



## OPEN ACCESS

## EDITED BY

Pedro Álvarez-Álvarez,  
University of Oviedo, Spain

## REVIEWED BY

Christer Bjorkman,  
Swedish University of Agricultural Sciences,  
Sweden

María Menéndez-Miguélez,  
Instituto Nacional de Investigación y  
Tecnología Agroalimentaria (INIA), Spain  
José Carlos Pérez-Girón,  
University of Granada, Spain

## \*CORRESPONDENCE

Carolina Mayoral

✉ c.mayoral@bham.ac.uk

A. Robert MacKenzie

✉ a.r.mackenzie@bham.ac.uk

RECEIVED 16 August 2023

ACCEPTED 25 September 2023

PUBLISHED 17 October 2023

## CITATION

Mayoral C, Ioni S, Luna E, Crowley LM,  
Hayward SAL, Sadler JP and MacKenzie AR  
(2023) Elevated CO<sub>2</sub> does not improve  
seedling performance in a naturally  
regenerated oak woodland exposed to biotic  
stressors.

*Front. For. Glob. Change* 6:1278409.

doi: 10.3389/ffgc.2023.1278409

## COPYRIGHT

© 2023 Mayoral, Ioni, Luna, Crowley, Hayward,  
Sadler and MacKenzie. This is an open-access  
article distributed under the terms of the  
[Creative Commons Attribution License  
\(CC BY\)](https://creativecommons.org/licenses/by/4.0/). The use, distribution or reproduction  
in other forums is permitted, provided the  
original author(s) and the copyright owner(s)  
are credited and that the original publication in  
this journal is cited, in accordance with  
accepted academic practice. No use,  
distribution or reproduction is permitted which  
does not comply with these terms.

# Elevated CO<sub>2</sub> does not improve seedling performance in a naturally regenerated oak woodland exposed to biotic stressors

Carolina Mayoral<sup>1,2\*</sup>, Susanna Ioni<sup>1,2</sup>, Estrella Luna<sup>1,2</sup>,  
Liam M. Crowley<sup>1,2,3</sup>, Scott A. L. Hayward<sup>1,2</sup>, Jon P. Sadler<sup>1,4</sup> and  
A. Robert MacKenzie<sup>1,4\*</sup>

<sup>1</sup>Birmingham Institute of Forest Research, University of Birmingham, Birmingham, United Kingdom,

<sup>2</sup>School of Biosciences, University of Birmingham, Birmingham, United Kingdom, <sup>3</sup>Department of Biology, University of Oxford, Oxford, United Kingdom, <sup>4</sup>School of Geography, Earth and Environmental Sciences, University of Birmingham, Birmingham, United Kingdom

**Introduction:** Insect herbivores and biotrophic pathogens are major stressors influencing natural regeneration in woodlands. Information on the effect of elevated CO<sub>2</sub> (eCO<sub>2</sub>) on plant-insect-pathogen interactions under natural conditions is lacking.

**Methods:** We studied the effects of eCO<sub>2</sub> on leaf-out phenology, as well as on levels of insect herbivory and powdery mildew (PM), i.e., reduction of leaf photosynthetic material. We then assessed the combined impacts of these biotic stressors and eCO<sub>2</sub> on seedling photosynthesis and growth. A total of 92 naturally recruited and 114 potted seedlings of 5 temperate tree species (*Quercus robur*, *Acer pseudoplatanus*, *Corylus avellana*, *Crataegus monogyna*, and *Ilex aquifolium*) within a mature oak woodland were studied.

**Results:** We found that eCO<sub>2</sub> advanced leaf-out phenology and was a significant explanatory variable for growth and physiological performance in potted seedlings. Potted oak seedlings experienced 11-fold higher tissue loss from insect herbivory than natural seedlings. The earliest leaf-flushing species, hawthorn, and the evergreen holly were resistant to insect attack and were not affected by PM. Oak was defoliated most but showed the highest regeneration capacity. Hazel was more resistant to PM infection than oak and sycamore. Despite being highly infected by PM, sycamore was less affected than oak. The more vigorous sycamore and oak seedlings suffered more severe PM disease.

**Conclusion:** No evidence emerged that eCO<sub>2</sub> enhances natural regeneration under biotic stress for any of the species studied.

## KEYWORDS

FACE forest, defoliation, trophic, phenology, regeneration, leaf physiology

## 1. Introduction

Insect herbivores and pathogenic fungi are considered the most common biotic stress agents in temperate forests (Jactel et al., 2009) and represent important barriers to seedling establishment. This combination of stress factors may also have serious consequences for forest biodiversity and ecosystem functions (Hasegawa et al., 2018; Pureswaran et al., 2018). Temperate forests are among the biggest sinks of emitted carbon (Le Quéré et al., 2018) in a world where anthropogenic activities have led to a 45% increase of atmospheric CO<sub>2</sub> concentration, and consequently a mean 0.85°C increase in global temperature since the Industrial Revolution. Although increasing CO<sub>2</sub> (“carbon fertilisation”) might result in an increase of the forest carbon sink (Ciais et al., 2014), the complex interplay between biotic stressors and climatic factors, combined with differing responses across tree species, makes it very difficult to predict overall consequences of climate change on forest health and forest carbon assimilation capacity (Jactel et al., 2019).

While investigating the direct impact of rising CO<sub>2</sub> on the ecophysiology and functional response of trees can be accomplished in controlled environment chambers (e.g., Liu et al., 2004; Aranda et al., 2020), or in open-top chambers (e.g., Norby et al., 1995; Vanaja et al., 2006), investigating ecosystem-level responses requires large-scale forest Free-Air CO<sub>2</sub> Enrichment (FACE) facilities (e.g., Karnosky et al., 2001). FACE facilities provide the most ecologically realistic environment to date for evaluating CO<sub>2</sub> effects on forest ecosystems. The first-generation of FACE experiments were established in temperate and young monocultures (Nowak et al., 2004; DOE, 2020) and substantially improved our understanding of the capacity of forests to sequester carbon under elevated CO<sub>2</sub> (eCO<sub>2</sub>) (Nowak et al., 2004; Ainsworth and Long, 2005; Norby et al., 2005; Norby and Zak, 2011; Medlyn et al., 2015). Second generation FACE experiments, the Eucalyptus free air carbon dioxide enrichment experiment facility (EucFACE) in Australia and the Birmingham Institute of Forest Research FACE Facility (BIFoR-FACE) in UK, have advanced this endeavour to include old-growth forests, allowing researchers to investigate whether eCO<sub>2</sub> alters a broader range of forest processes in the canopy, subcanopy, herb-and-regeneration layer (Gherlenda et al., 2016a; Jiang et al., 2020; Roberts et al., 2022), and soils (Martins et al., 2021). EucFACE was established in a eucalyptus forest (Duursma et al., 2016), whereas BIFoR-FACE, the site of the present study, is in a more complex and mixed canopy (Hart et al., 2020) dominated by English oak (*Quercus robur* L.) in the top canopy, and co-dominated in the understorey by other deciduous and evergreen broadleaves such as sycamore (*Acer pseudoplatanus* L.), hazel (*Corylus avellana* L.), hawthorn (*Crataegus monogyna* Jacq), and holly (*Ilex aquifolium* L.).

In temperate deciduous forests, leaf-chewing lepidopteran larvae are a key feeding guild, such larvae can remove significant amounts of photosynthetically active material early in the season (Schowalter et al., 1986). Large outbreaks of insect herbivores are commonly followed by newly developed leaves, prone to fungal infection causing diseases such as powdery mildew (PM) (e.g., *Erysiphe* spp.) (Field et al., 2019). PM diseases are caused by a large family of biotrophic fungal pathogens, comprising ca. A total of 900 species of the phylum Ascomycota growing superficially on plant surfaces (Takamatsu et al., 2013). During the

growing season, hyphae are produced on both upper and lower leaf surfaces, diverting plant photosynthates for the benefit of the fungus (Glawe, 2008). At early stages of PM infection, cells are not killed and photosynthesis can be stimulated, usually followed by a fast decline in activity leading to the desiccation of the leaf and subsequent death (Wang et al., 2014). Several species of PM infect oak, representing a serious threat to natural oak regeneration (Woodward et al., 1929; Lanier et al., 1976; Takamatsu et al., 2007). Other PMs infect sycamore (Talgo et al., 2011), hazel (Arzanlou et al., 2018), and hawthorn (Khairi and Preece, 1978). This disease is the most common foliar pathogen on oaks in Europe (Marçais and Desprez-Loustau, 2014) and represents a substantial biotic pressure on the species the BIFoR-FACE (Supplementary Figure 1).

It is expected that eCO<sub>2</sub> will stimulate leaf-level photosynthesis (Norby and Zak, 2011; Ryan, 2013). However, it is not clear whether this will result in a more attractive food source to flush-feeding insects or not, since palatability could result altered if the C:N ratio increases (Du et al., 2019) or if the plant invests more carbon in defensive compounds (Robinson et al., 2012). Stimulation of photosynthesis will potentially benefit biotrophic pathogens which prefer vigorous hosts to physiologically stressed plants (Jactel et al., 2012; Che-Castaldo et al., 2019; Field et al., 2019). In contrast, increased levels of biotic stress are associated with poorer status of leaf physiological parameters such as photosynthetic rates, fluorescence of chlorophylls, or nitrogen content (Moriondo et al., 2005; Daud et al., 2012; Watanabe et al., 2014; Madriaza et al., 2019). Thus, increased insect herbivory may negatively impact biotrophic pathogens. An additional complicating factor would be if eCO<sub>2</sub> stimulates photosynthesis and the development of the photosynthetic apparatus faster during the spring, due to more carbon availability (Linkosalo et al., 2017), thus advancing leaf-out phenology. This may entail important consequences for ecosystem structure and function in deciduous forests since leaf-out phenology defines the photosynthetically active period and the uptake of resources by deciduous trees (Polgar and Primack, 2011), and determines the duration of the growing season. While there is broad evidence that warming results in an advance of spring phenology (Chang et al., 2018; Wang et al., 2020a), little is known about the effect of eCO<sub>2</sub> on leaf-out phenology and leaf development after leaf flushing. Faster leaf development may also influence protective functions, for instance defence against insect herbivory (Damestoy et al., 2021) or late frost tolerance (Mayoral et al., 2015).

After bud burst, the leaves of deciduous species achieve full photosynthetic capacity at different times, depending on the species and the life stage. In temperate species, the lag from bud burst to full photosynthetic capacity has been found to range from around 2 weeks for *Betula pendula* (silver birch) up to 70 days for *Q. robur* (Morecroft et al., 2008; Linkosalo et al., 2017). A differential effect of eCO<sub>2</sub> on leaf-out phenology of coexisting tree species might, therefore, alter responses to spring abiotic and biotic stress (Pérez-Ramos et al., 2020) which in turn results in differential fitness and niche adaptation across forest tree species.

In this study, we examine the effects of eCO<sub>2</sub> on key aspects of both natural and potted seedlings performance (leaf-out phenology, growth, leaf photosynthetic parameters) and the interaction of these parameters with defoliating insects and PM, during a year with high pressure from these biotic factors. While natural seedlings emergence occurs under different growing

conditions (light, substrate, and water availability) following a random pattern, introducing potted seedlings within the FACE system allowed us for a more homogeneous growing condition to assess more accurately differences between species. Our objectives were to explore whether a higher concentration of CO<sub>2</sub> (1) would impact on leaf-out phenology and would differentially accelerate spring leaf development of coexisting tree species; (2) would attract generalist flush-feeding insects equally to tree species with different spring leaf development; (3) would impact on the defoliation intensity and the subsequent leaf regeneration capacity of the species; (4) would stimulate photosynthesis of the studied species and PM severity; and (5) would influence the growth of the species.

## 2. Materials and methods

### 2.1. Study site

Experiments were conducted in the BIFoR-FACE facilities (Hart et al., 2020), located in Mill Haft, Staffordshire, United Kingdom (52.8°N 2.30°W). The structure of this mature, temperate, oak dominated woodland is complex, comprising trees of multiple stature and age. The most representative understorey species, accompanying the canopy oaks (*Q. robur* L.), are sycamore (*A. pseudoplatanus* L.), hazel (*C. avellana* L.), hawthorn (*C. monogyna* Jacq), and holly (*I. aquifolium* L.) (Supplementary Table 1). The woodland is currently unmanaged although hazel was coppiced until about 30–40 years ago and has since regrown a dense, multi-stemmed canopy. The herb layer is dominated by bramble (*Rubus fruticosus* L.), ivy (*Hedera helix* L.), bluebell (*Hyacinthoides non-scripta* L.; see Crowley et al., 2021), and fern (*Dryopteris* sp.), with relatively high variability of the presence and abundance of each within the forest. Browsing and seed predation by vertebrates have been controlled since 2016, with deer excluded by fencing and squirrels numbers reduced (although not totally eliminated) by trapping (Bradwell et al., 2022).

Elevated CO<sub>2</sub> Fumigation began in April 2017 and will last until at least 2026. Six roughly cylindrical ring structures, as high as the tree canopy (c. 25 m) and 30 m wide support pipes that can deliver CO<sub>2</sub> in such a way that the woodland patches inside the array are immersed in eCO<sub>2</sub> to the treatment level from ground to top canopy, while the rest of the woodland remains largely unaffected (Hart et al., 2020). Three of these arrays are used to immerse patches in an atmosphere with CO<sub>2</sub> elevated 150 ppm above ambient (i.e., in 2019, to ~555 ppm). The other three arrays receive unaltered forest air and so remain at ambient (i.e., ~405 ppm, the global mean at the start of 2019)<sup>1</sup> (Supplementary Figure 1). Three non-infrastructure, ‘ghost,’ arrays with the same area were set up in order to assess the effect of the built infrastructure on the woodland. CO<sub>2</sub> fumigation is carried out during daylight hours of the growing season. Start-up dates for eCO<sub>2</sub> were 27th and 29th of March in 2018 and 2019, respectively, and the switch-off was on the 31st of October in both years.

### 2.2. Plant material

In June 2018, 121 naturally regenerated seedlings with a diameter below 1 cm at the root collar were identified across the experimental patches. We surveyed the entire accessible area at each array except areas occupied by dense bramble where the chance of finding tree regeneration was minimal due to low light levels. The least frequent species found were hazel and oak. The sample of 121 seedlings included the entire regeneration layer found in the experimental arrays for hazel and oak (12 and 65 seedlings/ha for hazel and oak, respectively, in 2018). Other species were more abundant and not all individuals were used for this study. There were not enough oak individuals in the ambient arrays and so we used seedlings in the ghost arrays to increase the sample size under ambient CO<sub>2</sub> conditions. In 2019, 86 of these seedlings were found (24 hawthorn, 5 hazel, 18 holly, 18 oak, and 22 sycamore). Six new oak seedlings were found in 2019 and these were added to increase sample size. Hazel regeneration was again limited in 2019 therefore no new seedlings were found. The exact sample size for different measurements varied throughout the study and is summarised in Table 1.

In July 2018, 120 (4 seedlings × 5 species × 6 arrays) 1-year-old, nursery-grown seedlings of the same 5 species (oak, sycamore, hazel, hawthorn, and holly) were introduced into the 6 infrastructure arrays (4 seedlings per species in each array). All were from the Midlands region of England (provenance area 403). Potted seedlings were used in order to more closely control the growing conditions, e.g., water nutrient and light availability. We used 7-litre (42 cm high) pots to minimise root growth constraints. All pots were filled with the same fertilised peat substrate and kept at field water capacity by means of self-irrigated trays (76 cm × 17 cm × 3.5 cm; length × width × height). The base tray provided a reservoir of 3.4 litres of water, and each tray contained 5 potted seedlings, one per species. In contrast to the naturally-regenerated seedlings, potted seedlings were grouped and placed at two different locations within each array (two trays per location with a total of 10 pots) under similar light environment on raised walkways to avoid competition with the herbaceous layer. Walkways were installed 30 cm above the ground surface (including in the ‘ghost’ arrays) to protect the experimental site from footfall (see Hart et al., 2020). Herbaceous plants growing inside of the pots were manually removed to avoid competition. One hazel and five hawthorns died during the experimental period in winter leaving a total of 114 potted seedlings (Table 1).

### 2.3. Leaf-out phenology and leaf development

Following to the temporal order of leaf events, we started monitoring the leaf phenology and leaf development of potted seedlings in late February to late April 2019 to investigate the effect of eCO<sub>2</sub> on the leaf-out phenology. We only used seedlings growing under similar conditions (potted seedlings) to avoid the impact of heterogenous growing conditions on leaf development of natural seedlings. We used a five-point scale of development: (I) seedlings with or without buds, buds usually covered by bracts; (II) seedlings with more developed buds, leaf primordia begin

<sup>1</sup> <https://www.esrl.noaa.gov/gmd/ccgg/trends/global.html>

TABLE 1 Number of seedlings of each species under ambient (A) and elevated (E) CO<sub>2</sub> concentration used to measure each variable (Leaf-out phenology, LLPM, LRC, PM, D, RGR, Gas exchange and F<sub>v</sub>/F<sub>m</sub>) at each measurement campaign.

		Leaf-out phenology, LLPM, LRC, PM, D, RGR		Gas exchange									F <sub>v</sub> /F <sub>m</sub>				
		Potted	Natural	Potted			Natural			Potted			Natural				
		Whole experiment		May 24 <sup>th</sup>	July 10 <sup>th</sup>	July 24 <sup>th</sup>	May 24 <sup>th</sup>	July 10 <sup>th</sup>	July 24 <sup>th</sup>	May 24 <sup>th</sup>	July 10 <sup>th</sup>	July 24 <sup>th</sup>	August 29 <sup>th</sup>	May 24 <sup>th</sup>	July 10 <sup>th</sup>	July 24 <sup>th</sup>	August 29 <sup>th</sup>
Hawthorn	A	8	13	7	8	8	9	8	10	7	8	8	8	0	8	4	12
Hawthorn	E	11	11	10	10	11	8	8	9	10	11	11	10	2	9	8	10
Hazel	A	12	0	11	10	9	0	0	0	11	10	9	9	0	0	0	0
Hazel	E	11	5	7	8	9	1	3	2	7	8	8	8	2	2	2	1
Holly	A	12	7	11	12	10	7	5	5	12	12	12	12	4	6	6	7
Holly	E	12	11	10	12	12	9	11	11	12	11	12	12	2	9	10	11
Oak	A	12	14	2	3	0	0	6	5	1	4	1	0	1	3	6	5
Oak	E	12	10	6	7	2	0	4	1	5	8	4	2	1	1	2	1
Sycamore	A	12	10	8	10	6	9	7	8	10	10	10	10	0	5	7	8
Sycamore	E	12	11	8	11	10	4	4	5	12	12	12	11	3	4	6	6
N		114	92	80	91	77	47	56	57	87	94	87	82	15	47	51	61

LLPM, loss of leaf photosynthetic material; LRC, leaf regeneration capacity; PM, powdery mildew; D, diameter; RGR, relative growth rate; F<sub>v</sub>/F<sub>m</sub>, maximum photochemical efficiency of photosystem II; N, total number of seedlings used to measure each variable at each campaign in 2019; Hawthorn, *Crataegus monogyna*; Hazel, *Corylus avellana*; Holly, *Ilex aquifolium*; Oak, *Quercus robur*; Sycamore, *Acer pseudoplatanus*. The bold values represent statistically significant differences.



to open; (III) seedlings with growing leaves, but not completely expanded or developed; (IV) seedlings with opened leaves but not completely developed; and (V) seedlings with fully opened and developed leaves.

## 2.4. Loss of leaf photosynthetic material caused by insect herbivory and powdery mildew (PM)

In the UK, insect herbivory and PM damage are usually temporally and spatially separated (Field et al., 2019). The majority of insect herbivore damage in the UK occurs in late spring while PM occurs in summer. We visually monitored the loss of leaf photosynthetic material (LLPM) caused by defoliating insects from April to May (Spring 2019) and/ or severe PM infection causing necrosis of foliar tissues from June to August (Summer 2019). LLPM was recorded by trained observers as a continuous variable percentage from 0 (no loss of leaf photosynthetic material) to 100% (complete loss of leaf photosynthetic material).

Secondary bud flushes and subsequent leaf growth caused by intense insect defoliation resulted in a negative change of LLPM in early summer with respect to the previous scoring campaign. This leaf regeneration capacity, LRC (eq. 1), was calculated as the maximum of minus one times the change in LLPM, or zero:

$$LRC_t = \max(LLPM_{t-1} - LLPM_t; 0) \quad (1)$$

To test for PM impact, we used a four-point ordinal scale of infection severity based on the % of total leaf surface showing symptoms of fungal mycelium: (“0”) infection ~ 0%; (“1”) infection <50%; (“2”) 51%–75% infected; (“3”) 76%–100% infected. This scale has previously been used to score PM by other authors (Bert et al., 2016; Field et al., 2019).

## 2.5. Leaf photosynthetic parameters

Leaf gas exchange measurements were taken on both natural and potted seedlings on 3 campaigns, representative of different phases of the 2019 spring/summer season: 24<sup>th</sup> May, intense defoliation; 10<sup>th</sup> July, early PM symptoms; and 24<sup>th</sup> July, late PM symptoms. Seedlings with leaves not fully developed, that were highly defoliated, or severely attacked by PM, were not measured. A portable open gas exchange system (LI-6400XT, LICOR, Lincoln, NE, USA) was used to measure net photosynthetic rates ( $A_n$ ) at a saturating light intensity of 700  $\mu\text{mol photons m}^{-2} \text{s}^{-1}$  (as determined by light response curves of net photosynthetic rates) and a flow rate of 300  $\mu\text{mol s}^{-1}$ . Reference  $\text{CO}_2$  was set at either 550 ppm for  $e\text{CO}_2$  arrays or 400 ppm for ambient  $\text{CO}_2$  arrays. Data were logged once stabilised. Maximum photochemical efficiency of photosystem II ( $F_v/F_m$ ) was measured using a portable Pocket PEA chlorophyll fluorimeter (Hansatech Instruments Ltd) in leaves after 30-min dark-adaptation. We conducted a 4<sup>th</sup>  $F_v/F_m$  campaign on the 29<sup>th</sup> of August because more seedlings were suitable for  $F_v/F_m$  measurements at that time, whereas they were not suitable to measure  $A_n$ .

## 2.6. Seedling growth

Two perpendicular diameters (D) and shoot length (H) were measured in three campaigns in 2019 on seedlings previously marked near the root collar. Campaign 1: 5th of May (natural seedlings), 14th of May (potted seedlings); campaign 2: 10th of July (natural and potted seedlings); and campaign 3: 29th of August (natural and potted seedlings).

Relative growth rates ( $\text{RGR}_d$ , %  $\text{day}^{-1}$ ) were calculated according to the following equation:

$$\text{RGR}_d = 100 \frac{1}{\bar{D}} \frac{d\bar{D}}{dt} \approx 100 \left( \frac{\log \bar{D}_2 - \log \bar{D}_1}{t_2 - t_1} \right) \quad (2)$$

where,  $\bar{D}_n$  is the average of the two perpendicular diameters (mm) at times  $t_n$  (days). The  $\text{RGR}_d$  values were calculated as a single finite difference over the whole experimental period ( $\text{RGR}_{98}$  over 98 days from May to August 2019).

## 2.7. Statistical analysis

All analyses were performed using PROC GLIMMIX, PROC MIXED SAS version 9.4 (SAS Institute Inc, 2016) and the glmmTMB (v1.1.2.3; Magnusson, 2021) package in R 4.1.2 (R Core Team, 2022). We used generalized linear mixed models (GLMMs) and linear mixed effects models (LMM) to investigate the drivers of leaf-out phenology and leaf development, photosynthetic parameters and seedling growth. Tukey’s *post-hoc* tests were used to assess the significance of differences between pairs of groups.

We used a GLMM to investigate the effect of  $\text{CO}_2$  concentration and time on the leaf-out phenology of the potted seedlings. We used a multinomial distribution with cumulative logit link function to model categorical variables.

We used GLMMs to explore differences between species and  $\text{CO}_2$  concentrations in loss of leaf photosynthetic material at two representative dates (24<sup>th</sup> of May: maximum defoliation; and 29<sup>th</sup> of August: maximum PM infection). To address overdispersion and/or excess of zeros we used negative binomial response distribution for potted seedlings (in SAS) and zero inflated negative binomial response distribution for natural seedlings (in R) in both cases with log link function. We log transformed seedling height to include it as a covariate.

The leaf regeneration capacity from secondary bud flushes was investigated from the 24<sup>th</sup> of May to the 20<sup>th</sup> of June in potted seedlings, when defoliating insects affected to a greater extent. We used GLMM to test for differences between species and treatments by computing all the LRC data together, irrespective of date of data capture. To address overdispersion and excess of zeros we used a zero inflated negative binomial response distribution with log link function.

To test for differences between species and  $\text{CO}_2$  concentration in the severity of PM infection throughout the experimental period we used generalized linear mixed models with a Gaussian distribution and identity link function.

We fitted LMM in SAS to explore variations in growth (D) over time, to analyse RGR for the whole experimental period (i.e.,  $\text{RGR}_{98}$ ) and to analyse leaf physiological variables  $A_n$  and  $F_v/F_m$ . Due to the impossibility of measuring immature leaves

or intensively defoliated leaves, sample sizes were small for some measurement campaigns, and for this reason  $A_n$  was not analysed in late August. To avoid computational problems,  $A_n$  and  $F_v/F_m$  were analysed separately at each measurement campaign. **Table 1** shows the number of seedlings used for the analysis of physiological variables at each campaign by treatment.

To simplify interactions between fixed effects, we separately explored differences between potted or natural seedlings and between species using two-way ANOVAs by means of the models specified above for each variable.

For each of the models described above, we specified an array-specific random effect if the corrected Akaike Information Criterion (AICC) was minimized and used Tukey's-Kramer's adjustment for multiple comparisons between unequal sample sizes.

Model residuals were visualised to check whether assumptions were satisfied. We used the package DHARMA when models were fit by R.

### 3. Results

Different models were used to explore the effect of  $eCO_2$  on multiple response variables. **Table 2** presents information about the type of model used to fit each response variable, the goodness of fit, and the statistical significance of the explanatory variables. **Table 2** is complemented by figures representing each studied variable at each corresponding subsection. The presentation of results is structured to align with the chronological order of leaf events observed during our study.

#### 3.1. Leaf-out phenology of introduced seedlings

The timing of leaf emergence was significantly different between species ( $P < 0.0001$ ; **Figure 1**). *Post hoc* pairwise comparison classified hawthorn in the group (A) as the earliest species to emerge, while oak leaves developed latest (C). Hazel and sycamore occupied an intermediate position (B). Due to the characteristics of this variable, we conducted subset analyses for each species (simple least squares means, **Supplementary Table 2**), we found a faster leaf development of oak under  $eCO_2$  compared to ambient  $CO_2$  concentration ( $P = 0.0209$ ), while the rest of the species did not show differences in leaf development between treatments.

#### 3.2. LLPM caused by insect herbivory

The two-way ANOVA showed that natural seedlings of oak, hazel and sycamore had lower total LLPM than potted seedlings ( $P < 0.0001$ ; **Figure 2**).

For natural seedlings, LLPM on the 24th of May 2019, corresponding with the insect herbivory peak before regrowth (**Figure 2**), was significantly different between species ( $P < 0.0085$ ), but not between  $CO_2$  treatments for any species. **Figure 2** also shows *post hoc* pairwise comparisons of means between species

represented by capital letters, where oak and sycamore are classified (in descending order) in a different group (A) to holly (B), while hawthorn and hazel occupied an overlapping position (AB).

For potted seedlings, LLPM again differed between species ( $P < 0.0001$ ), but not  $CO_2$  treatments. *Post hoc* pairwise comparisons between species, classified oak, hazel and sycamore in a different group (A), followed by hawthorn (B) and holly (C). LLPM increased 11-fold in potted oak compared to natural oak, reaching the maximum under ambient conditions (**Table 3**). On this campaign, mean LLPM in natural oak, sycamore and hazel seedlings ranged from 6.8% to 10% while in potted seedlings it ranged from 42.6% to 81.4%.

On the 29th of August, corresponding with the powdery mildew infection maximum, significant differences were found between species ( $P = 0.0003$ ), as well as a significant species  $\times$  treatment interaction in natural seedlings ( $P = 0.0136$ ). *Post hoc* pairwise comparisons between species, classified natural oak and sycamore in a different group (A) to holly (B), while hazel and hawthorn remained in an overlapping position (AB). The interaction effect was due to a slightly higher LLPM in hawthorn under ambient  $CO_2$  than under  $eCO_2$  but it was not significant in the *post hoc* pairwise comparisons (**Table 3**). Although pairwise comparisons did not show differences between treatments, natural seedlings, including oak, exhibited the largest LLPM in both May and August under  $eCO_2$ .

For potted seedlings, LLPM again differed between species ( $P < 0.0001$ ), but not  $CO_2$  treatments. On the 29th of August, potted seedlings demonstrated a similar grouping pattern in *post hoc* pairwise comparisons among species to that found on the 24th of May. We found a significant effect of the covariate plant height ( $P = 0.002$ ), the more vigorous seedlings showed higher impact of PM (**Supplementary Figure 2**).

#### 3.3. Leaf regeneration capacity against herbivores

We assessed the  $LRC_{27}$  in potted seedlings for two consecutive scoring periods (between 24th May and 20th June, a total of 27 days). We found differences between species ( $P < 0.0001$ ) but no differences between  $CO_2$  treatments (**Figure 3**). *Post hoc* comparison between species classified holly in a separate group (B) from all other species (A), while *post hoc* comparisons between species and treatments classified oak under both ambient and  $eCO_2$  as the most damage tolerant species in a separated group (**Table 3** and **Figure 3**).

#### 3.4. LLPM caused by PM

We assessed the severity of PM infection on natural and potted seedlings. Despite the stems of 4 potted hazels and 3 potted oaks being green, we could not assess the impact of PM due to complete defoliation through herbivory damage. Thus, fully defoliated seedlings and holly (this species did not show infection) were not included in the PM analyses. For naturally regenerated hazel seedlings, the lack of PM infection, and the small sample size under ambient conditions, caused convergence problems and so was also not included into the model fitting.

TABLE 2 Summary of results of the models.

Response variable	Date	Material	Explanatory variable	DF	F or Z-value	Pr > F or Z	Response distribution	Link function	Random effect	Pearson $\chi^2$ / DF (overdispersion)	AICC
Leaf-out phenology	Repeated	Potted	Species	3	53.65	<0.0001	Multinomial	cumulative	Yes	–	–
	Measures		Treatment	1	1.48	0.2241*		logit			
			Date	4	57.22	<0.0001					
			Spp × Treat	3	2.73	0.0437					
			Array		0.61	0.2712					
LLPM (%)	24-May	Natural	Species	4	37.5	0.0085	Zero inflated	log	No	0.18	339.7
			Treatment	1	0.02	0.4757	Negative				
			Spp × Treat	3	10.6	0.6124	Binomial				
LLPM (%)	24-May	Potted	Species	4	48.42	<0.0001	Negative	log	No	0.99	897.5
			Treatment	1	0.48	0.491	Binomial				
			Spp × Treat	4	0.25	0.9077					
			logH	1	0.84	0.3619					
LLPM (%)	29-August	Natural	Species	4	37.6	0.0003	Zero inflated	log	No	0.33	486.1
			Treatment	1	0.01	0.8952	Negative				
			Spp × Treat	3	10.7	0.0136	Binomial				
LLPM (%)	29-August	Potted	Species	4	47.27	<0.0001	Negative	log	No	0.79	892.3
			Treatment	1	0.03	0.874	Binomial				
			Spp × Treat	4	1.51	0.2051					
			logH (+)	1	9.93	0.0022					
Regeneration capacity	24-May	Potted	Species	4	33.6	<0.0001	Zero inflated	log	No	0.08	805.8
LRC	To		Treatment	1	0.76	0.3819	Negative				
	20-June		Spp × Treat	4	2.75	0.6005	Binomial				

(Continued)

TABLE 2 (Continued)

Response variable	Date	Material	Explanatory variable	DF	F or Z-value	Pr > F or Z	Response distribution	Link function	Random effect	Pearson $\chi^2$ / DF (overdispersion)	AICC
Powdery mildew	Repeated	Natural	Species	2	31.82	<0.0001	Gaussian	Identity	Yes	0.39	558.1
	Measures		Treatment	1	1.11	0.2939					
			Date	3	11.85	<0.0001					
			Spp $\times$ Treat	2	0.75	0.4753					
			Array		0.35	0.3631					
Powdery mildew	Repeated	Potted	Species	3	82.48	<0.0001	Gaussian	Identity	Yes	0.59	788
	Measures		Treatment	1	0.27	0.6047					
			Date	3	43.88	<0.0001					
			Spp $\times$ Treat	3	1.62	0.1858					
			Array		0.35	0.3631					
Diameter (mm)	Repeated	Natural	Species	4	31.88	<0.0001	-	-	No		499.2
	Measures		Treatment	1	0.69	0.4439					
			Date	2	16.07	0.0004					
			Spp $\times$ Treat	3	4.35	0.0427					
Diameter (mm)	Repeated	Potted	Species	4	59.25	<0.0001	-	-	No		840.5
	Measures		Treatment	1	29.07	0.0057					
			Date	2	13.94	0.0013					
			Spp $\times$ Treat	4	4.98	0.0084					
RGRd	24-May	Natural	Species	3	2.17	0.1041	-	-	No		<1019.6
	To		Treatment	1	1.51	0.2246					
	29-August		Spp $\times$ Treat	3	0.66	0.5832					
RGRd	24-May	Potted	Species	4	1.4	0.2397	-	-	No		<956.9
	To		Treatment	1	5.4	0.0223					
	29-August		Spp $\times$ Treat	4	1.21	0.3116					

(Continued)

TABLE 2 (Continued)

Response variable	Date	Material	Explanatory variable	DF	F or Z-value	Pr > F or Z	Response distribution	Link function	Random effect	Pearson $\chi^2$ / DF (overdispersion)	AICC
$A_n$	24-May	Natural	Species	3	1.55	0.2168	-	-	No		189.9
			Treatment	1	1.17	0.2853					
			Spp $\times$ Treat	2	7.83	<b>0.0014</b>					
$A_n$	24-May	Potted	Species	4	3.41	<b>0.0135</b>	-	-	Yes		315.2
			Treatment	1	0.41	0.5253					
			Spp $\times$ Treat	4	2.09	0.0917					
$A_n$	10-July	Natural	Species	4	4.32	<b>0.0047</b>	-	-	No		216.2
			Treatment	1	0.95	0.3344					
			Spp $\times$ Treat	3	0.4	0.752					
$A_n$	10-July	Potted	Species	4	2.68	<b>0.0372</b>	-	-	No		347.1
			Treatment	1	17.24	<b>&lt;0.0001</b>					
			Spp $\times$ Treat	4	1.14	0.3418					
$A_n$	24-July	Natural	Species	4	0.78	0.546	-	-	No		155.1
			Treatment	1	0.23	0.6311					
			Spp $\times$ Treat	3	0.92	0.4384					
$A_n$	24-July	Potted	Species	4	1.19	0.3231	-	-	Yes		213.5
			Treatment	1	0.76	0.3853					
			Spp $\times$ Treat	3	0.57	0.6339					
$F_v/F_m$	24-May	Natural	Species	3	0.71	0.5641	-	-			<46.4
			Treatment	1	0.14	0.7112					
			Spp $\times$ Treat	2	0.55	0.591					
$F_v/F_m$	24-May	Potted	Species	4	2.94	<b>0.0253</b>	-	-			<379.5
			Treatment	1	8.51	<b>0.0045</b>					
			Spp $\times$ Treat	4	0.92	0.4555					
$F_v/F_m$	10-July	Natural	Species	4	2.9	<b>0.0313</b>	-	-			<243.6
			Treatment	1	0.02	0.8841					
			Spp $\times$ Treat	3	0.09	0.9641					

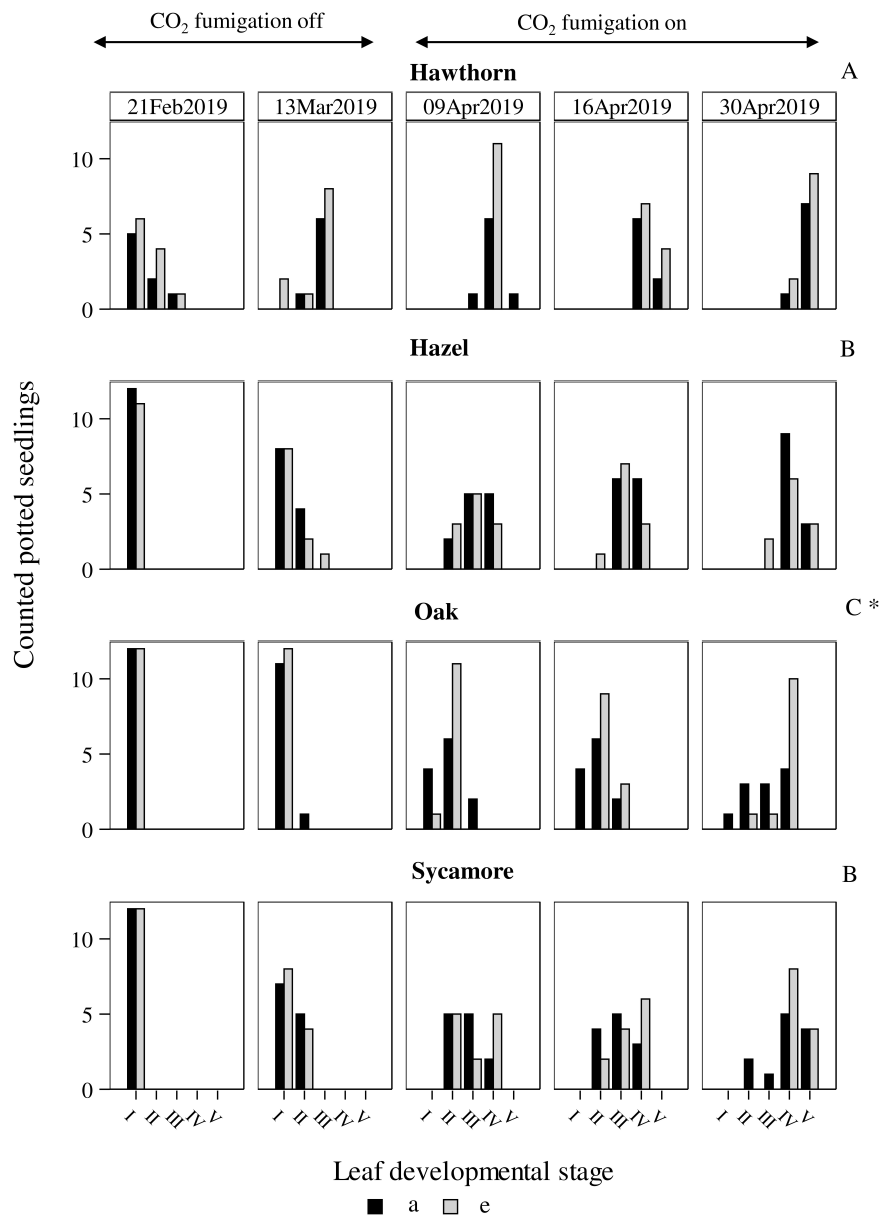
(Continued)



TABLE 2 (Continued)

Response variable	Date	Material	Explanatory variable	DF	F or Z-value	Pr > F or Z	Response distribution	Link function	Random effect	Pearson $\chi^2$ / DF (overdispersion)	AICC
$F_v/F_m$	10-July	Potted	Species	4	2.94	<b>0.0253</b>	-	-			<379.5
			Treatment	1	8.51	<u>0.0045</u>					
			Spp $\times$ Treat	4	0.92	0.4555					
$F_v/F_m$	24-July	Natural	Species	4	6.37	<b>0.0003</b>	-	-			<235.9
			Treatment	1	2.13	0.1511					
			Spp $\times$ Treat	3	0.79	0.507					
$F_v/F_m$	24-July	Potted	Species	4	5.01	<b>0.0012</b>	-	-			<373.6
			Treatment	1	0.21	0.6457					
			Spp $\times$ Treat	4	0.76	0.5529					
$F_v/F_m$	29-August	Natural	Species	4	10.31	<b>&lt;0.0001</b>	-	-			<291.6
			Treatment	1	2.32	0.1334					
			Spp $\times$ Treat	3	0.74	0.5326					
$F_v/F_m$	29-August	Potted	Species	4	11.66	<b>&lt;0.0001</b>	-	-			<415.8
			Treatment	1	0.74	0.3914					
			Spp $\times$ Treat	3	0.45	0.7193					

The bold values represent statistically significant differences. Underlined Pr > F represents significant differences between treatments. In leaf-out phenology the symbol \* represents significant differences between treatments in the simple least square analyse performed for each species.



**FIGURE 1**  
 Leaf development of potted seedlings under ambient (a) and elevated (e) CO<sub>2</sub> concentrations of the 4 deciduous species. Horizontal axes represent leaf developmental stages: from I -seedlings without developed buds- to V -seedlings with fully developed leaves-. Vertical axes represent number of seedlings counted in each developmental stage at each date. Capital letters represent differences between species. Symbol \* represents differences between treatments. Hawthorn: *Crataegus monogyna*; Hazel: *Corylus avellana*; Holly: *Ilex aquifolium*; Oak: *Quercus robur*; Sycamore: *Acer pseudoplatanus*.

Natural seedlings exhibited less severe PM infection than potted seedlings ( $P < 0.0001$ ). Natural hawthorn and hazel did not exhibit any visible symptom of PM at any scoring date, while potted seedlings did (Figure 4).

We found differences between species in PM levels on natural seedlings ( $P < 0.0001$ ), but no differences between CO<sub>2</sub> treatments (Figure 4). *Post hoc* comparisons between species for natural seedlings classified oak and sycamore as the most infected group (A) compared to hawthorn (B).

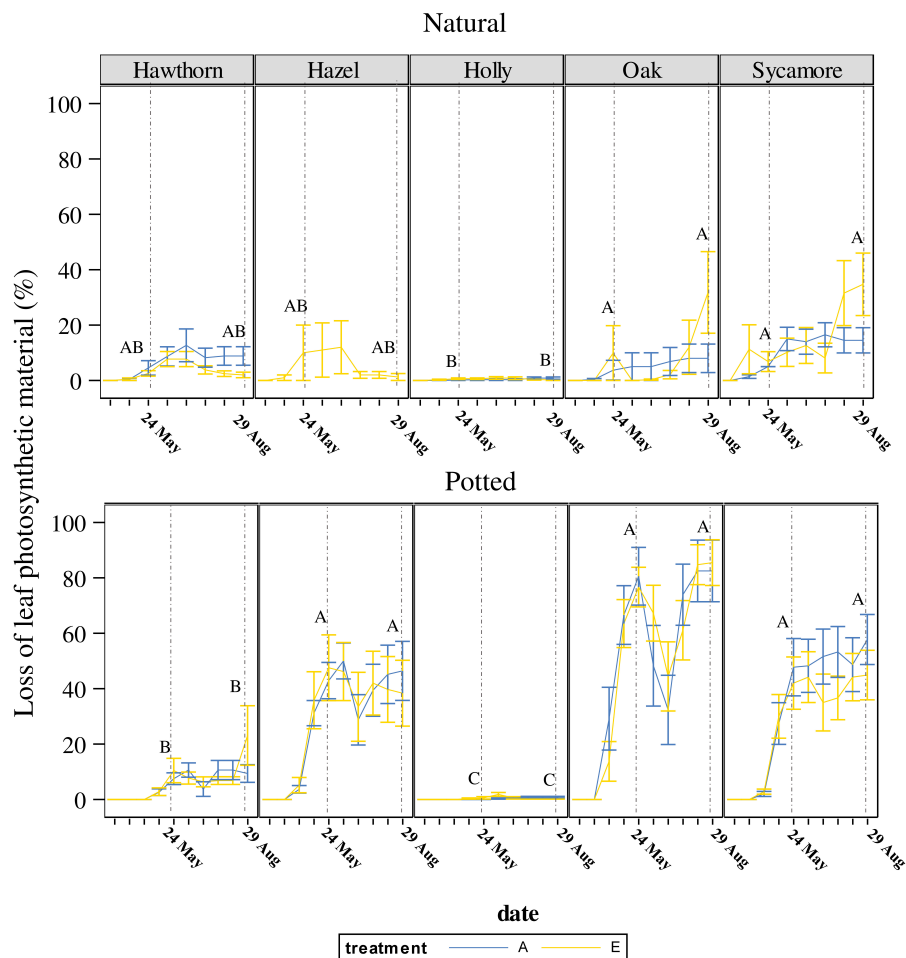
For the potted seedlings, there were no significant differences in PM between CO<sub>2</sub> treatments. *Post hoc* pairwise comparisons between species classified each species in a different group (in

descending order of infection) oak (A), sycamore (B), hawthorn (C), and hazel (D), while significant pair-wise overlaps occurred in the *post hoc* comparison between species and treatment (Table 3).

### 3.5. Photosynthetic performance

We did not find statistical differences between potted and naturally regenerated seedlings in  $A_n$  or  $F_v/F_m$ .

On the first campaign (24<sup>th</sup> May),  $A_n$  did not show differences between species or CO<sub>2</sub> treatments for either natural or potted seedlings. For this first campaign, the interaction



**FIGURE 2** Temporal variation in loss of leaf photosynthetic material (LLPM) throughout the experimental period under ambient (A) and elevated (E) CO<sub>2</sub> concentrations. Vertical solid lines represent standard error of the mean. Vertical dash lines represent key dates at which statistical analyses were conducted (May 24<sup>th</sup>: maximum defoliation; and August 29<sup>th</sup>: maximum powdery mildew infection). Capital letters represent the *post hoc* grouping by species in each analysed date. Different letters indicate differences between species for a same scoring campaign. Hawthorn: *Crataegus monogyna*; Hazel: *Corylus avellana*; Holly: *Ilex aquifolium*; Oak: *Quercus robur*; Sycamore: *Acer pseudoplatanus*.

species × treatment in natural seedlings was significant because, contrary to the other species, sycamore showed higher  $A_n$  under ambient CO<sub>2</sub> (Figure 5). A more sensitive classification by species and treatments (*Post hoc* pairwise comparison) is shown in Supplementary Table 3. On the second campaign (10th July), natural hawthorn and sycamore (A) showed statistically higher  $A_n$  than oak (B), while holly remained in an overlapping position (AB) (Figure 5). In the second campaign, potted hazel (A) showed significantly higher  $A_n$  than potted oak (B), the other species remained in an overlapping position (AB). There was a significant positive CO<sub>2</sub> effect on  $A_n$  in potted seedlings on the second campaign, corresponding to the most favourable measurement conditions. On the third campaign (29th August) there were no differences in  $A_n$  between species or treatments for both natural and potted seedlings. All seedlings showed lower  $A_n$  compared to the other measurement campaigns. Leaf temperature was higher on this campaign (c. 0.28°C) compared to the other campaigns (c. 21°C). We found a negative correlation between  $A_n$  and T under both eCO<sub>2</sub> and ambient conditions for the small range of leaf temperatures studied (data not shown).

Regarding photochemical efficiency as a measure of stress in plants, Figure 5 shows variations in  $F_v/F_m$  throughout the experimental period. On the first campaign, corresponding with the maximum defoliation intensity,  $F_v/F_m$  was below 0.80 in potted hazel, sycamore and oak as well as in natural hawthorn. Such values indicate that leaves were not fully developed and/or stressed (Maxwell and Johnson, 2000; Murchie and Lawson, 2013). *Post hoc* pairwise comparisons by species are represented in Figure 5. Overall, holly and hawthorn showed higher  $F_v/F_m$  than sycamore, oak and hazel in both potted and natural seedlings. A more detailed analysis by species and treatments is also shown in Supplementary Table 3.

### 3.6. Seedling growth performance: diameter and RGR<sub>d</sub>

The repeated-measures analyses showed differences in the natural seedling diameters between campaigns ( $P = 0.0004$ ), although there were no statistical differences between July and

TABLE 3 Mean and standard error of variables calculated by generalized linear mixed models.

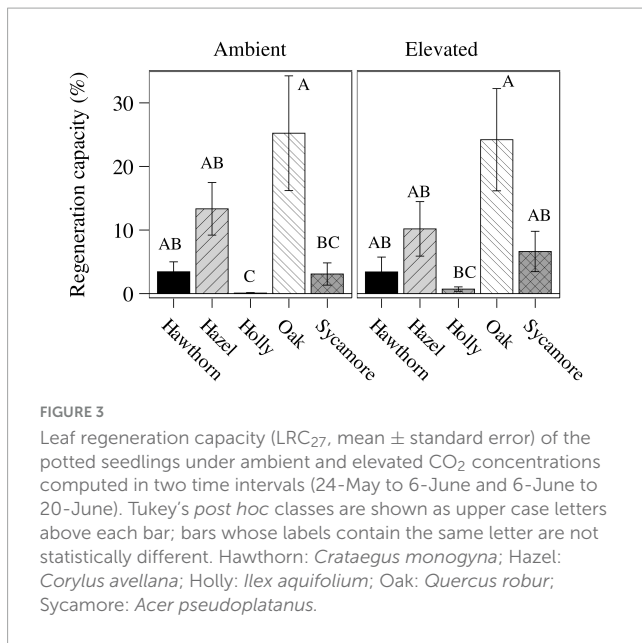
<i>Post hoc</i>	Species	Treat	t-ratio	Pr >  t	Mean	Stde	<i>Post hoc</i>	Species	Treat	t-value	Pr >  t	Mean	Stde
LLPM natural seedlings 24-May							LLPM potted seedlings 24-May						
A	Hazel	E	-1.131	0.2615	10.0	11.1	A	Oak	A	18.19	<0.0001	81.4	19.7
A	Oak	E	-1.612	0.1109	9.9	7.8	A	Oak	E	17.94	<0.0001	76.3	18.4
A	Sycamore	E	-2.181	0.032	6.8	5.1	A	Hazel	E	15.16	<0.0001	46.6	11.8
A	Hawthorn	A	-2.944	0.0042	4.5	3.0	A	Sycamore	A	15.83	<0.0001	47.5	11.6
A	Oak	A	-3.342	0.0013	3.7	2.5	A	Hazel	A	15.46	<0.0001	43.2	10.5
A	Sycamore	A	-2.464	0.0158	5.0	3.9	A	Sycamore	E	15.34	<0.0001	42.6	10.4
A	Hawthorn	E	-3.428	0.001	2.5	1.9	B	Hawthorn	E	8.24	<0.0001	9.8	2.7
A	Holly	E	-4.818	<0.0001	0.6	0.5	B	Hawthorn	A	6.41	<0.0001	7.6	2.4
A	Holly	A	-4.031	0.0001	0.1	0.2	C	Holly	E	-1.3	0.1983	0.6	0.3
							C	Holly	A	-1.98	0.0504	0.3	0.2
<i>Post hoc</i>	Species	Treat	t-ratio	Pr >  t	Mean	Stde	<i>Post hoc</i>	Species	Treat	t-value	Pr >  t	Mean	Stde
LLPM natural seedlings 29-August							LLPM potted seedlings 29-August						
A	Sycamore	E	-2.781	0.0067	34.7	18.3	A	Oak	E	17.28	<0.0001	80.0	20.3
A	Oak	E	-0.172	0.8636	31.8	17.6	AB	Oak	A	15.22	<0.0001	68.7	19.1
AB	Sycamore	A	-1.574	0.1193	14.5	8.1	AB	Sycamore	A	16.13	<0.0001	60.6	15.4
AB	Hawthorn	A	-2.781	0.0067	8.8	4.4	AB	Sycamore	E	14.9	<0.0001	56.9	15.4
AB	Oak	A	-3.092	0.0027	8.0	3.8	AB	Hazel	A	14.35	<0.0001	51.2	14.0
BC	Hawthorn	E	-5.027	<0.0001	2.0	1.1	AB	Hazel	E	13.59	<0.0001	44.9	12.6
BC	Hazel	E	-3.943	0.0002	1.3	1.2	BC	Hawthorn	E	9.88	<0.0001	21.1	6.5
BC	Holly	A	-4.771	<0.0001	0.9	0.7	C	Hawthorn	A	6.66	<0.0001	8.9	2.9
C	Holly	E	-5.961	<0.0001	0.2	0.2	D	Holly	A	-1.23	0.2202	0.6	0.3
							D	Holly	E	-2.6	0.0109	0.2	0.1

(Continued)

TABLE 3 Continued)

							<i>Post hoc</i>	Species	Treat	t-ratio	Pr >  t	Mean	Stde
Leaf regeneration capacity potted seedlings													
							A	Oak	A	-0.437	0.6628	25.2271	18.9166
							A	Oak	E	-0.513	0.6082	24.2084	17.3811
							AB	Hazel	A	-1.342	0.1809	13.3333	9.5861
							AB	Hazel	E	-1.643	0.1019	10.1819	7.6530
							AB	Sycamore	E	-2.308	0.0219	6.6250	4.7776
							AB	Hawthorn	A	-2.613	0.0096	3.4375	3.0531
							AB	Hawthorn	E	-3.074	0.0024	3.4091	2.5824
							BC	Sycamore	A	-3.346	0.0010	3.0833	2.2742
							BC	Holly	E	-5.154	<0.0001	0.7083	0.5360
							C	Holly	A	-5.999	<0.0001	0.0833	0.0839
<i>Post hoc</i>	Species	Treat	t-ratio	Pr -  t	Mean	Stde	<i>Post hoc</i>	Species	Treat	t-value	Pr -  t	Mean	Stde
Powdery mildew natural seedlings							Powdery mildew potted seedlings						
A	Oak	E	6.7	<0.0001	0.74	0.11	A	Oak	A	16.44	<0.0001	1.98	0.12
A	Sycamore	E	6.05	<0.0001	0.66	0.11	A	Oak	E	15.98	<0.0001	1.89	0.12
A	Oak	A	6.75	<0.0001	0.66	0.10	AB	Sycamore	E	13.44	<0.0001	1.52	0.11
AB	Sycamore	A	3.55	0.0005	0.42	0.12	BC	Sycamore	A	10.66	<0.0001	1.26	0.12
B	Hawthorn	A	<0.13	0.8963	<0.01	0.10	CD	Hawthorn	A	5.21	<0.0001	0.72	0.14
B	Hawthorn	E	<0.12	0.9044	<0.01	0.11	DE	Hawthorn	E	4.62	<0.0001	0.54	0.12
							DE	Hazel	A	2.33	0.0203	0.28	0.12
							E	Hazel	E	0.79	0.433	0.10	0.13





August (Figure 6, campaigns 2 and 3). *Post hoc* pairwise comparison (Supplementary Table 3) classified natural holly and sycamore into the largest diametric group A, while natural oak and hawthorn were classified into the group B. Hazel remained in an overlapping position (AB).

For potted seedling diameters, we found that species, treatment, date, and the interaction between species and treatment was highly significant ( $P < 0.008$ ; Figure 6). *Post hoc* pairwise comparisons (Supplementary Table 3) broadly followed those for the natural seedlings. Sycamore was classified as the most fast-growing species with the largest diameter (group A), followed by hazel (B), holly remained in an intermediate position between sycamore and hazel (AB), oak was classified into a different group (C), and finally hawthorn was the most slow-growing species, classified into the smallest diameter group (D).

The RGR<sub>98</sub> was analysed from May to August (Figure 7). We did not identify any significant CO<sub>2</sub> effect for the natural seedlings, but eCO<sub>2</sub> significantly increased growth rate for the potted seedlings ( $P = 0.0223$ ), driven to a large extent by the strong treatment effect on potted hawthorn and oak (Figure 7).

## 4. Discussion

### 4.1. Relationship between leaf-out phenology and herbivores

In contrast to Norby et al. (2003), who did not find a significant CO<sub>2</sub> effect on the leaf-out phenology of 14-year-old saplings in a deciduous FACE forest, we did find a stimulation by CO<sub>2</sub> of the leaf development of potted oak seedlings. Oak seedlings growing under eCO<sub>2</sub> reached more advanced developmental stages earlier than those growing under ambient conditions (Figure 1). Whilst we observed an eCO<sub>2</sub> treatment effect on plant phenology, the lack of an observed difference in herbivory suggests that eCO<sub>2</sub> does not affect palatability at this growth stage. Despite not analysing leaf

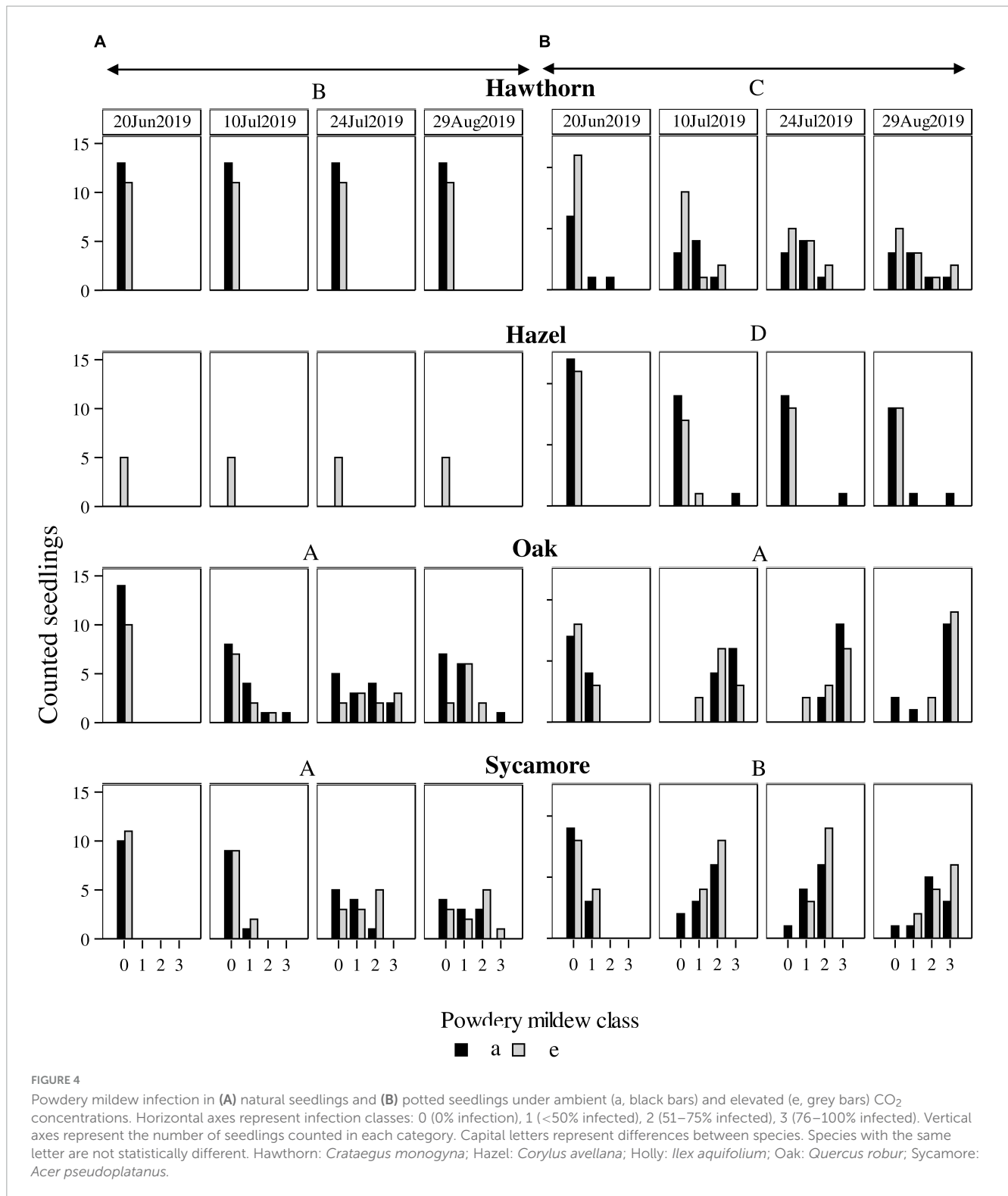
stoichiometry in this experiment due to lack of leaf material of some seedlings, the C:N ratios of adult oaks do not show an eCO<sub>2</sub> effect in the BIFoR-FACE forest (Gardner et al., 2022), supporting our results on invariability of the palatability under eCO<sub>2</sub>.

Of the four deciduous species, hawthorn exhibited the highest resistance to herbivory, showing low defoliation intensity in both natural (3.5%) and potted seedlings (9%) (Figure 2). One explanation is that the early leaf-out phenology of hawthorn was advantageous. Hawthorn leaf-out began significantly earlier, in February, compared to the other species, and leaves were fully expanded by April 2019 (Figure 1, top row), before the insect defoliation peak (Figure 2, left-most column). Oak flushing began significantly later, and even by late April oak leaves were not completely expanded in any of the experimental plots (Figure 1, second row and Figure 2, fourth column from left). As leaves age, they become physically tougher and contain greater concentrations of unpalatable defence compounds (Feeny, 1970; Polgar and Primack, 2011). Consequently, many defoliating herbivores feed primarily on young leaves, which supports the idea that leaf-out phenology influences the extent of herbivory experienced by these tree species. Another explanation is that even though the key defoliating herbivore species across the experimental site are generalists (a survey of insects took place between 2017 and 2019, see Crowley et al., 2023) they do not exert a consistent herbivory pressure on all the tree species (Hittenbeck et al., 2019). We are not able to discriminate between these possibilities.

### 4.2. Impact of insect defoliation on seedlings

We found no significant impact of eCO<sub>2</sub> on LLPM through insect herbivory for each tree species studied (Figure 2 and Table 2). These results agree with previous findings in Euc-FACE (Gherlenda et al., 2016b) and are consistent with other herbivory studies at BIFoR FACE that found no defining pattern of herbivory change under eCO<sub>2</sub> (data not shown). One possible explanation for this, is that all these studies were conducted during the first 3–4 years of fumigation of the BIFoR-FACE, and so impacts on herbivory may yet emerge. Our results show that oak potted seedlings experienced 11-fold higher LLPM due to insect herbivory than natural seedlings (Figure 2, May 24th data point). Indeed, within this experiment, the majority of the species growing in natural conditions were much better able to avoid or to defend themselves against herbivory. In this context, recent research has highlighted the role of interactions with root symbionts and other soil microbes in plants induced resistance against herbivore insects (e.g., Castaño et al., 2020). This complex soil ecosystem was lacking for potted seedlings and suggests delving deeper into these intricate relationships may offer valuable insights into the mechanisms driving the divergent herbivory responses. In addition, our results highlight clear differences in herbivory damage across tree species under all conditions tested, and identifying what factors contribute to these differences is key to predicting future winners and losers given that pest damage is expected to increase with global climate change (Guyot et al., 2016).

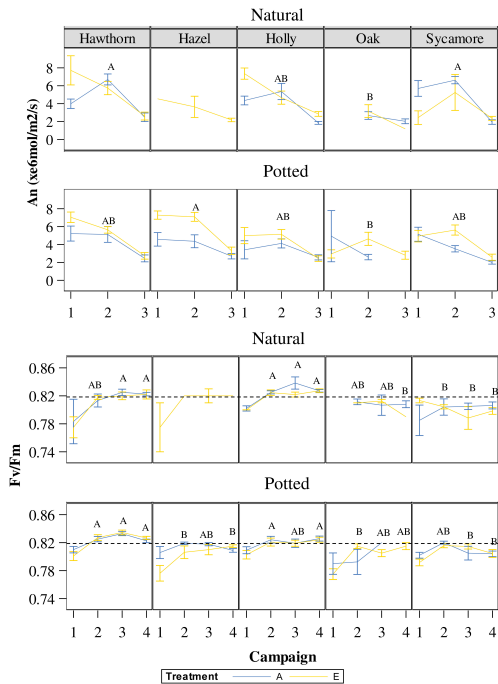
In this study, the only evergreen species, holly, did not show symptoms of herbivory or pathogen attack. Low palatability of



holly to feeding adult insects (Southwood, 1986) is consistent with the finding of the present study and we did not find literature supporting PM infection on holly. Leaf-mining larvae, known to be present in the BIFoR FACE site, are the main cause of LLPM in holly, however, damage by this feeding guild has been recorded mainly in more mature plants than those seedlings growing at the site, which may explain why no damage was recorded in this study.

### 4.3. Regeneration following insect defoliation

Species-specific responses such as compensatory growth to insect herbivory have been found within the BIFoR-FACE forest. Compensatory growth has often been reported as a tolerance strategy after defoliation (Massey et al., 2005; Fuenzalida et al., 2019;

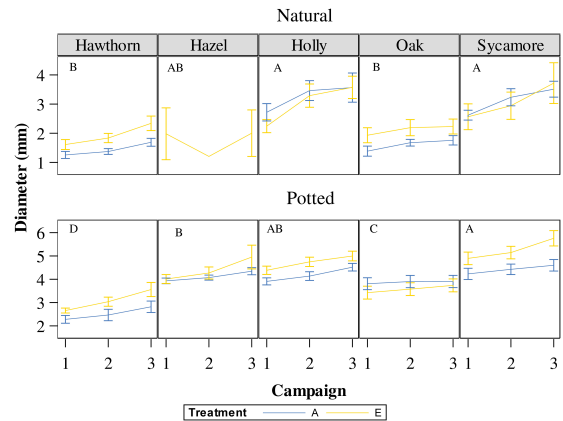


**FIGURE 5**  
Variations in  $A_n$  and  $F_v/F_m$  throughout the experimental period (campaign 1: 24-May; campaign 2: 10-July; campaign 3: 24-July; campaign 4: 29-August) under ambient CO<sub>2</sub> concentration (A) and elevated CO<sub>2</sub> concentration (E). Capital letters inside of the figure represent differences between species in the same campaign. Species whose labels contain the same letter in the same campaign are not statistically different. When the sample size of natural hazel or oak were too small they were not included in the multiple comparison tests. Hawthorn: *Crataegus monogyna*; Hazel: *Corylus avellana*; Holly: *Ilex aquifolium*; Oak: *Quercus robur*; Sycamore: *Acer pseudoplatanus*.

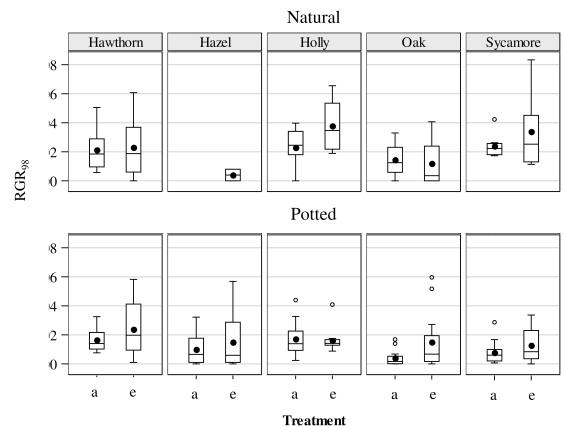
Wang et al., 2020b). Defoliation intensity and recovery time play a crucial role in compensatory growth capacity. Oak exhibited the most favourable response to defoliating insects, demonstrating an increase in leaf production through “lammas” shoots and a robust recovery of photosynthetic material (Figures 2, 3). We find no eCO<sub>2</sub> effect on compensatory growth (leaf production) for any species, although longer term assessments are required. Successive years of intense defoliation may strongly reduce compensation because of induced morphological changes such as vessel size or defects in secondary xylem (Anttonen et al., 2002). Previous studies have found lower survival rate and performance of younger seedlings subjected to experimental defoliation higher than 50% (Barton, 2016; Lorca et al., 2019).

#### 4.4. Impact of powdery mildew on seedlings

Within the BIFoR-FACE site, oak showed the most severe PM infection, followed by sycamore. This result is in agreement with the classical theory of the host density effect where higher abundance of a particular tree species is related to higher infection levels (Burdon and Chilvers, 1982). Our results are also consistent with Hantsch et al. (2013), who found a strong species identity effect



**FIGURE 6**  
Stem diameter of natural and potted seedlings under ambient CO<sub>2</sub> concentration (A) and elevated CO<sub>2</sub> concentration (E). Campaign 1: 5<sup>th</sup> of May in natural seedlings and 24<sup>th</sup> of May in potted seedlings; campaign 2: 10<sup>th</sup> of July; Campaign 3: on the 29<sup>th</sup> of August. Capital letters represent differences between species in the same campaign. Species whose labels contain the same letter in the same campaign are not statistically different. Hawthorn: *Crataegus monogyna*; Hazel: *Corylus avellana*; Holly: *Ilex aquifolium*; Oak: *Quercus robur*; Sycamore: *Acer pseudoplatanus*.



**FIGURE 7**  
Relative growth rates (RGR<sub>d</sub>, % day<sup>-1</sup>) for potted and natural seedlings, calculated as a single finite difference over the whole experimental period (97 days from May to August) under ambient (a) and elevated (e) CO<sub>2</sub> concentrations. Plotted are medians (horizontal line), means (black dots), interquartile range (box), upper and lower 5% Ilex (whiskers) and outliers (circles). Hawthorn: *Crataegus monogyna*; Hazel: *Corylus avellana*; Holly: *Ilex aquifolium*; Oak: *Quercus robur*; Sycamore: *Acer pseudoplatanus*.

in *Quercus* spp. in relation to specialized foliar biotrophic fungi in temperate forests, specifically that the presence of particular disease-prone species explained higher pathogen loads. Unlike other species, mature oak trees show PM infection every year, thus producing a dense pool of spores in the deciduous leaves at ground level to infect seedlings as they emerge the following spring.

The existing studies conducted to investigate the effect of elevated CO<sub>2</sub> on the severity of PM are contrasting. For instance, Sanchez-Lucas et al. (2023) found a significant CO<sub>2</sub> effect on the severity of PM infection in oak seedlings grown at very high CO<sub>2</sub>

concentration (1000 ppm) in controlled environment conditions. Whereas [Watanabe et al. \(2014\)](#), who studied photosynthetic traits in a FACE coppice forest of Japanese oak infected by PM, did find lower infection level under eCO<sub>2</sub>. We did not find a CO<sub>2</sub> effect on the severity of PM, however, we found that the more vigorous sycamore and oak seedlings suffered more severe disease (see slopes of these 2 species; [Supplementary Figure 2](#)), suggesting that the species in BIFoR FACE traded-off photosynthate between growth and defence differently.

#### 4.5. Effect of herbivory and PM infection on photosynthetic parameters and growth

The lowest net photosynthetic rates ( $A_n$ ) measured during our study were found in late July ([Figure 5](#)), at the same time as the maximum mildew infection. This pattern was observed in all species; therefore, we cannot differentiate whether the decreased  $A_n$  was due to PM or some environmental factor. The combination of decreased photosynthetic activity and loss of photosynthetic material may have critical consequences, especially for oak regeneration. Oak trees produce a part of their annual ring before budburst; therefore, a limited photosynthetic ability can compromise the reserve of carbohydrates required in the next spring ([Marçais and Desprez-Loustau, 2014](#)).

Photochemical efficiency ( $F_v/F_m$ ) is an indicator of the state of the photosynthetic machinery, and a reliable diagnostic indicator of responses to stress (e.g., [Maxwell and Johnson, 2000](#); [Molina-Montenegro et al., 2013](#)). Some authors have found a correlation between compensatory photosynthesis in response to leaf damage, for instance [Madriaza et al. \(2019\)](#) found an interesting compensation of  $F_v/F_m$  and relative growth rates after defoliation and propose the use of  $F_v/F_m$  as an indicator of tolerance to herbivory. In contrast, [Amagai and Kudo \(2019\)](#) found compensation of  $F_v/F_m$  after defoliation, but this effect did not influence shoot growth. Our results of  $F_v/F_m$  highlighted damage caused by biotic stress in oak, sycamore and hazel, which exhibited low  $F_v/F_m$  values throughout the experiment ([Figure 5](#)), while the most resistant species, holly and hawthorn, showed higher values of  $F_v/F_m$ . Therefore, our experiment did not reveal any indications of compensatory changes in  $F_v/F_m$  values. Nevertheless, it is important to note that this effect might have been mitigated by the influence of powdery mildew infection, a factor previously associated with decreased  $F_v/F_m$  values ([Wang et al., 2014](#)).

In contrast to sycamore, which exhibited the largest diameter of the 5 species, oak and hawthorn showed the smallest diameters at the beginning of the experiment and were characterised as the slowest-growing species. However, hawthorn, one of the species less affected by PM, grew in diameter throughout the experiment, while oak practically did not grow during the PM infection (from July to August) under both natural and potted conditions, likely a direct consequence of nutrient uptake by the biotrophic fungus ([Desprez-Loustau et al., 2014](#); [Figure 5](#)). In summary, when considering all the species examined in our study, our results for potted seedlings are consistent with those reported by [Watanabe et al. \(2014\)](#), confirming an increase in  $A_n$  on the second measurement campaign (10th of July), enhance diameter growth and higher

RGR<sub>98</sub> under elevated eCO<sub>2</sub> conditions, regardless of the presence of PM infection ([Table 2](#)).

## 5. Conclusion

The impacts of the interaction of pathogens and pests on complex forest communities have been previously investigated, e.g., under drought-stress situations ([Jactel et al., 2012](#)) or different combinations of tree species and increasing diversity ([Field et al., 2019](#)). Very few studies have assessed how CO<sub>2</sub> enrichment might alter this interaction under natural conditions in a natural forest ([Gherlenda et al., 2016a,b](#)), and no studies to date have investigated how it might impact regeneration of a temperate deciduous mixed woodland. Our results showed that the relationship between biotic stressors and seedlings was mainly driven by tree species and whether they were naturally established or potted. We did not find a CO<sub>2</sub> treatment effect on the defoliation intensity nor PM severity even when there was a positive eCO<sub>2</sub> effect on foliar growth, as in potted oak seedlings.

The main cause of reduction of leaf photosynthetic material in natural seedlings was PM, representing a leaf loss c. 34% in natural oak and sycamore ([Figure 2](#)), and up to 80% in potted oak ([Figure 2](#)). Several factors we cannot distinguish between might explain the lower defoliation in naturally regenerated seedlings: epigenetic factors; or the greater distances between seedlings, i.e., “dilution” in the wider environment cf. the potted seedlings ([Stastny and Agrawal, 2014](#)).

The earliest leaf flushing species, hawthorn, and the evergreen holly showed fewer symptoms of PM infection and defoliation by insects, therefore their regeneration capacity remains mostly unknown because they were not defoliated. Hazel (potted and natural) showed both good regeneration capacity after intense defoliation close to that of oak, and more resistance to PM infection than oak and sycamore. In the context and with respect to the pressures we have studied, these traits place the regeneration of hazel in a better position compared to oak. Sycamore was more resistant to defoliation and PM than oak. The same high regeneration capacity that confers oak the highest tolerance against insects (“lammas” shoots), also makes this species most vulnerable to PM, since the pathogen benefits from younger tissues ([Field et al., 2019](#)). We find no evidence that the regeneration of oak, the dominating species, will be improved under eCO<sub>2</sub>.

## Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

## Author contributions

CM: Methodology, Software, Validation, Writing – original draft, Writing – review and editing, Conceptualization, Data curation, Formal analysis, Investigation. SI: Writing – review and editing. EL: Writing – review and editing. LC: Writing – review

and editing. SH: Writing – review and editing. JS: Writing – review and editing. AM: Investigation, Funding acquisition, Resources, Writing – review and editing.

## Funding

The author(s) declare financial support was received for the research, authorship, and/or publication of this article. Authors acknowledge support from the UK Natural Environment Research Council [CM and AM: NE/S015833/1 (QUINTUS); EL, SH, and AM: NE/V021346/1]. The BIFoR-FACE facility acknowledges generous underpinning support from the JABBS Trust, Norbury Park Estate, The John Horseman Trust, Ecological Continuity Trust, and the University of Birmingham.

## Acknowledgments

We gratefully acknowledge assistance from Prof. Michael Tausz, the BIFoR-FACE operations team, and Silvia Bugallo, a student on the Erasmus+ program.

## References

- Ainsworth, E. A., and Long, S. P. (2005). What have we learned from 15 years of free-air CO<sub>2</sub> enrichment (FACE)? A meta-analytic review of the responses of photosynthesis, canopy properties and plant production to rising CO<sub>2</sub>. *N. Phytol.* 165, 351–372. doi: 10.1111/j.1469-8137.2004.01224.x
- Amagai, Y., and Kudo, G. (2019). Responses of photochemical efficiency and shoot growth of alpine dwarf-pine *Pinus pumila* to experimental warming, shading, and defoliation in Japan. *Alpine Bot.* 129, 43–52.
- Anttonen, S., Piispanen, R., Ovaska, J., Mutikainen, P., Saranpää, P., and Vapaavuori, E. (2002). Effects of defoliation on growth, biomass allocation, and wood properties of *Betula pendula* clones grown at different nutrient levels. *Can. J. For. Res.* 32, 498–508.
- Aranda, I., Cadahía, E., and de Simón, B. F. (2020). Leaf ecophysiological and metabolic response in *Quercus pyrenaica* Willd seedlings to moderate drought under enriched CO<sub>2</sub> atmosphere. *J. Plant Physiol.* 244:153083. doi: 10.1016/j.jplph.2019.153083
- Arzanlou, M., Torbati, M., and Golmohammadi, H. (2018). Powdery mildew on hazelnut (*Corylus avellana*) caused by *Erysiphe corylacearum* in Iran. *For. Pathol.* 48, e12450. doi: 10.1094/PDIS-10-20-2281-PDN
- Barton, K. E. (2016). Low tolerance to simulated herbivory in Hawaiian seedlings despite induced changes in photosynthesis and biomass allocation [Article]. *Ann. Bot.* 117, 1053–1062. doi: 10.1093/aob/mcw021
- Bert, D., Lasnier, J.-B., Capdevielle, X., Dugravot, A., and Desprez-Loustau, M.-L. (2016). Powdery mildew decreases the radial growth of oak trees with cumulative and delayed effects over years. *PLoS One* 11:e0155344. doi: 10.1371/journal.pone.0155344
- Bradwell, J., Foyer, C., and MacKenzie, R. (2022). *Forest innovation to tackle the climate and biodiversity emergencies*. Birmingham: University of Birmingham.
- Burdon, J. J., and Chilvers, G. A. (1982). Host density as a factor in plant disease ecology. *Annu. Rev. Phytopathol.* 20, 143–166.
- Castaño, C., Camarero, J. J., Zas, R., Sampedro, L., Bonet, J. A., Alday, J. G., et al. (2020). Insect defoliation is linked to a decrease in soil ectomycorrhizal biomass and shifts in needle endophytic communities. *Tree Physiol.* 40, 1712–1725. doi: 10.1093/treephys/tpaa104
- Chang, H., Han, S. H., An, J., Park, M. J., and Son, Y. (2018). Warming results in advanced spring phenology, delayed leaf fall, and developed abnormal shoots in *Pinus densiflora* seedlings. *Trees-Struct. Funct.* 32, 1473–1479.
- Che-Castaldo, C., Crisafulli, C. M., Bishop, J. G., Zipkin, E. F., and Fagan, W. F. (2019). Disentangling herbivore impacts in primary succession by refocusing the plant stress and vigor hypotheses on phenology. *Ecol. Monogr.* 89, e01389.
- Ciais, P., Sabine, C., Bala, G., Bopp, L., Brovkin, V., Canadell, J., et al. (2014). “Carbon and other biogeochemical cycles,” in *Climate change 2013: the physical science basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, (Cambridge: Cambridge University Press), 465–570.
- Crowley, L. M., Ivison, K., Enston, A., Garrett, D., Sadler, J. P., Pritchard, J., et al. (2023). A comparison of sampling methods and temporal patterns of arthropod abundance and diversity in a mature, temperate, Oak woodland. *Acta Oecol.* 118:103873.
- Crowley, L. M., Sadler, J. P., Pritchard, J., and Hayward, S. A. (2021). Elevated CO<sub>2</sub> Impacts on Plant–Pollinator Interactions: A Systematic Review and Free Air Carbon Enrichment Field Study. *Insects* 12:512. doi: 10.3390/insects12060512
- Domestoy, T., Moreira, X. Q., Jactel, H., Valdes-Correcher, E., Plomion, C., and Castagneryol, B. (2021). Growth and mortality of the oak processionary moth, *Thaumetopoea processionea*, on two oak species: direct and trait-mediated effects of host and neighbour species. *Entomol. Gener.* 41, 13–25.
- Daud, R. D., de Cássia Conforto, E., and Feres, R. J. F. (2012). Changes in leaf physiology caused by *Calacarus heveae* (Acari, Eriophyidae) on rubber tree. *Exp. Appl. Acarol.* 57, 127–137. doi: 10.1007/s10493-012-9552-y
- Desprez-Loustau, M.-L., Saint-Jean, G., Barrès, B., Dantec, C. F., and Dutech, C. (2014). Oak powdery mildew changes growth patterns in its host tree: host tolerance response and potential manipulation of host physiology by the parasite. *Ann. For. Sci.* 71, 563–573.
- DOE (2020). *US Department of energy free-Air CO<sub>2</sub> enrichment experiments: FACE results, lessons, and legacy*. London: DOE.
- Du, C. J., Wang, X. D., Zhang, M. Y., Jing, J., and Gao, Y. H. (2019). Effects of elevated CO<sub>2</sub> on plant C-N-P stoichiometry in terrestrial ecosystems: A meta-analysis. *Sci. Total Environ.* 650, 697–708. doi: 10.1016/j.scitotenv.2018.09.051
- Duursma, R. A., Gimeno, T. E., Boer, M. M., Crous, K. Y., Tjoelker, M. G., and Ellsworth, D. S. (2016). Canopy leaf area of a mature evergreen *Eucalyptus* woodland does not respond to elevated atmospheric [CO<sub>2</sub>] but tracks water availability. *Glob. Change Biol.* 22, 1666–1676.
- Feeny, P. (1970). Seasonal changes in oak leaf tannins and nutrients as a cause of spring feeding by winter moth caterpillars. *Ecology* 51, 565–581.
- Field, E., Schönrogge, K., Barsoum, N., Hector, A., and Gibbs, M. (2019). Individual tree traits shape insect and disease damage on oak in a climate-matching tree diversity experiment. *Ecol. Evol.* 9, 8524–8540. doi: 10.1002/ecc3.5357

## Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

## Publisher’s note

All claims expressed in this article are solely those of the authors and do not necessarily represent those of their affiliated organizations, or those of the publisher, the editors and the reviewers. Any product that may be evaluated in this article, or claim that may be made by its manufacturer, is not guaranteed or endorsed by the publisher.

## Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/ffgc.2023.1278409/full#supplementary-material>



- Fuenzalida, T. I., Hernández-Moreno, Á., and Piper, F. I. (2019). Secondary leaves of an outbreak-adapted tree species are both more resource acquisitive and more herbivore resistant than primary leaves. *Tree Physiol.* 39, 1499–1511. doi: 10.1093/treephys/tpz083
- Gardner, A., Ellsworth, D. S., Crous, K. Y., Pritchard, J., and MacKenzie, A. R. (2022). Is photosynthetic enhancement sustained through three years of elevated CO<sub>2</sub> exposure in 175-year-old *Quercus robur*? *Tree Physiol.* 42, 130–144. doi: 10.1093/treephys/tpab090
- Gherlenda, A. N., Esveld, J. L., Hall, A. A., Duursma, R. A., and Riegler, M. (2016a). Boom and bust: rapid feedback responses between insect outbreak dynamics and canopy leaf area impacted by rainfall and CO<sub>2</sub>. *Glob. Change Biol.* 22, 3632–3641. doi: 10.1111/gcb.13334
- Gherlenda, A. N., Moore, B. D., Haigh, A. M., Johnson, S. N., and Riegler, M. (2016b). Insect herbivory in a mature Eucalyptus woodland canopy depends on leaf phenology but not CO<sub>2</sub> enrichment. *BMC Ecol.* 16:47. doi: 10.1186/s12898-016-0102-z
- Glawe, D. A. (2008). The powdery mildews: a review of the world's most familiar (yet poorly known) plant pathogens. *Annu. Rev. Phytopathol.* 46, 27–51. doi: 10.1146/annurev.phyto.46.081407.104740
- Guyot, V., Castagnyrol, B., Vialatte, A., Deconchat, M., and Jactel, H. (2016). Tree diversity reduces pest damage in mature forests across Europe [Article]. *Biol. Lett.* 12:5. doi: 10.1098/rsbl.2015.1037
- Hantsch, L., Braun, U., Scherer-Lorenzen, M., and Bruehlheide, H. (2013). Species richness and species identity effects on occurrence of foliar fungal pathogens in a tree diversity experiment [Article]. *Ecosphere* 4:12. doi: 10.1890/es13-00103.1
- Hart, K. M., Curioni, G., Blaen, P., Harper, N. J., Miles, P., Lewin, K. F., et al. (2020). Characteristics of free air carbon dioxide enrichment of a northern temperate mature forest. *Glob. Change Biol.* 26, 1023–1037. doi: 10.1111/gcb.14786
- Hasegawa, S., Piñeiro, J., Ochoa-Hueso, R., Haigh, A. M., Rymer, P. D., Barnett, K. L., et al. (2018). Elevated CO<sub>2</sub> concentrations reduce C4 cover and decrease diversity of understorey plant community in a Eucalyptus woodland. *J. Ecol.* 106, 1483–1494.
- Hittenbeck, A., Bialozyt, R., and Schmidt, M. (2019). Modelling the population fluctuation of winter moth and mottled umber moth in central and northern Germany. *Forest Ecosystems* 6, 1–18.
- SAS Institute Inc (2016). *SAS® 9.4 Language Reference: Concepts*, Sixth Edn. Cary, NC: SAS Institute Inc.
- Jactel, H., Koricheva, J., and Castagnyrol, B. (2019). Responses of forest insect pests to climate change: not so simple. *Curr. Opin. Insect sci.* 35, 103–108.
- Jactel, H., Nicoll, B. C., Branco, M., Gonzalez-Olabarria, J. R., Grodzki, W., Långström, B., et al. (2009). The influences of forest stand management on biotic and abiotic risks of damage. *Ann. For. Sci.* 66, 701–701.
- Jactel, H., Petit, J., Desprez-Loustau, M. L., Delzon, S., Piou, D., Battisti, A., et al. (2012). Drought effects on damage by forest insects and pathogens: A meta-analysis. *Glob. Change Biol.* 18, 267–276.
- Jiang, M., Medlyn, B. E., Drake, J. E., Duursma, R. A., Anderson, I. C., Barton, C. V., et al. (2020). The fate of carbon in a mature forest under carbon dioxide enrichment. *Nature* 580, 227–231. doi: 10.1038/s41586-020-2128-9
- Karnosky, D., Gielen, B., Ceulemans, R., Schlesinger, W., Norby, R., Oksanen, E., et al. (2001). "FACE systems for studying the impacts of greenhouse gases on forest ecosystems," in *The impact of carbon dioxide and other greenhouse gases on forest ecosystems*, eds D. F. Karnosky, R. Ceulemans, G. E. Scarascia-Mugnozza, and J. L. Innes (Wallingford: CABI Publishing).
- Khairi, S., and Preece, T. (1978). Hawthorn powdery mildew: Overwintering mycelium in buds and the effect of clipping hedges on disease epidemiology. *Trans. Br. Mycol. Soc.* 71, 399–404.
- Lanier, L., Joly, P., Bondoux, P., and Bellemère, A. (1976). *Mycologie et pathologie forestieres-v. 1: Mycologie forestiere.-v. 2*. Paris: Masson.
- Le Quéré, C., Andrew, R. M., Friedlingstein, P., Sitch, S., Hauck, J., Pongratz, J., et al. (2018). Global Carbon Budget 2018. *Earth Syst. Sci. Data* 10, 2141–2194.
- Linkosalo, T., El-Khoury, H., Mäkipää, R., Pulkkinen, P., and Juurola, E. (2017). Increased atmospheric CO<sub>2</sub> concentration enhances the development of photosynthetic capacity beyond the temperature effect for silver birch in simulated future climate. *Scand. J. For. Res.* 32, 651–657.
- Liu, X., Kozovits, A. R., Grams, T. E., Blaschke, H., Rennenberg, H., and Matyssek, R. (2004). Competition modifies effects of enhanced ozone/carbon dioxide concentrations on carbohydrate and biomass accumulation in juvenile Norway spruce and European beech. *Tree Physiol.* 24, 1045–1055.
- Lorca, E. A., Ferreras, A. E., and Funes, G. (2019). Seed size and seedling ontogenetic stage as modulators of damage tolerance after simulated herbivory in a woody exotic species. *Austral. J. Bot.* 67, 159–164.
- Madriaza, K., Saldaña, A., Salgado-Luarte, C., Escobedo, V. M., and Gianoli, E. (2019). Chlorophyll Fluorescence May Predict Tolerance to Herbivory. *Int. J. Plant Sci.* 180, 81–85.
- Marçais, B., and Desprez-Loustau, M.-L. (2014). European oak powdery mildew: impact on trees, effects of environmental factors, and potential effects of climate change. *Ann. For. Sci.* 71, 633–642.
- Martins, C. S., Nazaries, L., Delgado-Baquerizo, M., Macdonald, C. A., Anderson, I. C., and Singh, B. K. (2021). Rainfall frequency and soil water availability regulate soil methane and nitrous oxide fluxes from a native forest exposed to elevated carbon dioxide. *Funct. Ecol.* 35, 1833–1847.
- Massey, F. P., Press, M. C., and Hartley, S. E. (2005). Long- and short-term induction of defences in seedlings of *Shorea leprosula* (Dipterocarpaceae): Support for the carbon: Nutrient balance hypothesis [Article]. *J. Trop. Ecol.* 21, 195–201. doi: 10.1017/s0266467404002111
- Maxwell, K., and Johnson, G. N. (2000). Chlorophyll fluorescence—a practical guide. *J. Exp. Bot.* 51, 659–668. doi: 10.1093/jxb/51.345.659
- Mayoral, C., Strimbeck, R., Sánchez-González, M., Calama, R., and Pardos, M. (2015). Dynamics of frost tolerance during regeneration in a mixed (pine-oak-juniper) Mediterranean forest. *Trees* 29, 1893–1906.
- Medlyn, B. E., Zaehle, S., De Kauwe, M. G., Walker, A. P., Dietze, M. C., Hanson, P. J., et al. (2015). Using ecosystem experiments to improve vegetation models. *Nat. Clim. Change* 5, 528–534.
- Molina-Montenegro, M. A., Torres-Díaz, C., Gallardo-Cerda, J., Leppe, M., and Gianoli, E. (2013). Seabirds modify El Niño effects on tree growth in a southern Pacific island. *Ecology* 94, 2415–2425.
- Morecroft, M. D., Stokes, V. J., Taylor, M. E., and Morison, J. I. (2008). Effects of climate and management history on the distribution and growth of sycamore (*Acer pseudoplatanus* L.) in southern British woodland in comparison to native competitors. *Forestry* 81, 59–74.
- Moriondo, M., Orlandini, S., Giuntoli, A., and Bindi, M. (2005). The effect of downy and powdery mildew on grapevine (*Vitis vinifera* L.) leaf gas exchange. *J. Phytopathol.* 153, 350–357.
- Murchie, E. H., and Lawson, T. (2013). Chlorophyll fluorescence analysis: a guide to good practice and understanding some new applications. *J. Exp. Bot.* 64, 3983–3998. doi: 10.1093/jxb/ert208
- Norby, R. J., and Zak, D. R. (2011). Ecological lessons from free-air CO<sub>2</sub> enrichment (FACE) experiments. *Annu. Rev. Ecol. Syst.* 42, 181–203.
- Norby, R. J., Jastrow, J. D., Miller, M. R., Matamala, R., Boutton, T. W., Rice, C. W., et al. (2005). Elevated atmospheric carbon dioxide increases soil carbon. *Glob. Change Biol.* 11, 2057–2064.
- Norby, R. J., Sholtis, J. D., Gunderson, C. A., and Jawdy, S. S. (2003). Leaf dynamics of a deciduous forest canopy: no response to elevated CO<sub>2</sub>. *Oecologia* 136, 574–584. doi: 10.1007/s00442-003-1296-2
- Norby, R. J., Wullschlegel, S. D., Gunderson, C. A., and Nietch, C. T. (1995). Increased growth efficiency of *Quercus alba* trees in a CO<sub>2</sub>-enriched atmosphere. *N. Phytol.* 131, 91–97. doi: 10.1111/j.1469-8137.1995.tb03058.x
- Nowak, R. S., Ellsworth, D. S., and Smith, S. D. (2004). Functional responses of plants to elevated atmospheric CO<sub>2</sub>—do photosynthetic and productivity data from FACE experiments support early predictions? *N. Phytol.* 162, 253–280.
- Pérez-Ramos, I. M., Cambrollé, J., Hidalgo-Galvez, M., Matias, L., Montero-Ramírez, A., Santolaya, S., et al. (2020). Phenological responses to climate change in communities of plants species with contrasting functional strategies. *Environ. Exp. Bot.* 170:103852.
- Polgar, C. A., and Primack, R. B. (2011). Leaf-out phenology of temperate woody plants: From trees to ecosystems. *N. Phytol.* 191, 926–941.
- Pureswaran, D. S., Roques, A., and Battisti, A. (2018). Forest insects and climate change. *Curr. For. Rep.* 4, 35–50.
- R Core Team (2022). *R: A language and environment for statistical computing*. Vienna: R Foundation for Statistical Computing.
- Roberts, A. J., Crowley, L. M., Sadler, J. P., Nguyen, T. T., Gardner, A. M., Hayward, S. A., et al. (2022). Effects of elevated atmospheric CO<sub>2</sub> concentration on insect herbivory and nutrient fluxes in a mature temperate Forest. *Forests* 13:998.
- Robinson, E. A., Ryan, G. D., and Newman, J. A. (2012). A meta-analytical review of the effects of elevated CO<sub>2</sub> on plant–arthropod interactions highlights the importance of interacting environmental and biological variables. *N. Phytol.* 194, 321–336. doi: 10.1111/j.1469-8137.2012.04074.x
- Ryan, M. G. (2013). Three decades of research at Flakaliden advancing whole-tree physiology, forest ecosystem and global change research. *Tree Physiol.* 33, 1123–1131. doi: 10.1093/treephys/tp100
- Sanchez-Lucas, R., Mayoral, C., Raw, M., Mousouraki, M., and Luna, E. (2023). Elevated CO<sub>2</sub> alters photosynthesis, growth and susceptibility to powdery mildew of oak seedlings. *Biochem. J.* 480, 1429–1443. doi: 10.1042/BCJ20230002
- Schwalter, T., Hargrove, W., and Crossley, D. (1986). Herbivory in forested ecosystems. *Annu. Rev. Entomol.* 31, 177–196.
- Southwood, S. R. (1986). *Insects and the plant surface*. New York City, NY: Edward Arnold.

- Stastny, M., and Agrawal, A. A. (2014). Love thy neighbor? Reciprocal impacts between plant community structure and insect herbivory in co-occurring Asteraceae [Article]. *Ecology* 95, 2904–2914. doi: 10.1890/13-1115.1
- Takamatsu, S., Braun, U., Limkaisang, S., Kom-Un, S., Sato, Y., and Cunnington, J. H. (2007). Phylogeny and taxonomy of the oak powdery mildew *Erysiphe alphitoides* sensu lato. *Mycol. Res.* 111, 809–826. doi: 10.1016/j.mycres.2007.05.013
- Takamatsu, S., Matsuda, S., and Grigaliunaite, B. (2013). Comprehensive phylogenetic analysis of the genus *Golovinomyces* (Ascomycota: Erysiphales) reveals close evolutionary relationships with its host plants. *Mycologia* 105, 1135–1152.
- Talgo, V., Sundheim, L., Gjarum, H., Herrero, M., Suthaparan, A., Toppe, B., et al. (2011). Powdery mildew on ornamental trees and shrubs in Norway. *Eur. J. Plant Sci. Biotechnol.* 5, 86–92.
- Vanaja, M., Maheswari, M., Ratnakumar, P., and Ramakrishna, Y. (2006). Monitoring and controlling of CO<sub>2</sub> concentrations in open top chambers for better understanding of plants response to elevated CO<sub>2</sub> levels. *Indian J. Radio Space Phys.* 35, 193–197.
- Wang, H. J., Wang, H., Ge, Q. S., and Dai, J. H. (2020a). The interactive effects of chilling, photoperiod, and forcing temperature on flowering phenology of temperate woody plants [Article]. *Front. Plant Sci.* 11:443. doi: 10.3389/fpls.2020.00443
- Wang, L.-F., Wang, M., and Zhang, Y. (2014). Effects of powdery mildew infection on chloroplast and mitochondrial functions in rubber tree. *Trop. Plant Pathol.* 39, 242–250.
- Wang, N., Zhao, M., Li, Q., Liu, X., Song, H., Peng, X., et al. (2020b). Effects of defoliation modalities on plant growth, leaf traits, and carbohydrate allocation in *Amorpha fruticosa* L. and *Robinia pseudoacacia* L. seedlings. *Ann. For. Sci.* 77:53.
- Watanabe, M., Kitaoka, S., Eguchi, N., Watanabe, Y., Satomura, T., Takagi, K., et al. (2014). Photosynthetic traits and growth of *Quercus mongolica* var. *crispula* sprouts attacked by powdery mildew under free-air CO<sub>2</sub> enrichment. *Eur. J. For. Res.* 133, 725–733.
- Woodward, R., Waldie, J., and Steven, H. (1929). Oak mildew and its control in forest nurseries. *Forestry* 3, 38–56. doi: 10.3390/plants11091248