



A Discontinuous Individual Growth Model of Swimming Crab *Portunus trituberculatus* and Its Application in the Nutrient Dynamic Simulation in an Intensive Mariculture Pond

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Environmental problems such as organic pollution and eutrophication caused by highly intensive mariculture activities constrain the sustainable and healthy development of industry. Therefore, it is necessary to quantify the nutrient dynamics of aquaculture animals in order to reduce the risk of environmental pollution. In this study, a discontinuous individual growth model of *Portunus trituberculatus* in an intensive mariculture pond of *P. trituberculatus*–*Penaeus japonicus*–*Sinonovacula constricta* was constructed based on a dynamic energy budget theory combined with the index of condition factor. This model better predicted the growth and molting behavior of *P. trituberculatus*, and an acceptable fit was obtained through model parameterization using the Add-my-Pet (AmP) method (mean relative error = 0.058, symmetric mean squared error = 0.007). Ten molts were simulated over 180 days and generally coincided with the recorded molt time points. Based on this model and *P. trituberculatus* populations, the dynamic processes of carbon, nitrogen, and phosphorus in ingestion, respiration, excretion, feces, residual feed, dead crabs, seeding, molt, and harvest were simulated. The carbon, nitrogen, and phosphorus ingested during the 180-day culture period were 4,938.57 kg ha⁻¹, 1,255.88 kg ha⁻¹, and 244.16 kg ha⁻¹, respectively. Carbon, nitrogen, and phosphorus removal by harvest accounted for 1.06%, 1.03% and 0.62% of the total ingestion, respectively, while carbon, nitrogen, and phosphorus removal by dead crabs accounted for 6.84%, 6.63%, and 4.04%, respectively, and carbon, nitrogen, and phosphorus released from residual feed into the water accounted for 41.43% of the total feed. The accurate simulation of molting behavior and nutrient dynamics in this study provides a theoretical basis for molting risk prevention and environmental stress assessment of *P. trituberculatus* and provides basic modules and data support for the construction of the integrated mariculture ecosystem model.

Keywords: intensive mariculture pond, nutrient dynamic simulation, discontinuous growth model, *Portunus trituberculatus* populations, molt, condition factor

INTRODUCTION

In order to secure fish supplies and reduce the global food crisis caused by the COVID-19 epidemic, intensive farming activities continue to increase around the world (Trottet et al., 2021). Intensive aquaculture involves high-density stocking and feeding of exogenous feed. The residual feed, feces, and excretion can have adverse effects on the aquaculture water bodies and the surrounding aquatic environment (e.g., water eutrophication, reduction of aquatic plant, and animal diversity) (Folke et al., 1988; Roy et al., 2020), and may also lead to food safety problems (e.g., drug residues and spread of pathogens) (Fan et al., 2011). Due to overfishing and resource decline in China, the degree of intensive mariculture of *P. trituberculatus* is increasing and the annual yield exceeds 100,000 tons (China Fishery Statistics Yearbook, 2022). In order to reduce the negative impact of intensive aquaculture activities on the environment, improve aquaculture yield and ecological benefits, and maintain sustainable development of the industry, it is necessary to construct a reasonable integrated multi-trophic aquaculture model for *P. trituberculatus* (Largo et al., 2016; Knowler et al., 2020).

As a method of long-term stable integrated aquaculture, polyculture of crab, shrimp, and shellfish reduces eutrophication and purifies water through shrimp and shellfish feeding on residual bait and feces, and shellfish filtering phytoplankton and suspended matter (Gao et al., 2008; Chang et al., 2020). In order to improve the complementarity between allotment species and main animals in terms of food and ecological niches, it is necessary to understand the growth of the animals and their impact on the environment during the aquaculture process (Reid et al., 2020). The assessment of the environmental impact of *P. trituberculatus* aquaculture has been reported in numerous papers (Zhang et al., 2015; Feng et al., 2018), but their feeding and metabolic activities have been neglected and nutrient dynamics have not been quantified well enough to accurately guide farm production. Continuous monitoring of the physiological dynamics of aquatic animals in the field is time-consuming and laborious. Because of this, individual growth models, which can reproduce the dynamics of each physiological activity in combination with environmental conditions, may be useful tools (Reid et al., 2020; Dong et al., 2022).

A dynamic energy budget (DEB) model is an individual growth model based on the physical and chemical properties of energy metabolism, which can reflect the universal laws of biological energy metabolism (Kooijman, 1986). The model not only quantifies the energy used for growth and the energy distribution throughout the life history stages (including shell and gonadal development) but can also be easily applied to the study of different species and waters (Ren and Ross, 2001; Kooijman, 2010; Sousa et al., 2010). DEB models are increasingly applied to aquatic bioenergetics and population dynamics studies to provide guidance for aquaculture and fisheries management (Spillman et al., 2008; Ren and Schiel, 2008; Orestis et al., 2019). They have been used to study some crustaceans, including the molting behavior of swimming crab

Liocarcinus depurator, which provides the basis for their application to *P. trituberculatus* in this study (Campos et al., 2009; Talbot et al., 2019; Yang et al., 2020). However, the previous DEB model for *P. trituberculatus* only simulated growth during a single molt period, which has limited application in real production (Talbot et al., 2019).

The purpose of this study was to establish a DEB model for *P. trituberculatus* and simulate its nutrient (carbon, nitrogen, and phosphorus) dynamics based on the model, with the goal of providing important information for environmental assessment and integrated mariculture model management of *P. trituberculatus* in an intensive mariculture pond. In addition, we discussed the possibility of applying the model to carbon sink value estimation and ecosystem modeling in the future.

MATERIAL AND METHODS

DEB Model

The DEB model describes the energy allocation process of biological growth and reproduction as having three parts: 1) structural body (V), which is related to body length, metabolism, and structural maintenance; 2) storage energy (E), which is the assimilated energy that first enters a reserve compartment; and 3) energy for development and reproduction (E_R), which determines reproduction (Kooijman, 1986). Individuals ingest food in an amount proportional to the surface area of the organism and it is then converted into reserves through digestion and absorption at a constant efficiency. Stored energy comes from these reserves and is used for growth, reproduction, and maintenance of life activities based on the k -rule (Kooijman, 2010). The maintenance of structural material takes priority over growth and growth stops when food is insufficient and reserves are low.

Most standard DEB models assume continuous growth in structural volume but crustaceans are molt-growing species that initiate molting when nutrient accumulation and growth and development reach a certain level. This makes application of the DEB model to crustaceans somewhat challenging. Two key pieces of information are required to understand crustacean molt dynamics: the molt increment (MI) (i.e., the magnitude of the increase in size at molt) and the intermolt period (IP) (i.e., the length of time between two successive molts) (Chang et al., 2012). The primary factor influencing crustacean molting is the number of nutrients accumulated in the body (Abuhagr et al., 2014), and this can be assessed *via* indicators such as body weight and condition factors (Roberto and Defeo, 2002; Sharawy et al., 2019). MI and IP can be obtained from the pre- and post-molt lengths predicted by the DEB model if a threshold for nutrient accumulation at the start of the molt is available. In this study, we used condition factor (CF) (i.e., the ratio of body weight to the third power of carapace width) as an indicator of nutrient accumulation in *P. trituberculatus* to initiate molting, extending the DEB model for carapace width growth simulation (**Table 1**). When CF does not reach the molting threshold (k_{CF}), body weight will increase with the aquaculture period but carapace width will not grow. When CF reaches the

TABLE 1 | Equations describing the discontinuous individual growth model of the swimming crab *Portunus trituberculatus*.

Definition	Equation
DEB standard model Temperature dependence	$k(T) = k_0 \cdot \exp\left\{\frac{T_A}{T_1} - \frac{T_A}{T}\right\}$
Ingestion rate	$J_x = k(T) \cdot \{J_{xm}\} \cdot f \cdot V^{2/3}$
Assimilation rate	$P_A = AE \cdot J_x = k(T) \cdot f \cdot \{P_{Am}\} \cdot V^{2/3}$
Catabolic rate	$P_C = k(T) \frac{E}{[E_G] + \kappa \cdot \frac{E}{V}} \left(\frac{[E_G] \cdot \{P_{Am}\} \cdot V^{2/3}}{[E_M]} + [P_M] \cdot V \right)$
Maintenance rate	$P_M = k(T) \cdot [P_M] \cdot V$
Maturity maintenance rate	$P_J = k(T) \cdot \min(V, V_p) \cdot [P_M] \cdot \left(\frac{1-\kappa}{\kappa}\right)$
Reserve dynamic	$\frac{dE}{dt} = P_A - P_C$
Reproductive reserve dynamic	$\frac{dE_R}{dt} = (1-\kappa) \cdot P_C - P_J$
Biovolume growth	$\frac{dV}{dt} = \kappa \cdot P_C - P_M$
Volume	$V = \frac{E_V}{[E_G]}$
Carapace width	$CW = \frac{\delta_m}{V^{1/3}}$
Wet weight	$WW = \frac{E}{\mu_E} + \frac{\kappa_R \cdot E_R}{\mu_E} + V \cdot \rho$
Molting model Condition factor	$CF = \frac{WW}{CW^3} \times 100\%$
Carapace width	$\begin{cases} CW_t = CW_{t-1} & \text{if } CF < k_{CF} \\ CW_t = CW_t & \text{if } CF \geq k_{CF} \end{cases}$

molting threshold, molting occurs and the carapace width grows to the width simulated by the DEB model. According to observed data, the k_{CF} was 45% for the first 40 days and 30% for days 40 – 180.

Model Parameterisation

The parameters of the swimming crab DEB model were estimated according to the Add-my-Pet (AmP) procedure (Marques et al., 2019). The data required for parameterization included zero-variate data, univariate data, and pseudo-data. Zero-variate and univariate data were obtained from previous studies and surveys.

The physiological rate in the DEB model is temperature dependent and follows the Arrhenius relationship. The Arrhenius temperature (T_A) can be obtained from physiological experimental data describing the relationship between respiration and temperature and was estimated to be 6270 K (Dai et al., 2014; see **Figure A1**). Zero variate data related to the development and reproduction of *P. trituberculatus* were mainly obtained from fishery surveys along the Chinese coast (Dong, 2012). Univariate data for carapace width- wet weight, carapace width-age, and wet weight-age were obtained from pond aquaculture observation (Gao et al., 2016; Che et al., 2019).

Based on the completeness scale proposed by Lika et al. (2011), we assigned a completeness score of 2.5 to the data available for *P. trituberculatus*. Mean relative error (MRE) and symmetric mean squared error (SMSE) were used to assess the overall goodness of fit. The goodness-of-fit of model predictions were assessed by estimating the relative error for

each zero-variate data point and univariate data set (Marques et al., 2019).

Model Application Experiment

The experimental *P. trituberculatus*-*Penaeus japonicus*-*Sinonovacula constricta* integrated multi-trophic aquaculture pond used in this study was located in Zhoushan City, Zhejiang Province, China (24°35'N, 112°7'E), and was 1.33 ha in area, with an average water depth of 1.2 m during the study period. *P. trituberculatus* with a carapace width of 0.74 ± 0.05 cm were stocked to a density of $7.5 \text{ kg}\cdot\text{ha}^{-1}$. The experiment was carried out over 180 d, from June 2020 to November 2021. The water temperature in the pond was recorded continuously by a water temperature recorder (HOBO-MX2201, America). During the experiment, iced trash fish were provided daily at 17:50 and the feed level was recorded. Water was changed 1-2 times per month. The salinity range was 14.5 to 19.0. The water temperatures and feed levels are shown in **Figure A2**.

Carbon (C) and nitrogen (N) content in crab feed and feces were determined using an elemental analyzer. The samples were digested using the $\text{HClO}_4\text{-H}_2\text{SO}_4$ method. The phosphorus (P) content in feed and feces was determined using a flow injection analyzer. The energy content of the crab feed was determined using a PARR1281 oxygen bomb calorimeter.

After seeding, the growth and molting of crabs in the pond were monitored twice daily and the molting time was recorded. After molt, 30 – 50 crabs were removed and their wet weight, carapace width, and shell weight were measured. The C, N, and P contents of ground and mixed crabs were analyzed using method described above.

Simulation of Nutrient Dynamics Associated With *P. trituberculatus* Aquaculture

The simulation of nutrients dynamics associated with *P. trituberculatus* included the dynamic processes of ingestion, respiration, excretion, and feces, as well as the nutrient content of seeding, residual feed, dead crab release, molt, and harvest. The simulation of dynamic physiological processes takes into account the general temperature dependence of chemical (enzymatic) processes (Gillooly et al., 2001). Nutrient quantification based on the DEB model was scaled up to the population level based on the aquaculture density of the swimming crabs. The population density of *P. trituberculatus* was calculated as:

$$\frac{dMN}{dt} = -\delta_r \cdot MN \quad (1)$$

where MN represents the density of the swimming crabs (ind ha^{-1}), and δ_r represents their mortality (0.0253 d^{-1}).

The ingestion rate (J_x) of *P. trituberculatus* has been calculated using the individual growth model (**Table 1**). The food intake of *P. trituberculatus* was calculated according to equation (2):

$$\text{Ingestion} = \frac{J_x}{\text{EN}_{\text{food}}} \cdot \text{MN} \cdot 10^{-6} \quad (2)$$

where J_x represents the ingestion rate, which is calculated based on the DEB model and EN_{food} represents the energy content of the crab feed (20.08 J mg^{-1}). The intake of C, N, and P by *P. trituberculatus* were calculated as ($\text{kg ha}^{-1} \text{ d}^{-1}$):

$$C_{\text{Inge}} = \text{Ingestion} \cdot C_{\text{food}} \quad (3)$$

$$N_{\text{Inge}} = \text{Ingestion} \cdot N_{\text{food}} \quad (4)$$

$$P_{\text{Inge}} = \text{Ingestion} \cdot P_{\text{food}} \quad (5)$$

where C_{food} , N_{food} , and P_{food} represent the C, N, and P contents of the crab feed (36.61%, 9.31%, and 1.81%, respectively).

In intensive aquaculture, exogenous feed is typically provided in large quantities and adjusted to the culture period. The unconsumed food was calculated according to equation (6):

$$\text{Food}_{\text{wast}} = \text{Food}_{\text{w}} - \text{Ingestion} \quad (6)$$

where Food_{w} represents the daily crab feeding level ($\text{kg d}^{-1} \text{ ha}^{-1}$). The C, N, and P content of unconsumed food ($\text{kg d}^{-1} \text{ ha}^{-1}$) were calculated as:

$$C_{\text{wast}} = \text{Food}_{\text{wast}} \cdot C_{\text{food}} \quad (7)$$

$$N_{\text{wast}} = \text{Food}_{\text{wast}} \cdot N_{\text{food}} \quad (8)$$

$$P_{\text{wast}} = \text{Food}_{\text{wast}} \cdot P_{\text{food}} \quad (9)$$

Swimming crabs also affect the environment through respiration and excretion. Respiration rate is proportional to the catabolic rate (P_c , J d^{-1}), (Pouvreau et al., 2006). The amount of C released through respiration by *P. trituberculatus* was calculated as ($\text{kg ha}^{-1} \text{ d}^{-1}$):

$$C_{\text{resp}} = \frac{P_c}{\eta} \cdot 0.375 \cdot \text{MN} \cdot 10^{-6} \quad (10)$$

where η is a constant for converting oxygen to energy equivalents and equals $14.3 \text{ J mg}^{-1} \text{ O}_2$ (Gnaiger and Forstner, 1983) and 0.375 is the ratio of molecular weights used to transform O_2 in C. The amount of N and P released through excretion was calculated as ($\text{kg ha}^{-1} \text{ d}^{-1}$):

$$N_{\text{excr}} = \text{JmgC}^{-1} \cdot \left(\frac{N}{C} \right)_{\text{crab}} \cdot (P_M + (1 - \kappa_R) \cdot [(1 - \kappa) \cdot P_C - P_I] + P_I + \frac{[E_G] - [E_V]}{[E_G]} \cdot (\kappa \cdot P_C - P_I)) \cdot \text{MN} \cdot 10^{-6} \quad (11)$$

$$P_{\text{excr}} = N_{\text{excr}} \cdot \left(\frac{P}{N} \right)_{\text{excr}} \quad (12)$$

where JmgC is the ratio of carbon to energy value of the crab feed (55 J mg C^{-1}), $\left(\frac{N}{C} \right)_{\text{crab}}$ represents the ratio of nitrogen and carbon in crab ($0.247 \text{ mg N mg C}^{-1}$) and $\left(\frac{P}{N} \right)_{\text{excr}}$ represents the ratio of phosphorus and nitrogen in excreta ($0.0653 \text{ mg P mg N}^{-1}$). *P. trituberculatus* feces is the main source of

nutrients in pond sediment. The fecal wastes were calculated as ($\text{kg ha}^{-1} \text{ d}^{-1}$):

$$\text{Egestion} = \frac{(1 - \text{ae}) \cdot J_x}{\text{EN}_{\text{food}}} \cdot \text{MN} \cdot 10^{-6} \quad (13)$$

where ae represents the assimilation efficiency (0.8). The amount of C, N, and P released into the pond by the crab feces was calculated as ($\text{kg ha}^{-1} \text{ d}^{-1}$):

$$C_{\text{egest}} = \text{Egestion} \cdot C_{\text{food}} \quad (14)$$

$$N_{\text{egest}} = \text{Egestion} \cdot N_{\text{food}} \quad (15)$$

$$P_{\text{egest}} = \text{Egestion} \cdot P_{\text{food}} \quad (16)$$

Dead crabs need to be removed daily to prevent them from remaining decomposing in the water column and causing deterioration of the water quality. Based on their mortality rate, the C, N, and P contents of dead *P. trituberculatus* were calculated as (kg ha^{-1}):

$$C_{\text{death}} = \text{WW} \cdot \delta_r \cdot C_{\text{crab}} \cdot \text{MN} \cdot 10^{-6} \quad (17)$$

$$N_{\text{death}} = \text{WW} \cdot \delta_r \cdot N_{\text{crab}} \cdot \text{MN} \cdot 10^{-6} \quad (18)$$

$$P_{\text{death}} = \text{WW} \cdot \delta_r \cdot P_{\text{crab}} \cdot \text{MN} \cdot 10^{-6} \quad (19)$$

where C_{crab} , N_{crab} , and P_{crab} represent the C, N, and P contents of the crab (6.16%, 1.52%, and 0.18%, respectively).

Nitrogen and phosphorus levels in the water are also influenced by *P. trituberculatus* molting. Ten molts were simulated in this study, and the amounts of N and P released during each molt were estimated using the following equations (kg ha^{-1}):

$$N_{\text{molt}} = \text{WW} \cdot \delta_s \cdot N_{\text{shell}} \cdot \text{MN} \cdot 10^{-6} \quad (20)$$

$$P_{\text{molt}} = \text{WW} \cdot \delta_s \cdot P_{\text{shell}} \cdot \text{MN} \cdot 10^{-6} \quad (21)$$

where δ_s represents the ratio of shell weight to body weight (9.82%). N_{shell} and P_{shell} represent the N and P contents of crab shells (7.15% and 0.55%, respectively).

The C, N, and P contents of *P. trituberculatus* seeding and harvest were calculated as (kg ha^{-1}):

$$C_{\text{seed}} = \text{MN}_0 \cdot \text{WW}_0 \cdot C_{\text{crab}} \cdot 10^{-6} \quad (22)$$

$$N_{\text{seed}} = \text{MN}_0 \cdot \text{WW}_0 \cdot N_{\text{crab}} \cdot 10^{-6} \quad (23)$$

$$P_{\text{seed}} = \text{MN}_0 \cdot \text{WW}_0 \cdot P_{\text{crab}} \cdot 10^{-6} \quad (24)$$

$$C_{\text{harv}} = \text{MN}_{180} \cdot \text{WW}_{180} \cdot C_{\text{crab}} \cdot 10^{-6} \quad (25)$$

$$N_{\text{harv}} = \text{MN}_{180} \cdot \text{WW}_{180} \cdot N_{\text{crab}} \cdot 10^{-6} \quad (26)$$

$$P_{\text{harv}} = \text{MN}_{180} \cdot \text{WW}_{180} \cdot P_{\text{crab}} \cdot 10^{-6} \quad (27)$$

where MN_0 and WW_0 represent the density and wet weight of crab at the time of seeding, respectively. MN_{180} and WW_{180} represent the density and wet weight of crab at the time of harvest, respectively.

RESULTS

Model Parameters

The results of the predicted zero-variate values for swimming crab based on the AmP procedure are listed in **Table 2**. The DEB parameters were estimated at a reference temperature of 20°C (**Table 3**). The parameter estimation resulted in acceptable goodness-of-fit with MRE = 0.058 and SMSE = 0.007. The model underestimated the age at birth and life span and overestimated the length at end acceleration and ultimate length. The predicted value of the ultimate reproduction rate was in general agreement with the observed value. (http://www.bio.vu.nl/thb/deb/deblab/add_my_pet/entries_web/Portunus_trituberculatus/Portunus_trituberculatus_res)

Model Validation

Simulated and observed growth conditions and the molting behavior of swimming crabs from June to November are shown in **Figure 1**. The simulation was generally consistent with the observed results, indicating that the model accurately simulated the variation in wet weight and carapace width of *P. trituberculatus*. Ten molts were simulated over 180 days and this was equal to the number of molts observed and generally coincided with the recorded molt time points. The simulated molting cycle time lengthened with each stage, from 4 days for juvenile crabs in stages I – II (Observed days: 4), 12 days for juveniles in stages V – VI (Observed days: 11), and 38 days for adults in stages X – XI (Observed days: 36) (**Figure 1A**).

Simulation of Carbon, Nitrogen, and Phosphorus Dynamics Associated With *P. trituberculatus* Aquaculture

The simulation of nutrients dynamics associated with *P. trituberculatus* aquaculture (ingestion, respiration, excretion and feces) is shown in **Figure 2**. The dynamics of C, N, and P involved showed an overall trend of rising and then falling, with a maximum in August. In addition, the trend fluctuated significantly in August due to the high temperature.

The C, N, and P released during molts are shown in **Table 4**. The highest N and P release occurred during the ninth molt, and prior to the last molt, N and P release increased with the number of molts. The total amount of nitrogen and phosphorus released during the last five molts was close to 90% of the total release.

Figure 3 shows the C, N, and P fluxes associated with *P. trituberculatus* aquaculture over the six months rearing cycle including seeding and harvest. The C, N, and P introduced at seeding were 0.46 kg ha⁻¹, 0.11 kg ha⁻¹ and 0.01 kg ha⁻¹, respectively.

TABLE 3 | The parameter values of the DEB model for swimming crab *Portunus trituberculatus* at 20°C.

symbol	value	unit	parameter
[E _G]	4445	J cm ⁻³	volume-specific costs for structure
[E _V]	3556	J cm ⁻³	volume-specific energy of structure
[E _M]	42113	J cm ⁻³	maximum storage density
k	0.99	–	fraction of catabolic flux to growth and maintenance
k _R	0.95	–	fraction of reproductive reserves
δ _M	0.4284	–	shape coefficient
{J _{Xm} }	6062	J cm ⁻² d ⁻¹	maximum feeding rate per unit body surface area
{P _{AM} }	4805	J cm ⁻² d ⁻¹	maximum surface area-specific assimilation rate
[P _M]	434	J cm ⁻³ d ⁻¹	volume-specific maintenance rate
T ₁	293	K	reference temperature
T _A	6270	K	Arrhenius temperature
k ₀	1	–	reference physiological reaction rate at 293K
μ _E	48200	J g ⁻¹	energy content of reserves
ρ	1	g cm ⁻³	volume-specific wet flesh weight

The C, N, and P ingested during the 180-day culture period were 4,938.57 kg ha⁻¹, 1,255.88 kg ha⁻¹, and 244.16 kg ha⁻¹, respectively, of which approximately 1.06%, 1.03%, and 0.62% were removed by harvest and 6.84%, 6.63%, and 4.04% were removed by dead crabs. Overall, 44.46% of the N and 28.43% of the P were released as excretion, feces, and molts. The C, N, and P released from the residual feed accounted for 41.43% of the total feed. The C, N, and P released into the water (excluding dead crabs and harvest) during the 180-day aquaculture period were 4,480.53 kg ha⁻¹, 1,446.60 kg ha⁻¹, and 242.09 kg ha⁻¹, respectively.

DISCUSSION

The Discontinuous Individual Growth Model

Traditionally, the parameterization of the DEB model relies on a large number of physiological experiments on individuals held in standard conditions including controlled temperature (Ren and Schiel, 2008; Serpa et al., 2013). As an economically important farmed crab in China, studies on *P. trituberculatus* have focused on biological surveys, genetic breeding, and farming techniques (Lv et al., 2014; Wang et al., 2018; Duan et al., 2021), lacking sufficient physiological information (e.g., starvation experiments). The AmP procedure allows us to estimate parameters for a species for which limited physiological data is available (Lika et al., 2011; Ren et al., 2020) because it allows us to estimate model parameters according to biological (zero-variate data) and growth characteristics (univariate data). Complete zero-variate data include age, length at birth, weight,

TABLE 2 | Zero-variate data used for estimating parameters of the swimming crab *Portunus trituberculatus* DEB model. Observed data and relative error (RE) are specified.

Symbol	Unit	Observation	Prediction	parameter	RE	Reference
a _b	d	7.0	6.5	age at birth	0.067	(Dong, 2012)
a _m	d	730	720	life span	0.014	(Dong, 2012)
L _j	cm	0.383	0.394	length at end acceleration	0.028	(Gao et al., 2016)
L _i	cm	25.0	26.2	ultimate length	0.047	(Wiki)
R _i	#d ⁻¹	7296	7154	ultimate reproduction rate	0.019	(Dong, 2012)

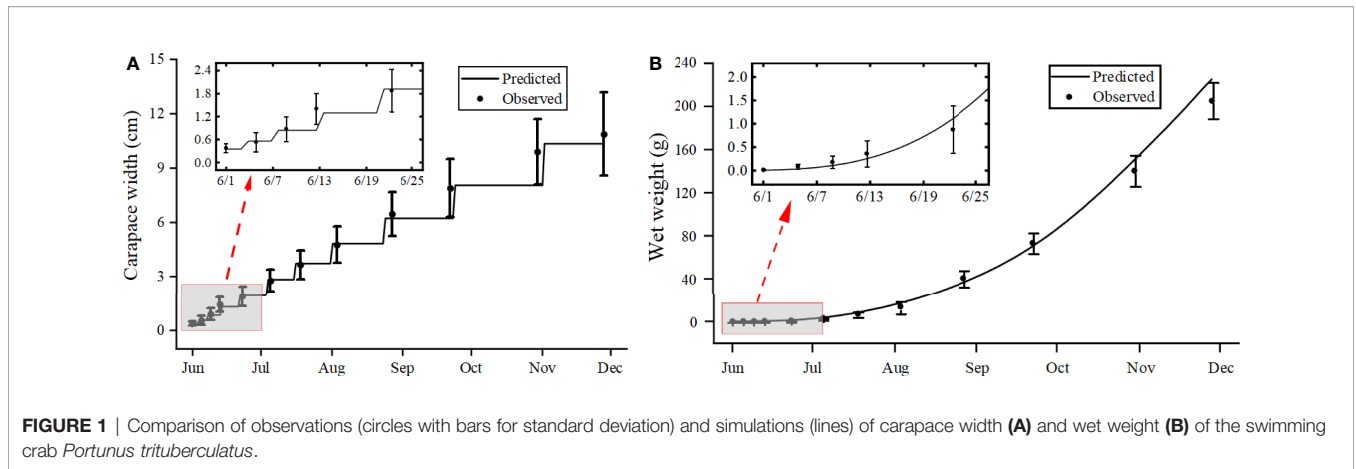


FIGURE 1 | Comparison of observations (circles with bars for standard deviation) and simulations (lines) of carapace width **(A)** and wet weight **(B)** of the swimming crab *Portunus trituberculatus*.

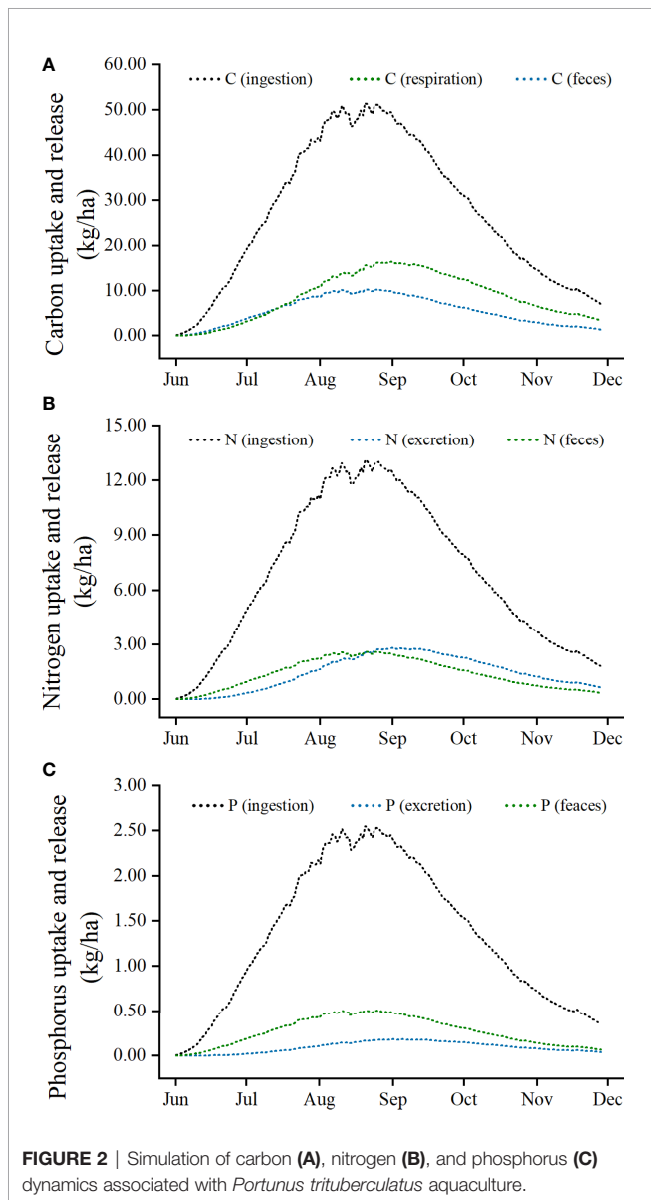


FIGURE 2 | Simulation of carbon **(A)**, nitrogen **(B)**, and phosphorus **(C)** dynamics associated with *Portunus trituberculatus* aquaculture.

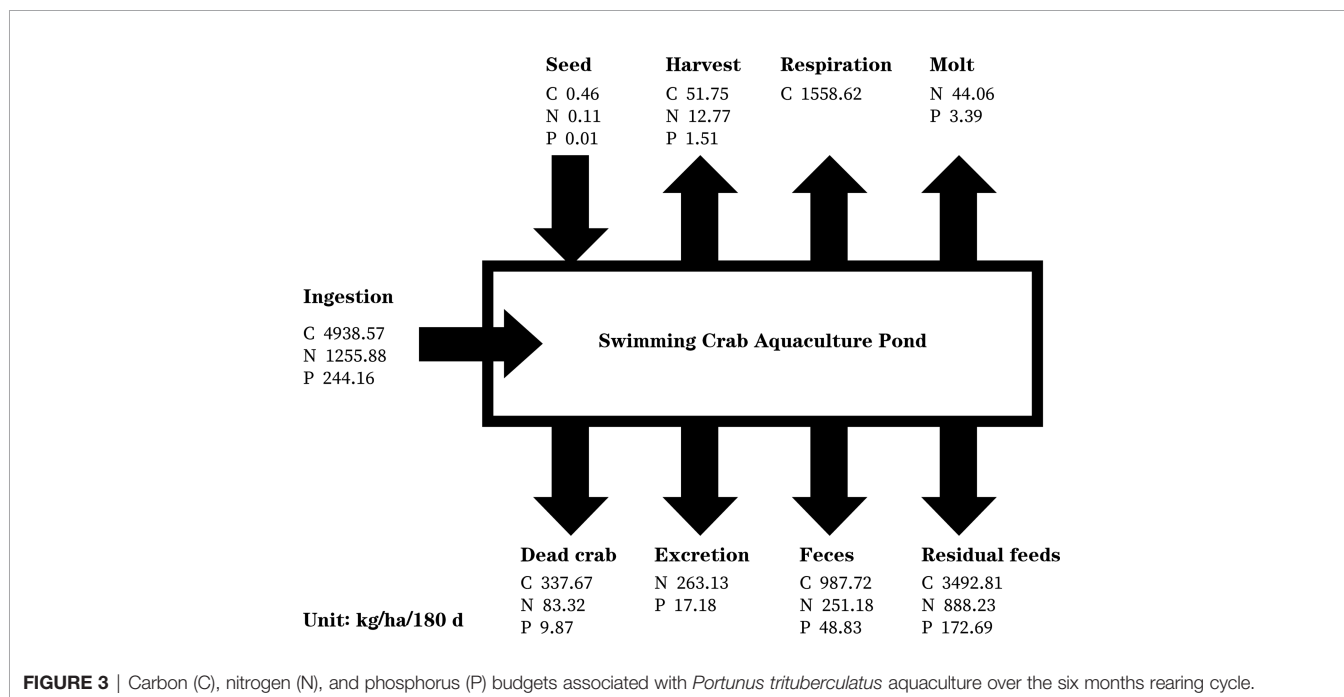
maturation temperature, lifespan, ultimate length, and ultimate weight. Inadequate accumulation of relevant knowledge may affect the accuracy of some of these parameters (Marques et al., 2018). Another advantage of the AmP method is that MRE and SMSE can be used to check the veracity of the data (Marques et al., 2019) and the simulated and predicted values of the parameters in this study show a high match (Table 2), indicating that the model parameters are reasonable and valid.

More than 1,000 papers have been published on DEB theory (https://www.zotero.org/groups/500643/deb_library/library), which is widely used in aquaculture for finfish, bivalves, shrimp (Cheng et al., 2018; Dambrine et al., 2020; Yang et al., 2020), and swimming crab *L. depurator* (Talbot et al., 2019). This study showed that the DEB model can be applied to *P. trituberculatus* in integrated aquaculture ponds (Figure 1). The DEB model predicts individual growth based on food and water temperature dynamics. In the DEB model, food condition f (range: 0 – 1) is used to represent food availability (density and mass), whereas typically the unstable food condition of bivalves, which is usually set to a fixed value in models of finfish, shrimp, and other cast-feeding farming animals, is used (Campos et al., 2009; Ren et al., 2020). This study included values of f ranging from the lowest food scenario ($f = 0$) to arbitrary feeding ($f = 1$), and the results show that the most growth occurred at $f = 1$, indicating adequate and high-quality feed in the experimental pond (Haberle et al., 2020). DEB theory uses the Arrhenius equation to describe the effect of changing temperature on the rate of individual physiological responses (Kooijman, 2010). *P. trituberculatus* is a eurythermic species but temperature changes have a large effect on its metabolism (Lu et al., 2015). Although the water temperature in the experimental ponds fluctuated within 10°C, the temperature dependence function of the *P. trituberculatus* DEB model showed a significant difference in August (high temperature: 29.1 – 30.70 °C) and November (low temperature: 21.70 – 24.40 °C) (Figure A3).

The molting period of crustaceans is often accompanied by unstable ingestion and metabolism, cannibalism, and other phenomena (Su et al., 2019). During the molting period, good feed and water quality and nutrient supplementation can help ensure a successful molt. This makes molting prediction an important prerequisite for crustacean aquaculture management (Lemos and Weissman, 2020; Liu et al., 2021). The application of

TABLE 4 | Carbon, nitrogen, and phosphorus released during the molting of *Portunus trituberculatus*.

Period	I – II	II – III	III – IV	IV – V	V – VI	VI – VII	VII – VIII	VIII – IX	IX – X	X – XI
Nitrogen release (g ha ⁻¹ 180d ⁻¹)	33	121	413	1166	2936	5067	7506	9352	9565	7900
Phosphorus release (g ha ⁻¹ 180d ⁻¹)	3	9	32	90	226	390	577	719	736	608



a discontinuous growth model based on mathematical functions and probability statistics to *P. trituberculatus* showed satisfactory results (Wang, 2017). The discontinuous growth and periodic loss of calcified structures due to molt make a growth modeling approach based on individual physiological and ecological characteristics inappropriate for crustaceans and most studies ignore molting behavior (Campos et al., 2009; Yang et al., 2020). In the swimming crab DEB model, Talbot et al. (2019) substituted wet weight for structural volume and established a logical relationship with physiological rates in the standard model to complete molt predictions within a single molt cycle. As a comprehensive index to determine the physiological and nutritional states of animals, the condition factor has been used as an important predictor of molt in crustaceans such as penaeid shrimp (*Farfantepenaeus aztecus*, *F. brasiliensis*, *F. duorarum*, and *F. notialis*), brown shrimp (*Crangon crangon*) and Chinese mitten crab (*Eriocheir sinensis*) (Roberto and Defeo, 2002; Chen et al., 2016; Sharawy et al., 2019). The results of this study showed that the growth model of *P. trituberculatus* could accurately predict molt (Figure 1A), but molt was also influenced by the environment (temperature, precipitation, etc.). Because these environmental factors are uncontrollable outside of experimental settings, the stability of the model needs further validation. In addition, shell mineralization in shrimp and crabs can play a role in carbon sequestration (Troell et al., 2009), and

the successful simulation of molting behavior can provide a reference for carbon sink value estimation of crustaceans.

Nutrient Dynamics Simulation

By using the ecologically complementary habits of different farming organisms and making full use of system materials and energy, the integrated aquaculture model helps to improve economic efficiency and reduce farming pollution (Troell et al., 2003). In China, the shrimp-crab and shrimp-crab-shellfish integrated aquaculture model has become the main farming model for *P. trituberculatus*. However, the existing studies on intensive ponds of *P. trituberculatus* have focused on the overall static carbon, nitrogen, and phosphorus budget and have not quantified nutrient dynamics associated with the aquaculture process (Dong et al., 2013; Zhang et al., 2016). Considering that the simulation of nutrient dynamics at the population level is more meaningful than at the individual level, this study combined the DEB model with the crab density to model the population dynamics of *P. trituberculatus* (Figure A3) and simulated the related nutrient dynamic processes (Figure 2). The patterns of carbon, nitrogen, and phosphorus uptake and release caused by the ingestion and metabolism of the swimming crabs in the experiment were determined by density changes and fluctuated in July and August due to the frequency of hot weather in summer that reduced physiological activity (Lu et al., 2015).

The assessment of the carbon, nitrogen, and phosphorus budget in the ecosystem is an effective method to evaluate nutrient utilization, energy conversion efficiency and pollution levels in aquaculture ponds (Guo et al., 2017). The results showed that $\leq 8\%$ of the carbon, nitrogen, and phosphorus in feed was converted to swimming crab (including harvested and dead crabs), with most of the remainder being released into the aquaculture environment in the form of excretion, feces, and molts. Meanwhile, 41% of the nutrients in the feed were not consumed by *P. trituberculatus* and was discharged into the aquaculture water environment in the form of residual feed, resulting in 4,480.53 kg ha⁻¹, 1,446.60 kg ha⁻¹ and 242.09 kg ha⁻¹ of carbon, nitrogen, and phosphorus being retained in the water. Aquaculture discharges can raise the nutrient load of nearby waters and lead to water eutrophication (Herbeck et al., 2013). We have combined Japanese shrimp and razor clams in the *P. trituberculatus* aquaculture system to improve food resource utilization, increase farm production, and reduce the negative environmental impact of farming activities. Of course, assessing the carrying capacity and proportion of farming organisms in an integrated aquaculture system requires long-term dynamic simulation and analysis. In a follow-up study, we will couple the simulation of *P. trituberculatus* nutrient dynamics to an ecosystem model for ecosystem-level assessment of carrying capacity and environmental pollution to provide scientific and precise management information for integrated aquaculture activities.

CONCLUSIONS

The discontinuous growth model established in this study will accurately relay the growth and molting behavior of *P. trituberculatus*, and provide support for a series of techniques such as seeding, feeding, water quality control, and molting management in intensive mariculture. The assessment of nutrient dynamics, in combination with population dynamics

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and physiological activities, can provide technical guidance for integrated aquaculture practices and help achieve environmental and economic sustainability. This model can not only be used independently for mariculture assessment but can also be combined into an ecological model of an integrated multi-trophic aquaculture ecosystem in order to assess the potential impacts of aquaculture on a larger spatial scale.

DATA AVAILABILITY STATEMENT

The original contributions presented in the study are included in the article/**Supplementary Material**. Further inquiries can be directed to the corresponding author.

AUTHOR CONTRIBUTIONS

SD, XX, and FL designed the experiments and wrote the manuscript. LY and HS performed the experiments. FW supervised and validated the manuscript. All authors contributed to the article and approved the submitted version.

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SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2022.918449/full#supplementary-material>

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