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## EDITED BY

Chenxi Li,  
Xi'an University of Architecture and Technology,  
China

## REVIEWED BY

Andreas Magerl,  
University of Natural Resources and Life  
Sciences Vienna, Austria  
Leonardo Clavijo,  
Universidad de la República, Uruguay

## \*CORRESPONDENCE

Nicolas Bélanger,  
✉ nicolas.belanger@teluq.ca

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# Opportunities and challenges to improve carbon and greenhouse gas budgets of the forest industry through better management of pulp and paper by-products

Sharlène Laberge<sup>1,2</sup>, Blandine Courcot<sup>1,2</sup>, Andréanne Lagarde<sup>1,2</sup>,  
Simon Lebel Desrosiers<sup>1,2</sup>, Karima Lafore<sup>1,2</sup>, Evelyne Thiffault<sup>3</sup>,  
Nelson Thiffault<sup>4,5</sup> and Nicolas Bélanger<sup>1,2</sup>\*

<sup>1</sup>Département Science et Technologie, Télug, Université du Québec, Québec, QC, Canada, <sup>2</sup>Centre d'étude de la Forêt, Université du Québec à Montréal, Montréal, QC, Canada, <sup>3</sup>Centre de Recherche sur les Matériaux Renouvelables, Département des Sciences du bois et de la forêt, Université Laval, Québec, QC, Canada, <sup>4</sup>Canadian Wood Fibre Centre, Canadian Forest Service, Natural Resources Canada, Québec, QC, Canada, <sup>5</sup>Centre d'étude de la forêt, Faculté de Foresterie, de Géographie et de Géomatique, Université Laval, Québec, QC, Canada

Developing land use strategies to optimize carbon sinks and improve carbon footprints involves proposing efficient nature-based solutions that industries and businesses can implement while considering financial and legislative constraints. The pulp and paper industry is associated with significant greenhouse gas (GHG) emissions, primarily due to the substantial carbon dioxide (CO<sub>2</sub>) footprint of its mills. Also, some forestry operations contribute to the release of carbon to the atmosphere in the form of CO<sub>2</sub> and methane (CH<sub>4</sub>). Conversely, this industry could potentially be a significant ally in the fight against climate change by favoring forestry practices that reduce carbon emissions and increase its sequestration, namely, by adding value to industrial by-products (e.g., biosolids) instead of treating them as wastes and landfilling them. Notably, the pulp and paper industry has been seeking alternative uses of its by-products, such as fertilizers to maximize tree growth. In this paper, we identify opportunities and challenges that exist for the pulp and paper industry in regard to recycling industrial by-products to: 1) lower GHG emissions directly at the mill and 2) improve its GHG budget by increasing carbon sequestration in forests and plantations. We illustrate our analyses by describing a case study of a pulp and paper mill in southern Quebec, Canada, that uses its biosolids and other by-products as fertilizers. This case study highlights that this strategy could not only contribute to the reduction of GHGs but could also create added value and improve economic returns of forest operations.

## KEYWORDS

biosolids, greenhouse gases, landfilling, methane, nature-based climate solutions, pulp and paper industry

## 1 Introduction

Other than innovative technologies such as carbon capture (Kätelhön et al., 2021) and utilization (Jeffry et al., 2021) as well as charge carrier dynamics (Sun et al., 2023), nature-based climate solutions (NCS) are proposed to reduce atmospheric greenhouse gases (GHGs) and mitigate climate change (Griscom et al., 2017; Seddon, 2022). Nature-based climate solutions include efforts that aim to increase the ability of ecosystems to sequester carbon and decrease GHG emissions. Applying NCS may include land-use practices such as protecting forests or proceeding to afforestation of abandoned farmlands, decommissioned mine sites or abandoned industrial fields (Bastin et al., 2019; Lewis et al., 2019; Kaarakka et al., 2021). Nature-based climate solutions may also include more subtle changes in cultural practices such as reducing the use of synthetic fertilizers and using biochar, lime, green crops and crop rotation, and mycorrhizae to reduce nitrous oxide (N<sub>2</sub>O) emissions, a potent GHG (Forster et al., 2021; Hassan et al., 2022). In agricultural contexts, implementing such NCS could also imply using less intensive tilling practices to reduce carbon loss from soils (Maia et al., 2022). Similarly, the forest industry can modify forest management practices with the goal of sequestering more carbon and lowering GHG emissions (Drever, 2021).

The pulp and paper industry, known for its energy-intensive operations (either as electricity, heat for steam production and fuel), is closely tied to a substantial carbon dioxide (CO<sub>2</sub>) footprint, contributing to 2% of direct industrial CO<sub>2</sub> emissions in 2022 (Szabó et al., 2009; Del Rio et al., 2022; IEA, 2023), even though this industry often self-generates its electricity and heat using its residues as fuel. In the US, for example, Tomberlin et al. (2020) estimated that one metric ton of paper creates a production weighted average of 942 kg of CO<sub>2</sub> equivalent, with 50% coming from fuels (Tomberlin et al., 2020). Globally in 2021, the pulp and paper industry emitted an average of 0.45 t of CO<sub>2</sub> per tonne of paper produced, whereas projections suggest emissions of 0.31 t CO<sub>2</sub> per tonne of paper by 2030 (IEA, 2022).

Moreover, the industry faces emissions from by-products such as paper biosolids (PBs) and de-inking sludges, with landfilling and incineration recognized as sources of CO<sub>2</sub> and methane (CH<sub>4</sub>) emissions (Faubert et al., 2016). With the aim of developing a circular economy and mitigating climate change more efficiently, the pulp and paper industry has been looking at alternatives to landfilling and incineration of PBs. Notably, the industry has been using its biosolids as a soil amendment for silvicultural, agricultural and ecological restoration purposes, as well as for biorefinery products such as wood adhesives and fillers, thermoplastic composites, and sorbent materials (Pervaiz and Sain, 2015; Bilodeau-Gauthier et al., 2022; Chen et al., 2023; Grimond et al., 2023). Recycling of PBs in silviculture could increase carbon sequestration by maximizing tree growth (Bilodeau-Gauthier et al., 2022) and promoting a stable form of carbon in soils (Khlifa et al., 2023), as well as reducing GHG emissions compared to synthetic fertilizers (Chen et al., 2023). Additionally, it could avoid a significant CO<sub>2</sub>-equivalent flux to the atmosphere associated with the landfilling or incineration of PBs. This avoidance could also be conducted in a cost-effective way, while fast-growing tree plantations receiving PBs could contribute to fibre supply and help conserve habitats and biodiversity (Himes et al., 2022). In this

perspective paper, we aim to identify opportunities and challenges for the pulp and paper industry in recycling its PBs to: 1) lower GHG emissions directly at the mill site, 2) improve its GHG budget by increasing carbon sequestration in forests and plantations, and 3) favor a forest land management scheme for the preservation of natural habitats and biodiversity.

## 2 The world problem of biosolids, with emphasis on paper biosolids

Quantifying global waste generation proves challenging due to varied definitions and methodologies. Current estimates indicate waste production at 19.8 billion tonnes annually, with 15% comprising biomass like solid wood, construction wood, paper and cardboard (Maalouf and Mavropoulos, 2023). Despite uncertainties, projections suggest a global increase to 28 and 46 billion tonnes per year by 2030 and 2050, respectively (Di Giacomo and Romano, 2022). Managing these wastes poses a significant environmental challenge, with 40% being landfilled, 19% recycled or composted, and 11% incinerated (Kaza et al., 2018). As the global population continues to grow, waste management emerges as one of the greatest environmental challenges of the 21st century (Vaverková, 2019).

Effective waste management should be guided by policies that promote reducing at the source, reusing, recycling and recovering (i.e., the 4R principle; Rada et al., 2018). However, more often than not, it involves landfilling or incineration. Because incineration is costly and a direct and significant source of GHGs, landfilling is the preferred method and is expected to increase in the coming decades. Nonetheless, the expansion of landfill sites brings environmental challenges, including increased hydraulic loads of leachates on adjacent watercourses, air quality issues (e.g., odors and suspended particles) and GHG emissions (Wang et al., 2014; Zhang et al., 2019; Siddiqua et al., 2022).

Sewage sludge, a by-product of municipal wastewater treatment, is termed municipal biosolids (MBs) when meeting certain criteria for land application. These must exhibit low levels of pathogens and other contaminants such as metals, although criteria vary from country to country (LeBlanc et al., 2009; Lowman et al., 2013; Popoola et al., 2023). Global production of biosolids is projected to reach 200 million tonnes annually by 2025 (Mohajerani et al., 2017), with the pulp and paper industry being a major contributor. In 2020, 34 countries accounted for 82% of the global paper and paperboard production (402 million tonnes) (FAO, 2023), set to increase to 550 million tonnes by 2050 (Mabee and Roy, 2003). As such, the production of PBs is expected to increase proportionally. In Quebec, Canada, for example, paper mill biosolids amounted to 1 million tonne in 2018 (RECYC-QUÉBEC, 2018). The pulp and paper production process generates wastes and by-products, of which 87% is classified as pulp and paper mill sludge, whereas the remaining is considered as impurities, waste chemicals and gaseous emissions (Turner et al., 2022).

The pulp and paper industry is regulated for wastewater discharge from pulp and paper mills into surface waters. Paper biosolids, similar to MBs, are disposed by landfilling, incineration, or are applied to land for various purposes (e.g., agriculture, forestry, ecological restoration). In addition to environmental impacts,

landfilling and incineration of biosolids face limited social acceptance and economic challenges, such as substantial fees. These fees aim, in part, to incentivize industries to adhere to the 4R principle before resorting to elimination (Primeau, 2014; Faubert et al., 2015). While landfilling is cost-effective compared to incineration, cost recovery options (e.g., biogas recovery for energy production) remains challenging. Consequently, the use of MBs and PBs as soil amendments or materials for reconstructing soils in restoration projects is gaining popularity.

### 3 Landfilling of paper biosolids

Landfilling of PBs is the most widely used disposal method, and the pulp and paper industry generally manages its own landfilling facilities (Haile et al., 2021). In addition to hazardous substances reaching adjacent watercourses and causing air quality issues (Haile et al., 2021; Siddiqua et al., 2022), landfilling sites generate large amounts of GHGs under various conditions (Wang et al., 2014; Zhang et al., 2019). However, estimates of the proportion of landfill sites dedicated to PB disposal relative to landfill sites as a whole and empirical estimates of GHG emissions from them are non-existent, making it difficult to assess their contribution to GHG emissions globally (Primeau, 2014; Faubert et al., 2016; MELCC, 2020). Faubert et al. (2016) suggested from modelling and theory that 2.69 tonnes of CO<sub>2</sub> and 0.24 tonne of CH<sub>4</sub> could be emitted to the atmosphere for every tonne of PBs that is landfilled. Methane emissions are a primary concern for climate change due to its high global warming potential (GWP) compared to CO<sub>2</sub>, i.e., 28 to 32 times higher over 100 years as initially suggested by Byrne and Goldblatt (2014). However, a new model (GWP\*) has recently been proposed to improve calculations of warming-equivalent emissions since GWP can overestimate the cumulative effect of short-lived climate pollutants, such as CH<sub>4</sub>, by not properly considering the short-term and long-term effects of these pollutants in the atmosphere (Cain et al., 2019; Lynch et al., 2020). Methane from landfilled PBs is considered the main GHG because most of CO<sub>2</sub> emissions could be offset by the main constituent of PBs, i.e., short lignin fibers derived from photosynthesis (Camberato et al., 2006). However, landfilled biosolids can also generate N<sub>2</sub>O through the cooxidation of ammonia by methanotrophic soil bacteria in cover soils (Zhang et al., 2009).

Methanogenesis occurs during the anaerobic mineralization of organic matter (Le Mer and Roger, 2001). Anaerobic conditions in landfill cells are created over time through the compaction of layers of organic matter (Olivier, 2013). Because PBs are rich in wood fibers and organic matter, they provide a significant carbon source for methanogenic microbes (Camberato et al., 2006). During the fermentation process, methanogenic microbes transform organic matter into CH<sub>4</sub> and some CO<sub>2</sub> (Le Mer and Roger, 2001). Landfill cells are typically covered with an engineered cap comprising a clay layer to limit water infiltration and gas exchange. This cap is further covered by a sandy mineral soil with low organic matter content (Handel et al., 1997; Fraser-McDonald et al., 2022). A portion of the CH<sub>4</sub> produced is then oxidized through methanotrophy to CO<sub>2</sub> and water in the cover soil layer, facilitated by its high oxygen levels (Boeckx et al., 1996; Le

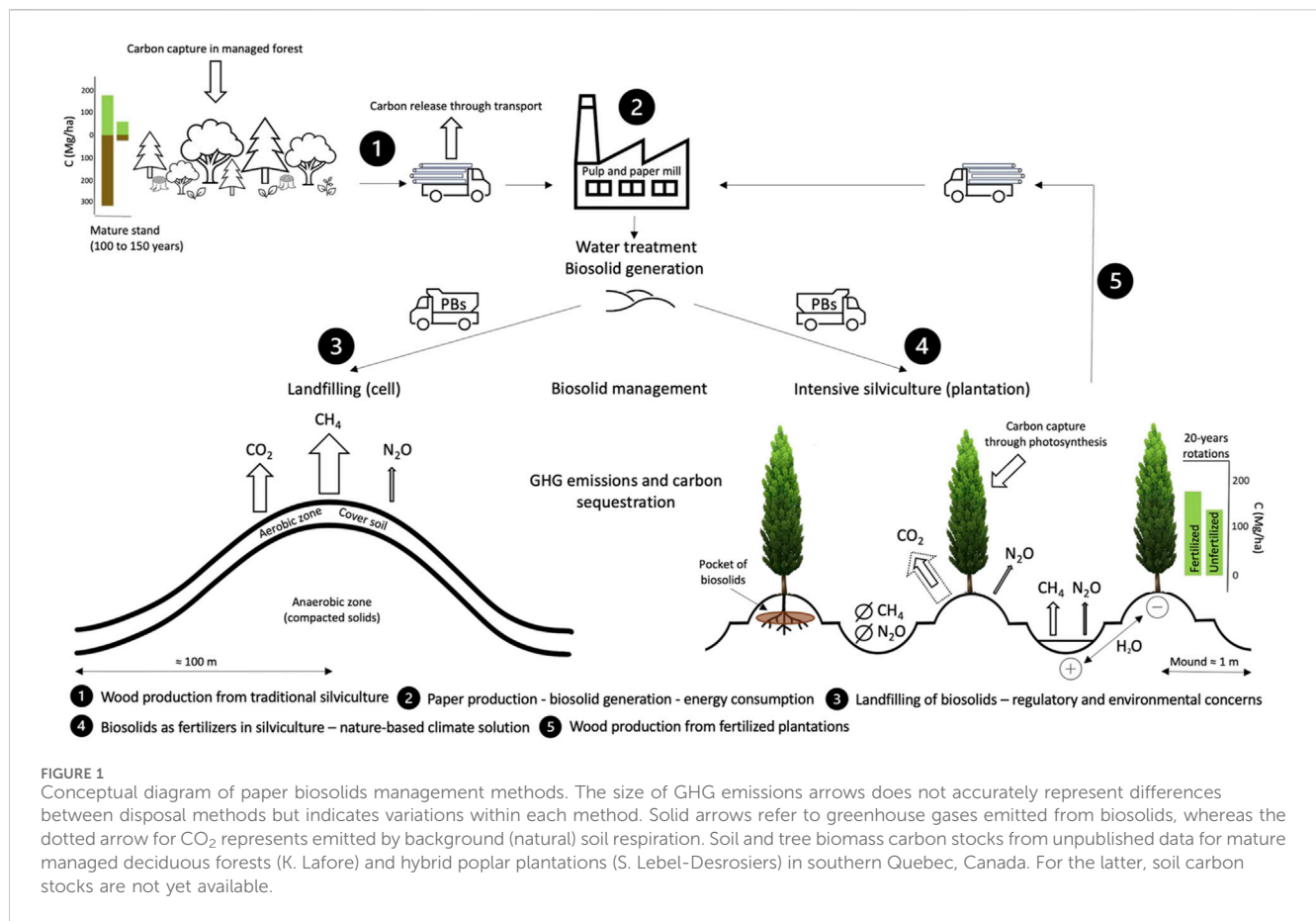
Mer and Roger, 2001) and low organic matter content (Handel et al., 1997). Methane that remains unaffected in the cover soil layer is released to the atmosphere by diffusion and plant gas exchange (Le Mer and Roger, 2001).

Cover soils thus aim to reduce CH<sub>4</sub> emissions by creating oxygenated conditions where CH<sub>4</sub> transforms into CO<sub>2</sub> and water before reaching the atmosphere. However, fissures, cracks and holes in the mineral cap often form, creating preferential channels for CH<sub>4</sub> from the organic layer to bypass oxidation and be released directly to the atmosphere (Schroth et al., 2012). Also, highly porous material can allow CH<sub>4</sub> to flow through the cover soil too quickly for oxidation to occur before reaching the atmosphere (Wang et al., 2022). The heterogeneous nature of the cover soil often leads to high spatial variability in CH<sub>4</sub> fluxes across a landfill site (Schroth et al., 2012). Estimates of CH<sub>4</sub> emissions from PB landfill sites, assuming that 10% of CH<sub>4</sub> is oxidized in the cover soil, may therefore be underestimated (Chanton et al., 2009; Schroth et al., 2012). However, membrane gas permeation has been used in recent years to channel biogas for energy use or to transform it into CO<sub>2</sub> through flaring (Makaruk et al., 2010; Chmielewski et al., 2019). To our knowledge, however, this method has only been used in municipal landfills.

### 4 Paper biosolids as fertilizers

Application of MBs and PBs to agricultural soils can positively impact soil fertility and crop yields (Gagnon and Ziadi, 2012; Lu et al., 2012; Ziadi et al., 2013; Abdi et al., 2016; Sharma et al., 2017). Other than providing nitrogen and phosphorus, PBs: 1) are rich in calcium, thus addressing soil acidity issues, 2) enhance soil structure by providing organic matter (or carbon), 3) increase nutrient cycling by improving soil microbial biomass and activity and 4) increase soil cation exchange capacity due to the added organic matter and increased pH. Paper biosolids are less enriched in nitrogen and phosphorus compared to MBs and contain fewer contaminants, thus reducing environmental risks (Charbonneau et al., 2001; Gagnon et al., 2013). The high fiber content of PBs also imparts slow-nitrogen-release fertilizer properties (Gagnon and Ziadi, 2012). In forestry, the use of PBs can be particularly significant for carbon sequestration. It has the potential to stimulate tree growth by improving foliar nutrition and, consequently, accelerating carbon capture through photosynthesis (Lteif et al., 2007; Rodriguez et al., 2018; Bilodeau-Gauthier et al., 2022). However, further work is necessary to assess how fertilization with PBs can influence carbon sequestration of forest plantations under different climates, site conditions (e.g., soils), site (mechanical) preparation and weed management. To our knowledge, no effort was made in synthesizing such data.

Fertilization with PBs can decrease GHG emissions when compared to synthetic fertilizers such as urea (Chen et al., 2023). Furthermore, the use of PBs in agriculture and forestry diverts biosolids from landfills, thereby reducing associated CH<sub>4</sub> emissions (Faubert et al., 2019). Considering these factors, repurposing PBs from the pulp and paper industry as fertilizers in silviculture emerges not only as a theoretical concept but also as a practical NCS. To illustrate this potential, we present a case study from



southern Quebec. This case study demonstrates how the purposeful use of PBs in silvicultural treatments can effectively contribute to sustainable waste management practices and significantly enhance carbon sequestration efforts (Figure 1).

In this case study, PBs are specifically used to fertilize fast-growing hybrid poplar (e.g., *Populus x canadensis* × *P. maximowiczii*) plantations, classified as an intensive silvicultural practice. Trees are planted on mound which are created with an excavator (Figure 1). Using allometric equations from the R software package “allodb” (Gonzalez-Akre et al., 2022), it was estimated that, over a 20-year period, plantations receiving industrial by-products could sequester up to 26.7% more carbon than unfertilized plantations (treated: 171 Mg C ha<sup>-1</sup>; unfertilized: 134 Mg C ha<sup>-1</sup>). Achieving a target of 200 m<sup>3</sup> ha<sup>-1</sup> for merchantable wood could be realized, on average, in about 20 years in fertilized plantations (compared to 140 m<sup>3</sup> ha<sup>-1</sup> in unfertilized plantations), allowing faster rotations and higher carbon sequestration in the long term. In comparison, wood biomass (aboveground) in twenty-one mature (~100–150 years) deciduous forests and mixedwood stands in southern Quebec holds between 38 and 184 Mg ha<sup>-1</sup> of carbon (K. Lafore, unpublished data).

It could be argued that this wood biomass is destined for paper, a product with a short life span and limited recycling cycles, resulting in fewer benefits in terms of carbon emission mitigation. Yet, augmenting yields in forest plantations destined for paper production also has the added benefit of relieving pressure on other forests for wood supply. This, in turn, leads to the

protection of larger swaths of forest landscapes supporting the provision of other ecosystem services (Wang et al., 2022).

Soil carbon stocks are expected to substantially increase after PB application due to their high organic carbon content and carbon:nitrogen ratio (Henry et al., 1994; Gagnon and Ziadi, 2012). At an application rate of about 125 Mg ha<sup>-1</sup>, the newly added carbon (approximately 50–60 Mg C ha<sup>-1</sup>) is anticipated to become a stable long-term storage in forest soils, similar to observations in ecological restoration projects, some being afforestation efforts, using MBs for soil reconstruction (Trlica and Teshima, 2011; Carbassa et al., 2020; Khelifa et al., 2023). These afforestation projects can also lead to significant tree growth and carbon sequestration as well (e.g., Grimond et al., 2023; Bélanger et al., 2024). Soil carbon improvement results from ‘pockets’ of PBs trapped at the base of the mounds during their creation and due to limited mixing, along with the aggregation of organic carbon to clay-mineral particulates. The stability of these carbon pools is largely influenced by the chemical signature of the carbon forms in the PBs, particularly the presence of more labile carbohydrates (Gagnon and Ziadi, 2022).

In the mid-term, increased tree growth is expected to further enhance soil carbon stocks via increased litter flux, including roots, to the soil (Thevathasan and Gordon, 1997; Arevalo et al., 2011), promoting forest floor development and carbon accumulation in mineral horizons. Preliminary data suggest that forest floor depth of the southern Quebec case study sites increased from an average of 1.7–5 cm in just 12 years for plantations receiving PBs, while no



change in forest floor depth was observed in unfertilized plantations over 14 years. The soil organic content of boreal, mixed and deciduous forest soils in Quebec is, on average, 44 Mg C ha<sup>-1</sup> (Tremblay and Ouimet, 2000). Less than 3% of the forests exhibited soil organic content (SOC) above 100 Mg C ha<sup>-1</sup> in this study. Similarly, a very large variation in soil carbon stocks (2–315 Mg ha<sup>-1</sup>) was measured for old-growth forest stands in southern Quebec (K. Lafore, unpublished data). Although estimating the amount of this new carbon is challenging, a forest floor in steady state in these temperate forests typically reaches a depth of 5–15 cm and is 25%–40% of the soil carbon stock. The expectation is that this steady state will be reached more rapidly in the plantations receiving PBs.

Another key, yet complex aspect to consider, is the potential reduction in GHG emissions by diverting PBs from landfills and applying them to plantations. It is anticipated that applying relatively small amounts of PBs at the soil surface, without or with very limited compaction and under well-aerated soil conditions, will significantly lower overall CH<sub>4</sub> and N<sub>2</sub>O emissions compared to landfilling. However, spatial variation in fluxes is expected to be large, correlating with field microtopography associated with mounding. Mounding creates a heterogeneous soil surface with: 1) dry zones characterized by lower litter accumulation (mounds), 2) unaltered areas with soil moisture levels and litter accumulation rates expected from a well-drained forest soil (ground level), and 3) wetter zones with substantial litter accumulation (depressions) that could contribute to CH<sub>4</sub> and N<sub>2</sub>O emissions. In fact, like methanogenesis, denitrification rates are higher in compacted and poorly drained soils (Beare et al., 2009). Considering that wet depressions represent about one-third of the planted landscape, GHG emissions under this microtopography should be assessed. Preliminary summer sampling at the case study site suggests that CH<sub>4</sub> fluxes are present in only some depressions and CO<sub>2</sub> fluxes were lower in (wet) depressions and on (dry) mounds (287–356 mg m<sup>-2</sup> h<sup>-1</sup>) than on unaltered soil (701 mg m<sup>-2</sup> h<sup>-1</sup>).

The values above suggest no significant change in CO<sub>2</sub> fluxes due to mounding or PB application, as they are within the range of those measured in comparable forest stands under the same soil temperatures (Bélangier et al., 2021). Similarly, our preliminary data suggest that PB application does not lead to the production of N<sub>2</sub>O across the mounding microtopography. Chen et al. (2023) observed the opposite after PB application in hybrid poplar plantations in Alberta, i.e., a 21% and 17% increase in CO<sub>2</sub> and N<sub>2</sub>O emissions, but no change in CH<sub>4</sub> emissions. These sites appear to show less microtopography (prepared by one-way cultivation). Measurements of the southern Quebec case study sites were taken 3 years after PB application, while fresh and wet biosolids are known to be significant sources of N<sub>2</sub>O (Roman-Perez and Hernandez-Ramirez, 2022). The duration of these large emissions is not well known, especially in forest soils; this deserves our full attention because the benefits that could be gained by reducing landfill CH<sub>4</sub> emissions and increasing carbon sequestration with the use of PBs in hybrid poplar plantations could be offset if this practice also causes N<sub>2</sub>O emissions for a prolonged period. Further monitoring of soil conditions and gas efflux is thus required to determine if carbon accumulates in the new soils and to assess GHG dynamics to fully evaluate their potential as a climate solution.

## 5 Conclusion

Management of waste materials from the pulp and paper industry often involves landfilling. However, our preliminary results offer a promising perspective for the enhanced recovery and recycling of PBs, leading to reduced GHG emissions and improved carbon sequestration when combined with appropriate silvicultural practices. As more field data are collected to better quantify landfill CH<sub>4</sub> emissions, diverting PBs from landfills to plantations could prove valuable for the forest sector to contribute to CH<sub>4</sub> emissions mitigation targets.

Despite the promising benefits of applying biosolids in forestry, challenges must be addressed. Environmental and regulatory concerns, including the potential for heavy metal accumulation and pathogen transmission, require careful management (Pöykiö et al., 2007; Lu et al., 2012). Among the pollutants of concern are perfluoroalkylated substances, synthetic chemicals such as perfluorooctane sulfonate and perfluorooctanoic acid (PFAS), which pose health risks (Kirk et al., 2018; Ehrlich et al., 2023; van Larebeke et al., 2023). Communities near landfill sites have raised public health concerns, including quality-of-life impacts due to odors (Lowman et al., 2013). Moreover, it is not yet clear how the wastewater treatment processes impact the concentration of PFAS in sludges and biosolids (Behnami et al., 2024; Gewurtz et al., 2024). Therefore, the development of guidelines and monitoring systems is essential to ensure that the use of biosolids adheres to environmental safety standards and does not adversely affect forest ecosystems or human health.

## Data availability statement

The datasets presented in this article are not readily available because this is a perspective paper only and we report preliminary data only. When full data are published, they will be made available publicly. Requests to access the datasets should be directed to nicolas.belanger@teluq.ca.

## Author contributions

SL: Conceptualization, Data curation, Formal Analysis, Investigation, Methodology, Writing—original draft, Writing—review and editing. BC: Conceptualization, Formal Analysis, Investigation, Methodology, Writing—original draft, Writing—review and editing. AL: Conceptualization, Investigation, Methodology, Writing—original draft, Writing—review and editing. SL: Conceptualization, Formal Analysis, Investigation, Methodology, Writing—original draft, Writing—review and editing. Data curation. KL: Data curation, Conceptualization, Formal Analysis, Investigation, Methodology, Writing—original draft, Writing—review and editing. ET: Funding acquisition, Project administration, Resources, Supervision, Conceptualization, Formal Analysis, Investigation, Methodology, Writing—original draft, Writing—review and editing. NT: Funding acquisition, Project administration, Resources, Supervision, Conceptualization, Formal Analysis, Investigation, Methodology, Writing—original draft, Writing—review and editing. NB: Validation,

Conceptualization, Investigation, Methodology, Writing—original draft, Writing—review and editing, Data curation, Formal Analysis, Funding acquisition, Project administration, Resources, Supervision.

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## Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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