



## OPEN ACCESS

## EDITED BY

Guoxiang Li,  
Nanjing Normal University, China

## REVIEWED BY

Kleoniki Natalia Petrou,  
Imperial College London, United Kingdom  
Maria Urbaniec,  
Kraków University of Economics, Poland

## \*CORRESPONDENCE

Zuzana Hajduová,  
✉ zuzana.hajduova@euba.sk

RECEIVED 19 July 2024

ACCEPTED 27 August 2024

PUBLISHED 04 September 2024

## CITATION

Lacko R, Hajduová Z and Dula R (2024)  
Development of environmental performance  
and circular economy in the European Union  
countries: the case of “post-2004” members.  
*Front. Environ. Sci.* 12:1467370.  
doi: 10.3389/fenvs.2024.1467370

## COPYRIGHT

© 2024 Lacko, Hajduová and Dula. This is an open-access article distributed under the terms of the [Creative Commons Attribution License \(CC BY\)](https://creativecommons.org/licenses/by/4.0/). The use, distribution or reproduction in other forums is permitted, provided the original author(s) and the copyright owner(s) are credited and that the original publication in this journal is cited, in accordance with accepted academic practice. No use, distribution or reproduction is permitted which does not comply with these terms.

# Development of environmental performance and circular economy in the European Union countries: the case of “post-2004” members

Roman Lacko<sup>1</sup>, Zuzana Hajduová<sup>2\*</sup> and Radúz Dula<sup>1</sup>

<sup>1</sup>Faculty of Commerce, University of Economics in Bratislava, Petržalka, Slovakia, <sup>2</sup>Faculty of Business Management, University of Economics in Bratislava, Petržalka, Slovakia

The study aims to compare the environmental efficiency within the group of European Union countries that joined the European Union in 2004 and later, and to identify the common circular economy determinants of efficiency. For this purpose, we performed Data Envelopment analysis and correlation analysis. We applied both constant and variable returns to scale models. Findings reveal that countries with a significant focus on services, particularly tourism, are more environmentally efficient. However, most countries are still heavily industry-oriented, with Bulgaria, Romania, and Croatia being the least efficient. The study also highlights the need for significant efforts to reduce greenhouse gas emissions. Interestingly, a positive correlation was found between resource productivity and the circular materials used, suggesting the importance of circular economy tools in improving the environment. Despite having a higher material and consumption footprint, these countries still produce a relatively high product and relatively low CO<sub>2</sub> emissions. Based on these analyses, we found regional differences and the need for setting cluster policies within the circular economy at the European Union level.

## KEYWORDS

environmental performance, European Union, circular economy, late-joiners, data envelopment analysis

## 1 Introduction

Creating policies and strategies by the European Union member states contributes to the dynamics of integration aimed at sustainable economic growth. The circular economy represents a change in how society interacts with nature, aiming to prevent the depletion of all resources, close material and energy loops, create sustainable business development, and improve the efficiency of circular economies as a pivotal factor in enhancing global GDP growth through circular means (European Commission, 2018). To achieve this, ecological innovations must be developed to protect natural capital, which can only be achieved by promoting renewable resources and expanding the use of materials through reuse and recycling. Companies that have already implemented sustainable environmental practices in conjunction with the circular economy have seen positive returns in terms of material cost savings, opening up new markets, increased turnover, and other economic benefits. The question of the circular economy of economies, precisely their efficiency, is a current topic evidenced by many literary sources. Using circular economy tools is often associated with

improving environmental efficiency (Ghisellini et al., 2016). However, these relationships may not generally apply to all entities of the global economy, which, in many cases, is caused by different social, economic, and political developments (Ferronato et al., 2019). These differences at the national and regional levels are only examined at a limited level in the existing literature. They often only deal with larger regional groupings, such as the European Union, OECD, or specific countries and their regions. Therefore, we want to contribute to filling the gap in research with this study. We will examine countries that lag significantly (even in convergence) behind the environmental performance of Western European countries in the most extensive research but are part of the same political grouping—the European Union. In the Literature Review section, we will examine the current perception of the issue at the global and local levels and set the goal of this study. The Research Methodology section will describe the methods used to thoroughly analyze the existing literature to help fulfill this study's objective. In the results section, we will focus on the specific position of selected countries in the field of environmental efficiency concerning the circular economy and possible reasons for ranking these countries. In the last part of the study, we will present implications, related proposals, limitations, and possibilities of future research direction based on the results of this study.

## 2 Literature review

Environmental policy can be defined as any measure by a government, corporation, public, or private organization relating to the effects of human activity on the environment, particularly those measures aimed at preventing or reducing the harmful effects of human activities on ecosystems. We characterize it as a worldview to understand the need to respect the environment and lead a life that does not destroy the possibility of its subsequent existence and, thus, human existence on Earth (Tobin, 2017). At the end of the 20th century, the European Union transition successfully transitioned to a common currency, the political division of Europe was eliminated, and the countries of Central and Eastern Europe applied for membership (Bertossi, 2011). The negatives of this period were the inability to eradicate the enormous unemployment of 15 million people out of work in figures, the internal market, which had weaknesses in trade in services between the Member States, and the low volume of research expenditure and the associated low level of innovation (Pollex and Lenschow, 2020). In 2000, the European Union drew the consequences of this situation and adopted the Lisbon Strategy, which was to have three pillars: economic, social, and ecological. At the Gothenburg Summit in 2001, an objective on quality of life and sustainable development was added under the green pillar, covering the management of natural resources, greenhouse gas emissions, and ultimately sustainable transport (Graziano and Vink, 2012).

The EU extends its policy to the Member States by harmonizing laws that preceded its accession to the EU. The change is due to the focus on the adaptation of the national administration, the organizational logic of national policy, policy-making, or, more generally, the changes that have taken place in the national political systems linked to European integration (Goetz and Hix, 2001). Thus, implementation depends on a good match between the

requirements of EU policies and existing national policies (Börzel and Risse, 2003). Good agreement is the extent to which compliance with the EU depends on consistency between the directive in question and the two national institutions, namely, the organization of interest groups and the legacy of national policy (Falkner, 2005). On the contrary, implementation will be weak if significant policy changes and reorganization of interest groups are needed. Parliaments that act as guardians of the *status quo* are critical and essential in this relationship (Mastenbroek and Kaeding, 2006).

Haverland questioned the importance of good correspondence between European provisions and national rules and practices in explaining the degree of national adaptation to European requirements. It bases its argument on analyses of the Packaging and Packaging Waste Directive in Germany, the Netherlands, and the United Kingdom (Haverland, 2000). We can conceptualize adaptation processes in response to Europeanization in two ways, resulting in a different emphasis on supporting factors. They point to two variants of the new institutionalism. It follows from the logic of rational institutionalism that Europeanism leads to domestic changes through the different empowerment of actors resulting from the redistribution of resources at the domestic level (Borzell, 2000). Implementation can be defined as the process of transposing and applying EU policies at the domestic level through national authorities (Melidis and Russel, 2020). Transposition means introducing supranational rules into national legal orders by adopting new laws, ratifications, or administrative measures (Domaradzki, 2019).

Environmental performance is often measured using methods that determine the environmental efficiency. Many research studies address this issue at the supranational level (Giannakitsidou et al., 2020; Matsumoto et al., 2020; Ríos and Picazo-Tadeo, 2021) as well as at the national or local level (Wolsink, 2010; Di Maria et al., 2020) and also combined (Sanz-Díaz et al., 2017). A large number of studies also deal with the European Union, and many of them indicate significant differences in the environmental efficiency of Central and Eastern European countries, which they believe is also caused by an insufficient level of efficiency of technological processes and related investments in innovation (Vlontzos et al., 2014). When looking at the research methodology, it is essential to note that the benchmark for less efficient countries is highly developed countries with a relatively high environmental awareness. This fact is, to some extent, limiting the perception of differences directly in the group of less developed countries. Suppose we want to achieve convergence of these countries. In that case, it is essential to realize the common characteristics of these countries but also differences that arise from geographical, political, and socio-economic realities (Adam and Tsarsitalidou, 2019), as was done, for example, for countries that joined the EU among the first authors of other studies (Kwon et al., 2017). Concerning environmental performance, in many cases, the question of the effects of circular economy tools also arises, where resource productivity, recycling rate, and use of renewable resources seem to be vital in improving the state of the environment even under EU conditions (Busu, 2019; Halkos and Petrou, 2019). Despite the efforts of EU authorities, there is still significant room for improvement in these policies and processes (Calisto Friant et al., 2021). In this contribution, we will examine common

TABLE 1 Description of variables.

|        | Variable name                           | Variable symbol | Units                  | Source   |
|--------|---|-----------------|------------------------|----------|
| Input  | Total Environmental Taxes               | Taxes           | mil. €                 | Eurostat |
|        | Total Employment                        | Empl            | Thousand persons       | Eurostat |
| Output | CO <sub>2</sub> emissions (undesirable) | CO <sub>2</sub> | Thousand tonnes        | Eurostat |
|        | Gross domestic product                  | GDP             | mil. € (market prices) | Eurostat |

Source: Own processing.

patterns and the possible impact of the circular economy on the environmental performance of selected countries.

The objective of this study is to evaluate the development of the environmental performance of EU countries that joined the EU structures after 2004 based on identified research gaps. We also want to verify the significance of the circular economy and its impact on changes in the environmental efficiency of selected countries. The research question of this study is, then, what are the differences in environmental efficiency among “late-joiners” when the efficiency frontier shifts lower? Are they increasing their environmental efficiency, or do they stagnate?

### 3 Research methodology

In this section, we will focus on the methods used to achieve the objective of this study. In some of the studies mentioned in the previous section, the Data Envelopment Analysis (DEA) method was used to evaluate environmental performance or efficiency. It is one of the most commonly used methods for measuring environmental performance, including for the EU, as confirmed by other current studies (Angelakoglou and Gaidajis, 2015; Giannakitsidou et al., 2020; Matsumoto et al., 2020; Musa et al., 2021; Ríos and Picazo-Tadeo, 2021; Dechezleprêtre et al., 2023).

Technical efficiency is commonly measured using linear programming models presented and modified by (Charnes et al., 1985; Färe et al., 1994; Cooper et al., 2007). In this study, we will use CCR (CRS) Equation 1 and BCC (VRS) Equation 2, input-oriented slack-based DEA models. Moreover, we will compute the excess of Emissions as undesirable output.

$$\begin{aligned}
 & \min_{\theta_B, \lambda} \theta_B \\
 & \text{s. t. } \theta_B X_o - X \lambda \geq 0 \\
 & Y \lambda \geq y_o \\
 & \lambda \geq 0.
 \end{aligned} \tag{1}$$

$$\begin{aligned}
 & \min_{\theta_B, \lambda} \theta_B \\
 & \text{s. t. } \theta_B X_o - X \lambda \geq 0 \\
 & Y \lambda \geq y_o \\
 & e \lambda = 1 \\
 & \lambda \geq 0.
 \end{aligned} \tag{2}$$

Table 1 describes the input and output variables used in your performance modeling.

We have collected data for our study from the Eurostat database (Eurostat, 2024). There are many studies in which the environmental DEA model was implemented. Inputs are commonly connected to labor, capital, and land (Alsaleh et al.,

2017) as the three main production factors. As capital, we decided to use environmental taxes, which we want to be as low as possible and should be used to motivate lower emissions and as capital input for environmental efficiency improvement. Total employment is the input related to labor (Woo et al., 2015; Toma et al., 2017). Because of the data unavailability, we decided to omit the factor land—using it would lead to data inconsistency. The most used inputs are the number of companies in selected industries or services, arable land, and others. Outputs are mostly connected to emissions factors, which are undesirable outputs (Zofio and Prieto, 2001; Wang et al., 2013; Halkos and Petrou, 2019), and Gross domestic product—which is the desirable output of production function (Madaleno et al., 2016; Cheng et al., 2018; Iram et al., 2020).

Because the research focused only on “late joiners” is missing, we decided to measure the environmental efficiency of countries that joined the EU in 2004 or later. These countries often have higher rates of industrial/services production ratio and could be less environmentally efficient. We will assess the environmental efficiency during the period from 2014 to 2022. This could also help us evaluate the effects of the COVID-19 pandemic. We collected data for a total of 13 countries, so we created a panel of 117 observations (Decision-making units—DMU). For the efficiency measurement DEA Window approach will be used (Wang et al., 2013; Farantos and Koutsoukis, 2022).

Table 2 presents descriptive statistics of data collected from the Eurostat database by individual country.

Many indicators of measuring circular economy were proposed (De Pascale et al., 2020), but only a few are accessible from online databases for the countries we have selected. We will use Pearson’s correlations to assess the correlation between environmental efficiency and circular economy indicators. The individual indicators of the circular economy were also collected from the Eurostat (2024) database. We will use these indicators:

- Circular material use rate—percentage
- Resource productivity—PPS per kilogram
- Material footprint—tonnes *per capita*
- Consumption footprint—single weighted score per inhabitant.

Using data from freely available databases provides a reliable proxy for conducting research in this study. Moreover, data availability makes it possible to continue and replicate similar research for the broader scientific community. Our sample of countries is representative as it includes all countries that joined the EU after 2004, and this is not the only common feature. They are also geographically, historically, and economically close, albeit with specific existing differences.

TABLE 2 Descriptive statistics of inputs and outputs of the DEA models.

| Country   | Var.            | Mean         | Std Dev     | Variance      | Minimum     | Maximum     |
|-----------|-----------------|--------------|-------------|---------------|-------------|-------------|
| Bulgaria  | CO <sub>2</sub> | -42942.91    | 3,601.34    | 12,969,682.85 | -47102.04   | -35190.45   |
|           | Taxes           | 1,859.46     | 876.7357842 | 768,665.64    | 1,220.69    | 4,095.16    |
|           | Emp             | 3,475.68     | 39.3497569  | 1,548.40      | 3,434.17    | 3,533.58    |
|           | GDP             | 58,475.06    | 13,504.83   | 182,380,494   | 43,024.70   | 85,800.70   |
| Croatia   | CO <sub>2</sub> | -12745.43    | 605.5019825 | 366,632.65    | -13684.70   | -11634.19   |
|           | Taxes           | 2,125.56     | 204.0721001 | 41,645.42     | 1,746.44    | 2,366.11    |
|           | Empl            | 1,648.49     | 50.5764911  | 2,557.98      | 1,574.99    | 1,724.44    |
|           | GDP             | 52,514.18    | 7,442.69    | 55,393,658.83 | 44,084.20   | 67,989.50   |
| Cyprus    | CO <sub>2</sub> | -5,549.98    | 205.4617165 | 42,214.52     | -5,868.11   | -5,246.70   |
|           | Taxes           | 569.1777778  | 34.2762592  | 1,174.86      | 519.5000000 | 613.1000000 |
|           | Emp             | 418.2722222  | 37.1123705  | 1,377.33      | 364.9400000 | 465.6700000 |
|           | GDP             | 21,599.71    | 3,365.69    | 11,327,873.96 | 17,482.80   | 27,777.00   |
| Czechia   | CO <sub>2</sub> | -81305.24    | 5,127.50    | 26,291,265.70 | -86095.35   | -71644.95   |
|           | Taxes           | 3,974.63     | 405.5883208 | 164,501.89    | 3,346.96    | 4,594.86    |
|           | Empl            | 5,284.64     | 99.4407725  | 9,888.47      | 5,094.54    | 5,385.11    |
|           | GDP             | 207,317.28   | 37,203.83   | 1,384,125,281 | 157,821.30  | 276,265.70  |
| Estonia   | CO <sub>2</sub> | -14801.47    | 4,359.94    | 19,009,110.51 | -19999.35   | -8,471.43   |
|           | Taxes           | 692.4177778  | 114.4824374 | 13,106.23     | 533.0700000 | 889.5500000 |
|           | Emp             | 657.0966667  | 16.3924243  | 268.7115750   | 627.7000000 | 684.0600000 |
|           | GDP             | 26,083.82    | 5,241.10    | 27,469,097.39 | 20,048.20   | 36,011.10   |
| Hungary   | CO <sub>2</sub> | -37274.85    | 1773.78     | 3,146,297.14  | -39850.66   | -34965.58   |
|           | Taxes           | 3,028.24     | 239.2414140 | 57,236.45     | 2,557.84    | 3,366.50    |
|           | Empl            | 4,530.28     | 157.2281836 | 24,720.70     | 4,222.60    | 4,695.55    |
|           | GDP             | 133,932.78   | 20,471.56   | 419,084,656   | 106,263.80  | 168,549.50  |
| Latvia    | CO <sub>2</sub> | -6,970.80    | 590.5730800 | 348,776.56    | -7,796.19   | -6,130.20   |
|           | Taxes           | 898.0866667  | 54.7108410  | 2,993.28      | 790.2500000 | 982.7300000 |
|           | Emp             | 898.5944444  | 12.8294789  | 164.5955278   | 876.0100000 | 916.9500000 |
|           | GDP             | 29,124.97    | 4,694.68    | 22,040,020.05 | 23,625.80   | 38,386.20   |
| Lithuania | CO <sub>2</sub> | -13,643.16   | 1,324.26    | 1,753,653.95  | -16,153.50  | -12,026.75  |
|           | Taxes           | 859.2600000  | 146.1531994 | 21,360.76     | 633.8800000 | 1,038.90    |
|           | Empl            | 1,366.29     | 31.5239675  | 993.7605278   | 1,319.30    | 1,431.91    |
|           | GDP             | 47,039.49    | 10,072.31   | 101,451,392   | 36,581.30   | 67,436.50   |
| Malta     | CO <sub>2</sub> | -1,585.80    | 310.8777242 | 96,644.96     | -2,364.76   | -1,316.12   |
|           | Taxes           | 290.5255556  | 30.9856019  | 960.1075278   | 240.6500000 | 347.8200000 |
|           | Emp             | 236.0655556  | 31.7967310  | 1,011.03      | 192.1200000 | 282.7600000 |
|           | GDP             | 12,741.48    | 2,749.97    | 7,562,350.41  | 8,751.10    | 17,432.30   |
| Poland    | CO <sub>2</sub> | -2,86,269.68 | 12,534.83   | 157,121,961   | -3,01982.35 | -270055.39  |
|           | Taxes           | 13,477.17    | 2,515.22    | 6,326,334.24  | 10,562.10   | 18,321.85   |
|           | Emp             | 16,456.00    | 433.0473415 | 187,530.00    | 15,862.00   | 17,369.00   |
|           | GDP             | 501,709.78   | 80,596.79   | 6,495,842,116 | 406,412.50  | 654,594.40  |
| Romania   | CO <sub>2</sub> | -62241.81    | 4,014.71    | 16,117,894.90 | -66239.88   | -54472.46   |
|           | Taxes           | 4,577.71     | 1,278.17    | 1,633,718.10  | 3,587.37    | 7,674.61    |
|           | Empl            | 8,731.06     | 85.3175417  | 7,279.08      | 8,579.20    | 8,828.50    |
|           | GDP             | 204,580.61   | 43,208.41   | 1,866,966,269 | 150,522.40  | 284,173.60  |
| Slovakia  | CO <sub>2</sub> | -28148.23    | 1975.47     | 3,902,487.69  | -30294.99   | -24972.69   |
|           | Taxes           | 2,230.08     | 241.5082336 | 58,326.23     | 1932.31     | 2,707.77    |
|           | Emp             | 2,502.18     | 69.7195342  | 4,860.81      | 2,363.05    | 2,583.64    |
|           | GDP             | 90,007.22    | 10,671.31   | 113,876,851   | 76,354.50   | 109,645.20  |
| Slovenia  | CO <sub>2</sub> | -10231.49    | 453.1709080 | 205,363.87    | -10866.94   | -9,517.17   |
|           | Taxes           | 1,585.81     | 97.1987011  | 9,447.59      | 1,452.74    | 1795.79     |
|           | Empl            | 1,014.15     | 54.3687509  | 2,955.96      | 935.4500000 | 1,090.67    |
|           | GDP             | 45,640.16    | 6,410.39    | 41,093,095.84 | 37,634.30   | 57,037.70   |

Source: own processing according to Eurostat (2024) data.

## 4 Results and discussion

This section will analyze the findings from the environmental efficiency modeling performed using the DEA method. Table 3 displays the results of the CRS and VRS environmental efficiency

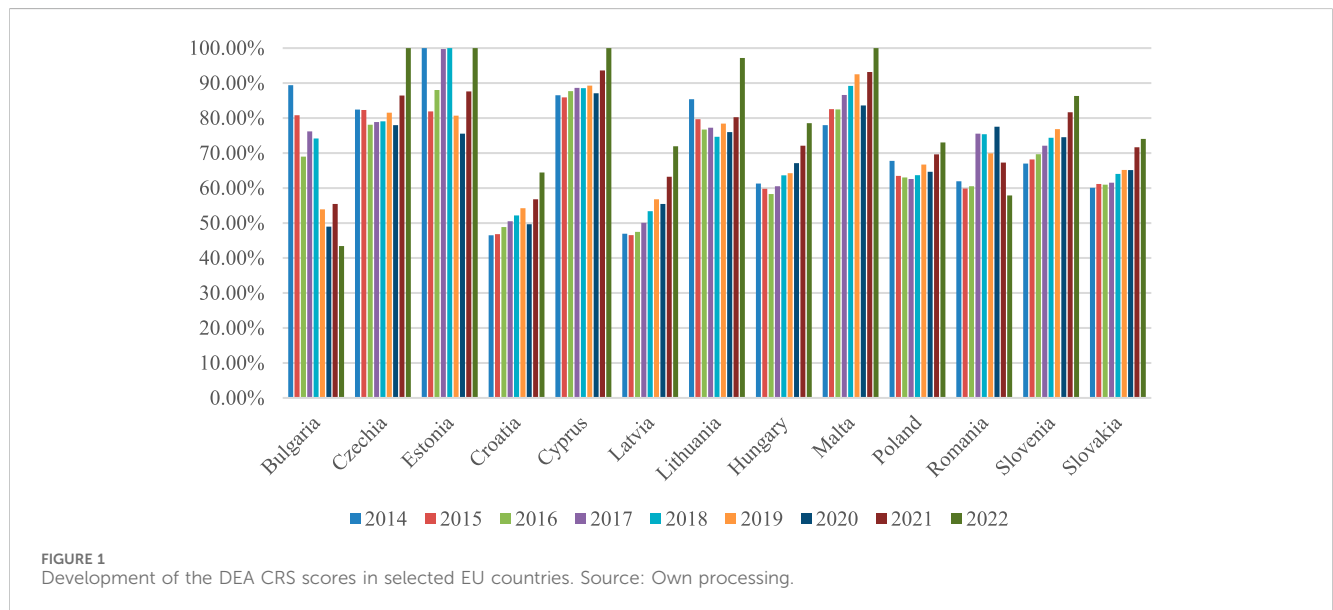
models. Given that a detailed table for each country in every year of the study would be too extensive, we provide only the essential descriptive characteristics of the efficiencies obtained.

Estonia, Cyprus, and Malta achieved the best average results in the CRS models. In the VRS models, the results are slightly different,

TABLE 3 Descriptive statistics of the DEA CRS and VRS efficiency results.

| Country   | CRS  |       |      |      | VRS  |       |      |      |
|-----------|------|-------|------|------|------|-------|------|------|
|           | Mean | StDev | Min  | Max  | Mean | StDev | Min  | Max  |
| Bulgaria  | 0.66 | 0.15  | 0.43 | 0.89 | 0.70 | 0.17  | 0.47 | 1.00 |
| Czechia   | 0.83 | 0.07  | 0.78 | 1.00 | 0.89 | 0.07  | 0.79 | 1.00 |
| Estonia   | 0.90 | 0.09  | 0.76 | 1.00 | 0.91 | 0.09  | 0.77 | 1.00 |
| Croatia   | 0.52 | 0.05  | 0.47 | 0.64 | 0.58 | 0.07  | 0.50 | 0.73 |
| Cyprus    | 0.90 | 0.04  | 0.86 | 1.00 | 0.93 | 0.03  | 0.89 | 1.00 |
| Latvia    | 0.55 | 0.08  | 0.47 | 0.72 | 0.57 | 0.09  | 0.47 | 0.77 |
| Lithuania | 0.81 | 0.07  | 0.75 | 0.97 | 0.87 | 0.07  | 0.80 | 1.00 |
| Hungary   | 0.65 | 0.06  | 0.58 | 0.79 | 0.66 | 0.06  | 0.59 | 0.79 |
| Malta     | 0.88 | 0.06  | 0.78 | 1.00 | 0.98 | 0.02  | 0.92 | 1.00 |
| Poland    | 0.66 | 0.03  | 0.63 | 0.73 | 0.99 | 0.02  | 0.95 | 1.00 |
| Romania   | 0.67 | 0.07  | 0.58 | 0.78 | 0.68 | 0.06  | 0.60 | 0.78 |
| Slovenia  | 0.75 | 0.06  | 0.67 | 0.86 | 0.80 | 0.08  | 0.71 | 0.96 |
| Slovakia  | 0.65 | 0.05  | 0.60 | 0.74 | 0.68 | 0.07  | 0.61 | 0.82 |

Source: own processing.



with Poland and Czechia achieving the highest efficiencies. However, countries that were highly efficient in the CRS models also performed well in the VRS models. Croatia, Latvia, Slovakia, and Hungary are the least efficient in the CRS and VRS models.

Overall, the CRS and VRS models do not show significant differences, indicating that the discrepancies between these two models are minimal. It is also important to note that countries with lower industrial intensity and a greater focus on services achieve better results. This service orientation helps them produce lower emissions at a comparable level of GDP. In other words, countries that focus more on the service sector can reduce

emissions more efficiently while maintaining economic growth. Focusing on the service sector helps these countries maintain lower emissions while achieving economic growth. In Figure 1, a comparison of environmental efficiency development is presented.

During the observed period, we recorded improvements in the values of technical environmental efficiency in most countries. When comparing the beginning and end of the observed period, we identified several interesting facts. The most significant decrease was recorded between 2014 and 2022 in the case of Bulgaria. Specifically, the efficiency decreased by 51.43%. A slight reduction in efficiency also occurred in Romania – 6.54%. In all

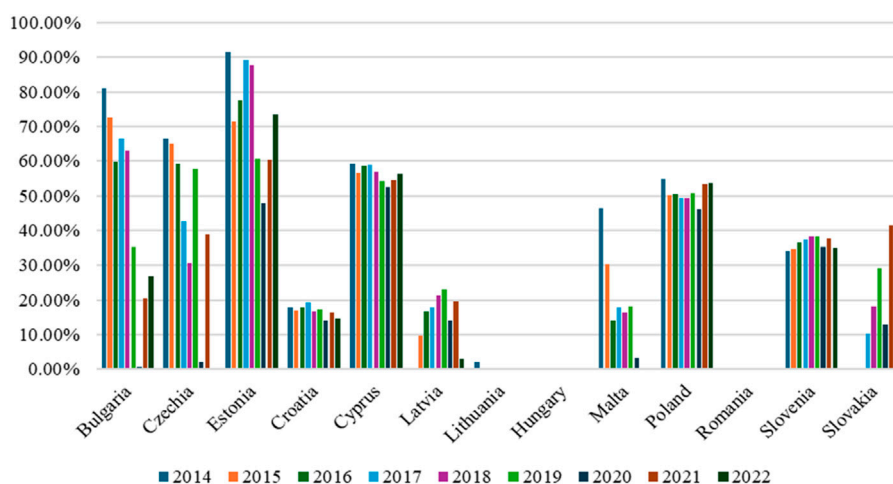


FIGURE 2

The excess CO<sub>2</sub> emissions in selected EU countries (DEA CRS assumption). Source: Own processing.

other countries, there was an increase in efficiency, with Latvia (53.2%), Croatia (38.52%), and Slovenia (28.86%) being among the strongest growing countries. The most efficient countries during the observed years are Malta, Cyprus, Estonia, and the Czech Republic. Malta and Cyprus are countries that are significantly oriented towards services, specifically tourism. We also examined the direct impact of the COVID-19 pandemic on the values of environmental efficiency. In most countries, there was a slight decrease of up to 10%. Romania was the only country with a minimal increase in efficiency values (10.9%). Figure 2 shows the slack values of excessive CO<sub>2</sub> production, meaning the percentage needed to reduce CO<sub>2</sub> emissions to achieve efficiency within the observed group of countries.

Within the scope of our results, we recorded a significant need to reduce CO<sub>2</sub> emissions in several countries. The most pronounced surpluses of CO<sub>2</sub> emissions were observed in Estonia, Cyprus, Poland, and other countries, highlighting the urgent need for targeted emission reduction strategies in these regions. These surpluses indicate areas where current environmental policies and practices may not meet the desired efficiency and sustainability targets, necessitating more robust and effective interventions.

Despite these challenges, some countries have demonstrated significant improvement in reducing their CO<sub>2</sub> emission surpluses over the years. For instance, Bulgaria has shown a remarkable reduction, with emission surpluses decreasing from approximately 81% in 2014 to 27% in 2022. This substantial improvement reflects successful policy implementation and a shift towards more sustainable practices. Similarly, Malta has made impressive strides, reducing its CO<sub>2</sub> emission surplus from nearly 46.5% to 0% within the same period. These examples underscore the potential for significant progress when appropriate measures are adopted and effectively executed.

On the other hand, countries such as Lithuania, Hungary, and Romania have maintained low or no CO<sub>2</sub> emission surpluses, positioning them as leaders in environmental efficiency within the post-2004 EU member states. Their success can serve as a

benchmark and provide valuable lessons for other countries striving to improve their environmental performance.

The efficiency results for the Variable Returns to Scale (VRS) model, which accounts for scale efficiency and offers a more tailored assessment of each country's performance, are shown in the following Figure 3. This figure illustrates the comparative efficiency scores, highlighting the countries leading in environmental performance and those requiring further improvement. By visualizing these results, stakeholders can better understand the relative efficiency of each country and identify critical areas for policy intervention and resource allocation to promote a more sustainable and circular economy across the European Union.

Similar results as in the case of the Constant Returns to Scale (CRS) model were also recorded in the case of the Variable Returns to Scale (VRS) model. Some countries consistently achieved high-efficiency values under both models, demonstrating solid environmental performance and effective resource management. These countries include Poland, Malta, Cyprus, Estonia, and the Czech Republic. Their consistent high efficiency indicates robust environmental policies and practices that have enabled them to maintain low CO<sub>2</sub> emissions relative to their economic output.

In contrast, some other countries showed an increasing trend in efficiency, reflecting gradual improvements in their environmental performance. Notable among these are Bulgaria and, to some extent, Romania. In Bulgaria, significant efforts have led to substantial gains in efficiency, showcasing the impact of targeted environmental policies and initiatives. Although showing less dramatic improvements, Romania also indicates positive trends toward better environmental management.

Figure 4 illustrates the necessary percentage reductions in CO<sub>2</sub> for each country. This figure provides a clear visual representation of the CO<sub>2</sub> reduction targets required to achieve optimal efficiency. It highlights the percentage reductions each country needs to align with best practices and reduce its environmental impact. By identifying these targets, the figure is a crucial tool for



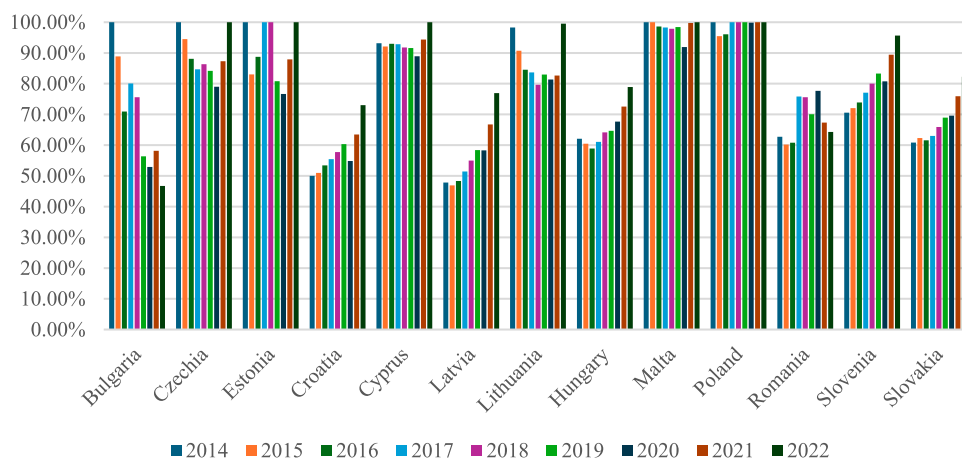


FIGURE 3 Development of the DEA VRS scores in selected EU countries. Source: Own processing

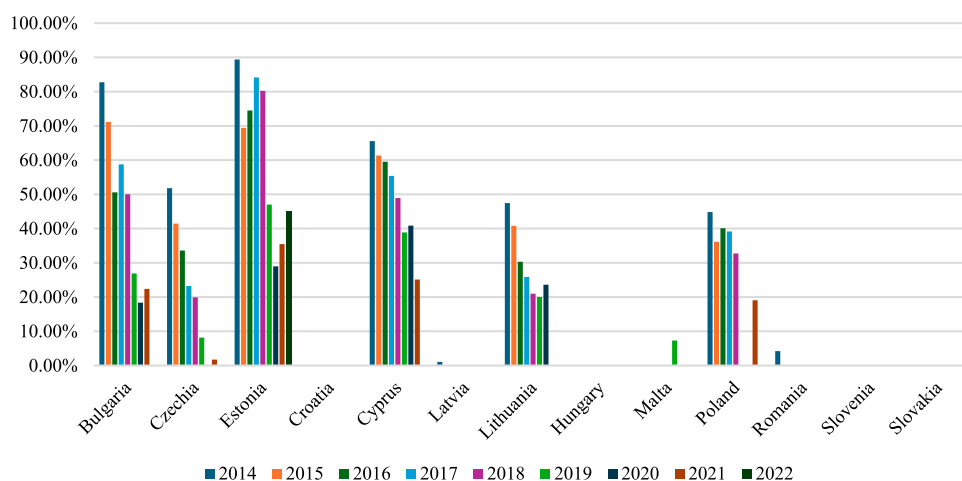


FIGURE 4 The excess CO<sub>2</sub> emissions in selected EU countries (DEA CRS assumption). Source: own processing

polymakers and ecological planners, enabling them to prioritize actions and allocate resources effectively to achieve sustainability goals.

The surpluses of CO<sub>2</sub> emissions in the Variable Returns to Scale (VRS) model are significantly lower than those in the Constant Returns to Scale (CRS) model. This indicates that the VRS model, which accounts for scale efficiency, reflects more realistic and achievable CO<sub>2</sub> reduction targets for the evaluated countries. The VRS model’s ability to provide a more nuanced understanding of efficiency variations due to scale is crucial in policy formulation and resource allocation for environmental management.

Estonia, Lithuania, Poland, and Bulgaria recorded more pronounced surpluses of CO<sub>2</sub> emissions. These countries have exhibited higher levels of inefficiency in CO<sub>2</sub> emissions relative to their peers, indicating a need for targeted interventions and policies to address these surpluses. Despite these high surpluses,

an essential finding of our study is that the CO<sub>2</sub> emission surpluses had a decreasing trend over the observed period. Despite room for improvement, these countries are on a positive trajectory toward better environmental performance and more efficient resource utilization.

Table 4 presents the correlation matrix between the individual input and output variables of the Data Envelopment Analysis (DEA) model and the circular economy indicators. This correlation matrix is pivotal as it provides insights into the relationships between resource inputs, economic outputs, and environmental outcomes. Understanding these correlations helps identify key leverage points for enhancing circular economy practices and improving overall environmental efficiency. The matrix elucidates how inputs (such as energy consumption and material usage) and outputs (such as GDP and waste generation) interact with circular economy indicators like recycling rates, waste reduction, and resource productivity. This

TABLE 4 Correlation matrix of individual variables, including circular economy indicators.

|                 | CO <sub>2</sub> | Taxes  | Empl  | GDP   | Circ   | Res   | Mat    | Cons   | CRS   | VRS  |
|-----------------|-----------------|--------|-------|-------|--------|-------|--------|--------|-------|------|
|                 |                 |        |       |       | MatUse | Prod  | Footpr | Footpr |       |      |
| CO <sub>2</sub> | 1.00            | 0.96   | 0.95  | 0.95  | 0.19   | -0.24 | 0.03   | -0.30  | -0.14 | 0.31 |
|                 |                 | <.0001 | 0.00  | 0.00  | 0.04   | 0.01  | 0.72   | 0.00   | 0.14  | 0.00 |
| Taxes           | 0.96            | 1.00   | 0.95  | 0.98  | 0.13   | -0.19 | 0.04   | -0.34  | -0.23 | 0.20 |
|                 | 0.00            |        | 0.00  | 0.00  | 0.15   | 0.04  | 0.64   | 0.00   | 0.01  | 0.03 |
| Empl            | 0.95            | 0.95   | 1.00  | 0.96  | 0.04   | -0.33 | 0.10   | -0.47  | -0.22 | 0.17 |
|                 | 0.00            | 0.00   |       | 0.00  | 0.70   | 0.00  | 0.28   | 0.00   | 0.02  | 0.07 |
| GDP             | 0.95            | 0.98   | 0.96  | 1.00  | 0.14   | -0.19 | 0.07   | -0.36  | -0.15 | 0.24 |
|                 | 0.00            | 0.00   | 0.00  |       | 0.14   | 0.04  | 0.46   | 0.00   | 0.11  | 0.01 |
| Circ<br>MatUse  | 0.19            | 0.13   | 0.04  | 0.14  | 1.00   | 0.27  | -0.07  | 0.29   | 0.31  | 0.37 |
|                 | 0.04            | 0.15   | 0.70  | 0.14  |        | 0.00  | 0.47   | 0.00   | 0.00  | 0.00 |
| Res<br>Prod     | -0.24           | -0.19  | -0.33 | -0.19 | 0.27   | 1.00  | -0.70  | 0.49   | 0.15  | 0.15 |
|                 | 0.01            | 0.04   | 0.00  | 0.04  | 0.00   |       | 0.00   | 0.00   | 0.12  | 0.12 |
| Mat<br>Footpr   | 0.03            | 0.04   | 0.10  | 0.07  | -0.07  | -0.70 | 1.00   | 0.03   | 0.29  | 0.16 |
|                 | 0.72            | 0.64   | 0.28  | 0.46  | 0.47   | 0.00  |        | 0.79   | 0.00  | 0.09 |
| Cons<br>Footpr  | -0.30           | -0.34  | -0.47 | -0.36 | 0.29   | 0.49  | 0.03   | 1.00   | 0.57  | 0.51 |
|                 | 0.00            | 0.00   | 0.00  | 0.00  | 0.00   | 0.00  | 0.79   |        | 0.00  | 0.00 |
| CRS             | -0.14           | -0.23  | -0.22 | -0.15 | 0.31   | 0.15  | 0.29   | 0.57   | 1.00  | 0.85 |
|                 | 0.14            | 0.01   | 0.02  | 0.11  | 0.00   | 0.12  | 0.00   | 0.00   |       | 0.00 |
| VRS             | 0.31            | 0.20   | 0.17  | 0.24  | 0.37   | 0.15  | 0.16   | 0.51   | 0.85  | 1.00 |
|                 | 0.00            | 0.03   | 0.07  | 0.01  | 0.00   | 0.12  | 0.09   | 0.00   | 0.00  |      |

Source: own processing.

comprehensive analysis aids policymakers and stakeholders in making informed decisions to foster sustainable development in the post-2004 EU member states.

The first row of the correlation pair shows the value of the Pearson correlation coefficient, while the second row displays its *p*-value at the 0.05 significance level. Generally, it can be stated that the use of circular economy materials and increasing resource productivity (statistically insignificant) correlates with higher values of environmental efficiency. An unexpected result is the positive correlation between Material (statistically insignificant in the case of the VRS model) and Consumption footprint. Given the research sample, we must point out that these correlation pairs do not indicate causality but only suggest possible situations that may play a specific role in shaping national and supranational EU policies. The unexpected existence of a positive correlation between environmental efficiency and the variables of Material Footprint and Consumption Footprint also indicates the specificity of these countries. One possible reason for this positive correlation could be that these countries are still dependent on heavy industry (automotive industry, engineering industry, raw materials processing), but increasing energy efficiency and reducing greenhouse gas emissions significantly contribute to improving

the situation in the field of the circular economy despite still high carbon footprint indicators.

## 5 Conclusion

In this section, we focus on evaluating the results of our study. Generally, we must highlight the importance and uniqueness of the chosen research object, where we compare countries with similar geopolitical and economic development. This is particularly important from the perspective of understanding benchmarking methods. In many cases (Vlontzos et al., 2014; Halkos and Petrou, 2019; Matsumoto et al., 2020), if these countries are also compared with Western EU countries, they are significantly undervalued, and the differences within this more minor (and still lagging) group of countries are not sufficiently highlighted. Using this research object, the benchmark becomes precisely the countries from this group.

The results of the DEA CRS and VRS models pointed out the differences within this group of countries. They indicate that countries that are significantly oriented towards services—in our case, tourism—are more environmentally efficient. However, it should be emphasized that tourism is also a significant producer



of greenhouse gas emissions, but in many cases, such as aviation emissions, are not included in the indicators. Most of the countries studied are still significantly oriented towards industry. This was also reflected in the results of DEA modeling. Countries such as Bulgaria, Romania, and Croatia are among the most inefficient. It should be noted that they joined the EU later than the other states in our sample. This reflects that even though they are significant countries in the field of tourism, their economy still has to be driven by industrial sectors and the extraction of mineral resources.

Another important finding is the level of emission surpluses that countries produce when converting equilibrium factors into their country's GDP. We found that most countries still need to significantly reduce environmental pollution from greenhouse gas emissions. This can be challenging to some extent, as industrial technologies pollute the environment the most and require significant expenditure on innovation. However, these are pretty extensively supported by the EU. The third important finding is the positive correlation between resource productivity and the use rate of circular materials, which indicates that using circular economy tools significantly improves the environment in these countries. A surprising finding is that the material and consumption footprint indicators positively correlate. This suggests that even though countries have a significantly higher material and consumption footprint, they still bring a relatively high product and relatively low CO<sub>2</sub> emissions at a given rate of inputs.

This study also has its limitations, which stem from the timing of the countries' entry into the EU, as the year of entry varies for some countries. This can delay the effects of the EU's convergence goals and policies on countries that joined the EU later. Another limitation is that some indicators in the field of the circular economy are inconsistent and, therefore, had to be excluded from the study's database, which, to some extent, impacted the examination of other specific processes and relationships.

Despite not being part of the main agenda when the EU was founded, environmental policy has become a key focus area. Based on the development of environmental policy, we can state that the EU has created extensive legislation for environmental protection. In the presented work, we focused on comparing the results in this area among the member countries that joined the EU at the same time in 2004. This study reflects significant differences in this development.

This area of research is widely discussed in the scientific sphere, which only stimulates the possibilities for future research directions. Since the environmental aspect is dominant in EU policies, it is

possible to expect broader databases that can expand the possibilities of examining and linking environmental efficiency with other economic, social, and political determinants.

## Data availability statement

The original contributions presented in the study are included in the article/supplementary material, further inquiries can be directed to the corresponding author.

## Author contributions

RL: Conceptualization, Formal Analysis, Supervision, Visualization, Writing—original draft, Writing—review and editing. ZH: Investigation, Methodology, Project administration, Resources, Writing—original draft. RD: Data curation, Funding acquisition, Software, Validation, Writing—review and editing.

## Funding

The author(s) declare that financial support was received for the research, authorship, and/or publication of this article. This study was supported by the national grant KEPA no. 020EU-4/2024 and VEGA no. 1/0109/24.

## Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

## Publisher's note

All claims expressed in this article are solely those of the authors and do not necessarily represent those of their affiliated organizations, or those of the publisher, the editors and the reviewers. Any product that may be evaluated in this article, or claim that may be made by its manufacturer, is not guaranteed or endorsed by the publisher.

## References

- Adam, A., and Tsarsitalidou, S. (2019). Environmental policy efficiency: measurement and determinants. *Econ. Gov.* 20 (1), 1–22. Available at: doi:10.1007/s10101-018-0219-y
- Alsaleh, M., Abdul-Rahim, A. S., and Mohd-Shahwahid, H. O. (2017). Determinants of technical efficiency in the bioenergy industry in the EU28 region. *Renew. Sustain. Energy Rev.* 78, 1331–1349. Available at: doi:10.1016/j.rser.2017.04.049
- Angelakoglou, K., and Gaidajis, G. (2015). A review of methods contributing to the assessment of the environmental sustainability of industrial systems. *J. Clean. Prod.* 108, 725–747. Available at: doi:10.1016/j.jclepro.2015.06.094
- Bertossi, C. (2011). National models of integration in Europe: a comparative and critical analysis. *Am. Behav. Sci.* 55 (12), 1561–1580. Available at: doi:10.1177/0002764211409560
- Börzel, T. A. (2000). Why there is no "southern problem". On environmental leaders and laggards in the European Union. *J. Eur. Public Policy* 7 (1), 141–162. Available at: doi:10.1080/135017600343313
- Börzel, T. A., and Risse, T. (2003). "Conceptualizing the domestic impact of Europe." in *The politics of europeanization*. Editors K. Featherstone and C. M. Radaelli 1st edn. (Oxford University Press/Oxford), 57–80. doi:10.1093/0199252092.003.0003
- Busu, M. (2019). Adopting circular economy at the European union level and its impact on economic growth. *Soc. Sci.* 8 (5), 159. Available at: doi:10.3390/socsci8050159
- Calisto Friant, M., Vermeulen, W. J. V., and Salomone, R. (2021). Analysing European Union circular economy policies: words versus actions. *Sustain. Prod. Consum.* 27, 337–353. Available at: doi:10.1016/j.spc.2020.11.001
- Charnes, A., Clark, C. T., Cooper, W. W., and Golany, B. (1985). A developmental study of data envelopment analysis in measuring the efficiency of maintenance units in the U.S. air forces. *Ann. Operations Res.* 2 (1), 95–112. Available at: doi:10.1007/BF01874734
- Cheng, S., Liu, W., and Lu, K. (2018). Economic growth effect and optimal carbon emissions under China's carbon emissions reduction policy: a time substitution DEA approach. *Sustainability* 10 (5), 1543. Available at: doi:10.3390/su10051543

- Cooper, W. W., Seiford, L. M., and Tone, K. (2007). *Data envelopment analysis: a comprehensive text with models, applications, references and DEA-solver software*. 2nd edn. New York, US: Springer US. Available at: <https://www.springer.com/la/book/9780387452814> (Accessed September 7, 2018).
- Dechezleprêtre, A., Nachtigall, D., and Venmans, F. (2023). The joint impact of the European Union emissions trading system on carbon emissions and economic performance. *J. Environ. Econ. Manag.* 118, 102758. Available at: doi:10.1016/j.jeem.2022.102758
- De Pascale, A., Arbolino, R., Szopik-Deczyńska, K., Limosani, M., and Ioppolo, G. (2020). A systematic review for measuring circular economy: the 61 indicators. *J. Clean. Prod.* 281, 124942. Available at: doi:10.1016/j.jclepro.2020.124942
- Di Maria, F., Sisani, F., Contini, S., Ghosh, S. K., and Mersky, R. L. (2020). Is the policy of the European Union in waste management sustainable? An assessment of the Italian context. *Waste Manag.* 103, 437–448. Available at: doi:10.1016/j.wasman.2020.01.005
- Domaradzki, S. (2019). Opportunistic legitimisation and de-Europeanisation as a reverse effect of Europeanisation. *Glob. Discourse* 9 (1), 221–244. Available at: doi:10.1332/204378919X15470487645475
- European Commission (2018). Communication from the commission to the European parliament, the council, the European economic and social committee and the committee of the regions. On a monitoring framework for the circular economy. Available at: <https://ec.europa.eu/environment/circular-economy/pdf/monitoring-framework.pdf> (Accessed October 20, 2019).
- Eurostat (2024). Database - Eurostat. Available at: <https://ec.europa.eu/eurostat/data/database> (Accessed March 10, 2024).
- Falkner, G. (2005). *Complying with Europe: EU harmonisation and soft law in the member states*. Cambridge, UK: Cambridge University Press.
- Farantos, G. I., and Koutsoukis, N.-S. (2022). Greek health system efficiency and productivity: a window DEA and Malmquist method measurement. *J. Future Sustain.* 2 (3), 113–124. Available at: doi:10.5267/j.jfs.2022.10.001
- Färe, R., Grosskopf, M., Norris, M., and Zhang, Z. (1994). “Productivity developments in Swedish hospitals: a malmquist output index approach,” in *Data envelopment analysis: theory, methodology, and applications*. Editor A. Charnes, (Dordrecht: Springer Netherlands), 253–272. Available at: doi:10.1007/978-94-011-0637-5\_13
- Ferronato, N., Rada, E. C., Gorrity Portillo, M. A., Cioca, L. I., Ragazzi, M., and Torretta, V. (2019). Introduction of the circular economy within developing regions: a comparative analysis of advantages and opportunities for waste valorization. *J. Environ. Manag.* 230, 366–378. Available at: doi:10.1016/j.jenvman.2018.09.095
- Ghisellini, P., Cialani, C., and Ulgiati, S. (2016). A review on circular economy: the expected transition to a balanced interplay of environmental and economic systems. *J. Clean. Prod.* 114, 11–32. Available at: doi:10.1016/j.jclepro.2015.09.007
- Giannakitsidou, O., Giannikos, I., and Chondrou, A. (2020). Ranking European countries on the basis of their environmental and circular economy performance: a DEA application in MSW. *Waste Manag.* 109, 181–191. Available at: doi:10.1016/j.wasman.2020.04.055
- Goetz K. H. and Hix S. (2001). *Europeanised politics? European integration and national political systems*. London, Portland, OR: Frank Cass
- Graziano, P. R., and Vink, M. P. (2012). “2. Europeanization: concept, theory, and methods,” in *The member states of the European union*. Editors P. R. Graziano, and M. P. Vink (Oxford University Press). Available at: doi:10.1093/hepl/9780199544837.003.0002
- Halkos, G., and Petrou, K. N. (2019). Assessing 28 EU member states’ environmental efficiency in national waste generation with DEA. *J. Clean. Prod.* 208, 509–521. Available at: doi:10.1016/j.jclepro.2018.10.145
- Haverland, M. (2000). National adaptation to European integration: the importance of institutional veto points. *J. Public Policy* 20 (1), 83–103. Available at: doi:10.1017/S0143814X00000763
- Iram, R., Zhang, J., Erdogan, S., Abbas, Q., and Mohsin, M. (2020). Economics of energy and environmental efficiency: evidence from OECD countries. *Environ. Sci. Pollut. Res.* 27 (4), 3858–3870. Available at: doi:10.1007/s11356-019-07020-x
- Kwon, D. S., Cho, J. H., and Sohn, S. Y. (2017). Comparison of technology efficiency for CO<sub>2</sub> emissions reduction among European countries based on DEA with decomposed factors. *J. Clean. Prod.* 151, 109–120. Available at: doi:10.1016/j.jclepro.2017.03.065
- Madaleno, M., Moutinho, V., and Robaina, M. (2016). Economic and environmental assessment: EU cross-country efficiency ranking analysis. *Energy Procedia* 106, 134–154. Available at: doi:10.1016/j.egypro.2016.12.111
- Masterbroek, E., and Kaeding, M. (2006). Europeanization beyond the goodness of fit: domestic politics in the forefront. *Comp. Eur. Polit.* 4 (4), 331–354. Available at: doi:10.1057/palgrave.cep.6110078
- Matsumoto, K., Makridou, G., and Doumpos, M. (2020). Evaluating environmental performance using data envelopment analysis: the case of European countries. *J. Clean. Prod.* 272, 122637. Available at: doi:10.1016/j.jclepro.2020.122637
- Melidis, M., and Russel, D. J. (2020). Environmental policy implementation during the economic crisis: an analysis of European member state “leader-laggard” dynamics. *J. Environ. Policy and Plan.* 22 (2), 198–210. Available at: doi:10.1080/1523908X.2020.1719051
- Musa, M. S., Jelilov, G., Iorember, P. T., and Usman, O. (2021). Effects of tourism, financial development, and renewable energy on environmental performance in EU-28: does institutional quality matter? *Environ. Sci. Pollut. Res.* 28 (38), 53328–53339. Available at: doi:10.1007/s11356-021-14450-z
- Pollex, J., and Lenschow, A. (2020). Many faces of dismantling: hiding policy change in non-legislative acts in EU environmental policy. *J. Eur. Public Policy* 27 (1), 20–40. Available at: doi:10.1080/13501763.2019.1574869
- Ríos, A.-M., and Picazo-Tadeo, A. J. (2021). Measuring environmental performance in the treatment of municipal solid waste: the case of the European Union-28. *Ecol. Indic.* 123, 107328. Available at: doi:10.1016/j.ecolind.2020.107328
- Sanz-Díaz, M. T., Velasco-Morente, F., Yñiguez, R., and Díaz-Calleja, E. (2017). An analysis of Spain’s global and environmental efficiency from a European Union perspective. *Energy Policy* 104, 183–193. Available at: doi:10.1016/j.enpol.2017.01.030
- Tobin, P. (2017). Leaders and laggards: climate policy ambition in developed states. *Glob. Environ. Polit.* 17 (4), 28–47. Available at: doi:10.1162/GLEP\_a\_00433
- Toma, P., Miglietta, P. P., Zurlini, G., Valente, D., and Petrosillo, I. (2017). A non-parametric bootstrap-data envelopment analysis approach for environmental policy planning and management of agricultural efficiency in EU countries. *Ecol. Indic.* 83, 132–143. Available at: doi:10.1016/j.ecolind.2017.07.049
- Vlontzos, G., Niavis, S., and Manos, B. (2014). A DEA approach for estimating the agricultural energy and environmental efficiency of EU countries. *Renew. Sustain. Energy Rev.* 40, 91–96. Available at: doi:10.1016/j.rser.2014.07.153
- Wang, K., Yu, S., and Zhang, W. (2013). China’s regional energy and environmental efficiency: a DEA window analysis based dynamic evaluation. *Math. Comput. Model.* 58 (5), 1117–1127. Available at: doi:10.1016/j.mcm.2011.11.067
- Wolsink, M. (2010). Contested environmental policy infrastructure: socio-political acceptance of renewable energy, water, and waste facilities. *Environ. Impact Assess. Rev.* 30 (5), 302–311. Available at: doi:10.1016/j.eiar.2010.01.001
- Woo, C., Chung, Y., Chun, D., Seo, H., and Hong, S. (2015). The static and dynamic environmental efficiency of renewable energy: a Malmquist index analysis of OECD countries. *Renew. Sustain. Energy Rev.* 47, 367–376. Available at: doi:10.1016/j.rser.2015.03.070
- Zofio, J. L., and Prieto, A. M. (2001). Environmental efficiency and regulatory standards: the case of CO<sub>2</sub> emissions from OECD industries. *Resour. Energy Econ.*, 21. doi:10.1016/S0928-7655(00)00030-0