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# Dynamic relationships between gross primary production and energy partitioning in three different ecosystems based on eddy covariance time series analysis

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Ecosystems are responsible for strong feedback processes that affect climate. The mechanisms and consequences of this feedback are uncertain and must be studied to evaluate their influence on global climate change. The main objective of this study is to assess the gross primary production (*GPP*) dynamics and the energy partitioning patterns in three different European forest ecosystems through time series analysis. The forest types are an Evergreen Needleleaf Forest in Finland (ENF\_FI), a Deciduous Broadleaf Forest in Denmark (DBF\_DK), and a Mediterranean Savanna Forest in Spain (SAV\_SP). Buys-Ballot tables were used to study the intra-annual variability of meteorological data, energy fluxes, and *GPP*, whereas the autocorrelation function was used to assess the inter-annual dynamics. Finally, the causality of *GPP* and energy fluxes was studied with Granger causality tests. The autocorrelation function of the *GPP*, meteorological variables, and energy fluxes revealed that the Mediterranean ecosystem is more irregular and shows lower memory in the long term than in the short term. On the other hand, the Granger causality tests showed that the vegetation feedback to the atmosphere was more noticeable in the ENF\_FI and the DBF\_DK in the short term, influencing latent and sensible heat fluxes. In conclusion, the impact of the vegetation on the atmosphere influences the energy partitioning in a different way depending on the vegetation type, which makes the study of the vegetation dynamics essential at the local scale to parameterize these processes with more detail and build improved global models.

## KEYWORDS

gross primary production, surface energy fluxes, Granger causality, flux tower, autocorrelation function, atmosphere-vegetation feedback, vegetation dynamics

## 1. Introduction

Vegetation exerts a strong influence on the interaction between the land surface and the atmospheric processes. Climate plays a key role in the carbon, water, and energy fluxes of ecosystems, and the ecosystem itself is also responsible for strong feedback processes that affect climate, as they modify the relative contribution of sensible and latent heat to the total energy of atmospheric air, and therefore directly influence the balance between evapotranspiration and heat exchange processes on the Earth's surface (Arora, 2002; Esau and Lyons, 2002; Forzieri et al., 2020). Thus, land cover and its changes directly affect this balance (Kueppers et al., 2007; Malhi et al., 2008; Shukla et al., 2019). The mechanisms and consequences of this feedback are still uncertain and must be studied to assess their influence on global climate change (Seneviratne and Stöckli, 2008; Schimel et al., 2015).

The role of vegetation in the interaction between atmospheric and land surface processes is misrepresented in land surface models (LSMs) (Lu and Kueppers, 2012; Williams et al., 2016). Significant improvements in understanding these interactions have been achieved when vegetation dynamics are considered (Williams and Torn, 2015; Ma et al., 2017). In this context, Lan et al. (2020) revealed the importance of considering vegetation dynamics through remote sensing observations to quantify the effect of greening on energy partitioning in China. In addition, understanding the land–atmosphere coupling at the local scale in different ecosystems is also needed to improve model performance (Lu and Kueppers, 2012).

Ecosystem traits have been shown to exert a significant effect on dynamic processes of energy, water, and carbon fluxes. In this sense, Wilson et al. (2002) evaluated the factors (physiological and atmospheric) controlling the energy partitioning for different ecosystems and climates during the warm season, showing clear differences among them. Other studies have used the leaf area index (*LAI*) as an indicator of the influence of vegetation on surface energy fluxes. In this sense, Ferguson et al. (2012) included the influence of *LAI*, as well as other factors, on land–atmosphere interaction at a global scale and concluded that the transition zones tend to have the strongest coupling, while energy-limited regions showed the weakest force. At an ecosystem scale, Wang et al. (2019) found high variability in the responses to energy partitioning between plant functional types and years along a river basin in China due to changes in *LAI* and soil water content (*SWC*). In addition, Hoek van Dijke et al. (2020) studied the link between *LAI* and surface fluxes revealing different correlations depending on the ecosystem type. However, they suggested that there is a limitation of using the *LAI* to model or extrapolate surface water and energy fluxes, especially in the Evergreen Needleleaf and Deciduous Broadleaf Forests because *LAI* accounts for vegetation structure but is not a direct measure of the canopy conductance. They concluded that their results cannot be used to study causality between *LAI* and surface fluxes.

Gross primary production (*GPP*) is more related to canopy conductance than *LAI* (Baldocchi et al., 2001b) and could be used also as a proxy for the influence of vegetation on surface fluxes. In this sense, at a global scale, Jung et al. (2011) have shown the connections between carbon and energy fluxes, with the largest

fluxes occurring in the humid tropics and the smallest fluxes in cold and dry environments. At an ecosystem scale, Williams and Torn (2015) studied the relationship between *GPP* and the evaporative fraction in different ecosystems, and Jamiyansharav et al. (2011) observed the connection between latent heat flux (*LE*) and green biomass in a shortgrass steppe.

At the local scale, the eddy covariance (*EC*) technique has been established as one of the best methods to measure carbon, water, and energy fluxes between ecosystems and the atmosphere (Baldocchi et al., 1988; Aubinet et al., 2012; Baldocchi, 2014). Flux towers provide *EC* measurements at the ecosystem level with the option to investigate different time scales (Baldocchi et al., 2001a; Kang et al., 2014). Although flux tower measurements are representative of a limited area windward (depending on sensor height, wind, and stability), they are appropriate because fluxes variability are representative of larger spatial scales because of the spatial coherence of climate anomalies (Nemani et al., 2003; Ciais et al., 2005).

The high temporal resolution of *EC* measurements makes statistical time series analysis (*TSA*) an excellent method to analyze and study these data. *TSA* in its temporal domain (Hamilton, 1994; Box et al., 2015) provides tools and methodologies to model and forecast the dynamics of any variable based on its temporal pattern, which is not other than the history of the variable itself (Huesca et al., 2009). Particularly, the autocorrelation function (*ACF*) has been already used for different environmental purposes. Dornelas et al. (2013) quantified biodiversity changes with the *ACF*. Within an agricultural context, Setiawan et al. (2014) and Tornos et al. (2015) assessed the dynamics of several spectral indices in relation to rice management practices and hydroperiod, whereas Recuero et al. (2019) identified different crop-fallow rotation practices. Within forested areas, the use of this tool has not been fully exploited. In this sense, Huesca et al. (2015) studied broadleaf forests' temporal dynamics in Navarra (Spain). Therefore, in this research, *TSA* is proposed to analyze and model seasonal dynamics of forest ecosystem properties to study the variability of their *GPP* and their energy partitioning, and, over time, to improve models to plan their sustainable management. In addition, *TSA* could be useful to assess the causality between land–atmosphere processes (Krich et al., 2020). In this sense, Granger causality tests (Granger, 1969) could be a first step to assessing land–atmosphere feedback. Granger causality tests have been previously used in forest ecosystems to study the causality between soil respiration and its meteorological and ecological drivers (Detto et al., 2013). In terms of vegetation and climate, Jiang et al. (2015) used this test to observe the impacts of vegetation change on the local surface climate in China, and Papagiannopoulou et al. (2017) used non-linear Granger causality tests to investigate climate-vegetation dynamics.

The general objective of this research is to evaluate the dynamics of gross primary production (*GPP*) and energy partition patterns in three different European forest ecosystems through time series analysis. The three ecosystems were selected according to their contrasting differences in terms of functioning and climate. The specific objectives are as follows: (1) to evaluate the annual and inter-annual dynamics of *GPP*, meteorological variables, and energy fluxes and (2) to evaluate the causality in the Granger

direction between *GPP* and energy fluxes in the very short term (daily frequency).

## 2. Materials and methods

### 2.1. Study area

Three forests in Europe have been selected for this study (Figure 1 and Table 1). The first one is known as Hyytiälä and is in the south of Finland at 240 17' 36.9" E and 610 50' 48.5" N at an altitude of 181 m above sea level. It is an Evergreen Needleleaf Forest (ENF\_FI) of Scots pine (*Pinus sylvestris* L.). The climate at this site is boreal with a mean annual temperature of 2.9°C and a mean annual precipitation of 709 mm. The soil is composed of sandy and coarse silty glacial till (Suní et al., 2003).

The second one, known as Sorø, is located in the east of Denmark at 110 38' 40.7" E and 550 29' 9.1" N at an altitude of 40 m above sea level. It is a Deciduous Broadleaf Forest (DBF\_DK) dominated by the European beech (*Fagus sylvatica* L.), with small scattered stands of Norway spruce (*Picea abies* L. Karst.), as well as single trees of other conifers such as European larch (*Larix decidua* Mill.). The climate at this site is temperate maritime with a mean annual temperature of 8.2°C and a mean annual precipitation of 660 mm. The soils are brown soils classified as either Alfisols or Mollisols with a 10–40 cm deep organic layer (Pilegaard et al., 2011).

The third one is known as Las Majadas del Tiétar located in the west of Spain (SAV\_SP) at 5°46'29"W and 39°56'26" N at an altitude of ~260 m above sea level. It is Savannah Mediterranean Forest (SAV) known in Spanish as dehesa and is dominated by sclerophyllous holm oaks (*Quercus ilex* L. ssp. *ballota* (Desf.) Samp.), with an understory in which Mediterranean grassland species predominate. The climate at this site has typical Mediterranean characteristics with a mean annual temperature of 16.7°C and a mean annual rainfall of 572 mm, with a marked seasonality. The soil is classified as Dystric Cambisol (Casals et al., 2009).

### 2.2. Flux tower data

In each site, there is an eddy covariance flux tower. The flux towers of Hyytiälä (ENF\_FI) and Sorø (DBF\_DK) are integrated into the global network FLUXNET (<https://fluxnet.fluxdata.org/>), and the tower of Las Majadas (SAV\_SP) is integrated into the European fluxnet database cluster (<https://www.europe-fluxdata.eu/>).

The study period of Hyytiälä and Sorø was from 1996 to 2015 and from 2004 to 2011 in Las Majadas. The following pre-processed daily variables for gap-filling and filtering (Papale and Valentini, 2003; Reichstein et al., 2005; Papale et al., 2006) were used (Table 2) as follows: (1) air temperature ( $T_a$ , °C), (2) atmospheric pressure ( $P$ , hPa), (3) wind speed ( $WS$ ,  $m\ s^{-1}$ ), (4) vapor pressure deficit ( $VPD$ , hPa), (5) precipitation ( $PPT$ ,  $mm\ day^{-1}$ ), (6) soil water content ( $SWC$ , % vol), (7) latent heat flux ( $LE$ ,  $W\ m^{-2}$ ), (8) sensible heat flux ( $H$ ,  $W\ m^{-2}$ ), and (9) gross primary productivity ( $GPP$ ,  $g\ C\ m^{-2}\ day^{-1}$ ).

Atmospheric pressure was used to estimate the mixing ratio ( $r$ ,  $g\ H_2O\ kg\ dry\ air^{-1}$ ) as follows:

$$r = \varepsilon \frac{e}{(P - e)} \quad (1)$$

where  $\varepsilon$  is the ratio of molar masses of vapor and dry air (622),  $e$  is the vapor pressure derived from the difference between  $e_s$  (the saturation vapor pressure derived from Clausius–Clapeyron equation) and  $VPD$ , and  $P$  is the atmospheric pressure.

Carbon (net ecosystem exchange) and energy (sensible and latent heat) fluxes were estimated through the eddy covariance method (Baldocchi, 2003, 2014). In ENF\_FI and DBF\_DK,  $GPP$  was calculated according to the daytime method (Lasslop et al., 2010), while in SAV\_SP,  $GPP$  was calculated as the difference between the ecosystem respiration ( $Reco$ ,  $g\ C\ m^{-2}\ day^{-1}$ ), which was estimated according to the short-term temperature response of night-time fluxes (Papale and Valentini, 2003; Reichstein et al., 2005) and net ecosystem exchange ( $NEE$ ,  $g\ C\ m^{-2}\ day^{-1}$ ) (Equation 2).

$$GPP = Reco - NEE \quad (2)$$

### 2.3. Statistical methods

First, all variables were analyzed through a univariate time series analysis. Buys-Ballot tables (Buys-Ballot, 1847), i.e., an average year, were used to study the intra-annual variability of meteorological data, energy fluxes, and  $GPP$ . The autocorrelation function was applied to assess the dynamics of the variables' time series allowing the identification of repetitive patterns, periodicity components, and temporal dependency in the time series by a set of correlation coefficients that quantify the relationship between observations at different time intervals (lags) (Box et al., 2015). In our case, our lags correspond to a daily frequency.

Second, the causality of  $GPP$  and energy fluxes was studied with Granger causality tests (Granger, 1969), using the alternative tests introduced by Geweke (1984). According to Granger (1969), "...  $Y_t$  is causing  $X_t$  ( $Y_t \rightarrow X_t$ ) if we are better able to predict  $X_t$  using all available information than if the information apart from  $Y_t$  had been used". Thus, the Granger causality tests are used to verify the usefulness of one variable to forecast another. The hypothesis of the test is as follows:

The null hypothesis ( $H_0$ ): "past values (at different lags, in our case daily lags) of a variable  $Y_t$  cannot help to predict other variable  $X_t$ ".

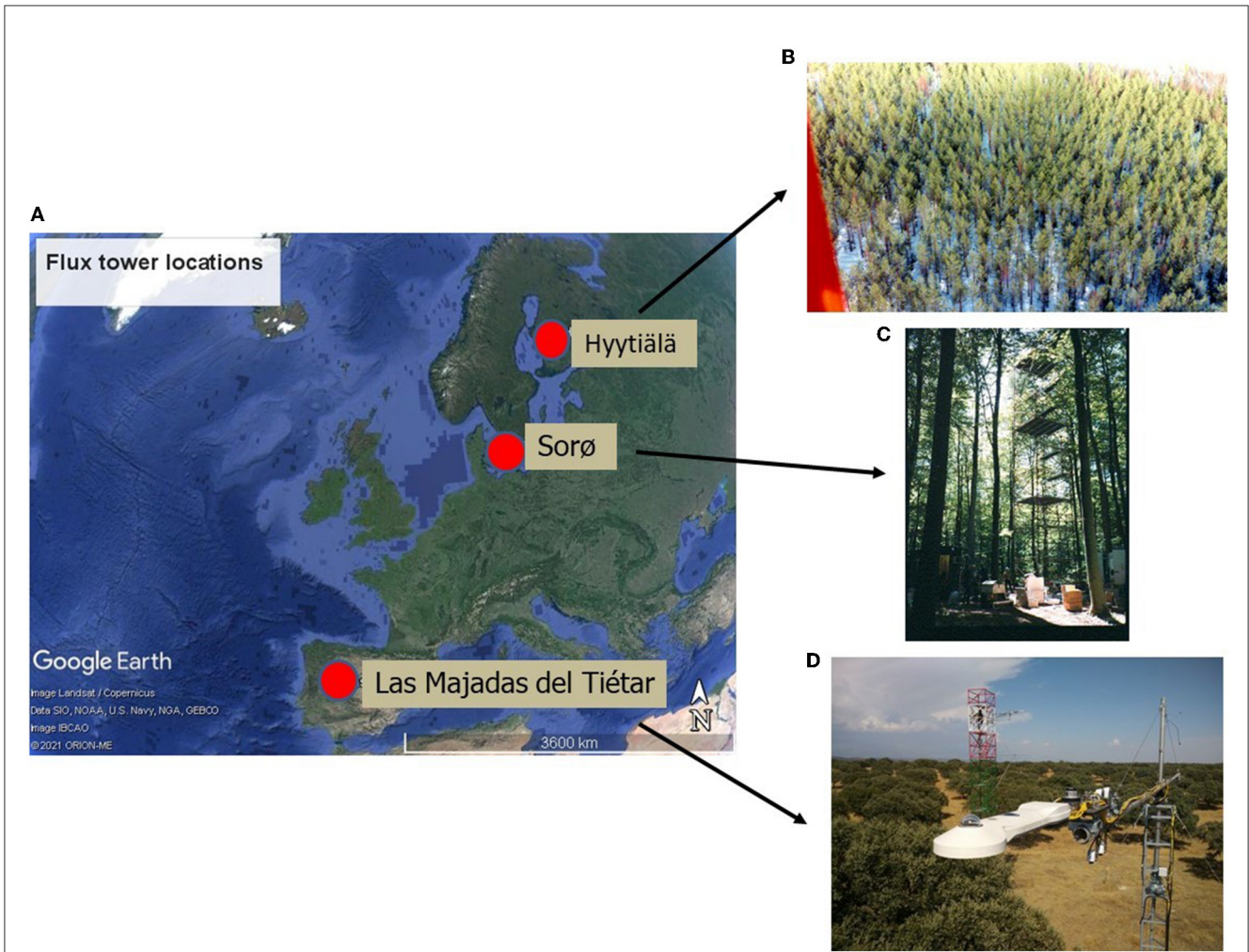
The alternative hypothesis ( $H_1$ ): "past values (at different lags, in our case daily lags) of a variable  $Y_t$  can help to predict other variable  $X_t$ ".

The test is based on the following ordinary least squares (OLS) regression model:

$$X_t = \alpha_0 + \sum_{j=1}^m \alpha_j X_{t-j} + \sum_{j=1}^m \beta_j Y_{t-j} + \varepsilon_t \quad (3)$$

where  $\alpha_j$  and  $\beta_j$  are the regression coefficients, and  $\varepsilon_t$  is the error term. Thus, the test is based on the null hypothesis as follows:





**FIGURE 1** Location of the three European ecosystems (A): Evergreen needle leaf forest of Finland (ENF\_FI) (B), Deciduous Broadleaf Forest of Denmark (DBF\_DK) (C) and Mediterranean dehesa of Spain (SAV\_SP) (D). Images from CEAM (2021) and Fluxnet (2021). Map in (A) from © Google Earth (Landsat/Copernicus).

**TABLE 1** Site characteristics of the three flux tower sites (Tmean: mean annual temperature in °C, PPT mean: mean annual rainfall in mm).

Site	Tmean (°C)	PPTmean (mm)	Soil type	Plant functional type	Study period (observations)	References
Hyytiälä (ENF_FI)	2.9	709	Haplic podzol	ENF	1996–2015 (6940)	Suni et al., 2003
Sorø (DBF_DK)	8.2	660	Mollisol	DBF	1996–2015 (6940)	Pilegaard et al., 2001
Las Majadas del Tiétar (SAV_SP)	16.7	572	Cambisol	SAV	2004–2011 (2922)	Casals et al., 2009

$$H_0: \beta_1 = \beta_2 = \dots = \beta_m = 0 \text{ (m = lags)}$$

Being:

If we estimate the following restricted equation also by OLS:

$$X_t = \alpha_t + \sum_{j=1}^m \gamma_j X_{t-j} + e_t \tag{4}$$

$$RSS1 = \sum_{j=1}^m \hat{\epsilon}_t^2 \tag{6}$$

$$RSS0 = \sum_{j=1}^m \hat{\epsilon}_t^2 \tag{7}$$

Then, the statistic used for the test is an *F*-statistic based on comparing the sum of square residuals of both models as follows:

$$f = \frac{\frac{(RSS0 - RSS1)}{m}}{\frac{(RSS1)}{(T - 2m - 1)}} \sim F_{m, T - 2m - 1} \tag{5}$$

The statistical analysis was computed through Statgraphics 18 (StatPoint Technologies Inc., VA, United States), Eviews 10 University edition (IHS Global Inc., CA, United States), and SAS software (SAS 9.4 Software, SAS Institute Inc., NC, United States).

TABLE 2 Variables obtained from the flux tower sites.

Variable	Units	Method
Air Temperature, $T_a$	°C	Direct measurement
Atmospheric pressure, $P$	hPa	Direct measurement
Wind speed, $WS$	$m\ s^{-1}$	Direct measurement
Vapor pressure deficit, $VPD$	hPa	Derived from $T_a$ and Relative Humidity
Precipitation, $PPT$	$mm\ day^{-1}$	Direct measurement
Soil water content, $SWC$	% vol	Direct measurement
Latent Heat, $LE$	$W\ m^{-2}$	Eddy covariance
Sensible Heat, $H$	$W\ m^{-2}$	Eddy covariance
Gross primary productivity, $GPP$	$g\ C\ m^{-2}\ day^{-1}$	Derived from NEE and Reco (SAV_SP) and using the day-time method (ENF_FI and DBF_DK).

### 3. Results

#### 3.1. Average annual patterns of the variables

Figure 2 shows the average annual time series of the three GPP (daily values), i.e., Buys-Ballot. The DBF of Denmark (GPP\_DK) and the ENF of Finland (GPP\_FI) showed a clear annual cycle with maximum values in June ( $\sim 16\ g\ C\ m^{-2}\ day^{-1}$ ) and July ( $\sim 9\ g\ C\ m^{-2}\ day^{-1}$ ), respectively, and minimum values close to  $0\ g\ C\ m^{-2}\ day^{-1}$  in winter. On the other hand, daily GPP in the dehesa of Spain (GPP\_SP) showed minimum values at the end of the summer ( $\sim 1\ g\ C\ m^{-2}\ day^{-1}$ ) and winter ( $\sim 2\ g\ C\ m^{-2}\ day^{-1}$ ), and maximum values in spring ( $\sim 7\ g\ C\ m^{-2}\ day^{-1}$ ) and late autumn ( $\sim 2\ g\ C\ m^{-2}\ day^{-1}$ ).

Gross primary production (GPP) in Finland and Denmark showed an accumulated annual value of  $\sim 1,175$  and  $1,940\ g\ C\ m^{-2}\ year^{-1}$ , respectively, throughout the study period while GPP in Spain exhibited an average annual value of  $\sim 1,120\ g\ C\ m^{-2}\ year^{-1}$ .

Figure 3 shows the intra-annual dynamics of meteorological variables and energy fluxes for the three ecosystems.

The Evergreen Needleleaf Forest of Finland, in a boreal climate, showed minimum values of air temperature (below  $-5^\circ C$ ) and precipitation ( $\sim 1.5\ mm\ day^{-1}$ ) during winter and maximum values of both variables in summer ( $\sim 17^\circ C$  and  $5.5\ mm\ day^{-1}$ ). SWC was relatively constant ( $\sim 25\%$ ) during winter but had a maximum at the end of April after snow melt and a minimum ( $\sim 20\%$ ) during summer, with the maxima temperature and GPP directly related to the evapotranspiration, although PPT was also higher during this season. The wind speed was  $\sim 3.5\ m\ s^{-1}$  all year, except in summer which had lower values of  $\sim 3\ m\ s^{-1}$ . Due to the low temperatures, LE, VPD, GPP (the three are close to zero), H ( $\sim 12\ W\ m^{-2}$ ), and  $r$  ( $\sim 2\ g\ kg^{-1}$ ) were minimum during winter. In spring, all these variables increased, especially sensible heat which reached its maximum in mid-May ( $H\sim 76\ W\ m^{-2}$ ). Then, the rest of the variables presented their maxima in summer ( $T_a\sim 17^\circ C$ ,

$VPD\sim 8\ hPa$ ,  $LE\sim 75\ W\ m^{-2}$ , and  $r\sim 8.5\ g\ kg^{-1}$ ). Subsequently, they began to decrease at the end of summer and during autumn.

The Deciduous Broadleaf Forest of Denmark is located in a temperate climate and had minimum  $T_a$  values in winter (higher than  $0^\circ C$ ), and PPT values were relatively constant all year. As a consequence, VPD ( $\sim 0.5\ hPa$ ), LE ( $\sim 0\ W\ m^{-2}$ ), H ( $\sim 45\ W\ m^{-2}$ ), and  $r$  ( $\sim 4\ g\ kg^{-1}$ ) were also low during winter.  $T_a$  began to increase at the end of March, as also happened in Finland; later VPD, LE, and  $r$  increased at the same time as GPP did, reaching their maximum values around mid-July ( $T_a\sim 18^\circ C$ ,  $VPD\sim 6\ hPa$ ,  $LE\sim 105\ W\ m^{-2}$  and  $r\sim 10\ g\ kg^{-1}$ ). H also increased from March, but its maximum value ( $H\sim 75\ W\ m^{-2}$ ) was reached in May. At the end of summer, all variables began to decrease (H begin to decrease earlier) and continued decreasing during autumn. Wind speed was higher than in Finland,  $\sim 5.8\ m\ s^{-1}$  during autumn and winter and  $\sim 4.5\ m\ s^{-1}$  in spring and summer.

The savanna ecosystem of Spain, located in a Mediterranean climate, is characterized by higher temperatures ( $\sim 27^\circ C$ ) and a strong drought during summer (PPT  $\sim 0\ mm\ day^{-1}$ , SWC around 5%).  $T_a$  was also at its minimum during winter (higher than  $5^\circ C$ ), and subsequently, VPD,  $r$ , LE, and H were also low during winter ( $VPD\sim 1.5\ hPa$ ,  $r\sim 6\ g\ kg^{-1}$ ,  $LE\sim 15\ W\ m^{-2}$ , and  $H\sim 6\ W\ m^{-2}$ ).  $T_a$  began to increase earlier than in Finland and Denmark, in early March. At the same time, GPP,  $r$ , LE, and H also began to increase. LE reached its maximum at the end of April ( $LE\sim 95\ W\ m^{-2}$ ), before the maximum of  $r$  which occurred in mid-June ( $r\sim 8.5\ g\ kg^{-1}$ ) and those of VPD and H occurred at the end of July ( $VPD\sim 28\ hPa$ ,  $H\sim 100\ W\ m^{-2}$ ). At the end of the summer, LE reached its minimum ( $LE\sim 14\ W\ m^{-2}$ ) and then showed a recovery and a second maximum in autumn ( $LE\sim 40\ W\ m^{-2}$ ) directly related to the behavior shown in GPP in Figure 2 and in accordance with the marked seasonality in water availability. Wind speed was  $\sim 2.5\ m\ s^{-1}$  all year, except in winter which had lower values of  $\sim 2\ m\ s^{-1}$ .

#### 3.2. Assessment of the variable's dynamics

The autocorrelation function (ACF) of the daily GPP had a clear annual pattern in the three ecosystems (Figure 4). The absolute values of GPP\_DK and GPP\_FI were higher than those of GPP\_SP in all lags, with the greatest differences being observed at 180 lags,  $\sim 6$  months, in the case of negative values. Moreover, GPP\_SP showed lower autocorrelation values at the annual lags (lags 365, 730), with less significant values after the second year.

The autocorrelation function for daily meteorological variables and energy fluxes in the three ecosystems is presented in Figure 5. All variables showed a clear annual pattern in the three ecosystems, i.e., the same pattern was found every 365 lags. While in the ENF\_FI and DBF\_DK, autocorrelation values of the variables at 1 year term were similar to those at the second year term, in Spain they were lower, i.e., lower memory at the long term. The autocorrelation values of  $T_a$  were very similar in the three ecosystems at all lags. VPD in the savanna of Spain showed higher ACF values at short, medium, and annual terms than the ENF\_FI and the DBF\_DK which had similar values. H showed higher ACF values in Spain than in Finland in a short term. However, ACF values in Finland

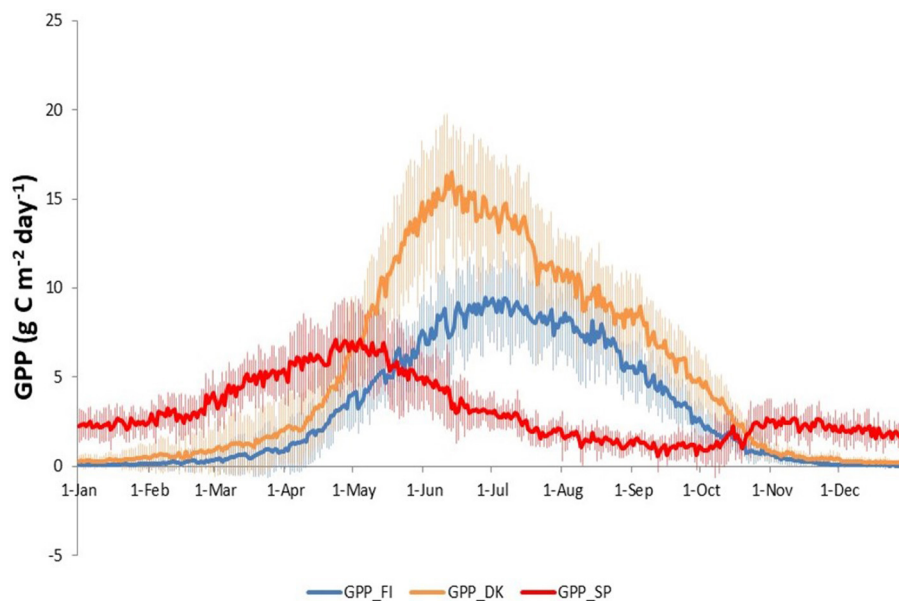


FIGURE 2

Annual average, i.e., Buys-Ballot, of daily GPP time series in three European ecosystems: Evergreen Needleleaf Forest of Finland (ENF\_FI, blue line), Deciduous Broadleaf Forest of Denmark (DBF\_DK, orange line) and Mediterranean Savanna ecosystem of Spain (SAV\_SP, red line).

were similar to those in Spain at an annual term (lag 365) and higher at a second-year term (lag 730). Denmark presented the lowest values during all lags. The mixing ratio,  $r$ , showed higher ACF values in Finland and Denmark than in Spain during all lags, indicating a higher seasonality. Although very similar values of the ACF of  $LE$  were obtained in the three ecosystems in the short term, those of the annual lags of Denmark and Finland were greater than those of Spain. In the three ecosystems,  $PPT$  and  $WS$  only showed significant values of the ACF in the short term. Although ACF values were lower and not significant, an annual pattern is suggested in the three ecosystems.  $SWC$  showed the strongest differences between the three ecosystems. While in Spain, ACF values of  $SWC$  revealed a clear annual pattern, values in Finland suggested the presence of two cycles, and in Denmark showed a less marked annual cycle.

### 3.3. Energy partitioning during the growing period

Figure 6 shows the Bowen ratio ( $\beta = H/LE$ ) (a), the ratio between  $GPP$  vs. the sum of sensible heat and latent heat (b), the ratio between  $GPP$  and sensible heat (c), and the ratio between  $GPP$  and latent heat (d), for daily values of the average year during the growing period (April to October for the northern ecosystems and all year for the savanna ecosystem). This period was chosen to select the period when potential solar radiation and temperature were high, avoiding unusual values in these ratios (Wilson et al., 2002).

The savanna ecosystem showed the highest Bowen ratio and the lowest calculated ratio. Mediterranean ecosystems present higher radiation and a larger amount of  $H+LE$ , which is associated with

a lower primary production efficiency in terms of total energy ( $GPP/(H+LE)$ , panel b), and in terms of sensible heat flux ( $GPP/H$ , panel c), with practically null ratios and that could even become negative due to anomalous values in the average year, especially in winter. The evapotranspiration efficiency was also low ( $GPP/LE$ , panel d).

On the contrary, both northern ecosystems showed similar production efficiencies in terms of total energy ( $GPP/(H+LE)$ , panel b) and evapotranspiration ( $GPP/LE$ , panel d). However, the Deciduous Broadleaf Forest of Denmark showed a lower Bowen ratio, which leads to a lower  $GPP/H$  ratio. This is related to a larger amount of latent heat in relation to sensible heat in DBF\_DK and is associated with the larger plant activity in this forest.

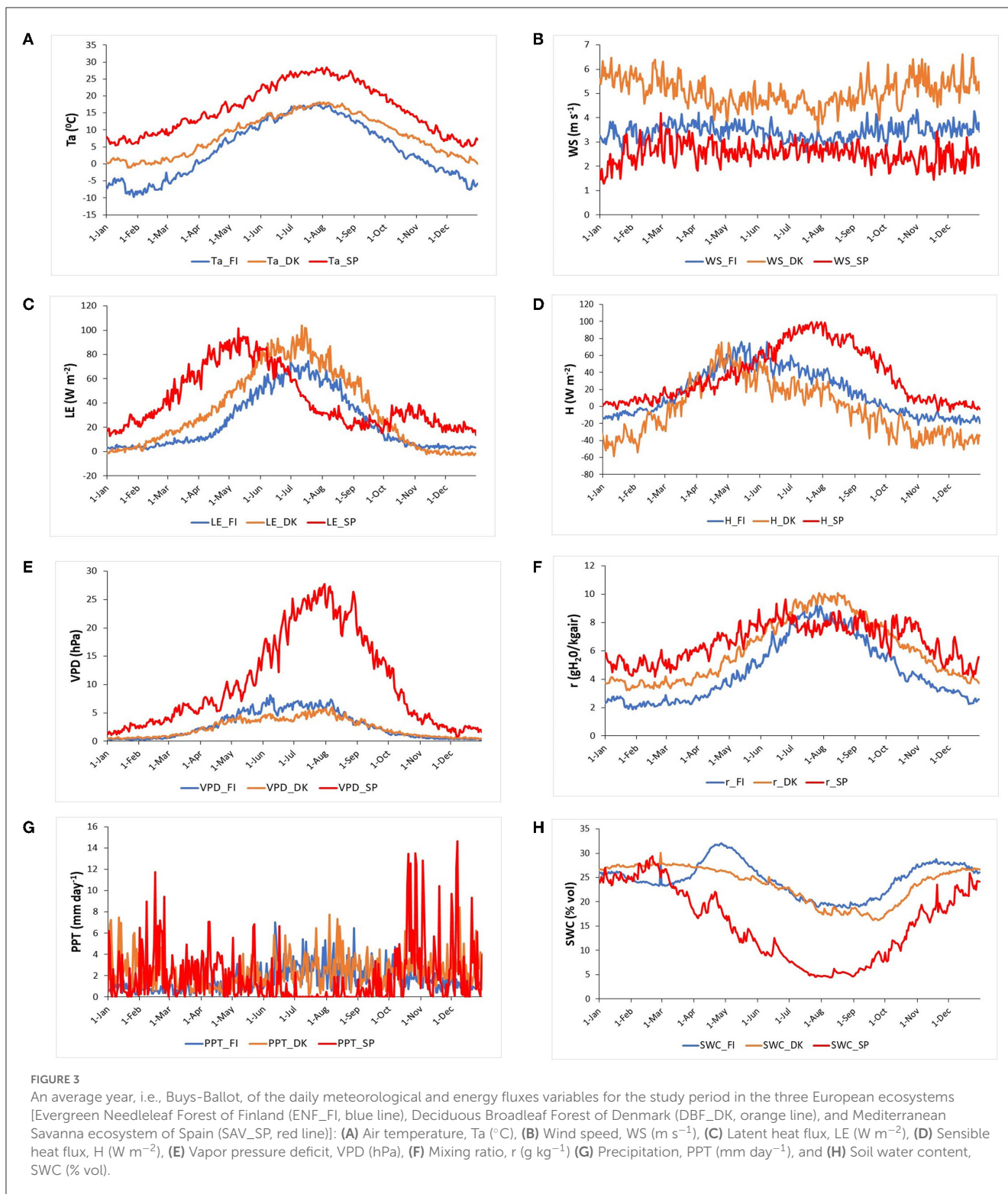
### 3.4. Granger causality tests between GPP and energy fluxes

Figure 7 shows the results of the Granger causality tests between  $GPP$  and  $LE$  in both directions ( $GPP \Rightarrow LE$ ) and between  $GPP$  and  $H$  in both directions ( $GPP \Rightarrow H$ ), with the null hypothesis ( $H_0$ ): “the left-side variable does not Granger-cause the right-side variable.”

Significant F-test statistics were found when  $LE$  caused  $GPP$  in the three ecosystems for the first lag, and then values decreased for the following lags. In Finland, F-test values were higher than in Denmark. While in both ecosystems, F-test statistics were significant until lag 10 (except for lag 2 in Finland), and in Spain, F-test statistics were not significant at a 5% confidence level from lag 6 to lag 10.

Significant and higher F-test statistics were found when  $GPP$  caused  $LE$  in the three ecosystems, especially in Denmark and Finland with very high F-test values

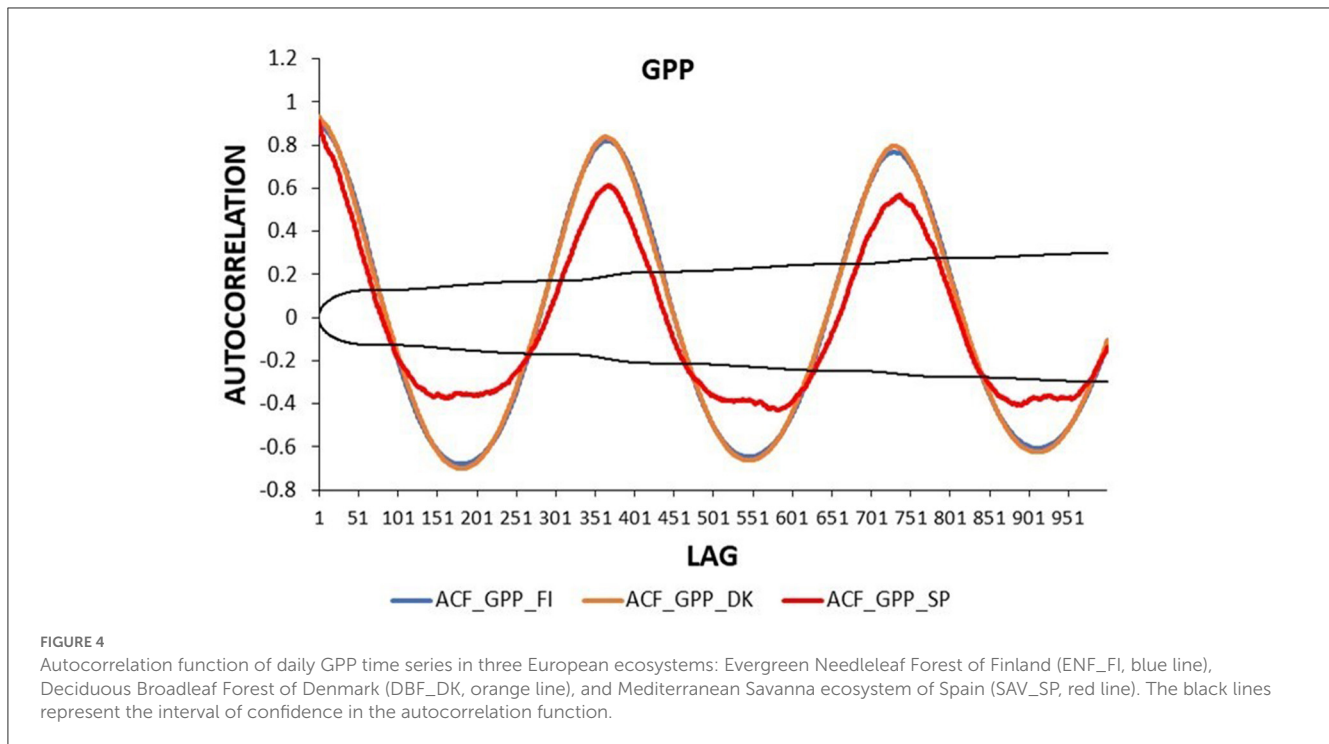




decreasing from  $\sim 500$  for the first lag to 30 in lag 6 but having significant values until lag 10. In Spain,  $F$ -test values also showed a decreasing trend but with much lower values ranging approximately between 63 in the first lag and 6 in lag 10.

Significant and higher  $F$ -test statistics were found when  $H$  caused  $GPP$  than in the opposite direction in Finland and Spain.

In fact, when  $GPP$  caused  $H$  in Spain,  $F$ -test statistics were only significant at a 5% confidence level for the first two lags. In this sense,  $H$  caused  $GPP$  and  $F$ -test values were higher in Finland than in Spain and showed a decreasing trend from lag 1 to lag 10. Meanwhile, in Denmark, higher  $F$ -test statistics were found when  $GPP$  caused  $H$  at a very short term (3 lags). Then,  $H$  caused slightly more  $GPP$  than backward.



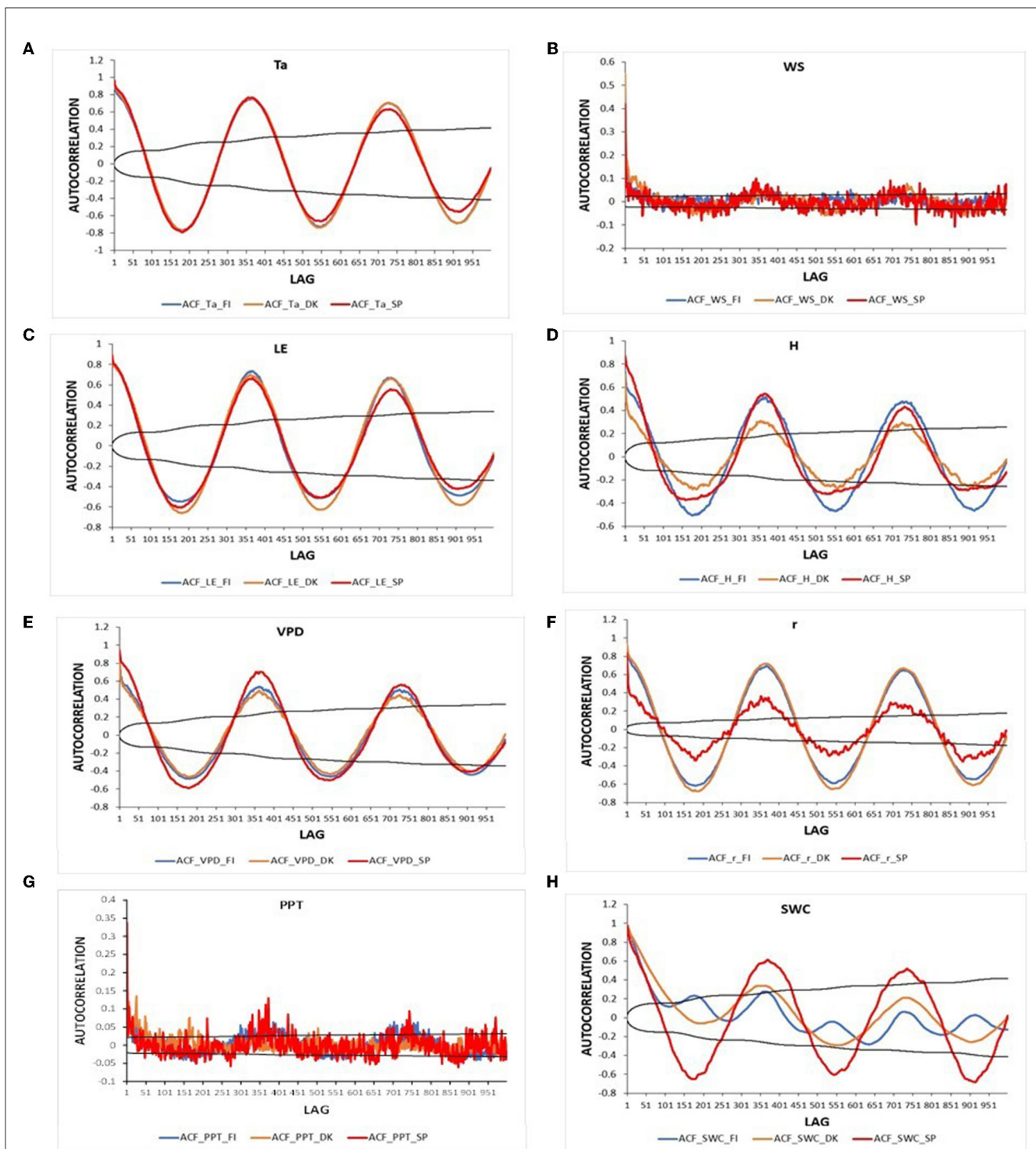
## 4. Discussion

The analyses presented served to quantify the importance of the amount of carbon fixed every year by these different types of ecosystems (Beer et al., 2010; Pan et al., 2011), and the results, in terms of the quantity of annual GPP, were consistent with other studies. Lagergren et al. (2008) showed a GPP mean annual value of  $\sim 1,612 \text{ g C m}^{-2} \text{ year}^{-1}$  in Denmark and  $1,047 \text{ g C m}^{-2} \text{ year}^{-1}$  in Finland at the same flux tower sites for the period between 2000 and 2005, while Cicuéndez et al. (2015) showed an average annual value  $\sim 1,091 \text{ g C m}^{-2} \text{ year}^{-1}$  for the 2004–2008 period in the savanna of Spain. In terms of memory, i.e., autocorrelation values in the Mediterranean ecosystem due to the inter-annual variability in SWC which is responsible for the variability in gross primary production and which, in fact, is modulating the variability of  $LE$  and  $r$ , the two latter variables had less memory at long term than in Finland and Denmark. In contrast, due to very high VPD and  $H$  values every summer, the autocorrelation function revealed a more marked annual cycle in these variables than in temperate and boreal climates.

Different patterns in the annual dynamics of the energy partitioning into sensible and latent heat were found between the three sites. In the northern sites (ENF\_FI and DBF\_DK), most of the available energy was converted into sensible heat at the beginning of the spring, while in the Mediterranean Savanna ecosystem (SAV\_SP), this component was dominant in summer, after the growing period (Figure 3). At higher latitudes, low temperatures which caused low VPD values and the lack of vegetative activity resulted in all the energy dedicated to sensible heat, whose maximum was reached before the maximum of temperature and solar radiation. This large sensible heat

component had a significant effect on temperature increase that enhanced the beginning of photosynthetic activity and therefore GPP. This has been confirmed by the Granger causality tests which showed that  $H$  is statistically causing GPP in both ecosystems (ENF\_FI and DBF\_DK) probably through its effect on temperature. In fact, the ENF\_FI showed a higher GPP/ $H$  ratio indicating that the ecosystem is more efficient when radiation was being spent in raising temperature because when optimal temperature conditions were reached, coniferous trees began to produce. According to Mäkelä et al. (2006), the thermal growing season in this site started between 18 April and 7 May. Regarding the DBF\_DK, leaf emergence occurred approximately at the end of April, and leaf expansion lasted  $\sim 1$  month (Wang et al., 2005). Other researchers have found that temperature and day length were limiting factors in this ecosystem during winter, controlling leaf emergence (Basler and Korner, 2014; Dolschak et al., 2019). It seems that the bud-break of beech could coincide with the maximum of sensible heat (Wilson and Baldocchi, 2000). Probably, the first increment in GPP was caused by the growth of herbaceous and shrub species of the ecosystem which are adapted to grow before the shading of beech leaves (Pilegaard et al., 2001; Brunet et al., 2010; Paul-Limoges et al., 2017). It was also probable that coniferous, which represents 20% of the footprint area (Pilegaard et al., 2001), could begin to fix carbon before beech trees (Pilegaard et al., 2001). After this phase, in both ecosystems (DBF\_DK and ENF\_FI), an increase in transpiration together with higher VPD values (because of higher temperatures) resulted in large and increasing values of the latent heat component that were shown to be caused in a Granger sense by GPP, with high F-test Granger values. In addition, in Denmark, the expansion of LAI led to more shade on the ground which entailed a reduction of changes in surface temperature and subsequently affected  $H$ . Thus, GPP affected  $H$  more than backward at a very short term (3 lags).

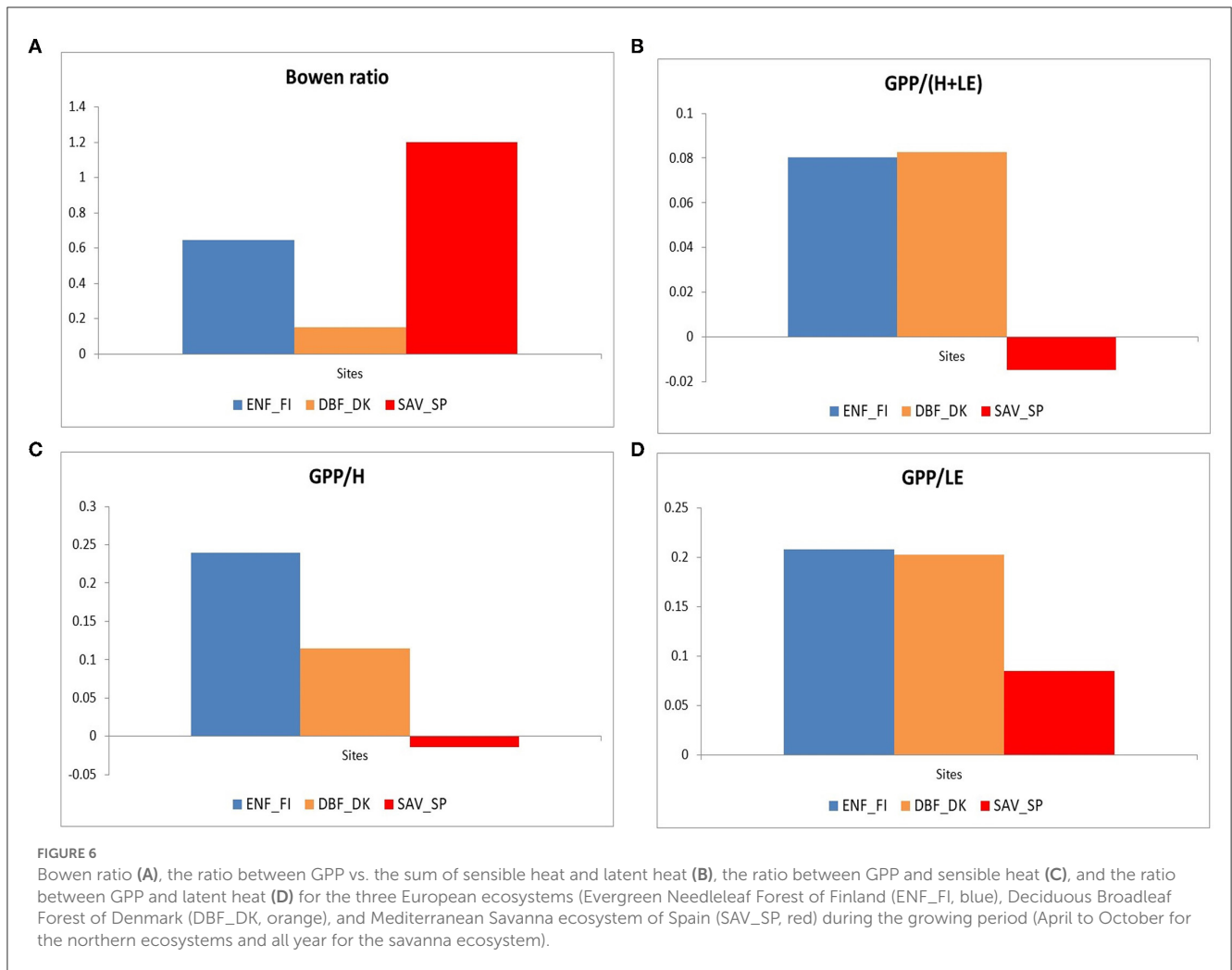




**FIGURE 5** Autocorrelation of the daily meteorological and energy fluxes variables for the study period in the three European ecosystems [Evergreen Needleleaf Forest of Finland (ENF\_FI, blue line), Deciduous Broadleaf Forest of Denmark (DBF\_DK, orange line), and Mediterranean Savanna ecosystem of Spain (SAV\_SP, red line)]: **(A)** Air temperature (Ta) (°C), **(B)** Wind speed (WS) (m s<sup>-1</sup>), **(C)** Latent heat flux (LE) (W m<sup>-2</sup>), **(D)** Sensible heat flux (H) (W m<sup>-2</sup>), **(E)** Vapor pressure deficit (VPD) (hPa), **(F)** Mixing ratio (r) (g kg<sup>-1</sup>) **(G)** Precipitation (PPT) (mm day<sup>-1</sup>), and **(H)** Soil water content (SWC) (% vol). The black lines represent the interval of confidence in the autocorrelation values.

In northern ecosystems, the net radiation continued increasing during the summer, and they are adapted to have the maximum *GPP* when solar radiation is maximum taking advantage of the optimal conditions for photosynthetic activity and showing clearly

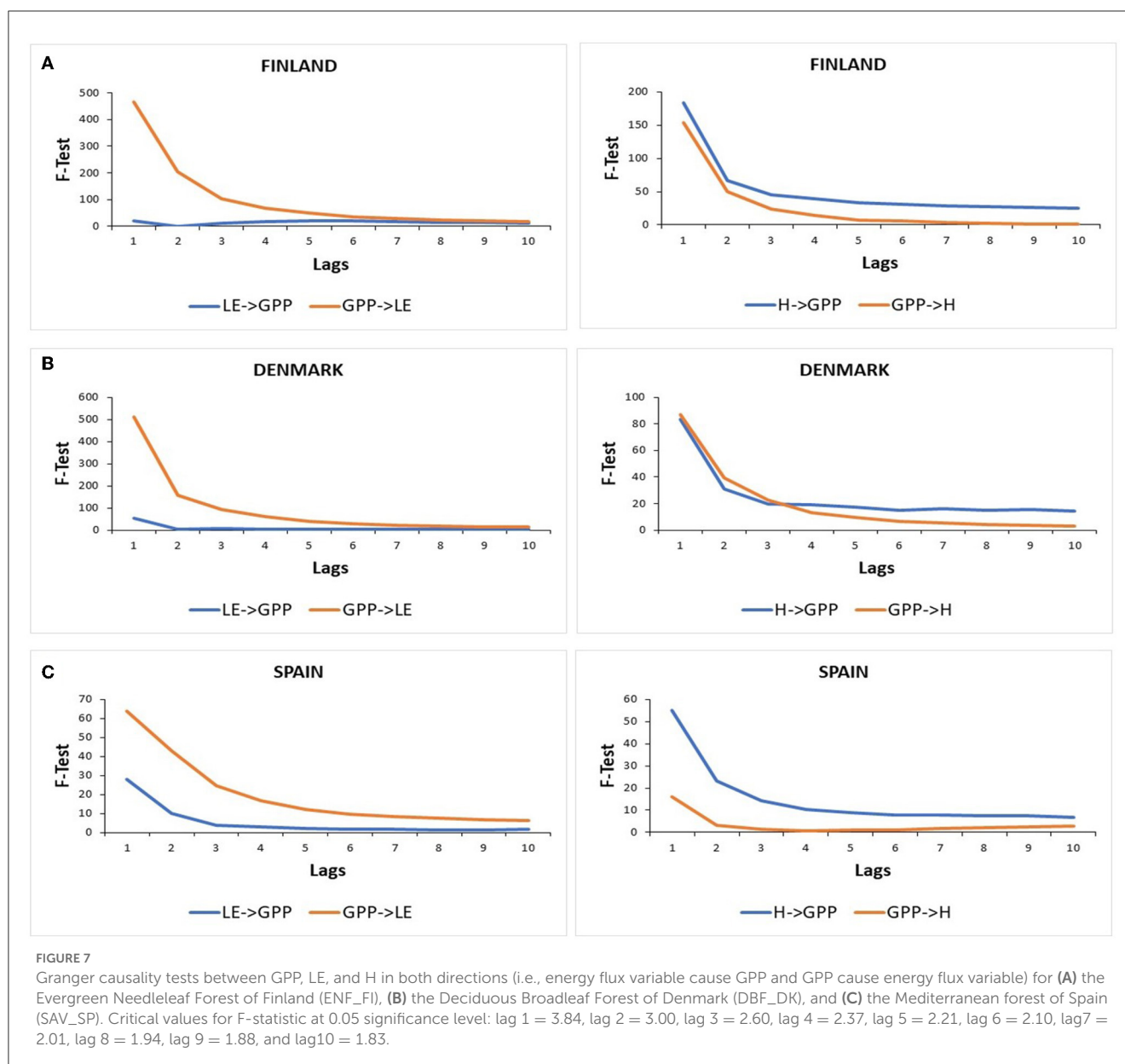
higher values of *GPP* during this period than the water-limited ecosystems (SAV\_SP). In fact, tracheid production in coniferous at boreal sites (ENF\_FI) occurred around the time of maximum day length and not during the maximum temperature period (Rossi



et al., 2006; Kulmala et al., 2017). This fact could explain the delay between the maximum *GPP* at the end of June, and the maximum temperature, *LE*, and *r* at the end of July. However, *GPP* remained approximately high and constant during summer. In contrast, in DBF\_DK, *GPP* reached a clear maximum at the end of June remaining relatively high for 1 month when *LE* reached its maximum, with high transpiration rates (Wang et al., 2019), and then began to decrease clearly by the end of July.

However, in the savanna ecosystem of Spain, the temperature was limited only during a short period in winter, and in general the tree and the herbaceous layer could produce during this season (Chiariello, 1989; Xu and Baldocchi, 2004; Ma et al., 2007); then, as soon as water was available, the energy was converted into the latent heat component. Water availability is the main limiting factor in this type of ecosystem (Ross et al., 2012; Gilabert et al., 2015; Gómez-Giraldez et al., 2018). In this respect, Jung et al. (2011) showed the importance of the water cycle in the inter-annual carbon fluxes dynamics in semiarid regions. Therefore, in spring, the latent heat was maximum before the maximum temperature and solar radiation. Sridhar et al. (2019) showed this evolution in other semiarid regions. Vegetation in Mediterranean ecosystems, especially grasslands, is adapted to grow earlier in the spring to take advantage of lower temperatures, lower *VPD*, and higher

*SWC* than the following months (Evans and Young, 1989; Milla et al., 2010; Cicuéndez et al., 2015; Román-Cascón et al., 2020). During this period, *GPP* was strongly coupled to latent heat and *SWC*. Thus, Granger causality tests indicated that *GPP* statistically caused latent heat in the savanna ecosystem, especially due to the *GPP* of the herbaceous layer which contributes more to the latent heat than the tree layer during spring (El-Madany et al., 2018), especially in savanna ecosystem with a low tree cover. However, F-test values were lower than in the energy-limited regions because the savanna showed lower *GPP* values, with a lower *LAI* and less soil water content, and subsequently lower evapotranspiration rates. In summer, the sensible heat was at its maximum, producing high temperatures, while the latent heat reached its minimum values due to the decrease in *GPP* and the evapotranspiration rate of the species of the herbaceous layer which dried quickly due to high *Ta* and *VPD* values. The tree layer can continue to produce a slightly longer period than the herbaceous one because its deeper roots can draw water from the deeper layers of the soil (Joffre and Rambal, 1988; Cubera and Moreno, 2007; Pereira et al., 2007), although during the summer, the stomata will close due to the high *VPD* and *ETP* values associated with very high temperatures and lower values of *SWC*. In addition, the mixing ratio decreased also during summer reaching lower values than in the ENF\_FI and DBF\_DK.



After summer, in the DBF\_DK and the ENF\_FI, the net radiation became lower and sensible heat decreased, leading to a decrease in temperature which entailed a reduction in the *GPP* and the evapotranspiration process, causing a sharp decrease also in latent heat. This period coincided with the browning and fall of leaves in Denmark and the end of the growing season in the coniferous forest of Finland. Then, in winter, sensible heat became even more negative, this fact is related to the higher atmospheric stability conditions reached with cold air trapped close to the surface, and latent heat was nearly zero due to null photosynthetic activity with null *GPP* values. In addition, in these northern ecosystems, the mixing ratio was much lower than in mid-latitude areas due to their low temperatures that cause air masses with much less contribution to evaporation owing to their high relative humidity. On the other hand, in SAV\_SP, during autumn, owing to rainfall events and the recovery of the soil water content, *GPP*

increased (Cicuéndez et al., 2015; El-Madany et al., 2018), and subsequently, the latent heat also showed this recovery, leading to reduce sensible heat, which also decreased owing to less solar heating of the surface. Therefore, in this ecosystem, the latent heat was closely linked to an increase in soil moisture since water is limiting for plants. Thus, in other studies, large soil moisture–latent heat correlations were usually found (Román-Cascón et al., 2021), indicating a stronger land–atmosphere coupling in these transitional zones (Ferguson et al., 2012).

The northern ecosystems are more efficient in terms of water and energy (lower Bowen ratio, higher water use efficiency ( $GPP/LE$ ), higher  $GPP/(H+LE)$ , and higher  $GPP/H$ ) (Figure 6). Although northern ecosystems showed similar ratios, it has to be remarked that during the growing season, the coniferous forest of Finland showed higher net radiation due to a lower albedo and a higher Bowen ratio than the deciduous forest due to higher

surface resistance of leaves of coniferous trees and that evaporation is closer to equilibrium at deciduous forests (Wilson et al., 2002; Román-Cascón et al., 2020). Hoek van Dijke et al. (2020) suggested that using the *LAI* to model or extrapolate surface fluxes of water and energy is possible in savannas, grasslands, and Evergreen Broadleaf Forests, but it is limited in DBF and ENF. It seems that using *GPP* instead of *LAI* could improve the results in these ecosystems. This study is a first step to compare the influence of vegetation on the energy fluxes partitioning in clearly contrasted ecosystems located in different climates. However, this study should be extended to similar ecosystems in other areas to obtain more robust conclusions. In addition, it will be interesting to study other types of ecosystems to compare and obtain new results. In terms of energy partitioning, the role of soil heat flux (*G*) in these ecosystems should also be studied, although it is quite smaller in proportion to latent and sensible fluxes, it may have an effect during the growing season (results not shown).

## 5. Conclusion

Differences in gross primary production, and latent and sensible heat fluxes at the three different ecosystems analyzed were due to different plant functional types and their interaction with the meteorological variables in each climate. Temperature and solar radiation were the main limiting factors in the northern ecosystems, while water availability was a determinant of growth in the Mediterranean ecosystem. Northern ecosystems were more efficient in terms of water and energy during the growing period (lower Bowen ratio, higher water use efficiency, and higher *GPP/H*) than Mediterranean ecosystems.

The *GPP* of the three ecosystems showed a dynamic relationship with the energy fluxes partitioning and water fluxes dynamics confirming that *GPP* could be used also as a proxy of the influence of vegetation on surface fluxes. In the northern sites, most of the available energy was converted into sensible heat at the beginning of the spring, while in the Mediterranean ecosystems, this component was dominant after the growing period. Meanwhile, in the conifer forests of Finland, the latent heat flux was coupled to *GPP* during all growing seasons due to the factor of temperature, while in Denmark it began to be strongly coupled when leaf emergence occurred. In Spain, latent heat flux was coupled to *GPP* during all growing seasons but due to water availability.

The different vegetation types analyzed provide feedback to the atmosphere, influencing the energy partitioning (latent heat and sensible heat fluxes) in a different way. Granger causality tests indicated that *GPP* was causing latent heat more strongly in the northern ecosystems than in the savanna ecosystem, due to higher *LAI*s with higher soil water content, which entailed more transpiration rates, especially in the Deciduous Forest. However, sensible heat flux statically causes *GPP* more than backward, especially in Finland and Spain due to the fact that temperature regulated the carbon fluxes but with an inverse pattern in both ecosystems. In Denmark, a forest with a more closed canopy, the expansion of *LAI* led to more shade on the ground which entailed a reduction of changes in surface temperature and subsequently affected *H* among other factors. It is well known that it is essential to

include vegetation dynamics to improve atmospheric models. This study is the first step to quantify the intensity and the dominant direction of feedback processes between the ecosystem and the atmospheric boundary layer in different forests, especially focusing on the turbulent heat fluxes (*LE* and *H*).

## Data availability statement

The original contributions presented in the study are included in the article/supplementary material, further inquiries can be directed to the corresponding authors.

## Author contributions

VC, JL, and AP-O: conceptualization. VC, JL, VS-G, LR, and CS: methodology. JL and AP-O: software and resources. VC, JL, and VS-G: formal analysis. VC, JL, VS-G, CR-C, CY, and AP-O: investigation. VC, CR-C, CY, and AP-O: writing—original draft preparation. VC, CR-C, LR, CS, CY, and AP-O: writing—review and editing. AP-O: supervision. CY and AP-O: funding acquisition. All authors have read and agreed to the published version of the manuscript.

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## Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.



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