



# Mussel Byssal Attachment Weakened by Anthropogenic Noise

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The increasing underwater noise generated by anthropogenic activities has been widely recognized as a significant and pervasive pollution in the marine environment. Marine mussels are a family of sessile bivalves that attach to solid surfaces via the byssal threads. They are widely distributed along worldwide coastal areas and are of great ecological and socio-economic importance. Studies found that anthropogenic noise negatively affected many biological processes and/or functions of marine organisms. However, to date, the potential impacts of anthropogenic noise on mussel byssal attachment remain unknown. Here, the thick shell mussels *Mytilus coruscus* were exposed to an ambient underwater condition ( $\sim 50$  dB re 1  $\mu$ Pa) or the playbacks of pile-driving noise ( $\sim 70$  or  $\sim 100$  dB re 1  $\mu$ Pa) for 10 days. Results showed that the noise significantly reduced the secretion of byssal threads (e.g., diameter and volume) and weakened their mechanical performances (e.g., strength, extensibility, breaking stress, toughness and failure location), leading to a 16.95–44.50% decrease in mussel byssal attachment strength. The noise also significantly down-regulated the genes expressions of seven structural proteins (e.g., mfp-1, mfp-2, mfp-3, mfp-6, preCOL-P, preCOL-NG, and preCOL-D) of byssal threads, probably mediating the weakened byssal attachment. Given the essential functions of strong byssal attachment, the findings demonstrate that the increasing underwater anthropogenic noise are posing a great threat to mussel population, mussel-bed community and mussel aquaculture industry. We thus suggest that future work is required to deepen our understanding of the impacts of anthropogenic noise on marine invertebrates, especially these with limited locomotion ability, like bivalves.

**Keywords:** anthropogenic noise, mussel, byssal thread, attachment, mechanical performance, gene expression

## INTRODUCTION

The underwater noise generated by anthropogenic activities such as shipping, oil and gas exploration, and the installation of renewable energy devices, has now been widely recognized as a significant and pervasive pollution in the marine environment (Duarte et al., 2021). Anthropogenic noise undoubtedly alters the acoustic signature of marine ecosystems and poses a great threat to marine organisms (Jerem and Mathews, 2021). What's worse, given the growing levels of human activities, the underwater anthropogenic noise levels likely continue to increase in the foreseeable

future (Duarte et al., 2021). Over the past decades, the potential impacts of anthropogenic noise on marine organisms have raised global attention (Peng et al., 2015; Popper et al., 2020). Studies demonstrate that a wide variety of biological processes and physiological functions of marine organisms can be negatively affected by anthropogenic noise, such as acoustic communication (Di Iorio and Clark, 2010; Alves et al., 2021), auditory sensitivity (Kastelein et al., 2018; Vieira et al., 2021), foraging behavior (Purser and Radford, 2011; Wale et al., 2013), antipredator behavior (Simpson et al., 2015; Kok et al., 2021), reproduction (de Jong et al., 2018; Smott et al., 2018) and embryonic development (de Soto et al., 2013; Nedelec et al., 2015). However, to date, most studies were conducted on marine mammals and fishes (Erbe et al., 2018; Popper and Hawkins, 2019), and the effects of anthropogenic noise on marine invertebrates, especially bivalves, received much less attention (Di Franco et al., 2020; Wale et al., 2021).

Marine bivalves, one class of benthic invertebrate, are able to detect substrate-borne and water-borne sound through contact with both substrate and surrounding water (Roberts et al., 2015; Vazzana et al., 2016; Charifi et al., 2017). Among marine invertebrates, bivalves are likely to be particularly vulnerable to anthropogenic noise, since their benthic habit and limited locomotion ability make them unable to escape from a noise area (Day et al., 2017; Vazzana et al., 2020). More importantly, marine bivalves are of great ecological and socio-economic importance, such as: (1) using as a source of human food, (2) increasing biodiversity of local marine ecosystem through creating habitats for other organisms, (3) improving seawater quality by filtering the water and removing particulates within, (4) attenuating global warming and ocean acidification through carbon fixation, and (5) mitigating ocean eutrophication through filtering large quantities of organic matter from the water column and the production of biodeposits (Vaughn and Hoellein, 2018; Suplicy, 2020; van der Schatte Olivier et al., 2020). Limited available studies show that anthropogenic noise disrupts the larval development of the New Zealand scallop *Pecten novaezelandiae* (de Soto et al., 2013), causes physiological harm and/or alters behavior in adult marine bivalves, including the razor clam *Sinonovacula constricta* (Peng et al., 2016), the Mediterranean mussel *Mytilus galloprovincialis* (Vazzana et al., 2020), the blood clam *Tegillarca granosa* (Shi et al., 2019), the blue mussel *M. edulis* (Wale et al., 2019), and the scallop *P. fumatus* (Day et al., 2017). However, there are species-specific and ontogenetic variations of noise-induced changes (Wale et al., 2021). Despite the current advance, the full extent to which anthropogenic noise affects marine bivalves is still poorly understood. Regarding their critical ecological roles and socio-economic importance, it is urgent to deepen our understanding of the effects of anthropogenic noise on marine bivalves.

Marine mussels are a family of sessile bivalve species that are widely distributed along worldwide coastal areas (Zhao et al., 2020). They attach to a wide array of substrata with their strong holdfast structure, the byssal threads (Waite, 1983). The byssal thread includes three parts: (1) the proximal region, which has a corrugated surface and is highly extensible; (2) the distal region, that is smoother, stiffer and approximately twice the length of the proximal region; and (3) the adhesive

plaque, which adheres to the substrate (Bell and Gosline, 1996). Marine mussels can also attach to one another via byssal threads, and thereby form mussel aggregates or mussel beds, thus improving stability and protection against predation and environmental perturbations, increasing fertilization success, and providing habitats for other organisms (Borthagaray and Carranza, 2007; Liu et al., 2011; Christensen et al., 2015; Kong et al., 2019). A strong byssal attachment is thus essential for mussel survival, self-defense, reproduction and their ecological functions (Bandara et al., 2013; Sui et al., 2017; Shang et al., 2019). Hence, any potential impacts of anthropogenic noise on byssal attachment could have fundamental ecological implications for marine mussels and ecosystems. Given this, an understanding of noise effects in this regard is crucial. However, to date, the potential impacts of anthropogenic noise on mussel byssal attachment remain unknown.

Due to its ecological and economical importance, and wide distribution along the coastal areas of China, Korea and Japan (Li et al., 2015; Sui et al., 2015), the thick shell mussel *M. coruscus* was chosen and used as a testing organism in this study. A laboratory-based experiment was performed to investigate the effects of anthropogenic noise (specifically playback of pile-driving noise, an impulsive sound source) on byssal thread number, morphology and mechanical performance, and thereby the overall byssal attachment strength of mussel individual. Importantly, the expression levels of seven key genes encoding structural proteins of byssal threads, including four mussel foot proteins (mfp-1, mfp-2, mfp-3, and mfp-6) and three precursor collagen proteins (preCOL-P, preCOL-NG, and preCOL-D), were also determined to reveal the underlying molecular responses of marine mussels to anthropogenic noise.

## MATERIALS AND METHODS

### Animal Collection and Acclimation

Adult thick shell mussels *M. coruscus* with similar size (shell length of  $102.1 \pm 5.5$  mm) were collected from Shengsi Islands, Zhejiang, China ( $30^{\circ}73' N$  and  $122^{\circ}45' E$ ) and directly transported to the Qingjiang Station of Zhejiang Mariculture Research Institute. The mussels were transported in air, mimicking the natural condition during low tide. Following carefully cleaning of epibionts without damaging their shells, the mussels were acclimated in filtered and UV-irradiated natural seawater (temperature of  $22.5 \pm 0.4^{\circ}C$ , pH of  $8.09 \pm 0.06$ , and salinity of  $23.8 \pm 0.4$  PSU) with continuous aeration for 2 weeks. The mussels were fed twice daily (e.g., at 08:00~08:30 and 19:00~19:30) with the microalgae *Platymonas subcordiformis* to satiation. Excess food and feces were removed daily with seawater change (e.g., at 18:00~19:00), in which two thirds of the seawater was removed by siphoning, followed by refilling with filtered and UV-irradiated seawater.

### Experimental Design and Anthropogenic Noise Exposure

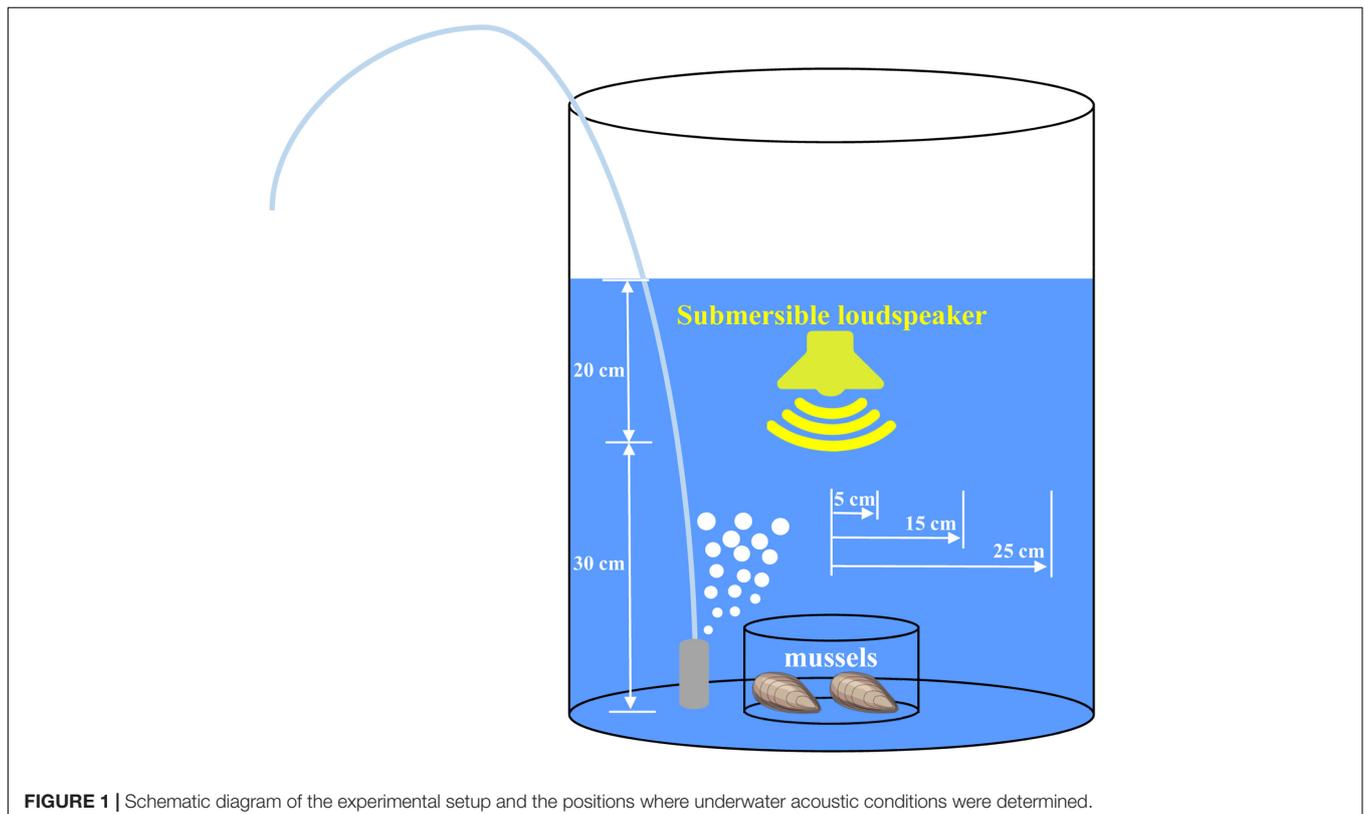
Following acclimation, mussels were randomly assigned to a control and two anthropogenic noise input groups. The ambient

condition without any additional anthropogenic sound input served as the control, while the underwater sound pressure levels of the two anthropogenic noise input groups were set at  $\sim 70$  and  $\sim 100$  dB re  $1 \mu\text{Pa}$  to mimic the underwater conditions with different degrees of anthropogenic sound, according to previous studies (Hildebrand, 2009; Slabbekoorn et al., 2010; Peng et al., 2016). The pile-driving sound recorded previously (Solan et al., 2016) was used as the source of anthropogenic noise input in this study. For the two anthropogenic noise input groups, mussels were exposed to playbacks of the pile-driving noise at respective underwater sound pressure levels. The experiment was performed with circular buckets (diameter of 60 cm, depth of 70 cm) filled with filtered and UV-irradiated natural seawater with a depth of 50 cm. For all groups, the circular bucket had a circular glass chamber (diameter of 20 cm, depth of 10 cm) centered on its floor, and each contained two individual mussels (**Figure 1**). The reason why each chamber contained only two mussels was to avoid the formation of mussel aggregate and its potential effects on sound propagation and the underwater sound pressure level.

A submersible loudspeaker (UW-30, Electro-Voice, Indiana, United States; frequency response 0.1–10 kHz; impedance 8 ohms; power-handling capacity 30 watts) was positioned in the center of the circular tank at a depth of 20 cm below the water surface and oriented toward the floor to generate the desired acoustic conditions (**Figure 1**). The submersible loudspeaker was driven using a power amplifier (AV-296, SAST, Guangdong, China; power-handling capacity 150 watts). The action acoustic conditions were

determined using the acoustics recording unit, which included a bioacoustics recorder (Song Meter SM2+, Wildlife Acoustics, MA, United States; 96 kHz sampling rate), and a calibrated omni-directional hydrophone (HIT-96-MIN, High Tech; Long Beach, Mississippi, United States; flat frequency response 0.02–30 kHz, sensitivity  $-164$  dB re  $1 \text{ V}/\mu\text{Pa}$ ). To test whether there were any variations of the underwater sound pressure levels at different positions, the hydrophone was placed near the bottom of the circular bucket at 5 cm, 15 cm, and 25 cm away from the center (**Figure 1**). Acquired acoustic data were analyzed with the Soundscape Analysis Software SACS V1.0 (Register number: 2014SR216788), and underwater sound pressure levels were obtained following the described method (Peng et al., 2016). During the entire experimental period, mussels were fed twice daily, and seawater was continuously aerated and renewed daily as described above. The seawater parameters, including temperature, pH and salinity, were kept as described above. Since the control and two anthropogenic noise input groups were conducted simultaneously in a laboratory, the experimental circular buckets were equipped with anti-vibration shielding to avoid the potential effects of the high sound pressure level noise on the acoustic conditions of other groups. The experiment plan was replicated five times with different individuals for each time and under identical experimental conditions. The experiment lasted for 10 days, and no animal mortality was observed.

Following exposure, one of the two mussels in the circular glass chamber was randomly selected, dissected and immediately frozen in liquid nitrogen for further gene expression analysis.



**FIGURE 1** | Schematic diagram of the experimental setup and the positions where underwater acoustic conditions were determined.

After carefully removing the existing byssal threads without damaging foot organ, the other mussel was allowed to regrow for 24 h to secrete new byssal threads. During the 24 h incubation, the experimental conditions were identical with those of the 10-day experimental period.

## Byssal Thread Collection and Quantitative Analysis

After regrowing for 24 h, the full-length byssal threads of mussels were carefully collected according to the method described previously (Zhao et al., 2017). Briefly, newly secreted byssal threads were cautiously removed from the substrate one by one. Mussels were dissected by severing the adductor muscles. The full-length byssal threads were then carefully excised at their attachment points to foot. After collection, the number of newly secreted byssal threads per mussel was manually counted. The length (mm) of each byssal thread was determined using a digital Vernier caliper with a precision of 0.01 mm. Digital image of each byssal thread was obtained using a CCD camera mounted on a microscope (BX53, Olympus, Japan). Diameters of three independent locations of the byssal thread were measured through image analysis using the free-access software ImageJ Version 1.51j8<sup>1</sup>, and the mean value of them was used as the diameter ( $d$ ;  $\mu\text{m}$ ) of the byssal thread. Byssal thread was assumed to be cylindrical (Bell and Gosline, 1996), and the cross-sectional area ( $\text{mm}^2$ ) of each byssal thread was calculated from the obtained diameter using the equation: cross-sectional area =  $\pi \times (d/2)^2$ . The volume ( $\text{mm}^3$ ) of each byssal thread was calculated by multiplying the cross-sectional area by the length.

## Mechanical Property Analysis of Byssal Thread

After quantitative analysis, mechanical properties of byssal thread were determined using a universal testing machine (AGS-J, Shimadzu, Japan) following the methods described previously (Zhao et al., 2017). The adhesive plaque was glued to a metal stub secured in the lower platform of the testing machine, while the proximal end of byssal thread was sandwiched between cardboard and further clamped at the upper grip of the testing machine. Tensile test of byssal thread was conducted in air at room temperature at a loading rate of 10 mm/min until failure occurred, mimicking the natural air exposure condition during low tide.

The breaking force (N), also named thread strength, was determined as the force required to break a single byssal thread. The breaking strain (%), also called thread extensibility was measured as the extension at failure, divided by the unstressed thread length. The breaking stress ( $\text{N}/\text{mm}^2$ ) was calculated as the value of breaking force divided by cross-sectional area. The toughness ( $\text{J}/\text{cm}^3$ ) was measured as the energy absorbed per unit volume until thread failure occurred. The location of failure (e.g., proximal, distal or adhesive plaque) was noted, and failure occurrence at each location was expressed as the percentage of the number of threads tested. When failure occurred at the

grip, the corresponding data were discarded from analysis to avoid underestimating the actual mechanical properties (Bell and Gosline, 1996; Moeser and Carrington, 2006).

According to previous studies (Bell and Gosline, 1996; Zhao et al., 2017; Shi et al., 2020), the overall mussel attachment strength was estimated as the potential maximal attachment strength (N) with the formula: attachment = average thread strength  $\times$  thread number. Prior to the estimation, the average thread strength was calculated as the mean value of all thread strengths for a given mussel individual.

## RNA Isolation, Real-Time PCR and Gene Expression Analysis

Following dissection, the foot tissue was immediately frozen in liquid nitrogen and stored at  $-80^\circ\text{C}$  upon use. Total RNA was extracted using TRIzol Reagent (Invitrogen, 15596018) following the manufacturer's protocol. DNA contamination in the RNA sample was removed by treating with DNase I (Invitrogen, 18047019). The RNA quality was checked by 1.0% formaldehyde-denatured agarose gel electrophoresis. The RNA concentration was quantified using a NanoDrop 1000 spectrophotometer (Thermo Fisher Scientific). The cDNA was synthesized with the high-quality total RNA using the M-MLV First Strand Kit (Invitrogen, C28025-032) according to the manufacturer's protocol. Real-time PCR was conducted using a CFX96 Real-Time System (Bio-Rad, United States) with a total reaction volume of 10  $\mu\text{L}$  containing 5  $\mu\text{L}$  of SsoFast EvaGreen Supermix (Bio-Rad, 1725201AP), 0.5  $\mu\text{L}$  of each primer (10  $\mu\text{M}$ ), 2  $\mu\text{L}$  of cDNA template and 2  $\mu\text{L}$  of double-distilled water. The cycle conditions were  $95^\circ\text{C}$  for 5 min, 40 cycles of  $94^\circ\text{C}$  for 20 s,  $61^\circ\text{C}$  for 20 s, and  $72^\circ\text{C}$  for 20 s. The specificity of each amplification reaction was verified by analyzing the melting curve. The relative gene expression level was calculated using the  $2^{-\Delta\Delta\text{Ct}}$  method (Livak and Schmittgen, 2001) with the *18S rRNA* gene as the internal reference. The expression of *18S rRNA* has been verified to be stable under ambient and different noise exposure conditions. The gene-specific primers using in this study were listed in Table 1.

## Statistical Analysis

One-way ANOVAs were performed to show the effects of position (i.e., distance to the center of the concentric circle vertically below the sound source) on the underwater sound

**TABLE 1** | Gene-specific primers used in real-time PCR.

Gene	Forward primer (5'–3')	Reverse primer (5'–3')
<i>18S rRNA</i>	CCTTGCTGCTCTTGATTGA	GAACTACGACGGTATCTGAT
<i>mfp-1</i>	TGGTACAATCAAGAACTG	AGAGAAGGATGAGAACGAAT
<i>mfp-2</i>	CGGTACAGAAAGCATCAT	CATCCTCATCGTCGTCATAT
<i>mfp-3</i>	TTTGCTGGCTTTAGTCCTT	ACCGCTATTCCATCCCTTA
<i>mfp-6</i>	CGGTGATTATGATTACAGAGG	GAAGACAGCATCCAGCAT
<i>preCOL-P</i>	GATCTTCACATGCATCAGC	CAGTCCACCTCCTAAAC
<i>preCOL-NG</i>	ACAAGGACCACAAGGAGAA	ACACCACCAACACCAGTT
<i>preCOL-D</i>	ACCAAGAGGAGATAGAGGAG	GGCTGTCTGAGGTCTTC

<sup>1</sup><https://imagej.nih.gov/ij/>

pressure level, and the effects of anthropogenic noise on byssal thread production, morphology, mechanical parameters and gene expression levels. *Post-hoc* Tukey's multiple tests were conducted to compare differences among groups. Prior to analysis, homogeneity of variance and normality of data were verified with the Levene's test and Shapiro-Wilk's tests, respectively. Chi-square test was performed to compare failure occurrences at each location. Data are presented as the mean  $\pm$  SD, and a *p*-value less than 0.05 was considered to indicate statistically significance.

## RESULTS

### The Underwater Sound Pressure Level

One-way ANOVAs showed that the distance to the center of the concentric circle vertically below the sound source had no significant effect on the underwater sound pressure levels near the bottom of the circular bucket, indicating no significant difference between the underwater sound pressure levels at 5, 15, and 25 cm away from the center (Table 2). As shown in Table 2, the underwater sound pressure level of the control group was  $\sim$ 55 dB re 1  $\mu$ Pa that was much lower than those of the two anthropogenic noise input groups. Additionally, the actual underwater sound pressure levels of the two anthropogenic noise input groups were generally consistent with the nominal values that set at  $\sim$ 70 and  $\sim$ 100 dB re 1  $\mu$ Pa. Given this situation, experimental results were therefore analyzed and reported with nominal underwater sound pressure levels.

### Byssal Thread Production and Morphology

The average number of newly secreted byssal threads per mussel was slightly reduced from 46.6 (100%) to 43.4 (93.13%) and 41 (87.98%) after the 10-day exposure to the 70 and 100 dB re 1  $\mu$ Pa anthropogenic noise, but no significant effect was observed [ $F_{(2,12)} = 3.47$ ,  $p = 0.06$ ] (Figure 2A). The length of newly secreted byssal thread was also not significantly affected after the 10-day exposure [ $F_{(2,12)} = 2.9$ ,  $p = 0.09$ ] (Figure 2B). However, a significant effect of anthropogenic noise on the diameter of the byssal thread was detected [ $F_{(2,12)} = 6.71$ ,  $p = 0.01$ ]. The diameter of mussels exposed to 70 and 100 dB re 1  $\mu$ Pa noise

decreased to approximately 97.18 and 93.50% of that of the control, respectively (Figure 2C). Compared with the control, the average volume of the byssal thread was significantly declined to about 87.09 and 80.40% of that of the control upon 70 and 100 dB re 1  $\mu$ Pa noise exposure, respectively [ $F_{(2,12)} = 11.22$ ,  $p = 0.002$ ] (Figure 2D).

### Mechanical Properties of Byssal Threads

The breaking force of byssal thread was significantly reduced for mussels exposed to the 70 and 100 dB re 1  $\mu$ Pa anthropogenic noise [ $F_{(2, 12)} = 66.67$ ,  $p = 3.17 \times 10^{-7}$ ], which were approximately 89.27 and 63.41% of that of the control (Figure 3A). The breaking stress of byssal thread was significantly decreased after the 10-day exposure [ $F_{(2,12)} = 21.62$ ,  $p = 1.05 \times 10^{-4}$ ]. Compared to control, the breaking stress was declined by about 5.57 and 27.84% for mussels exposed to 70 and 100 dB re 1  $\mu$ Pa anthropogenic noise, respectively (Figure 3B). The breaking strain of byssal thread was also significantly reduced by anthropogenic noise exposure [ $F_{(2,12)} = 39.58$ ,  $p = 5.20 \times 10^{-6}$ ]. The breaking strain of mussels exposed to 70 and 100 dB re 1  $\mu$ Pa anthropogenic noise were significantly decreased to approximately 80.27 and 65.93% of that of the control, respectively (Figure 3C). Similarly, the toughness of byssal thread was significantly affected by anthropogenic noise exposure as well [ $F_{(2,12)} = 75.70$ ,  $p = 1.57 \times 10^{-7}$ ]. The toughness of the 70 and 100 dB re 1  $\mu$ Pa anthropogenic noise exposure groups were remarkably declined to approximately 76.06 and 48.07% of that of the control, respectively (Figure 3D).

### Location of Byssal Thread Failure

As shown in the Figure 4, the thread failure occurred at either the proximal region or the adhesive plaque. The occurrence of thread failure located in the adhesive plaque was significantly increased by anthropogenic noise exposure ( $\chi^2 = 35.92$ ,  $df = 2$ ,  $p < 0.0001$ ). For the control group, only 24% thread failure occurred in the adhesive plaque, while it was increased to approximately 34 and 64% for the 70 and 100 dB re 1  $\mu$ Pa anthropogenic noise exposure groups, respectively.

### The Overall Mussel Attachment Strength

The overall mussel attachment strength was significantly weakened by anthropogenic noise exposure [ $F_{(2,12)} = 64.31$ ,  $p = 3.86 \times 10^{-7}$ ] (Figure 5). Compared to the control, the potential maximal attachment strength of mussels treated with 70 and 100 dB re 1  $\mu$ Pa anthropogenic noise was significantly decreased by 16.95 and 44.50%, respectively.

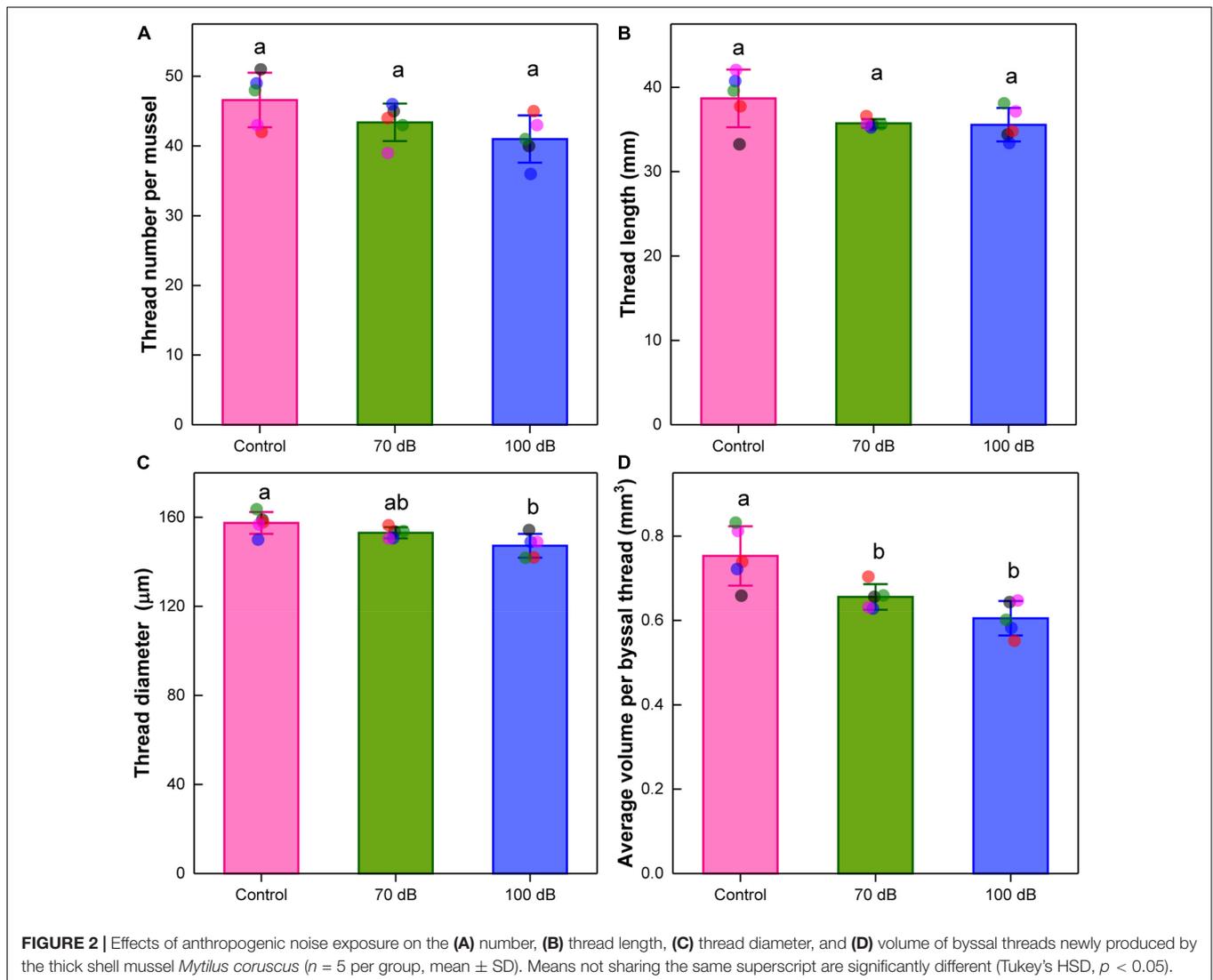
### Gene Expressions of Byssal Proteins

One-way ANOVAs showed that the expression levels of *mfp-1*, *mfp-2*, *mfp-3*, and *mfp-6* were significantly affected by anthropogenic noise exposure (Figures 6A–D). Compared to the control, the expression levels of *mfp-1* were significantly decreased by 34.81 and 46.28% for mussels exposed to 70 and 100 dB re 1  $\mu$ Pa anthropogenic noise, respectively (Figure 6A). Although the fold-change varied, the expression patterns of *mfp-2*, *mfp-3*, and *mfp-6* were similar to that of

**TABLE 2** | The underwater sound pressure levels (dB re 1  $\mu$ Pa) near the bottom of the circular bucket at 5, 15, and 25 cm away from the center of the concentric circle vertically below the sound source ( $n = 10$  per distance, mean  $\pm$  SD).

Group	Underwater sound pressure levels at distances to the center		
	5 cm	15 cm	25 cm
Control	55.02 $\pm$ 2.60 <sup>a</sup>	55.19 $\pm$ 2.73 <sup>a</sup>	55.03 $\pm$ 1.95 <sup>a</sup>
70 dB re 1 $\mu$ Pa noise input	74.16 $\pm$ 1.77 <sup>a</sup>	74.15 $\pm$ 2.44 <sup>a</sup>	73.95 $\pm$ 1.70 <sup>a</sup>
100 dB re 1 $\mu$ Pa noise input	101.82 $\pm$ 2.61 <sup>a</sup>	101.09 $\pm$ 1.96 <sup>a</sup>	100.62 $\pm$ 2.15 <sup>a</sup>

Means not sharing the same superscript are significantly different (Tukey's HSD,  $p < 0.05$ ).

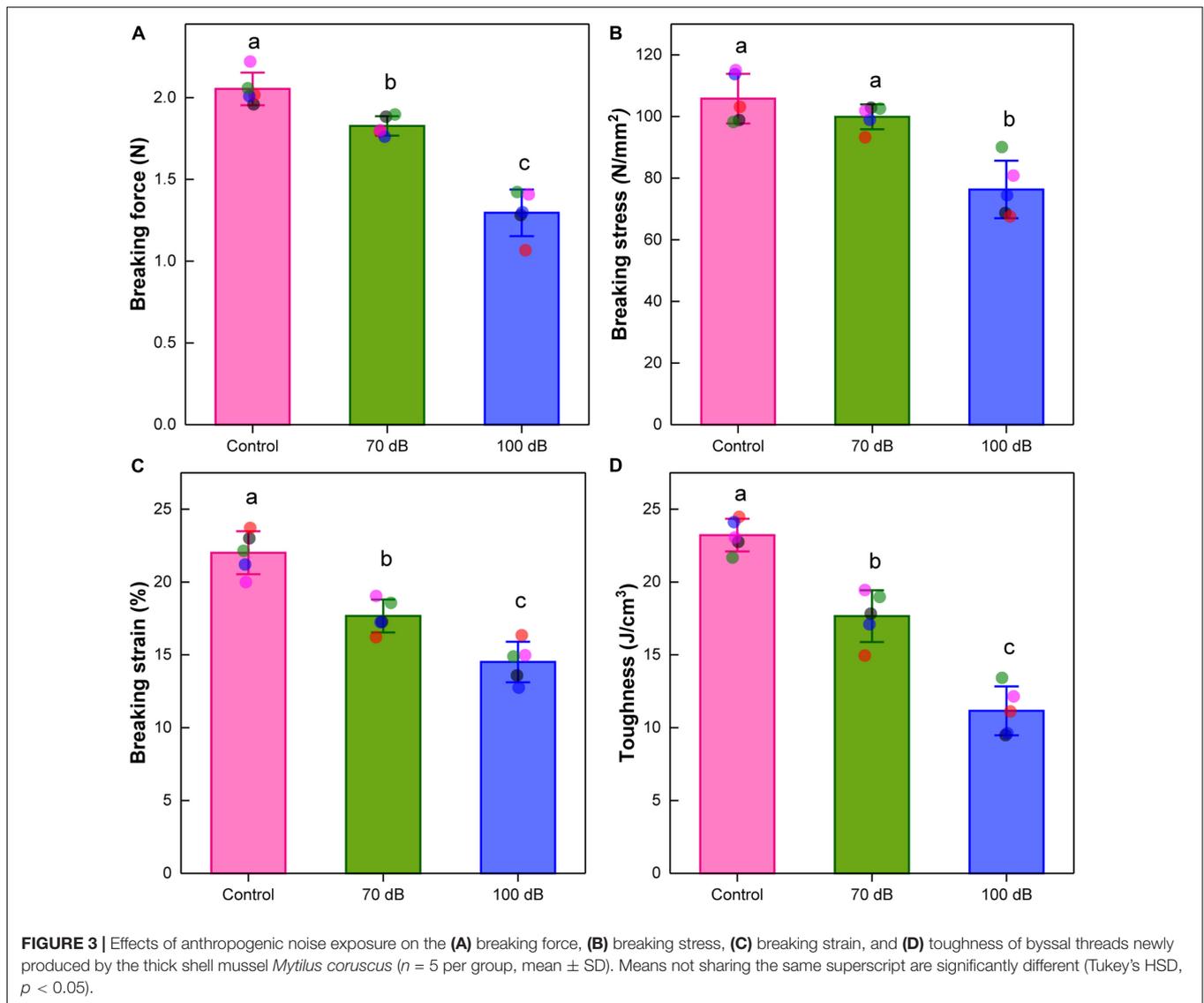


*mfp-1*, showing the down-regulation upon anthropogenic noise exposure (Figures 6B–D). Significant effects of anthropogenic noise on the expression levels of *preCOL-P*, *preCOL-NG*, and *preCOL-D* were observed as well (Figures 6E–G). The expression levels of *preCOL-P* of mussels exposed to 70 and 100 dB re 1  $\mu\text{Pa}$  anthropogenic noise were significantly reduced to 48.81 and 28.95% of that of control, respectively (Figure 6E). Similarly, for mussels treated by 70 and 100 dB re 1  $\mu\text{Pa}$  anthropogenic noise, the expression levels of *preCOL-NG* were, respectively, decreased to 93.69 and 48.47%, and the expression levels of *preCOL-D* were, respectively, down-regulated to 68.16 and 50.07%, compared to those of control (Figures 6F,G).

## DISCUSSION

Our results suggest that exposure to playbacks of the pile-driving noise had significant negative effects on byssal thread production and mechanical performances of the thick shell

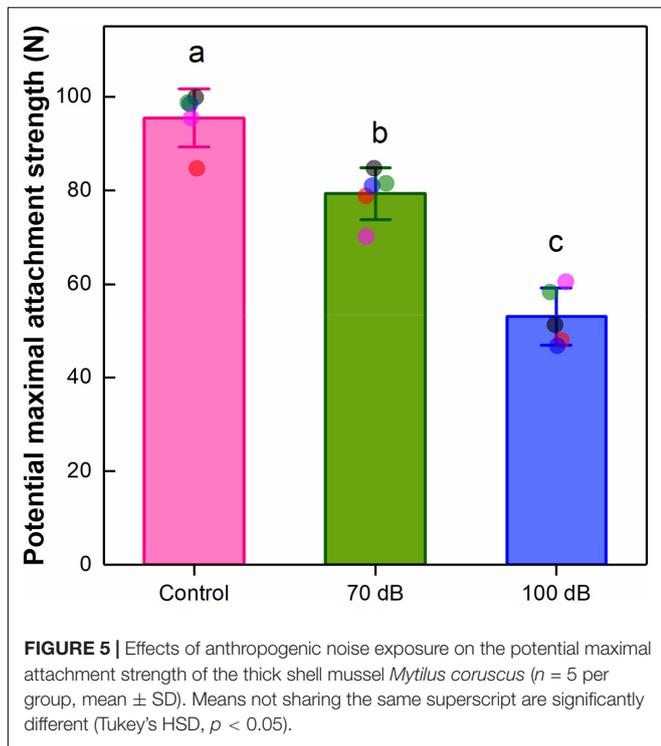
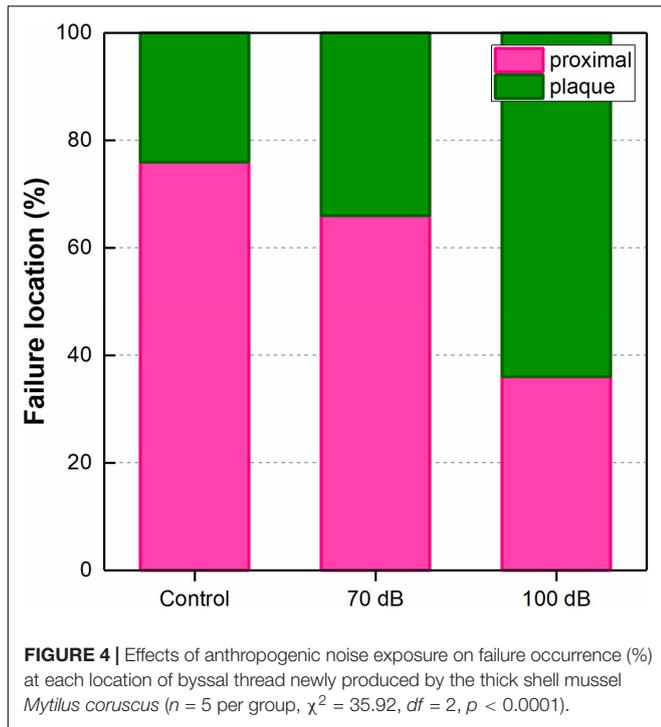
mussel *M. coruscus*, leading to reductions in the diameter, average volume, strength (i.e., breaking force), extensibility (i.e., breaking strain), breaking stress and toughness of individual byssal thread. Additionally, byssal threads of control mussels were commonly broken at the proximal region, while those of noise-treated mussels were largely fractured at the adhesive plaque, indicating the weakest part of byssal thread shifted from the proximal region to the adhesive plaque upon exposure to playbacks of the pile-driving noise. These noise-induced changes led to a 16.95–44.50% decrease in the overall byssal attachment strength, greatly limiting mussels' ability to attach firmly on substratum. On the one hand, the weakened byssal attachment would undoubtedly increase dislodgement risk of mussel individuals in natural habitats and commercial suspension cultures (O'Donnell et al., 2013), and subsequently make them more vulnerable to predators and environmental turbulence (e.g., waves) at the individual level. On the other hand, the weakened byssal attachment might also hinder the formation and maintenance of mussel beds, and in turn impair the recruitment, anti-predation capability



and ecological functions of mussel population. We therefore conclude that anthropogenic noise (e.g., the pile-driving noise) has negative impacts on both mussel individuals and population through hampering byssal attachment, which in turn pose a great threat to mussel bed communities and the mussel aquaculture industry.

The noise-induced decline in byssal thread production (i.e., the decrease in the diameter and volume of byssal thread, and the down-regulation of genes encoding structural proteins of byssal thread) might be due to limited energy availability. It has been shown that the energetic cost of byssal thread production is substantial, requiring 8–10% of a mussel's total energy expenditure (Hawkins and Bayne, 1985; Lurman et al., 2013; Roberts et al., 2021). However, studies have suggested that anthropogenic noise adversely affected the feeding behavior and metabolism activity of marine bivalves, including mussels, which seems to result in energy deficiency (Peng et al., 2016; Spiga et al., 2016; Shi et al., 2019; Wale et al., 2019; Vazzana et al., 2020). For

example, exposure to ship noise playbacks led to a 12% reduction in oxygen consumption and an 84% decrease in filtration rate of the blue mussel *M. edulis* (Wale et al., 2019). Under this circumstance, energy availability could be limited and mussels must allocate a relatively few energy to biological processes that are less critical for individual survival (Lachance et al., 2011; Shang et al., 2021). We suggest that upon exposure to the pile-driving noise, mussels tended to limit the energy allocation to byssal attachment, thus leading to the down-regulation of genes expression levels of byssal proteins and the decrease in byssal thread diameter and volume of the thick shell mussel *M. coruscus* in this study. It is worth noting that there was no significant alteration in byssal thread number and length. However, these were not opposite to the results described above. As known, to attach firmly on the substrate, mussel individuals must produce enough byssal threads angled away from the shell to the ideal places where they want to anchor (Bell and Gosline, 1996). However, the down-regulation of genes expression levels

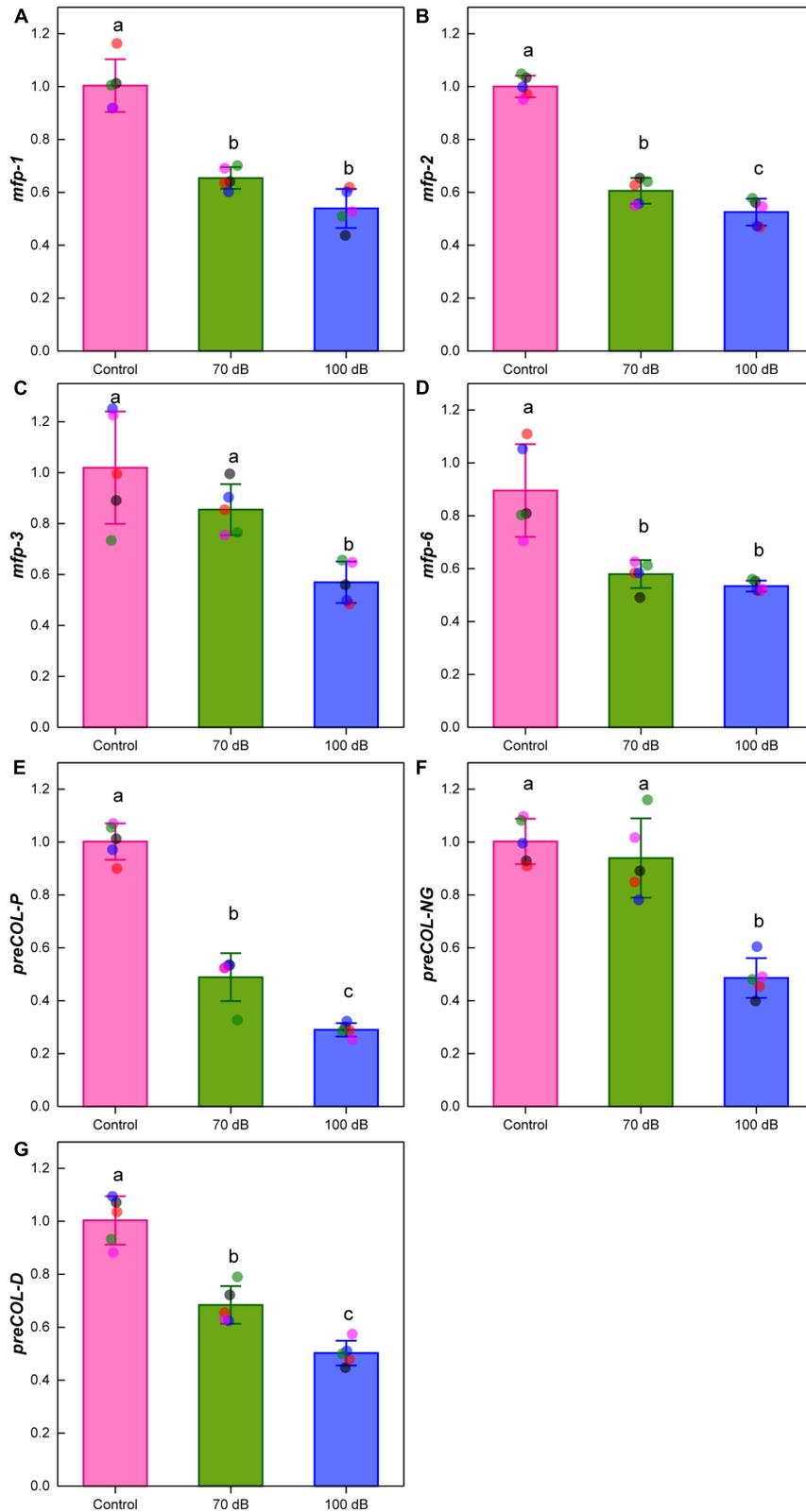


of byssal proteins and the decrease in thread volume suggest that less byssal proteins were synthesized, indicating there was insufficient materials for mussel individuals to produce byssal threads like before. Under this circumstance, decreasing thread diameter but keeping thread number and length should

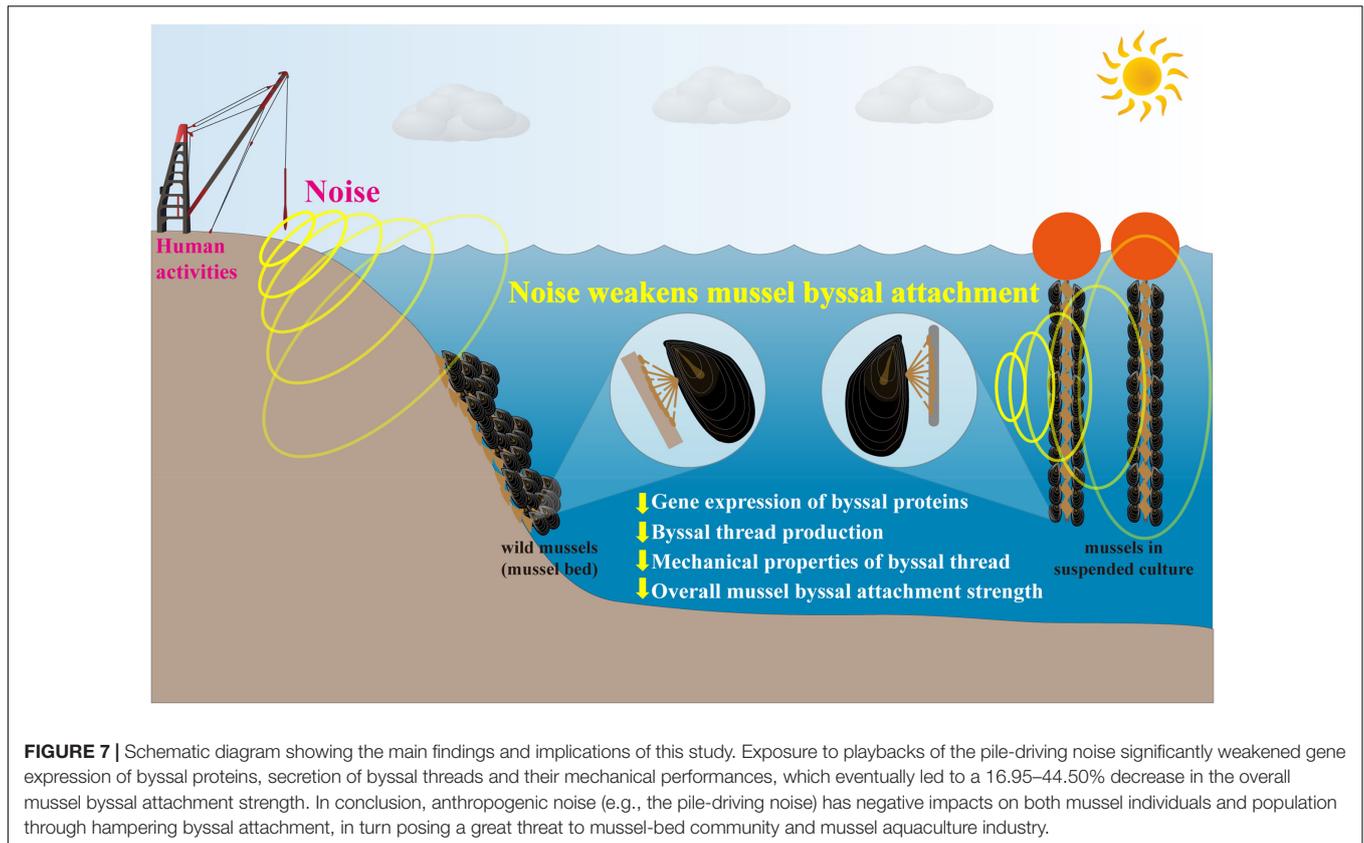
be an adaptive strategy to meet the lowest requirements for mussel attachment.

The noise-induced reduction in mechanical performances of byssal thread can be attributed to a series of reasons. Morphologically, the decrease in mechanical properties could be firstly owing to the morphological alterations of byssal thread. For example, the reduction in strength (i.e., breaking force), extensibility (i.e., breaking strain) and toughness might result from the thinner individual byssal thread. However, given the negative correlation between breaking stress and thread diameter reducing. This suggests that except morphological alterations, there must be other reasons contributing to the reduction in mechanical performances of byssal threads. Our results show that the mRNA levels of genes encoding structural proteins of byssal threads, including four mussel foot proteins (mfp-1, mfp-2, mfp-3, and mfp-6) and three precursor collagen proteins (preCOL-P, preCOL-NG, and preCOL-D), were significantly down-regulated upon exposure to the pile-driving noise. As known, the down-regulation of genes expression levels may not guarantee the reduction in translation of byssal proteins, but taken together with the decrease in thread volume, we suggest that levels of the four mussel foot proteins (mfp-1, mfp-2, mfp-3, and mfp-6) and the three precursor collagen proteins (preCOL-P, preCOL-NG, and preCOL-D) were, to some extent, down-regulated. It has been proven that the protective cuticle of byssal thread is comprised solely of mfp-1, while the inner collagenous core is mainly made of preCOL-P, preCOL-NG, and preCOL-D (~81% of the dry weight) (Silverman and Roberto, 2010). The other identified mussel foot proteins, including mfp-2, mfp-3, and mfp-6, are confined exclusively to the adhesive plaque (Lee et al., 2011). The preCOL-P is essential for the remarkable extensibility at the proximal region, and the preCOL-D provides the excellent stiffness property at the distal region of the byssal thread (Bandara et al., 2013). The preCOL-NG, distributed uniformly throughout the proximal and distal region, is considered to mediate the interaction between preCOL-P and preCOL-D, and thereby combines the remarkable extensibility and high strength properties of byssal thread (Bandara et al., 2013). These evidences show that the four mussel foot proteins (mfp-1, mfp-2, mfp-3, and mfp-6) and the three precursor collagen proteins (preCOL-P, preCOL-NG, and preCOL-D) are critical for the mechanical performances of byssal threads. Regarding this, the down-regulation of structural proteins of byssal threads should be one reason underlying the weakened mechanical properties and the shift of the weakest part of byssal thread.

It has been revealed that the mechanoreceptors within mussel foot participate in byssal threads secretion and adhesion (LaCourse and Northrop, 1978; Amini et al., 2017). Briefly, once mussel foot probes an ideal substrate, the mechanoreceptors and their downstream signal transduction pathways would be activated. The mussel then starts to fabricate byssal threads and subsequently attaches on the substrate. Importantly, the mechanoreceptors are also highly sensitive to mechanical waterborne vibration or acoustic particle motion and thus essential for mussel sound detection (Roberts et al., 2015;



**FIGURE 6 |** Effects of anthropogenic noise exposure on the expression levels of genes encoding structural proteins of byssal threads, including (A) mfp-1, (B) mfp-2, (C) mfp-3, (D) mfp-6, (E) preCOL-P, (F) preCOL-NG and (G) preCOL-D in the thick shell mussel *Mytilus coruscus* ( $n = 5$  per group, mean  $\pm$  SD). Means not sharing the same superscript are significantly different (Tukey's HSD,  $p < 0.05$ ).



**FIGURE 7 |** Schematic diagram showing the main findings and implications of this study. Exposure to playbacks of the pile-driving noise significantly weakened gene expression of byssal proteins, secretion of byssal threads and their mechanical performances, which eventually led to a 16.95–44.50% decrease in the overall mussel byssal attachment strength. In conclusion, anthropogenic noise (e.g., the pile-driving noise) has negative impacts on both mussel individuals and population through hampering byssal attachment, in turn posing a great threat to mussel-bed community and mussel aquaculture industry.

Vazzana et al., 2016). These suggest that mechanoreceptors mediate both mussel byssal attachment and sound detection. Thus, it is reasonable to infer that there might be any cross-talk between the two signal transduction pathways. In this way, the biological function of the mechanoreceptors in byssal attachment may be interfered by the coexisting noise. In turn, the process of byssal threads secretion and adhesion was disrupted, finally leading to the observed reduction in byssal thread production and mechanical performances in this study. In a word, the dysfunction, if any, of the mechanoreceptors might be one reason mediating the noise-induced alteration in mussel byssal attachment. Of course, empirical evidences are required to validate this assumption.

## CONCLUSION

In conclusion, this study found that 10-day continuous exposure to playbacks of the pile-driving noise significantly weakened gene expression of byssal proteins, secretion of byssal threads and their mechanical performances, which eventually led to a 16.95–44.50% decrease in the overall mussel byssal attachment strength (Figure 7). Given the essential functions of strong byssal attachment, these findings suggest that the increasing underwater anthropogenic noise is posing a great threat to mussel population, mussel-bed community and mussel aquaculture industry. We thus suggest that future work is needed to enhance our understanding of the impacts of anthropogenic noise on

marine invertebrates, especially these with limited locomotion ability, like bivalves.

## DATA AVAILABILITY STATEMENT

The original contributions presented in the study are included in the article/supplementary material, further inquiries can be directed to the corresponding author/s.

## AUTHOR CONTRIBUTIONS

XZ and GL conceived and designed this study. XZ and SS performed the whole experiments and collected the data. XZ wrote the manuscript, while other authors reviewed and revised the manuscript. GL provided oversight of the project. All authors gave final approval for publication and contributed to data analysis and interpretation.

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