



Effectiveness of Large-Scale Marine Protected Areas in the Atlantic Ocean for Reducing Fishing Activities

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The implementation of large-scale marine protected areas (MPAs) depends upon scarce conservation resources, while their effects on biodiversity conservation are rarely assessed to date. Quantitative evaluations are necessary to assess the effectiveness of large-scale MPAs in enhancing ecosystem resilience, protecting biodiversity, and mitigating expanding threats. In this study, the effectiveness of large-scale MPAs, which are remotely managed and in offshore areas of the southwestern Atlantic Ocean (Brazil), was assessed concerning the occurrence of fishing activities within their boundaries before and after their implementation. Two sets of MPAs surrounding the São Pedro and São Paulo archipelago (SPSP) and the Trindade-Martin Vaz Islands (TMV) were established in early 2018, each comprising one no-take (i.e., fully-protected) and one multiple-use (i.e., partially-protected) area. For this assessment, I used satellite detections of Vessel Monitoring System transmission to quantify the fishing pressure (i.e., “likely fishing days”) from commercial fisheries spanning 5 years (2015–2019). I then derived three metrics – fishing area, intensity, and density – to compare fishing activity within each MPA and year. The results showed that the effectiveness of the multiple-use MPAs was variable and contrasting, with SPSP experiencing a reduction in the fishing intensity and area and TMV experiencing an increase in both measures. An inverted pattern was evident for the no-take MPAs: while the one in the SPSP region experienced an increase in the fishing density after its establishment following a squeeze factor, the no-take MPA in the TMV region observed a decrease in the fishing density when comparing years before and after MPA implementation. These outputs can support managers in planning the implementation of further conservation strategies, such as monitoring and enforcement plans, and the analyses here also contribute to enhancing our understanding on the implications and challenges of adopting large-scale MPAs in the offshore environment as a high-profile strategy of ocean conservation.

Keywords: large marine protected areas, marine conservation, conservation assessment, commercial fishing, vessel tracking, vessel monitoring system, Brazil, fisheries

INTRODUCTION

Fishing is a leading cause of disturbances in the marine realm with consequences such as trophic cascade (Mumby et al., 2006; Shears et al., 2008) and loss of habitats (Kaiser et al., 2002; Lundquist et al., 2018). Moreover, overfishing in targeted and by-catch fisheries causes population decline in several species of the megafauna, including sharks, seabirds, and marine mammals (Dulvy et al., 2014; White et al., 2017). The global expansion of fisheries to meet the demand for fisheries resource extraction and the continued development of gear technology both have expanded and intensified the activity, with fishing occurring even in the remotest parts of the ocean (Sala et al., 2018).

Globally, fishing activity requires robust management measures to mitigate its impacts on marine biodiversity. Although there are several frameworks concerned with improving fisheries management (e.g., Booth et al., 2020), the designation of marine protected areas (MPAs) is the most applied tool to maintain biodiversity and fisheries at a sustainable level (Lubchenco and Grorud-Colvert, 2015). Several countries, including Brazil, are making compelling cases for historical progress toward achieving the international targets for marine protection under the Convention on Biological Diversity and the Sustainable Development Goals (Friedlander et al., 2016; Magris and Pressey, 2018; Claudet et al., 2021). As a consequence, recent years have also seen an increase in the development of large-scale MPAs (i.e., larger than 100,000 km²) over offshore and deep areas (Boonzaier and Pauly, 2016), following an *ad hoc*, opportunistic process (O'Leary et al., 2018). However, for these conservation efforts to drive outcomes for biodiversity, they must translate into significant mitigation of human impacts, particularly derived from fishing activity. This requires overcoming the monitoring and enforcement challenges associated with the large-scale governance of dynamic and remote seascapes (De Santo, 2013; Brooks et al., 2019).

Overall, there are several ways of measuring MPA effectiveness, and several frameworks have been proposed (Pomeroy et al., 2005; Pajaro et al., 2010; Zupan et al., 2018a). Percentages of an area under protection (i.e., MPA coverage), although commonly used, are misleading indicators of conservation success (Roberts et al., 2018). Indicators of MPA management effectiveness are intended to show how well MPAs are working towards their objectives (Pajaro et al., 2010), but they are usually evaluated using only managers' perceptions of good governance and MPA impacts (de Oliveira Júnior et al., 2021). Improvements of ecological conditions, such as the increase in species abundance, are seen as more accurate determinants of MPA effectiveness, but evaluations of offshore MPAs in remote areas are difficult due to data paucity and budgetary constraints related to the development of monitoring programs in such areas (Ban et al., 2017). Finally, a quantitative assessment of how well MPAs can abate the threatening processes provides an alternative, practical assessment of effectiveness (Zupan et al., 2018a) until detailed post-implementation monitoring data have been collected.

Vessel Monitoring System (VMS) data can help fill the gap in effectiveness assessments of large-scale, offshore MPAs and are widely used to evaluate fishing activity (Chang and Yuan,

2014; Delfour-Samama and Leboeuf, 2014; Rowlands et al., 2019). VMS can track vessel movements in near real-time using satellite transponders. Although the system is not tamper-proof (Appleby et al., 2018), it might be the only tool available to assess patterns of fishing activity and provides historical valuable information such as the vessel's identity, position, and associated fishing gear. These data can thus provide a unique baseline for determining whether MPAs are effective at reducing threats in the absence of other monitoring tools. Here, I used a long-term, large dataset tracking the movements of commercial fishing vessels before and after the two of the largest MPAs in the southern Atlantic Ocean (within Brazil's exclusive economic zone) were established – i.e., between the years 2015 and 2019 – to provide evidence of their effectiveness at reducing fishing pressure.

METHODS

Case Study Description

To meet global MPA commitments and in recognition of the relatively poor development of protected areas associated with the marine realm in Brazil, the Ministry of the Environment declared two sets of large-scale MPAs in the Southwestern Atlantic Ocean in early 2018: (i) two MPAs surrounding the São Pedro and São Paulo Archipelago (SPSP), which is formed by rocky islands in the mid-Equatorial North Atlantic Ocean (0°55'N; 29°20'W), distant about 1,000 km from the mainland; and (ii) two MPAs surrounding the Trindade Island and the Martin Vaz Archipelago (TMV), which is formed by the emerged part of the Vitória – Trindade submarine chain in the southwestern tropical Atlantic Ocean (29°18'S; 20°30'W), distant 1,160 km from the mainland (see **Supplementary Figure 1** for detailed zoomed views of both regions). These islands have among the highest fish biomass across Brazilian reefs (Morais et al., 2017), notable endemism (Simon et al., 2013; Pinheiro et al., 2020), and globally threatened fauna (Almeida et al., 2011; Duarte-Neto et al., 2012). Despite their biodiversity significance, both regions are also threatened by commercial fisheries and climate change (Magris et al., 2020). The MPAs comprise the territorial sea and exclusive economic zones of the islands.

The no-take MPA at the SPSP region (i.e., considered to be fully protected, and referring to the IUCN category III) was created to protect the southern portions of the archipelago and seamounts of the Mid-Atlantic Ridge, with a total size of 47,263.18 km². The multiple-use MPA (i.e., considered to be partially protected, and referring to the IUCN category IV) embraces the no-take one, including the majority of the small islands and a large open-ocean area, with a size of 407,052.36 km². The no-take MPA at the TMV region (same IUCN category as the no-take MPA at the SPSP region) was created to protect portions of the shallow reef habitats and the terrestrial environment, with a size of 67,696.71 km². This no-take is also nested within a multiple-use MPA of the same category as the SPSP described above, with a size of 402,377.1 km². Although all these MPAs have not been fully implemented (i.e., they have not elaborated their management plans), they correspond to about 95% of the total marine area protected in Brazil.

Commercial Fishing Activity

I used a 5-year dataset (January 2015–December 2019) of the spatial distribution of commercial fishing activity entering the study regions. The dataset was obtained from the processed VMS data provided by the National Program for tracking fishing vessels in Brazil (i.e., PREPS). The movement of fishing vessels is remotely tracked using a transponder, which transmits signals of vessel's position and behavior via satellite to ground stations on an hourly basis. To identify the behavior of vessels (e.g., navigating, fishing, and mooring), the signals are automatically processed based on spatial movement patterns and speed. I filtered out those records not associated with fishing activity and then included in the analysis only the positions by which vessels are very likely fishing. I identified a total of 1,844,902 transmitted signals that were associated with 152 active vessels and indicative of fishing operations within the study regions over the studied period. VMS is legally required for all fishing vessels larger than 15 m in Brazil, which is suitable for assessing fishing pressure in remote, offshore areas.

By using a database of fishing gears associated with each vessel, I could obtain more details about the fishing operations. For example, I found that most of the fishing operations were associated with pelagic longline (i.e., >80%), although I also registered other fishing gears such as bait boat – pole-and-line fishing, pelagic handline, and bottom trawl (registered exclusively for the TMV region). This information requires certain caution because the type of fishing gear associated with each vessel can be modified through the renewing process of fishing licenses, without being automatically updated into the system.

I collapsed the data points from all transmitted signals for each vessel into single days to derive a metric of fishing activity (i.e., “likely fishing days”) and accumulated this value for all vessels per 10×10 km grid cell within each year assessed. I also assigned the metrics to each no-take or multiple-use MPA by overlying the MPA boundaries and fishing data. I extracted the MPA shapefiles from the dataset held by the Brazilian Ministry of Environment¹. To determine the spatial similarity of total fishing days within each MPA among years, I calculated the Kendall correlation coefficient. This coefficient is a pairwise statistic that measures the degree of agreement among years.

Lastly, I summarized the following measures of fishing pressure within each MPA and year: (i) the total number of cells with fishing days >1 as a proxy of “fishing area”; (ii) the sum of fishing days as a proxy of “fishing intensity”; and (iii) the quotient of the total number of fishing days and the fishing area as a measure of “fishing density.” Following White et al. (2020), I sought to partially control for changes in fishing pressure not related to the modification of the protection status of the study regions. For this last set of analyses, I compared each measure of fishing pressure calculated as above against the same metrics associated with cells randomly selected across Brazil's EEZ, and limited to the corresponding total size of one set of large-scale MPAs (i.e., 455,000 km²). I generated the random selection of cells as described in Magris et al. (2020). I restricted the cells selection within other areas of Brazil's EEZ because international

waters can have different fisheries management regulations. I excluded fully-protected MPAs from the random selection as they might be effective at restricting fishing activities within their boundaries. I also allowed coastal areas to be selected because the commercial fishing fleet using the assessed gears is widely distributed across the entire Brazil's EEZ (Magris et al., 2020).

RESULTS

Considering the whole period from 2015 to 2019, I identified 28,226 total days of fishing activity in the SPSP region (93% of them within the area of the multiple-use MPA and about 7% within the no-take one) and 54,164 in the TMV region (82% of them within the area of the multiple-use MPA and about 18% within the no-take one).

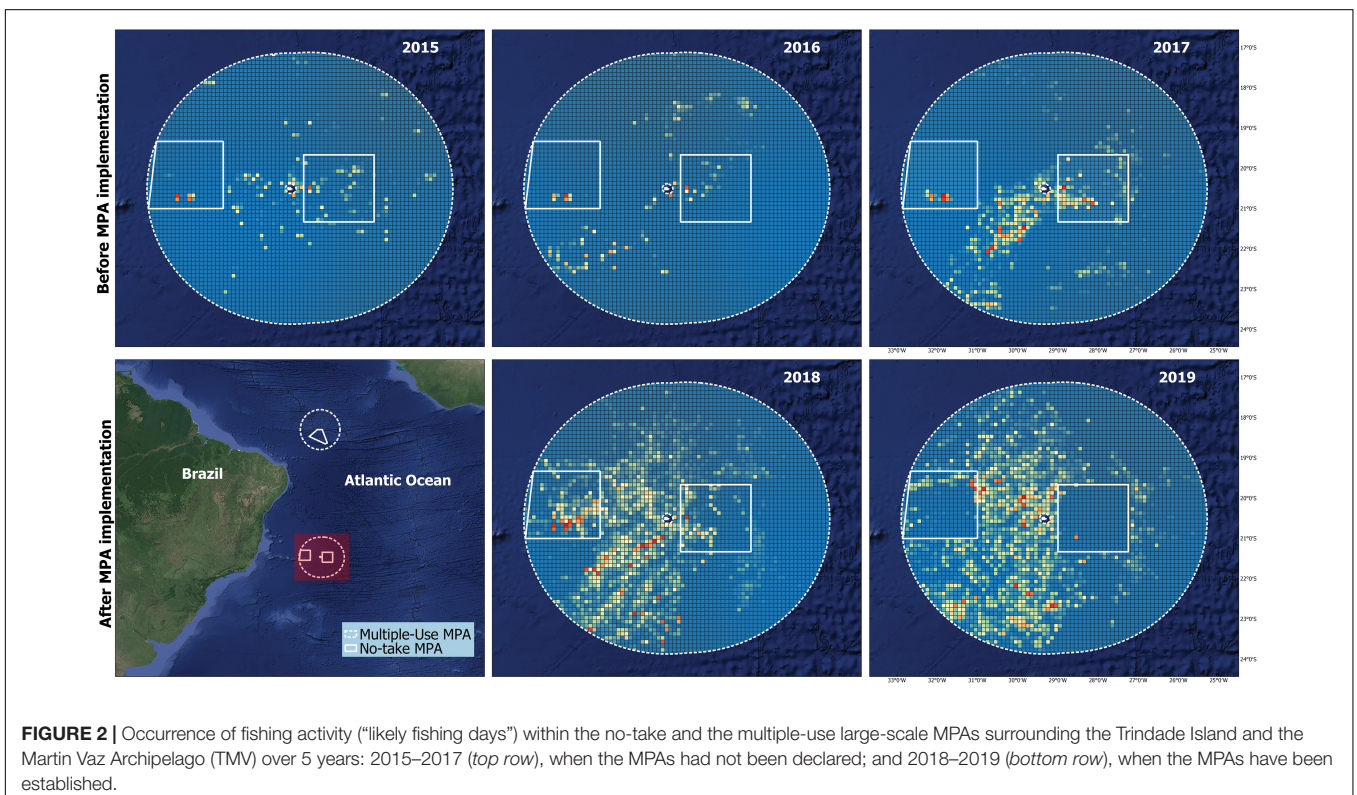
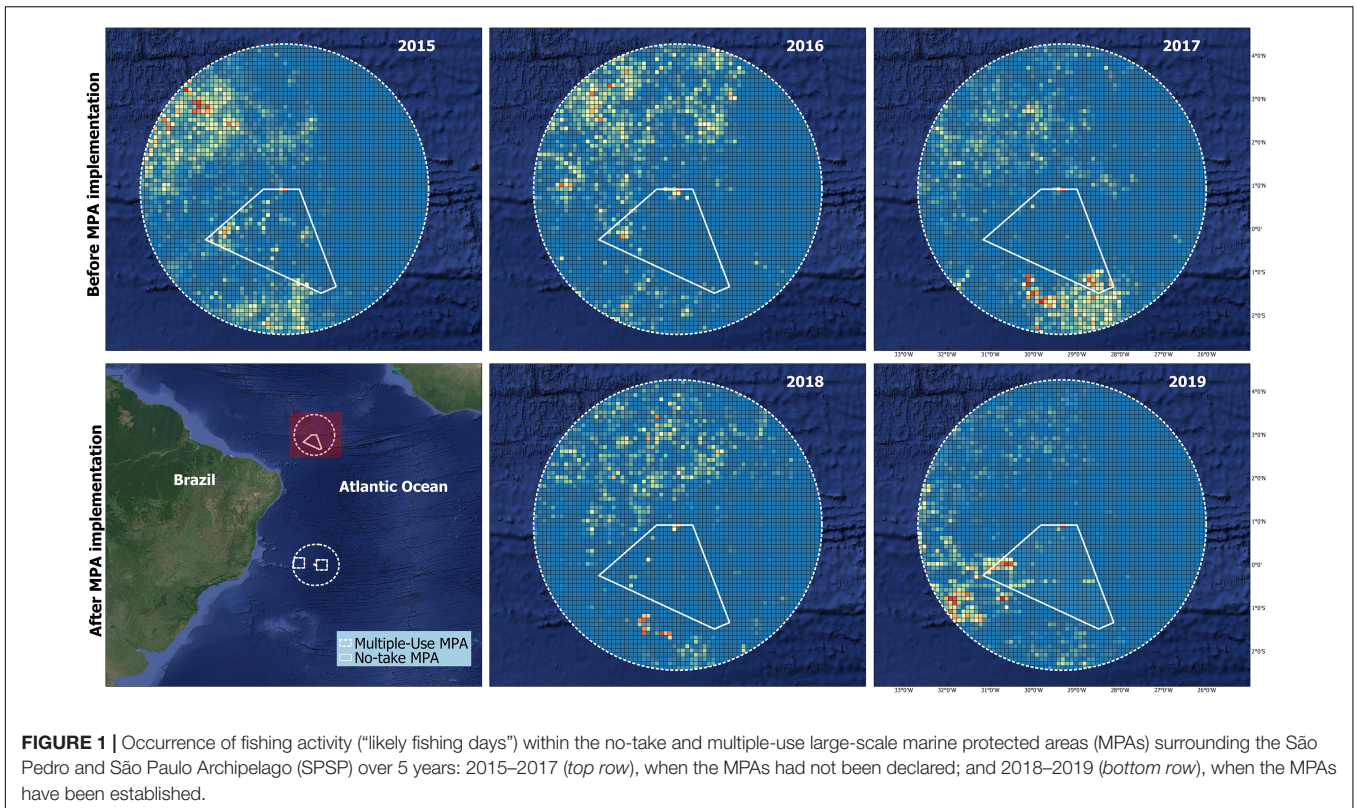
For the SPSP region (**Figure 1**), I recorded a hotspot of likely fishing days on the northwestern portion of the multiple-use MPA, between the years 2015 and 2016, and on the southern portion of this MPA for the year 2017. Hotspots of fishing within the no-take MPA followed the same spatial pattern of the multiple-use one for the years prior to MPA establishment. After MPA creation, hotspots of fishing activity were well distributed in 2018, and more spatially concentrated on the western portions of both MPAs in 2019.

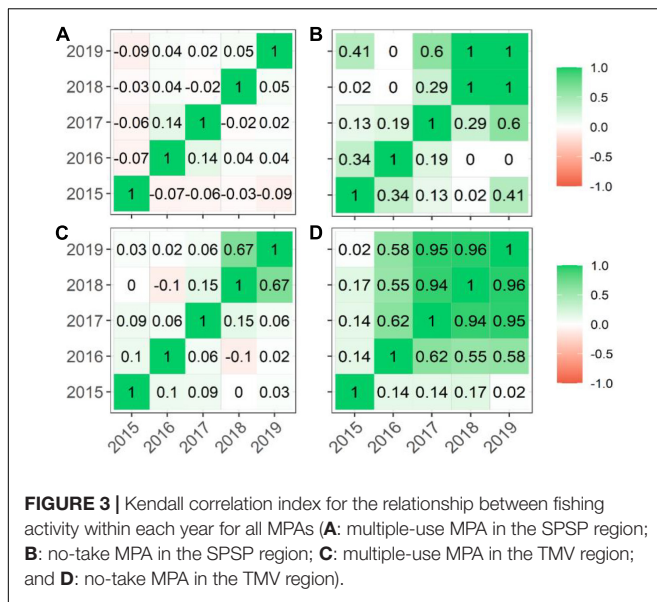
For the TMV region (**Figure 2**), I identified that hotspots of fishing activity clustered on the central parts of the region in 2017 and were well dispersed in the previous years. After MPA establishment, large areas of the multiple-use MPA could be identified as hotspots of fishing activity, mainly in its left half, closer to the mainland. While hotspots of fishing activity were identified within no-take MPAs in 2018, fishing activity was substantially reduced in 2019 for this MPA.

When I performed the correlation matrix analysis, three results emerged (**Figure 3**): (i) there was no agreement between the fishing activities occurring at each year within the multiple-use MPA in the SPSP region (Kendall coefficient: -0.09 – 0.14 ; **Figure 3A**); (ii) there was a strong agreement between the fishing activities occurring at several years before and after MPA establishment within the no-take MPA in the TMV region (Kendall coefficient: 0.94 – 0.96 ; **Figure 3D**); and (iii) there was only a substantial agreement between the fishing activities occurring in the years 2018–2019 within the no-take MPA in the SPSP region (Kendall coefficient = 1; **Figure 3B**), and within the multiple-use MPA in the TMV region (Kendall coefficient = 0.67 ; **Figure 3C**). While the first two cases imply that these specific MPAs might have little influence on the spatial patterns of fishing activity, the second situation indicates that the creation of those MPAs might have affected spatial patterns of fishing activity.

Analysis of fishing pressure within each MPA (**Figure 4**; top panels) revealed that the amount of fished area was reduced after MPA establishment for the SPSP region, which was not the case for the TMV region. The observed reduction in the SPSP region was followed by a decrease in the fishing intensity within the multiple-use MPA (middle panel) and an increase in the fishing density within the no-take MPA after their establishment (bottom panel). This was because fishing activity became more

¹<http://www.mma.gov.br/areas-protetidas/cadastro-nacional-de-ucs>



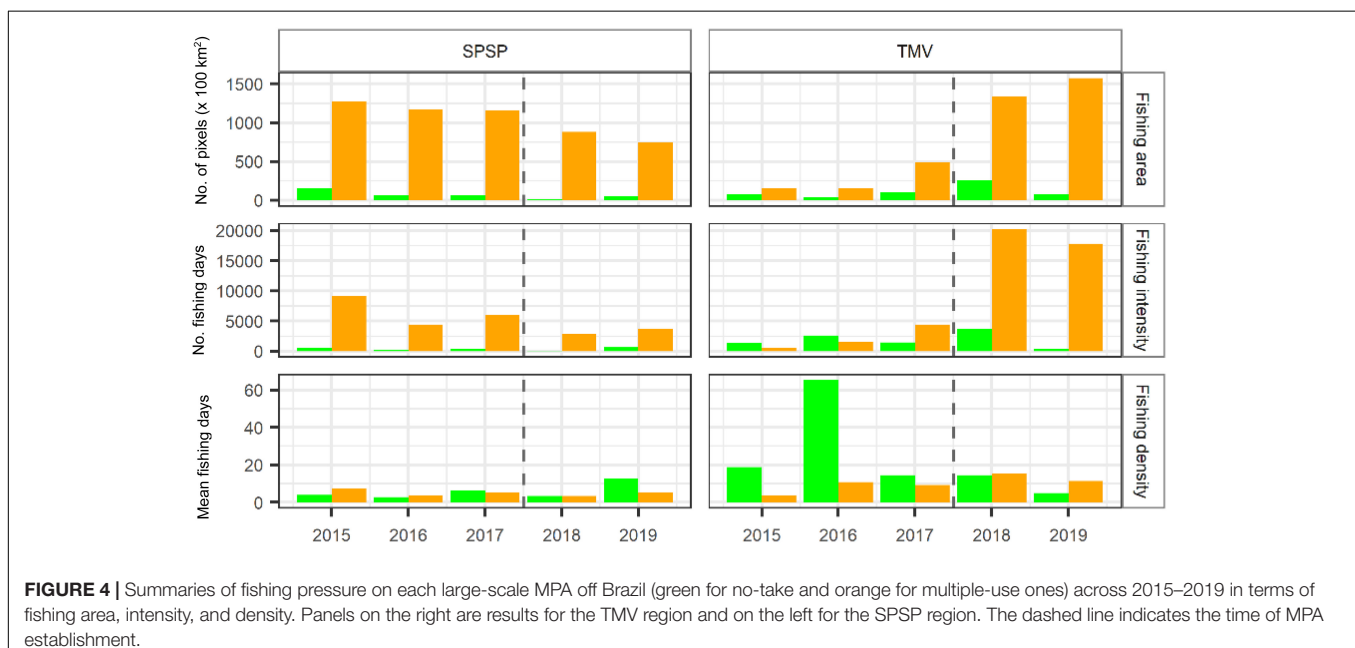


concentrated over smaller areas of this MPA. Overall, I also found a large increase in the fishing intensity within the multiple-use MPA for TMV (middle panel), which was also reported for the no-take MPA at least for the first year after MPA establishment (2018). When assessing the measure of fishing density for this region, I found that there was a small reduction of this measure in the no-take MPA. It was noticeable that fishing density seems not to change with MPA establishment for the multiple-use MPAs in both regions. I identified comparable and high levels of fishing pressure on the random areas that did not restrict commercial fisheries throughout the time assessed, regardless of the measure of fishing pressure used (**Supplementary Figures 2–4**).

DISCUSSION

The results presented here provide the first assessment of the conservation effectiveness of large-scale MPAs in the Southwestern Atlantic Ocean and contribute to the ongoing discussion about the benefits of this conservation strategy to mitigate threats from fishing (O’Leary et al., 2018). Results indicated that the effectiveness of the large-scale MPAs was variable and depended on the measure of fishing pressure used. Overall, there was a reduction in the fishing area and intensity in the SPSP region, but fishing became particularly intense over smaller areas (“squeeze factor”), particularly within the no-take MPA. On the other hand, while fishing area and intensity increased for the TMV region, a reduction in the fishing density was observed because the activity became spread over larger areas. More positively, a reduction in all measures of fishing pressure became apparent in 2019 for the no-take MPA in TMV. Fishing pressure is thus significant within these large-scale MPAs and monitoring and enforcement efforts to effectively promote their reduction over time needs to be encouraged.

I estimated that between 3 and 10% of the no-take MPA in the SPSP region, and between 12 and 38% of the no-take MPA in the TMV region remained potentially fished. This supports the existence of illegal fishing even in the remote places of the ocean as previously identified (Arias et al., 2016). Although there is some uncertainty in the VMS data to provide evidence of the magnitude of fishing activity as a result of the need to combine this technology with other forms of evidence gathering (Appleby et al., 2018), this is unlikely to change this result significantly. With the challenges associated with patrolling offshore and remote areas in the ocean, a more realistic approach to build evidence of illegal fishing would be to combine different data sources that are sufficient to lead to a prosecution, making enforcement effective.



As previously suggested (Magris and Pressey, 2018), the effects of multiple-use, large-scale MPAs have been marginal, at least in the short term. Indeed, multiple-use MPAs have been claimed to have a limited impact on biodiversity conservation (Giakoumi et al., 2017; Zupan et al., 2018b) when assessed in terms of improving biodiversity conditions *in situ*. On the other hand, some have argued that their contribution to ocean conservation would be to prevent mining expansion in the future (Giglio et al., 2018; Miller et al., 2018). While these conservation outcomes are not realized, several management recommendations could be derived for this category of MPAs to improve its effectiveness in the present. Areas identified as more important for biodiversity within their boundaries (Magris et al., 2020; Vilar et al., 2020) could be targeted for more strict fishing regulations through a zoning process. Moreover, fishing activity could be particularly required to adopt practices that reduce the risk of fishing mortality (Booth et al., 2019). Strengthening regulations and establishing adequate governance are key ingredients for increasing conservation benefits expected from effective MPAs.

Recent evidence has suggested that large-scale MPAs maintain fishing levels at a low level (White et al., 2020). The results presented here do not support this pattern at least for those MPAs affording partial protection. Multiple-use MPAs did not interfere in the spatial patterns of fishing activity over the time assessed and, in some instances, fishing intensity within no-take MPAs had even increased shortly after the creation of MPAs. These contrasting findings can be explained by the intrinsic difference between the sources of fishing detection systems (VMS *versus* Automatic Information Systems – AIS). At least in Brazil, the AIS system misses a considerable fraction of fishing vessels, rendering assessments based on that system misleading. For example, tracking the global footprint of fisheries using AIS across the national waters off Brazil, as well as other exclusive economic zones, has shown minimal fishing effort within this area (Kroodsmas et al., 2018), which is a misreport of the activity.

A major challenge to quantifying the conservation effectiveness of large-scale MPAs is the dynamic context in which threats operate over vast areas and data availability. The assessment of the threat reduction capacity of these MPAs might be influenced by other environmental conditions such as ocean currents, temperature, and distance from the mainland. Ongoing efforts to gather and analyze data for their influence on the occurrence of fishing activity will possibly result in the revised estimates of the conservation effectiveness of large-scale

MPAs, allowing more comprehensive assessments of their role in reducing fishing pressure.

Though there are venues for further development and refinement, this study constitutes an important first step in quantifying the effects of large-scale MPAs off Brazil. The case study highlights that, unlike other regions, fishing activity remains operating within multiple-use MPAs, and that avoiding illegal fishing within no-take MPAs is an urgent need. While remote sensing technologies provide spatially and temporally continuous assessment of fishing activities, it would need to be combined with other evidence-based tools on fishing effort for increasing existing levels of compliance and enforcement. The variations in fishing pressure among MPA types over time affirm the dynamic nature of managing offshore marine systems.

DATA AVAILABILITY STATEMENT

The processed data supporting the conclusions of this article will be made available by the authors, without undue reservation.

AUTHOR CONTRIBUTIONS

RAM was the sole author of this manuscript, developed the research question and protocol, conducted data collection and analysis, and wrote the content of the manuscript.

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SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2021.711011/full#supplementary-material>

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