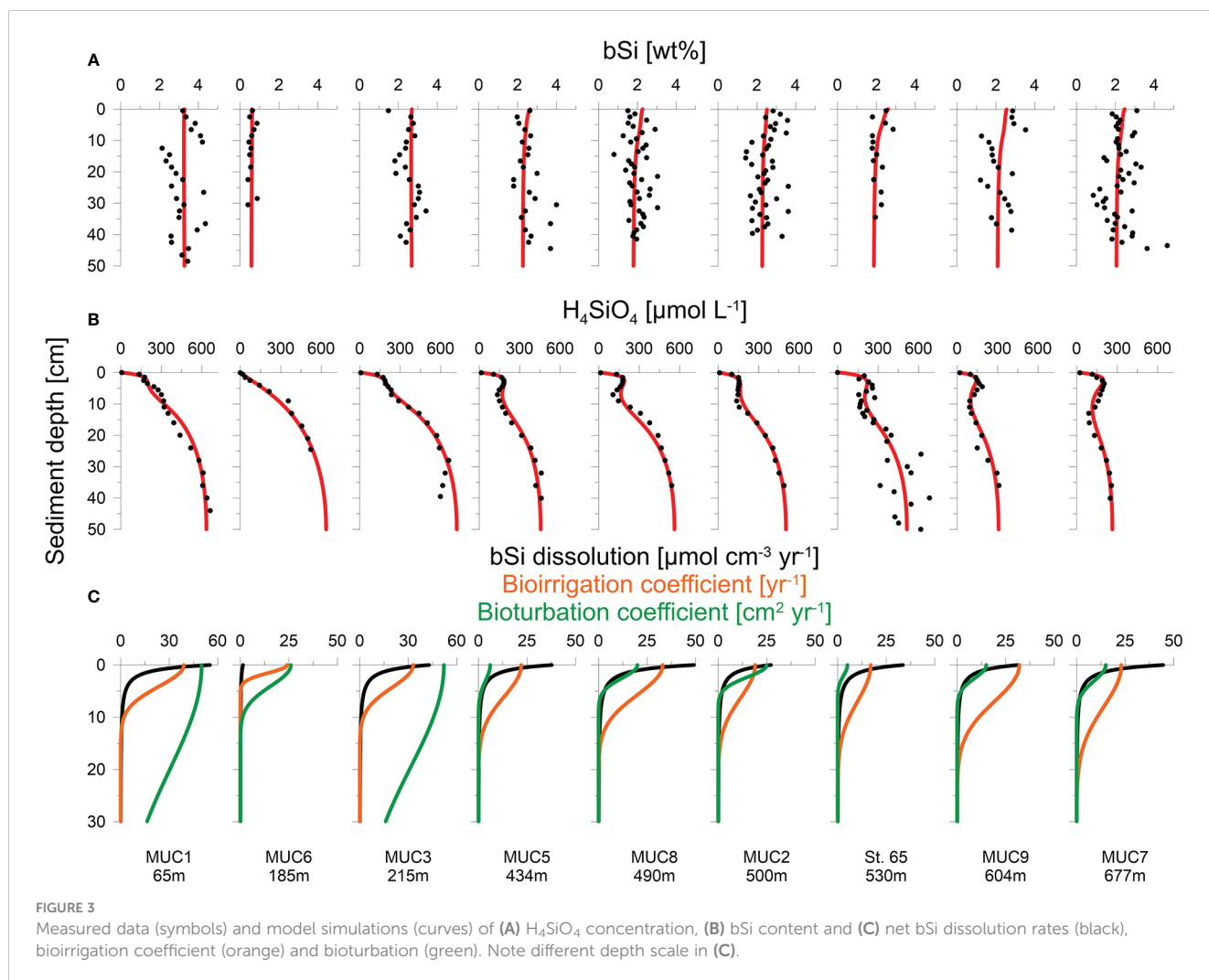


FIGURE 2
Measured data (symbols) and model simulations (curves) of ²¹⁰Pb_{ex}.

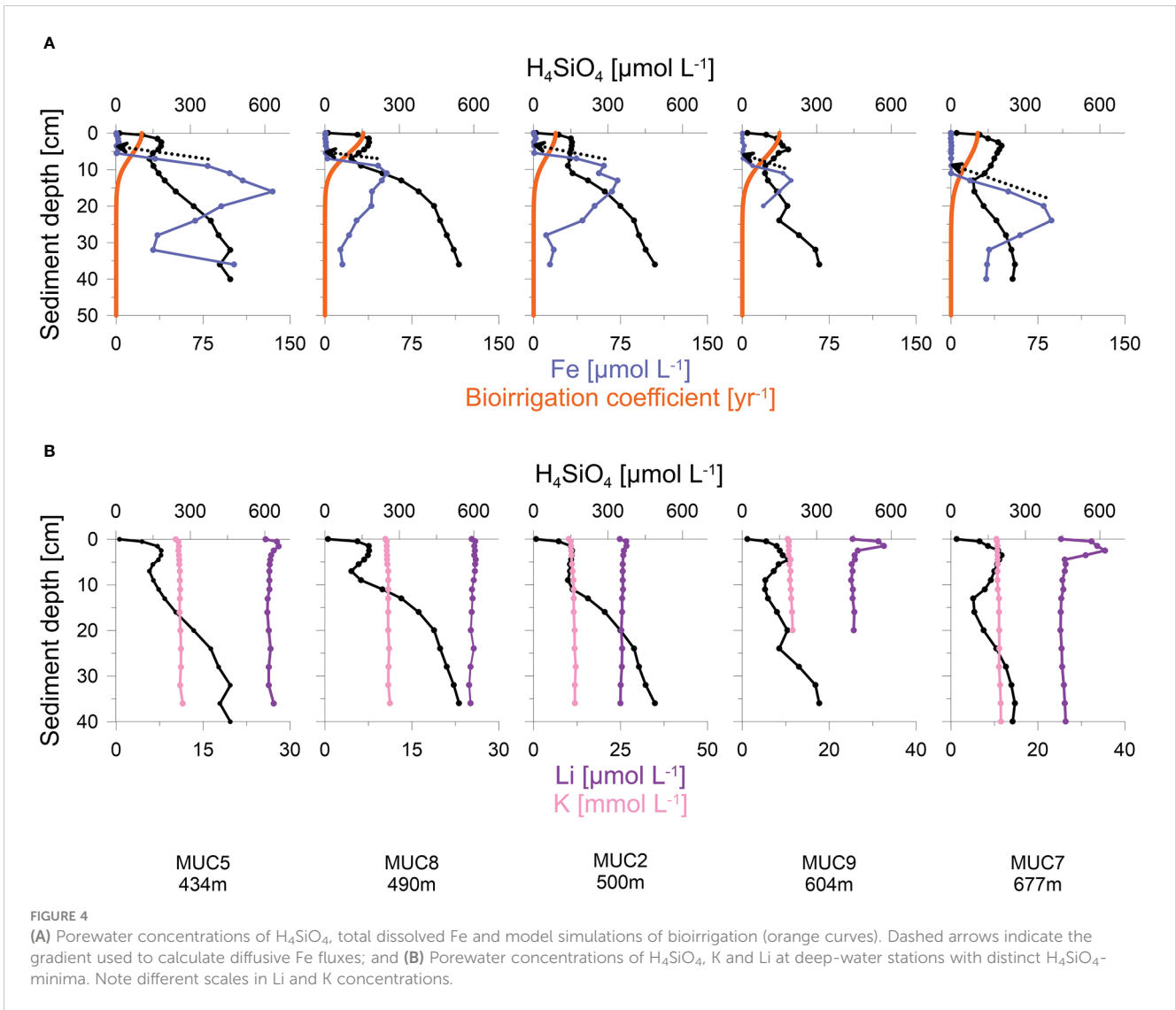


on average ca. 1100 mmol m⁻² yr⁻¹, is buried in the sediment. The other half dissolves and is subsequently returned to the water column. This calculation excludes the sandy station MUC6 and shallow stations MUC1 and MUC3, where non-exponential ²¹⁰Pb_{ex} profiles indicate substantial sediment mixing and more uncertainty in calculated bSi rain rates. The unusually intense and deep mixing rates required to simulate the data at MUC1 and MUC3 compared to other sites could be attributed to bottom trawling, which is frequent in the MUC1 region at water depths < 200 m (Sköld et al., 2018). Piles of sediment alongside furrows created by dragging trawl doors through the sediment can reach 1 to 2 dm in height (Bradshaw et al., 2021) and it may be that sediment cores taken at stations MUC1 and MUC3 penetrated such sediment piles during sampling. We find this explanation more convincing than deep bioturbation since no visual evidence of macrofauna was observed below ca. 10 cm depth during the slicing of the sediment cores. Furthermore, if the bioturbation parameters were set to values similar to those predicted for the undisturbed sites, sedimentation rates of 1 to 2 cm yr⁻¹ would be required to fit the ²¹⁰Pb_{ex} data (results not shown); a magnitude that seems unlikely for the Skagerrak. For a seafloor surface area of the Skagerrak basin below 200 m of about 10700 km², and assuming a mean bSi burial

flux of 550 mmol m⁻² yr⁻¹, we calculated a mean annual bSi burial flux of 6 · 10⁹ mol yr⁻¹.

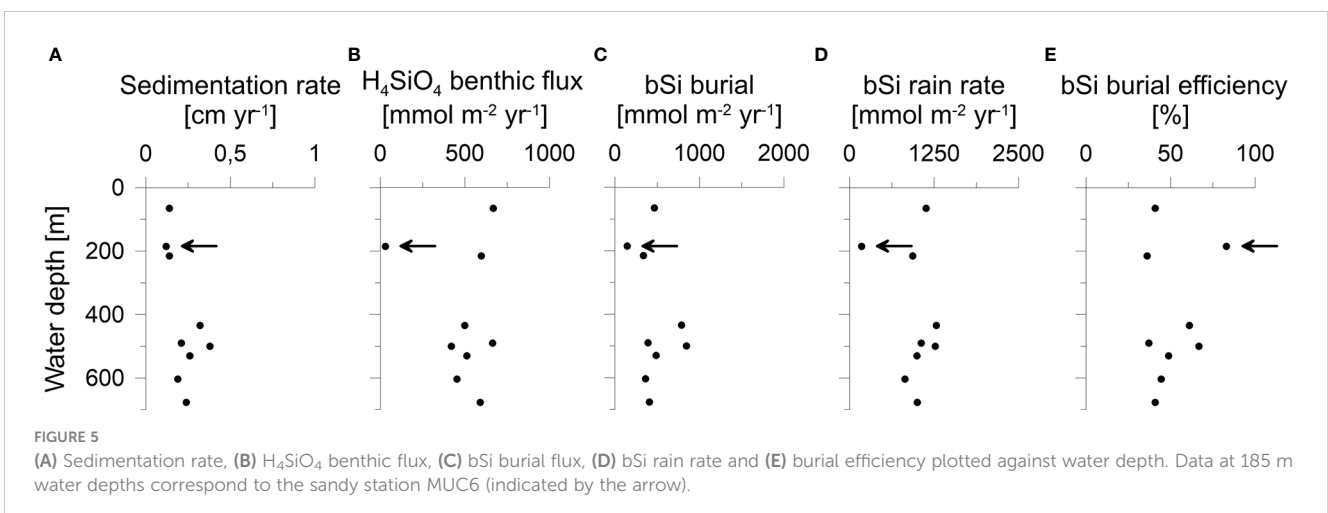
Ongoing dissolution of bSi leads to the accumulation of H₄SiO₄ in porewaters and drives H₄SiO₄ fluxes to the bottom water. Scanning electron micrographs have revealed evidence of chemical erosion of bSi in Skagerrak sediments (Meyenburg and Liebezeit, 1993). According to our model, the bulk of net bSi dissolution occurs in the upper 5 centimetres (Figure 3C). A steep down-core decline in the dissolution rate is mirrored by the steep increase in H₄SiO₄ concentrations below the sediment-water interface. This is not evident in the bSi profiles that show high variability, possibly due to interannual changes in bulk sediment supply and/or bSi rain rate. Our bSi contents are lower than ca. 6 wt % measured in a single sediment sample from the top 50 cm at 325 m water depth in the Skagerrak (Bohrmann, 1986). Meyenburg and Liebezeit (1993) and Fengler et al. (1994) reported 0.1 - 0.6 wt% bSi at water depths of 80 and 183 m, which is similar to the sandy station MUC6, although the sediment type in their data was not reported.

No obvious trends between Si fluxes and individual environmental factors were observed at our stations, despite sizeable differences in water depth. Aside from H₄SiO₄ concentration and temperature in the



water column, water depth typically controls the amount of bSi that reaches the seafloor (Nelson et al., 1995; Ragueneau et al., 2000; Tréguer and de la Rocha, 2013), with lower fluxes expected at deeper stations. However, apart from the sandy station MUC6, our data reveal

relatively little variability in Si fluxes across stations (Figure 5; Table 2). Additional input of bSi to the deeper sites from the North Sea may partly explain this observation (see section 5.3). We observed a higher bSi burial efficiency at MUC6 where the sediment was coarser, which



could point toward an additional control by grain size (Table 1; Figure 5E). Similar trends of burial efficiency and grain size were reported in sediments of the Helgoland mud area (Oehler et al., 2015b); a sedimentary depocenter in the southern North Sea.

To frame bSi cycling in the Skagerrak in a global context, we compared our bulk fluxes to sites compiled by Dale et al. (2021), most of which are from the deep-sea (>3000 m water depth). In general, the Skagerrak bSi burial data aligns well with the global trend (Figure 6) and extends the existing database towards continental margin sites of higher burial efficiencies and burial fluxes. The Skagerrak data is in the range of other continental margins, such as the Peru shelf (Froelich et al., 1988), the Helgoland mud area (Oehler et al., 2015b) and the Guaymas Basin (Geilert et al., 2020). Variations in bSi cycling among continental margins likely stem from a combination of different regional settings, i.e. water depth, surface productivity, sedimentation rates and sediment composition.

5.2 Controls on H_4SiO_4 porewater concentrations

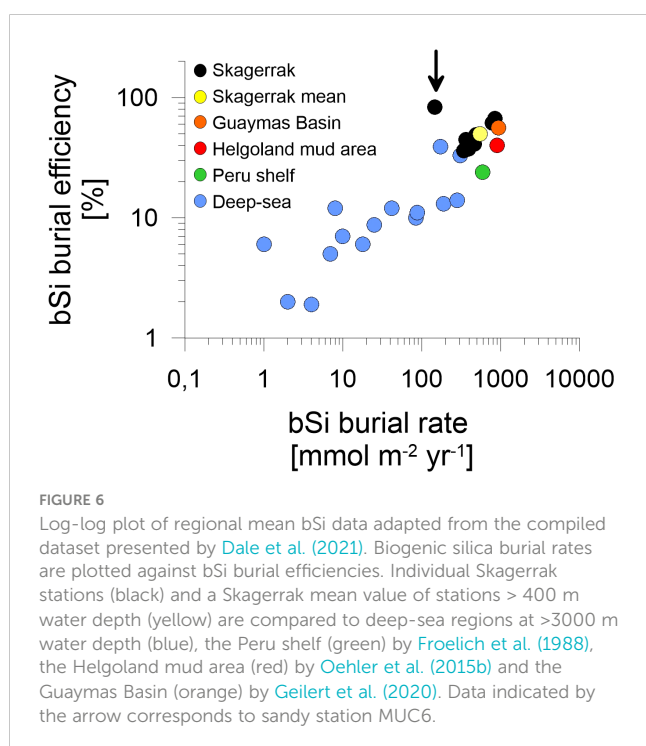
The rapid downcore increase in H_4SiO_4 to a constant asymptotic concentration observed in the Skagerrak is typical for fine-grained marine sediments (Hurd, 1973; Schink et al., 1974; McManus et al., 1995; Rabouille et al., 1997; Khalil et al., 2007; Oehler et al., 2015b; Ehlert et al., 2016). The tendency for a decrease in the maximum concentration with increasing water depth has been attributed to the proportion of lithogenic material present in the sediment, whereby the release of dissolved aluminium from clay minerals and subsequent adsorption or incorporation into the particulate silica phase lowers bSi solubility (Van Beusekom et al., 1997; Van Cappellen and Qiu, 1997; Dixit et al., 2001; Dixit and

Van Cappellen, 2003). At our sites, lithogenic material changed only slightly with water depth (Table 1). Consequently, inhibition of bSi dissolution by aluminium may not be a major control on H_4SiO_4 distributions in our sediment cores, although this conclusion would be strengthened with data on dissolved aluminium concentrations.

Especially at the deeper stations (>400 m), H_4SiO_4 porewater profiles were characterized by decreases or local minima of H_4SiO_4 concentrations between 2 - 15 cm (Figure 3B). The model predicts that the H_4SiO_4 minima are caused by more intense removal of H_4SiO_4 in discrete sediment layers by bioirrigation versus H_4SiO_4 added by bSi dissolution. It further predicts more intense bioirrigation at the deeper stations where the minima are more pronounced (Figure 3C). This is supported by reports that deep-water Skagerrak sediments are preferably colonized by deeply penetrating deposit feeders, whereas suspension and interface feeders are more abundant at shallower depths (Rosenberg et al., 1996; Dauwe et al., 1998). However, our results differ from Kristensen et al. (2018), who observed decreasing bioirrigation intensity with water depth. In addition, our modelled bioirrigation extends deeper than the zone where dissolved Fe is below the detection limit (Figure 4A). Given that bioirrigation enhances the transport of oxidants into the sediment, and that dissolved Fe is rapidly oxidized, this may indicate that our model overestimates the depth of bioirrigation. Hence, it might be possible that the H_4SiO_4 minima observed at deeper sites are not caused by deep-reaching bioirrigation. Instead, high rates of H_4SiO_4 removal by adsorption or precipitation might lower H_4SiO_4 concentrations in distinct sediment layers.

High sedimentary manganese and iron oxide contents have been reported for the Skagerrak, especially in deep waters (Canfield et al., 1993a; Thamdrup et al., 1994). Experimental studies have demonstrated that H_4SiO_4 can adsorb to iron oxides (Huang, 1975; Sigg and Stumm, 1981; Swedlund, 1999; Davis et al., 2001; Davis et al., 2002), suggesting that this process is a potential sink for H_4SiO_4 in Skagerrak sediments. Since the depths of local H_4SiO_4 minima and dissolved iron removal coincide (Figure 4A), this sink may be relevant for the formation of observed H_4SiO_4 minima. In order to estimate the significance of this process, potential H_4SiO_4 adsorption fluxes were calculated from the diffusive flux of Fe and then compared to H_4SiO_4 removal by bioirrigation integrated over the depth interval centred on the H_4SiO_4 minima, typically spanning 5 cm (Figure 4A). Under the assumption that all dissolved Fe is consumed by iron oxide formation, the diffusive flux of Fe (Eq. 2) was calculated at the depth where Fe increases. With a diffusion coefficient (D_{sed}) of $121 \text{ cm}^2 \text{ yr}^{-1}$ (Schulz and Zabel, 2006), dissolved Fe removal fluxes of $6 - 14 \text{ mmol m}^{-2} \text{ yr}^{-1}$ were calculated for the stations shown in Figure 4. Given mean H_4SiO_4 concentrations at the depth of H_4SiO_4 minima of ca. $150 \mu\text{mol L}^{-1}$ and assuming a pH value of 8, a molar Si/Fe sorption density on iron oxides of 0.2 is expected (Davis et al., 2002). Multiplying the Fe flux by the molar Si/Fe ratio yields potential H_4SiO_4 adsorption fluxes of $1 - 3 \text{ mmol m}^{-2} \text{ yr}^{-1}$. This corresponds to only 2 - 6% of bioirrigation fluxes integrated over the depth of H_4SiO_4 minima. It illustrates that H_4SiO_4 adsorption to iron oxides may be a feasible but probably not a dominant pathway of local H_4SiO_4 removal in Skagerrak sediments.

During authigenic clay formation, H_4SiO_4 is consumed in the presence of dissolved cations (e.g. K and Li) and weathered clay



minerals, a process referred to as reverse weathering (Mackenzie et al., 1981; Mackenzie and Kump, 1995; Loucaides et al., 2010; Ehlert et al., 2016). Reverse weathering occurs in anoxic marine sediments rich in biogenic opal and metal hydroxides (Michalopoulos and Aller, 1995; Michalopoulos et al., 2000) but has not been investigated in the Skagerrak. Previous studies have demonstrated reverse weathering to be a dominant process altering H_4SiO_4 porewater profiles on continental margins (Michalopoulos and Aller, 1995; Michalopoulos and Aller, 2004; Wallmann et al., 2008; Ehlert et al., 2016; Spiegel et al., 2021). Furthermore, the competition between bSi dissolution and reprecipitation during reverse weathering can create features similar to the H_4SiO_4 minima in our dataset (Spiegel et al., 2021). Since dissolved cations are consumed during reverse weathering, we compared the porewater distribution of H_4SiO_4 to dissolved K and Li concentrations (Figure 4B). The concentration profiles of K and Li are near-constant with sediment depth and, thus, do not show obvious indications for reverse weathering, i.e. uptake in the same horizon as H_4SiO_4 minima. However, we note that the effect of reverse weathering on dissolved cation concentrations, especially for K (~10 mM), might be too low relative to bulk porewater concentrations to result in distinct porewater signatures.

Since our model describes a continuous decline in net bSi dissolution with sediment depth, we are currently unable to resolve processes that operate in narrow depth intervals where the H_4SiO_4 porewater plateaus are observed. Hence, explicitly describing adsorption, precipitation and aluminium inhibition could lead to a better fit to the data and consequently to a more detailed description of the benthic bSi cycle in the Skagerrak. A model that includes these additional sinks may also be able to reproduce the data using shallower penetration depths of bioirrigation that are more consistent with those that have been previously reported by Kristensen et al. (2018). Further empirical data on these processes in Skagerrak sediments and elsewhere would benefit future bSi model descriptions. Since their parameterization is uncertain on the basis of our dataset and yet to be investigated in the Skagerrak, we reiterate that our simple model currently only provides net bSi dissolution rates.

5.3 Effect of lateral sediment supply on bSi/TOC ratios

DeMaster (2002) calculated global bSi burial on continental margins of $1.8 - 2.8 \cdot 10^{12} \text{ mol yr}^{-1}$ based on organic carbon burial and a mean molar bSi/TOC burial ratio of 0.6. However, DeMaster (2002) and subsequent budgets (Tréguer and de la Rocha, 2013; DeMaster, 2019) excluded continental margins dominated by high lithogenic loads, such as the Skagerrak, in their calculations. Rahman et al. (2016) and Rahman et al. (2017) reported bSi burial fluxes based on ^{32}Si isotopes in tropical and subtropical deltaic regions dominated by significant lithogenic inputs. Their results, when incorporated into more recent global Si budgets, point toward a greater significance of margins in total bSi burial (Tréguer et al., 2021). Since we have no ^{32}Si data, we compared our findings to previous approaches utilizing sedimentary bSi/TOC molar ratios.

Based on our field data (Table 1), we calculated a mean bSi/TOC molar ratio of 0.22 in Skagerrak sediments, which is distinctively lower than the mean ratio of 0.6 applied by DeMaster (2002). Other continental margin settings with high lithogenic inputs and where bSi was determined by the same method (Müller and Schneider, 1993) showed molar ratios of 1.0 for the Guaymas basin (Geilert et al., 2020), 0.52 for the Helgoland mud area (Oehler et al., 2015a; Oehler et al., 2015b) and 0.55 for the Amazon shelf (Spiegel et al., 2021). Dale et al. (2021) reported much lower ratios of 0.14 for the Peruvian oxygen minimum zone, which was attributed to a lack of bioturbation and enhanced bSi dissolution in undersaturated surface sediment layers. Tréguer et al. (2021) reported ratios between 2.4 and 11 in tropical and subtropical deltaic systems. The wide range in molar bSi/TOC ratios, presumably related to the differing diagenetic regimes, emphasizes the need to further study continental margins to better constrain global bSi burial.

Low bSi/TOC ratios in the Skagerrak may be confounded by the long transit time of sediments on the order of several months from the central North Sea and 1 - 2 years from the southern North Sea to the Skagerrak (Hainbucher et al., 1987; Dahlgaard et al., 1995; Salomon et al., 1995). Since seawater is undersaturated with respect to bSi, we hypothesize repeated resuspension and deposition during particle transport leads to more extensive bSi dissolution relative to POC. This may set the Skagerrak apart from most other continental margins with regard to the coupling between bSi and TOC. Similar regions need to be investigated in order to determine whether the Skagerrak is a unique environment with regards to bSi cycling or is emblematic of continental margins exhibiting protracted transport pathways.

6 Conclusions

In this study, we present bSi cycling in Skagerrak sediments based on geochemical data and modelling. Approximately $1100 \text{ mmol m}^{-2} \text{ yr}^{-1}$ of bSi annually rains onto the seafloor at the sites investigated, of which half dissolves in sediments and is reintroduced into the water column and half is permanently buried. Biogenic silica cycling shows some spatial variability across stations, which can probably be explained by differences in water depth, sedimentation rate and sediment grain size. At deep-water sites, we observed distinct minima or plateaus in the porewater profiles of H_4SiO_4 that are likely caused by deep-reaching bioirrigation rather than H_4SiO_4 removal by adsorption to iron oxide minerals. The contribution of reverse weathering to these minima remains uncertain. Biogenic silica cycling in the Skagerrak generally aligns with the global trends and shows comparable burial fluxes and burial efficiencies to other margins. However, Skagerrak sediments are characterized by distinctively lower molar bSi/TOC ratios compared to most other continental margins, which we suggest is due to decoupling between bSi and organic carbon during the long travel time of suspended matter transported from the North Sea. It is currently unclear whether Si cycling in the Skagerrak behaves similarly to other depocenters where sediment is first transported over large distances before it finally accumulates on the seafloor. The model presented here simulates bSi cycling and quantifies net bSi dissolution,

bioirrigation and bioturbation in Skagerrak sediments with few adjustable parameters but does not yet fully resolve the mechanisms behind H_4SiO_4 release and removal that operate in distinct and narrow depth intervals. Further work on the parameterization of bSi adsorption, reprecipitation and inhibition of bSi dissolution by aluminium would improve the predicted bSi kinetics. Our results contribute to the global dataset on bSi cycling and demonstrate the importance of continuing investigations in different regions, especially on continental margins, to better understand regional variability in bSi cycling and to constrain global Si budgets.

Data availability statement

The original contributions presented in the study are included in the article/Supplementary Material. Further inquiries can be directed to the corresponding author.

Author contributions

TS processed most of the samples in the laboratory and wrote the manuscript. TS, KW and AD developed the numerical model used in the study and contributed with discussions. NL helped with cruise planning, sample preparation in the laboratory and contributed with discussions. MS coordinated the research cruise Al561 and contributed with discussions. SS coordinated the biogeochemical observatory (BIGO) deployments and helped with evaluating the BIGO data. HK helped with the ^{210}Pb analyses and contributed with discussions. AP provided the porewater data for station 65. SL provided the grain-size data and helped with the interpretation. All authors contributed to the paper by drafting, reading, finalizing, and approving the text. All authors contributed to the article and approved the submitted version.

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Conflict of interest

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Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2023.1141448/full#supplementary-material>

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