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# Estimation of hourly one square kilometer fine particulate matter concentration over Thailand using aerosol optical depth

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In the recent years, concentration of fine particulate matter that are 2.5 microns or less in diameter (PM<sub>2.5</sub>) in Thailand has consistently exceeded the national ambient air quality standard. Currently, the measurement of PM<sub>2.5</sub> concentration relies on air quality monitoring stations operated by the Pollution Control Department of Thailand (PCD). However, these stations are insufficient, particularly in rural areas, where agricultural open burning are major sources of pollution after harvesting period. This study aims to enhance the monitoring of PM<sub>2.5</sub> concentration by leveraging cost-effective technologies. We propose the integration of satellite data, specifically Aerosol Optical Depth (AOD) from Multi-Angle Atmospheric Correction (MAIAC) product and Himawari-8 satellites, with the Weather Research and Forecasting Model (WRF) data, to provide supplementary data to the ground-based monitoring. Hourly 5 × 5 km<sup>2</sup> AOD data from Himawari-8 were downscaled to a high-resolution of 1 × 1 km<sup>2</sup>, leveraging the AOD distribution pattern of the concurrent MAIAC product using eXtreme Gradient Boosting (XGBoost) model. Notably, during Thailand's rainy season (May to August), the study observed a relative reduction in the training model's R-square value. This phenomenon is attributed to temporal discrepancies between Himawari-8 and the MAIAC products during this period. The predictive models of PM<sub>2.5</sub> concentrations with the identification of pertinent variables through Pearson's correlation analysis and recursive feature elimination, driven by the robust XGBoost model. Subsequently, the downscaled AOD, wind speed, temperature, and pressure were identified as predictors for the estimation of hourly PM<sub>2.5</sub> concentration. This comprehensive approach enabled the projection of PM<sub>2.5</sub> levels across Thailand, encompassing over 600,000 grids at 1 × 1 km<sup>2</sup> resolution. The developed models, thus, offer a valuable tool for robust and high-resolution PM<sub>2.5</sub> concentration estimation, presenting significant implications for air quality monitoring and management in Thailand.

## KEYWORDS

PM<sub>2.5</sub> concentration, aerosol optical depth, machine learning, Himawari, MAIAC

## 1 Introduction

In recent decades, the global community has experienced a growing concern over the emission of air pollution resulting from human activities, such as transportation, industry, biomass burning. The World Health Organization (WHO) reported in 2014 that air pollution was accountable for approximately 7 million premature deaths worldwide

(Amnuaylojaroen et al., 2020). Prolonged exposure to elevated levels of air pollution poses diverse health risks, with a particular emphasis on the impact of PM<sub>2.5</sub> (fine particulate matter with a diameter of 2.5 micron or less), a significant air pollutant that profoundly affects human health and wellbeing (Lelieveld et al., 2013; Amnuaylojaroen et al., 2019). PM<sub>2.5</sub> possesses the ability to deeply penetrate the respiratory tract and enter the lungs, leading to impaired lung function and the exacerbation of medical conditions such as asthma and heart disease (Tsai et al., 2000; Vichit-Vadakan et al., 2001; Jinsart et al., 2002).

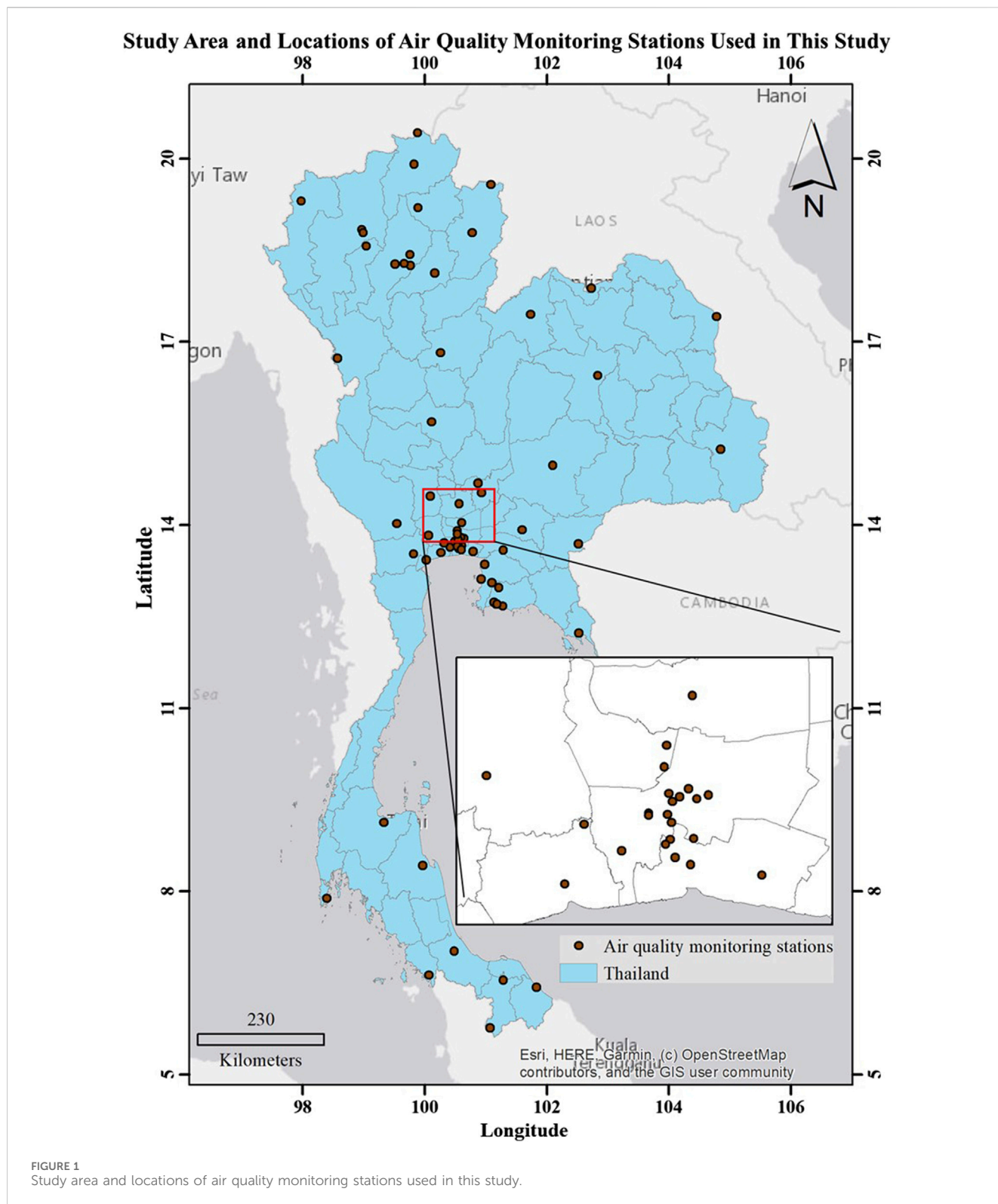
Southeast Asia, notably during dry season, faces recurrent annual air pollution problems predominantly attributed to biomass combustion (Yin et al., 2019). The extensive burning of biomass and other human activities have substantially contributed to the deterioration of air quality in Southeast Asia (Lee et al., 2018; Lee et al., 2019). Furthermore, unfavorable meteorological and geographical conditions also contribute to the air pollution challenges experienced in this region. Northern Thailand, in particular, characterized by its mountainous areas and surrounding rice fields, encounters escalating issues compounded by traffic congestion and the practice of burning stubble prior to the rainy season. As a result, air pollution is accumulated in narrow valleys (Oanh et al., 2006; Amnuaylawjarurn et al., 2010; Oanh and Leelasakultum, 2011; Lee et al., 2019).

To measure the concentration of fine particulate matter or PM<sub>2.5</sub>, it commonly relies on ambient air quality monitoring stations. However, these methods possess inherent limitations (such as costs of the equipment, monitoring and maintenance, availability of technical staffs) and may not provide comprehensive coverage of air quality across all areas (Qu et al., 2017). Consequently, estimating fine particulate matter concentration in the area with limited monitoring stations has emerged as a critical area of research. Estimation methods for PM<sub>2.5</sub> concentration can be broadly categorized into two types: those based on ground-level monitors and those that utilize satellite-based data. Ground-level monitor-based estimation approaches encompass techniques, such as land use regression models, generalized additive mixed models, hierarchical models, and geostatistical interpolation. On another hand, satellite-based estimation methods depend on remote sensing techniques (Qu et al., 2017; Zhang G. et al., 2018). Integrating remote sensing data from satellites, such as AOD with local measurement of meteorological parameters is increasingly being adopted in air quality monitoring practices (Hoff and Christopher, 2009). Machine learning and statistical models are frequently combined with satellite data to estimate PM<sub>2.5</sub> concentrations at a finer spatial resolution (Zhang T. et al., 2018; Joharestani et al., 2019; Xie et al., 2019; Yao et al., 2019; Yang et al., 2020). Previous studies have attempted to capture the relationship between PM<sub>2.5</sub> and satellite-retrieved AOD data using regression models (Bai et al., 2016; Yao et al., 2019). Random forest (RF) and XGBoost have been employed to predict PM<sub>2.5</sub> levels in many studies (Hu et al., 2017; Xiao et al., 2018).

Previous studies have predominantly concentrated on estimating ground-level PM<sub>2.5</sub> concentrations using AOD from different satellites. Specifically, a spatial resolution of

10 × 10 km<sup>2</sup> MODIS AOD has conventionally been employed to derive ambient concentrations of PM<sub>2.5</sub> on regional scales (Ma et al., 2014; Kong et al., 2016). Recently, a 3 × 3 km<sup>2</sup> MODIS AOD has been applied to delve into finer spatial details at urban levels, contributing to population-based PM<sub>2.5</sub> exposure and health effect studies (Xie et al., 2015). Moreover, most PM<sub>2.5</sub> estimations are typically presented at spatial resolutions of 3–10 km<sup>2</sup> to provide air quality in the study areas (Hongthong et al., 2022). The AOD from the MAIAC product represents a cutting-edge technique that combines pixel and image-based processing with time series analysis. The high-resolution (1 × 1 km<sup>2</sup>) AOD dataset minimizes spatiotemporal heterogeneities, thereby enhancing the overall accuracy of ground-level PM concentration estimates (Lyapustin et al., 2011). Another notable AOD product is Himawari-8 AOD, providing a high temporal resolution AOD product which is useful for investigating diurnal variations in air pollution with a spatial-temporal resolution of 5 × 5 km<sup>2</sup>. Recent research in China has started estimating real-time hourly ground-level PM<sub>2.5</sub> using the Himawari-8 AOD product (Chen J. et al., 2019; Sun et al., 2021; Xu et al., 2021). However, few studies have been conducted in Thailand, where air quality monitoring stations are limited (Peng-in et al., 2022).

In Thailand, prior studies have predominantly focused on estimating PM<sub>2.5</sub> concentrations in specific regions, particularly in Chiang Mai province and the Northern region. The estimation of PM<sub>2.5</sub> concentrations in Chiang Mai province utilized AOD data from MODIS with a spatial resolution of 10 × 10 km<sup>2</sup> and meteorological data from Thailand's Pollution Control Department through a multiple linear regression model. However, the study faced limitations related to low spatial resolution and the constrained spatial distribution of meteorological data from air quality monitoring stations (Kanabkaew, 2013). Other studies have also centered on estimating PM<sub>2.5</sub> concentrations based on the AOD of the MODIS-Terra platform, which provides a spatial resolution of 10 × 10 km<sup>2</sup>. These studies employed multiple or multivariate linear regression techniques for PM<sub>2.5</sub> concentration estimation (Kanabkaew, 2013; Wei et al., 2019; Amnuaylojaroen, 2022; Wongnakae et al., 2023). Another AOD product used for estimating particulate matter is the MAIAC product, offering a higher spatial resolution than other products. Hongthong et al. (2023) utilized this product to estimate PM<sub>10</sub> concentration in some provinces of Northern Thailand, using the data to assess the attributed respiratory disease burden. However, this study identified limitations in predicting PM<sub>10</sub> concentration, attributed to the availability of weather data only at monitoring stations and the MAIAC-AOD being available only once per day (Hongthong et al., 2023). Additionally, predicting PM<sub>2.5</sub> is challenging as it is influenced by different factors, including weather conditions and environmental seasonality, all of which can significantly affect the regression models used for prediction (Amnuaylojaroen, 2022). In this study, our focus was to develop the method to enhance the spatial and temporal resolutions of AOD using products from two satellites, and incorporating weather parameters from the



WRF to improve ground-level  $PM_{2.5}$  estimation in Thailand. This study holds significance not only in elucidating the high-resolution spatial distribution of  $PM_{2.5}$  concentration on an hourly basis, but also in its potential applications, such as solar energy assessment (Luo et al., 2019), aerosol data assimilation (Zhang et al., 2021).

## 2 Data and methods

### 2.1 Study area

The study was conducted to estimate hourly  $1 \times 1 \text{ km}^2$   $PM_{2.5}$  concentration over Thailand. Figure 1 illustrates the study area,

TABLE1 Summary of Data Collection in this Study.

| Data sources                           | Parameters   | Time period             | Resolution   |
|--|--|-------------------------|--|
| Himawari-8 satellite                   | AOD  | 2017–2021 (09:00–16:00) | - Hourly average data<br>- Spatial resolution of $5 \times 5 \text{ km}^2$ |
| MAIAC data (MODIS)                     | AOD  | 2017–2021               | - Daily average<br>- Spatial resolution of $1 \times 1 \text{ km}^2$       |
| WRF model                              | - Pressure (hPa)   | 2018–2021               | - Hourly average data  |
|  | - Temperature ( $^{\circ}\text{C}$ )                         |                         | - Spatial resolution of $6 \times 6 \text{ km}^2$                          |
|  | - Wind speed (m/s)   |                         |  |
|  | - Cumulative rainfall (mm)                                   |                         |  |
| Ambient air quality monitoring station | PM <sub>2.5</sub> concentration ( $\mu\text{g}/\text{m}^3$ ) | 2018–2021               | - Hourly average data  |
|  |  |                         | - Point measurement  |

including the locations of ambient air quality monitoring stations used in this study. These ambient air quality monitoring stations are operating by the Thailand Pollution Control Department (PCD). Across the region, there are 73 monitoring stations that monitor different air quality parameters, i.e., carbon monoxide, nitrogen oxide, sulfur dioxide, ozone, and particulate matter (PM<sub>2.5</sub> and PM<sub>10</sub>), as well as weather parameters, including wind speed, wind direction, relative humidity, and temperature from 2018 to 2021. However, it should be noted that not all pollutant and meteorological parameters are measured at every station.

## 2.2 Data collection

Data in this study were collected from four sources as follows:

The first dataset used in this study was the AOD from the Himawari-8 satellite, known for its remarkable ability to capture full-disk observations at 10-min intervals. It is also renowned for its precise detection and mapping of volcanic ash and aerosols, as indicated in previous studies (Marchese et al., 2018; Gao et al., 2021; Fu et al., 2023). The study used the spatial of  $5 \times 5 \text{ km}^2$  and a temporal resolution of 1 h, making them suitable for our research purposes for hourly monitoring data.

The second dataset, Multi-Angle Atmospheric Correction (MAIAC), derives its data from the Moderate Resolution Imaging Spectroradiometer (MODIS). This algorithm relies on time series analysis to determine spectral surface reflectance, a critical factor in aerosol retrieval. MAIAC's AOD retrievals have played a pivotal role in estimating ground-level PM distributions and supporting epidemiological studies on air pollution, as highlighted in existing research (Xiao et al., 2017). For our study, we utilized a spatial resolution of  $1 \times 1 \text{ km}^2$  and a temporal resolution of 1 day.

The third data was hourly average PM<sub>2.5</sub> concentration from 2018 to 2021 which were collected from 69 ambient air quality monitoring stations operated by PCD (69 stations from all 73 stations have PM<sub>2.5</sub> monitoring equipment). From all 69 monitoring stations, 23 monitoring stations were in Bangkok Metropolitan Region. There were 10 stations in central Thailand,

5 stations in northeast Thailand, 9 stations in east Thailand, 14 stations in north Thailand and 8 stations in south Thailand.

Lastly, meteorological data were extracted from the WRF model provided by the Thai Meteorological Department (TMD). The WRF data includes weather parameters, such as air pressure, temperature, wind speed, wind direction, accumulated rainfall. These parameters were averaged on an hourly basis and had a spatial resolution of  $6 \times 6 \text{ km}^2$ .

In conclusion, data collected in this study are summarized in Table 1.

## 2.3 Data processing

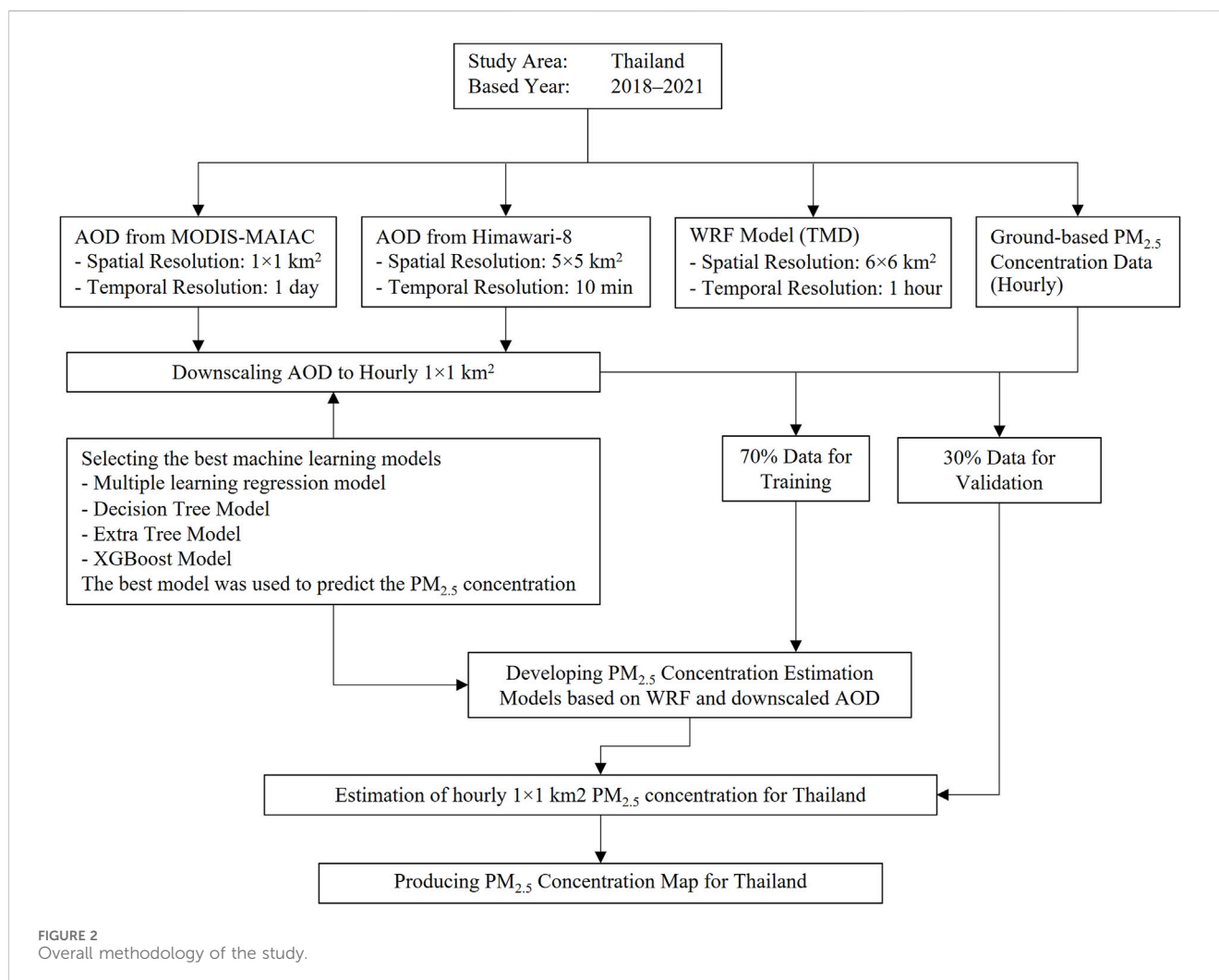
This study used Python programming to process the data. The main workflow encompassed downscaling hourly Himawari-8 AOD data from  $5 \times 5 \text{ km}^2$  to  $1 \times 1 \text{ km}^2$  resolution. Then, machine learning models were employed to predict PM<sub>2.5</sub> concentration using downscaled AOD and meteorological parameters. Figure 2 shows the workflow in this study.

### 2.3.1 Aerosol optical depth from Himawari-8

Aerosol Optical Depth from Himawari-8 with a spatial resolution of  $5 \times 5 \text{ km}^2$  and temporal resolution of 1 h from 9:00 to 16:00 were downloaded in the format of netCDF. To process the data, the rioarray library in Python was employed to convert the data format into a raster format. Additionally, the geopandas library was used to extract the data and save it in a CSV file for further analysis.

### 2.3.2 Aerosol optical depth from MAIAC

The MAIAC data, obtained from the MODIS sensors on Terra and Aqua satellites. The product used in this study is MCD19A2, which is provided in TIFF file format with a daily average and a spatial resolution of  $1 \times 1 \text{ km}^2$ . To process the data, Python's rasterio library was utilized to extract the relevant information. Subsequently, the geopandas library was employed to save the extracted data in a CSV file format, enabling further analysis of the data.



### 2.3.3 WRF model data

The WRF model data, acquired from the Thai Meteorological Department, is provided in CSV file format. In this study, the pandas and geopandas libraries were employed to extract the data and save it into a CSV file with a format consistent with the Himawari-8 and MAIAC data. This standardized format allows seamless integration and further analysis of the data, enabling efficient utilization in this study.

## 2.4 Machine learning models

In the course of this investigation, five distinct machine learning algorithms were scrutinized, with the optimal model subsequently chosen for the downscaling of AOD and the formulation of a  $PM_{2.5}$  prediction model. The five machine learning models under consideration include.

### 2.4.1 Recursive feature elimination

Recursive Feature Elimination (RFE) is a crucial feature selection technique employed to identify and remove the weakest features within a dataset. Its primary objective is to determine the optimal set of features for a given dataset (Granitto et al., 2006; Yan and Zhang, 2015). RFE iteratively eliminates features until the desired number of features

remains. The ranking of features is assessed by the RFE model, which systematically removes features during each iteration to address issues like collinearity and dependencies in the model (Ketu, 2022).

### 2.4.2 Multiple linear regression

Multiple Linear Regression (MLR) is a widely used statistical model for investigating the relationship between a continuous response variable and one or more predictor variables, which can be either continuous or categorical. MLR is a parametric model that assumes a normal distribution, constant variance, and a linear relationship between the response and predictor variables (Buya et al., 2023).

### 2.4.3 Decision tree model

The decision tree algorithm is represented as a tree structure, which can be binary or non-binary. Each non-leaf node corresponds to an attribute test, and every branch represents an attribute's possible outcomes within specified boundaries. Leaf nodes contain categorical values. The process of the decision tree involves classifying characteristic attributes starting from the root node and evaluating their values based on the selected output branches until a leaf node is reached, which determines the final category (Srivastava et al., 1999). To bring order to unstructured data and regularize the dataset, three



common methods are employed in support of decision trees: information gain, gain ratio, and Gini impurity (Zuo et al., 2020).

#### 2.4.4 Extra trees model

The Extra-trees model is also a tree-based ensemble learning method based on the bagging technique and construct from multiple decision trees, where each tree is generated by bootstrap sampling from the training dataset (Breiman, 2001; Hu et al., 2017; Chen et al., 2018), but introduces additional randomness in selecting features and splitting the points from all data samples in the tree-building process (Geurts et al., 2006; Wei et al., 2020; Wei et al., 2021).

#### 2.4.5 eXtreme gradient boosting model (XGBoost)

XGBoost is a gradient-boosting technique that improves performance and speed using a tree-based ensemble machine learning algorithm (Chen and Guestrin, 2016). Gradient boosting minimizes the loss function by sequentially adding weak learners via gradient descent optimization. This approach relies on three fundamental components: a loss function to measure predictive accuracy, a weak learner that may not classify perfectly but is better than random guessing, and an additive model that progressively integrates decision trees (Chen T. et al., 2019).

### 2.5 Development of high resolution AOD

In this study, machine learning models were utilized to merge AOD data obtained from two satellites, Himawari-8 and MAIAC, aiming to achieve AOD products with enhanced spatial and temporal resolutions. The MAIAC data provides high spatial resolution, but was available only once per day. Conversely, the Himawari-8 data provides hourly AOD products, but low spatial resolution of  $5 \times 5 \text{ km}^2$ . To reconcile these differences, the hourly Himawari-8 AOD data from 9:00 to 16:00 was downscaled from  $5 \times 5 \text{ km}^2$  to  $1 \times 1 \text{ km}^2$  using Python programming, aligned with the center point of the MAIAC product.

### 2.6 Prediction of ground-based $\text{PM}_{2.5}$ concentration

Hourly high spatial resolution AODs from two satellites generated from the downscaling method (section 2.5) with meteorological data from WRF model were used to develop models to estimate hourly  $\text{PM}_{2.5}$  concentration during the study period (2018–2021). The developed models were separated by month. Machine learning models explored to predict  $\text{PM}_{2.5}$  concentration in this study were linear regression, decision tree regressor, extra tree regressor, random forest regressor and extreme gradient boosting regressor models. Available data were divided into a 70% training set and a 30% validation set for model development and evaluation, respectively.

#### 2.6.1 Validation of the $\text{PM}_{2.5}$ prediction models

This study utilized machine learning models to estimate  $\text{PM}_{2.5}$  concentrations in the study area during 2018–2021, incorporating inputs from hourly AOD and WRF results. The performance of the  $\text{PM}_{2.5}$  estimation models was assessed by comparing the model

TABLE 2 Statistical methods and criteria for checking model performance.

| Statistics             | Equation  | Criteria     |
|------------------------|---|--------------|
| R-square               | $R^2 = 1 - \frac{\sum (y_{\text{observed}} - y_{\text{predicted}})^2}{\sum (y_{\text{observed}} - y_{\text{mean observed}})^2}$ | Close to one |
| Mean absolute error    | $\text{MAE} = \frac{1}{n} \sum  y_{\text{observed}} - y_{\text{predicted}} $  | Low value    |
| Root mean square error | $\text{RMSE} = \sqrt{\frac{1}{n} \sum (y_{\text{observed}} - y_{\text{predicted}})^2}$  | Low value    |

outputs with the ground-based  $\text{PM}_{2.5}$  concentrations from PCD's monitoring stations using statistical parameters, including r-square, mean absolute error, and root mean square error. The evaluation criteria for each statistical parameter can be found in Table 2.

### 2.7 Developing an hourly $1 \times 1 \text{ km}^2$ $\text{PM}_{2.5}$ concentration for Thailand

The validated models were used with Python programming libraries to estimate hourly  $\text{PM}_{2.5}$  concentration in a  $1 \times 1 \text{ km}^2$  grid over Thailand.

## 3 Results

### 3.1 Data completeness analysis

Data completeness analysis was conducted for AOD data from Himawari-8 and MAIAC, meteorological data from WRF model, and  $\text{PM}_{2.5}$  concentration data from PCD. The results are summarized in Table 3.

From Table 3, AOD data from MAIAC and Himawari were available between 5% and 45% of the total data during 2017–2021 due to cloud cover.  $\text{PM}_{2.5}$  concentration data obtained from air quality monitoring stations is available between 25% and 99% from 69 stations during 2018–2021 due to availability of  $\text{PM}_{2.5}$  monitoring equipment, data screening, potential equipment issues, etc. In contrast, meteorological data from WRF model data show 100% completeness (air pressure, temperature, wind speed, and accumulated rainfall) from 2018 to 2021. Therefore, this study used the data available based on the hours that data were complete to train the  $\text{PM}_{2.5}$  concentration estimation models.

### 3.2 Ground-based $\text{PM}_{2.5}$ monitoring data

The hourly pattern of  $\text{PM}_{2.5}$  concentrations remained remarkably consistent on an annual basis. From December to April of every year, the hourly  $\text{PM}_{2.5}$  concentrations consistently rise to very high levels. These cycles of elevated  $\text{PM}_{2.5}$  concentrations occur consistently on an annual basis, with the specific duration of high concentrations varying slightly from one region to another. Figure 2 shows hourly  $\text{PM}_{2.5}$  concentration monitored by 69 air quality monitoring stations around Thailand (average by region—BMR, central, north, northeast, east and south).

TABLE 3 Completeness of AOD, meteorological and PM<sub>2.5</sub> concentration data.

| Data type                   | Parameters   | Data completeness (%) | Time period |
|-----------------------------|--|-----------------------|-------------|
| MAIAC data                  | AOD  | 13–19                 | 2017–2021   |
| Himawari-8 data             | AOD  | 5–45                  | 2017–2021   |
| WRF data                    | Air pressure at sea level, temperature at ground level, wind speed at 10 m above the ground, precipitation | 100                   | 2018–2021   |
| Air quality monitoring data | PM <sub>2.5</sub> concentration  | 25–99                 | 2018–2021   |

### 3.3 Weather Research and Forecasting Model validation

The WRF data, sourced from the TMD, underwent rigorous validation against hourly average meteorological data from each air quality monitoring station operated by the PCD throughout Thailand from 2018 to 2021. The received WRF data has a spatial resolution of  $6 \times 6 \text{ km}^2$  and a temporal resolution of 1 h. The validation process for TMD-WRF revealed the following results: a correlation coefficient (R) of 0.7 for temperature, 0.5 for wind speed, and 0.1 for pressure and precipitation. While the correlation coefficients for pressure and precipitation were relatively low, our primary focus in this study lies on developing the methodology based on the available data in Thailand. Thus, with improvements in the WRF prediction data provided by the TMD, the estimation of PM<sub>2.5</sub> concentration is expected to enhance significantly.

### 3.4 Hourly $1 \times 1 \text{ km}^2$ AOD

This study employed various machine learning models, including linear regression, decision tree, extra tree regression, and extreme gradient boosting models, to downscaling  $5 \times 5 \text{ km}^2$  AOD measurements from Himawari-8 to  $1 \times 1 \text{ km}^2$  using AOD pattern from MAIAC. The training dataset consisted of approximately 8.5–250 million data points (differences based on data availability in each month), while the testing dataset contained around 3 to 102 million data points. The  $R^2$  values for both training and testing ranged from 0.04 to 0.58 (Figure 3). Notably, the  $R^2$  value was found to be lowest during the rainy seasons (June to August) due to the time gap between Himawari-8 and MAIAC data, which the satellite path time changes during this period. Additionally, number of available data for training and testing during rainy season was lower compared to other seasons. However, the results demonstrated that the three models, excluding linear regression, exhibited similar performance in downscaling the AOD data to  $1 \times 1 \text{ km}^2$ . In this study, the extreme gradient boosting regressor model (XGBRegressor) was chosen to downscale AOD due to its less resource requirement (processing time). Moreover, the XGBRegressor is the ensemble learning technique that was developed based on many models including the decision tree model to help reducing overfitting that can lead to more stable

and accurate predictions. Moreover, XGB has outperformed various statistical models in previous studies (Gupta and Christopher, 2009; Xiao et al., 2018).

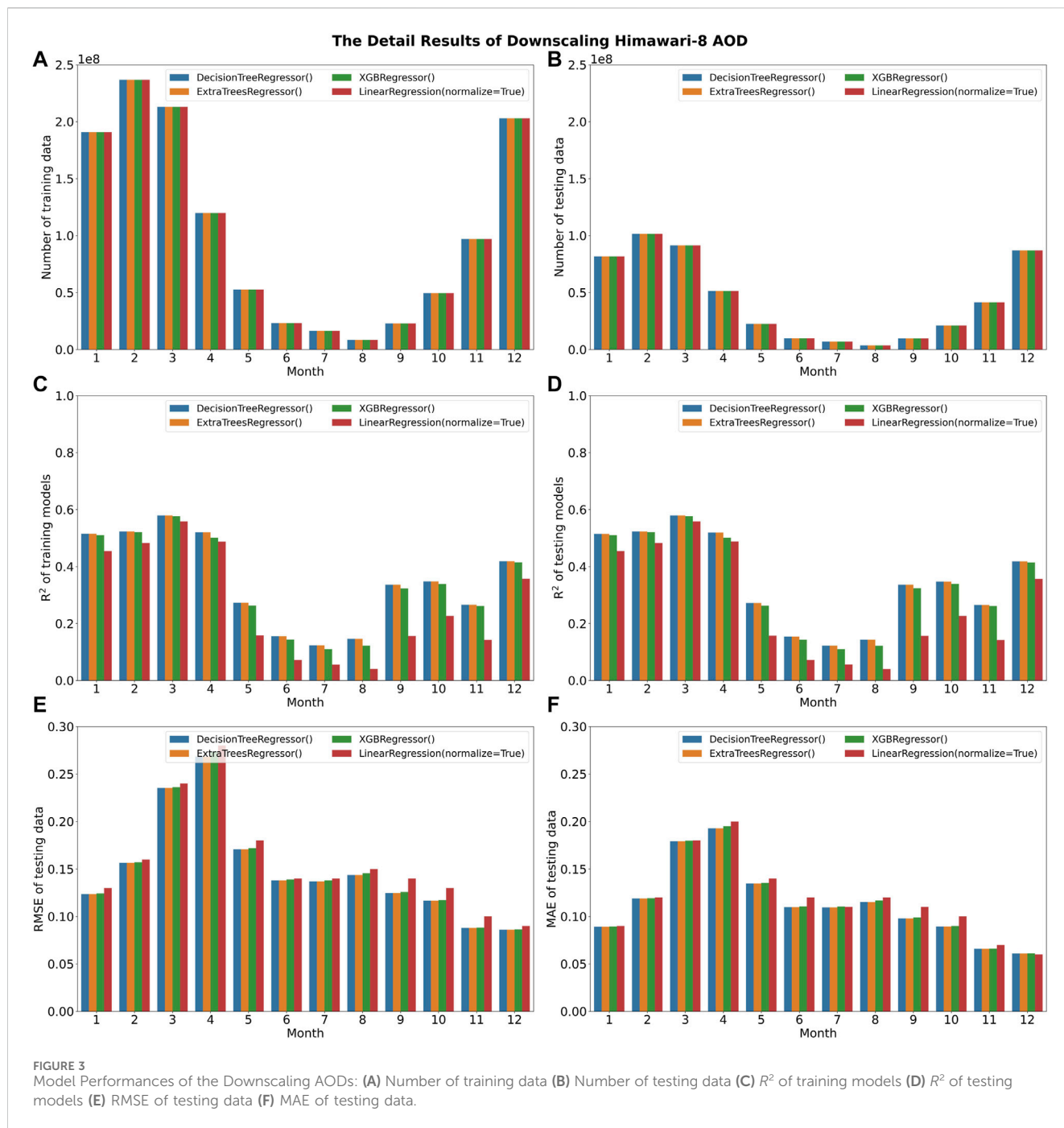
After obtaining the downscaling models, the Python programming was used to downscale AOD, as presented in Figure 4.

In Figure 4, a noteworthy observation is that the AOD values from MAIAC and Himawari-8 exceed those obtained through the downscaled AOD within the middle part of Thailand. This divergence can be attributed to the comprehensive spatiotemporal scope of the model developed for this study, which encompasses data from 2017 to 2021, providing a holistic perspective on AOD dynamics throughout Thailand. A key contributor to the relatively lower AOD values in this specific region is the pronounced presence of very high AOD, primarily due to the influence of anthropogenic aerosols (Luo et al., 2019) where this area is correspond to regions with extensive agricultural land use. This influence is particularly significant during December, coinciding with Thailand's harvest season, marked by various agricultural activities that release elevated aerosols into the atmosphere. Incorporating this information enriches our understanding of the complex interplay of factors affecting AOD, encompassing both natural and human-induced elements.

### 3.5 Hourly PM<sub>2.5</sub> concentration estimation

In this study, Pearson's coefficient was used to assess the correlation among meteorological parameters, downscaled AOD and PM<sub>2.5</sub> concentration from air quality monitoring data. Additionally, RFE was employed for the XGBRegressor and other models to identify key important variables for predicting PM<sub>2.5</sub> concentration.

The analysis using Pearson's coefficient (Figure 5) revealed that the highest correlation with PM<sub>2.5</sub> concentration was found with the downscaled AOD, followed by wind speed, pressure, precipitation, and temperature. However, the RFE analysis indicated that the important variables were the downscaled AOD, wind speed, temperature, pressure, and precipitation, in this order. Interestingly, when the precipitation data was excluded in predicting PM<sub>2.5</sub> concentrations, there was no significant difference in the model performance. Thus, this study decided to use four variables, namely, downscaled AOD (Himawari\_xgb),



wind speed, temperature, and pressure, for predicting  $PM_{2.5}$  concentration.

### 3.6 Performance of the $PM_{2.5}$ estimation model

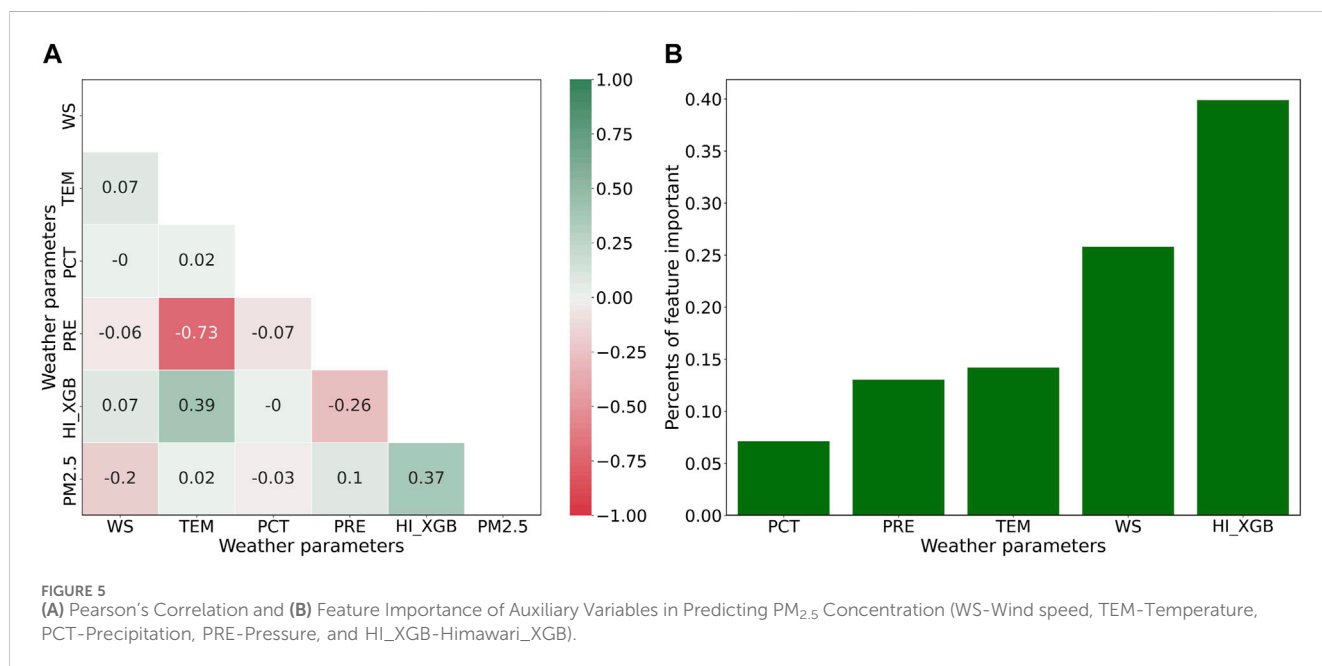
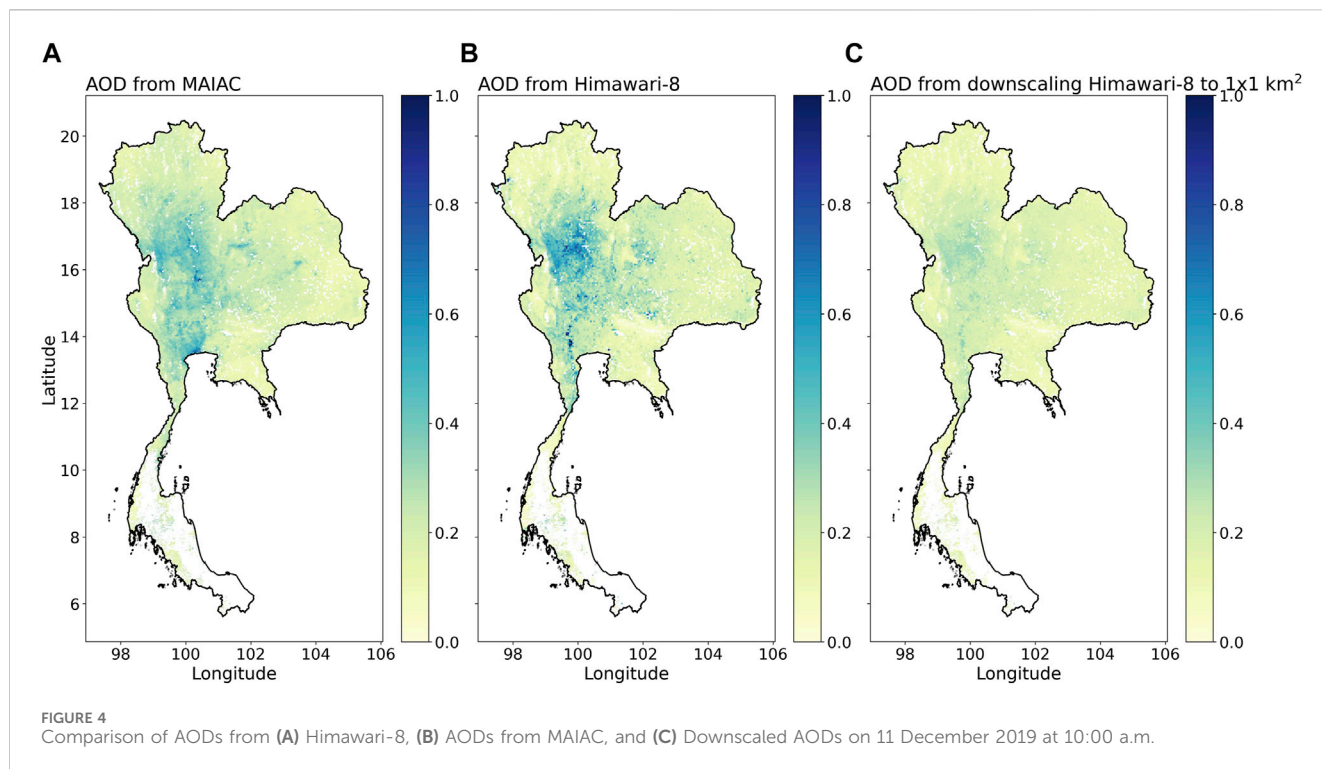
The performance evaluation of monthly  $PM_{2.5}$  concentration estimation models using XGBRegressor revealed that the  $R^2$  of the testing set ranges from 0.20 to 0.91. Moreover, the RMSE indicated that the prediction models had an error range of 7.07–36.93  $\mu\text{g}/\text{m}^3$ . Figure 6

presents monthly performances of  $PM_{2.5}$  estimation model using XGBRegressor.

### 3.7 Comparison of $PM_{2.5}$ concentration from monitoring station and model estimation

$PM_{2.5}$  concentration estimated from models in this study were compared with those from monitoring stations in Thailand on the corresponding grids where the monitoring stations were located. In total, this study compared 7,583 hourly data points from 2018 to 2021. The regression coefficient, based on the

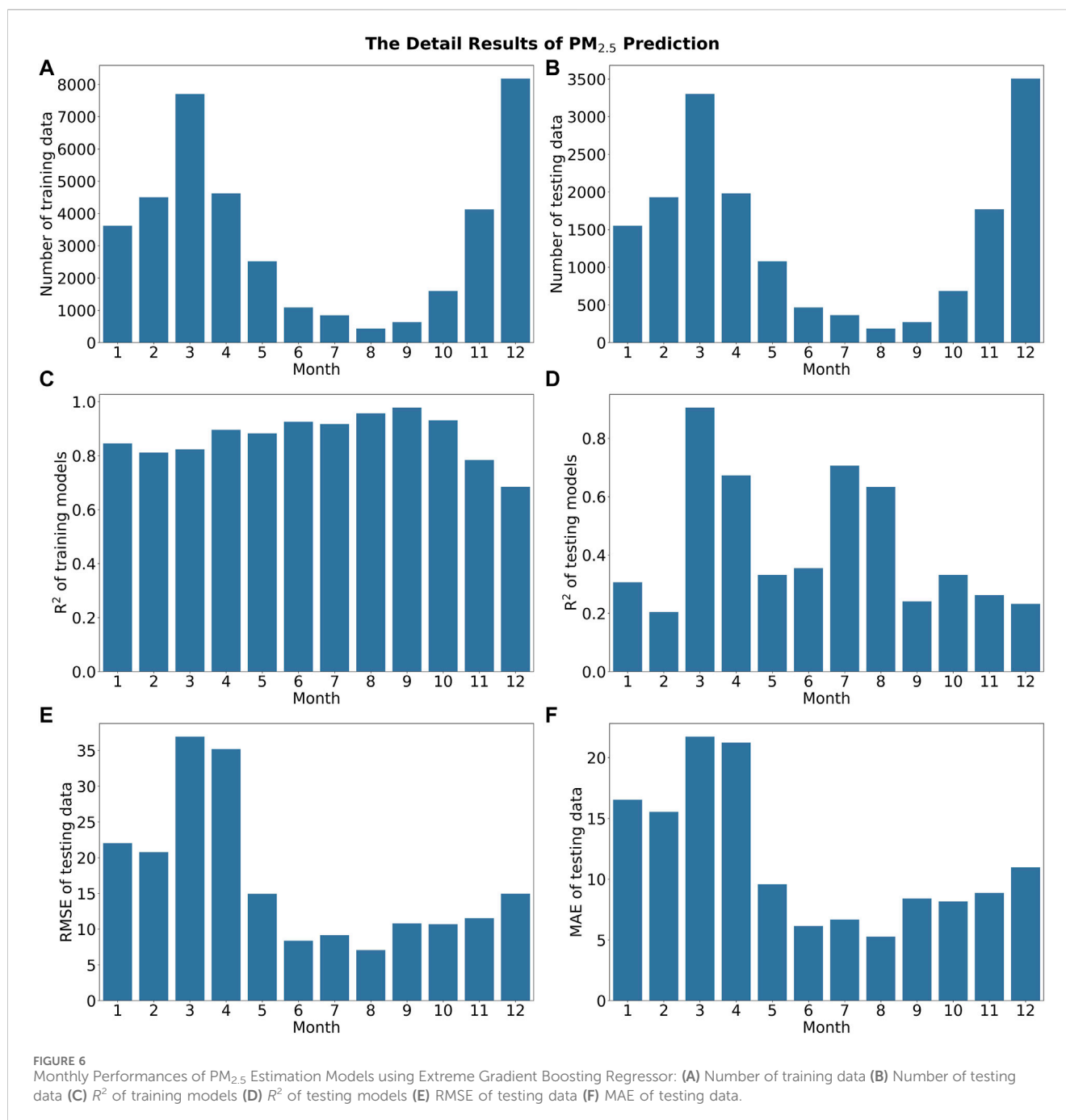




Pearson's coefficient method, was found to be approximately 0.80. However, the results indicated that the estimated  $PM_{2.5}$  concentration were generally higher than the monitored  $PM_{2.5}$  concentration (Figure 7).

Figure 8 displays the timeseries comparison between  $PM_{2.5}$  concentration from monitoring stations and model estimation. It is evident that the estimated values consistently exceeded the monitored values during March to April. This disparity can be attributed to the influence of elevated ambient temperatures, which can affect the AOD

data due to increased photochemical reactions. Additionally, during this period, there was a significant incidence of open burning of agricultural residues in the study area (Figure 9), further contributing to this observed difference. These effects contribute to higher AOD than other periods (Shen et al., 2018; Xian et al., 2022). Moreover, it is important to acknowledge that open burning has a direct impact on  $PM_{2.5}$  concentrations since it also emits gaseous pollutants, such as  $SO_2$ ,  $NO_2$ ,  $CO$ , all of which are correlated with both  $PM_{2.5}$  concentration and AOD (Amnuaylojaroen, 2022).



### 3.8 High spatial and temporal resolution PM<sub>2.5</sub> concentration map for Thailand

PM<sub>2.5</sub> concentration maps of Thailand were generated using a 1 × 1 km<sup>2</sup> grid based on MAIAC cells (625,057 grids covering Thailand). After model estimation was made for each grid, PM<sub>2.5</sub> concentrations were computed by averaging the data on an hourly, daily, monthly, and yearly basis (Figure 10). The spatial distribution of PM<sub>2.5</sub> concentrations revealed that a significant number of grids had missing data due to cloud interference, particularly data in the hourly-average format (Figure 10A). However, by aggregating the

data on a daily, monthly, and yearly basis, more data are available, providing a more comprehensive view of the PM<sub>2.5</sub> concentrations over Thailand.

Figure 10C monthly average (December 2019) PM<sub>2.5</sub> concentration: Significant increase in PM<sub>2.5</sub> concentration during the dry season, particularly in the northern region of Thailand, was observed. This temporal pattern aligns with the findings presented in Figure 11 during December 2019. During the dry season, multiple sources contribute to elevate PM<sub>2.5</sub> levels, including open burning and forest fires. While open burning and forest fire are widely recognized as the primary

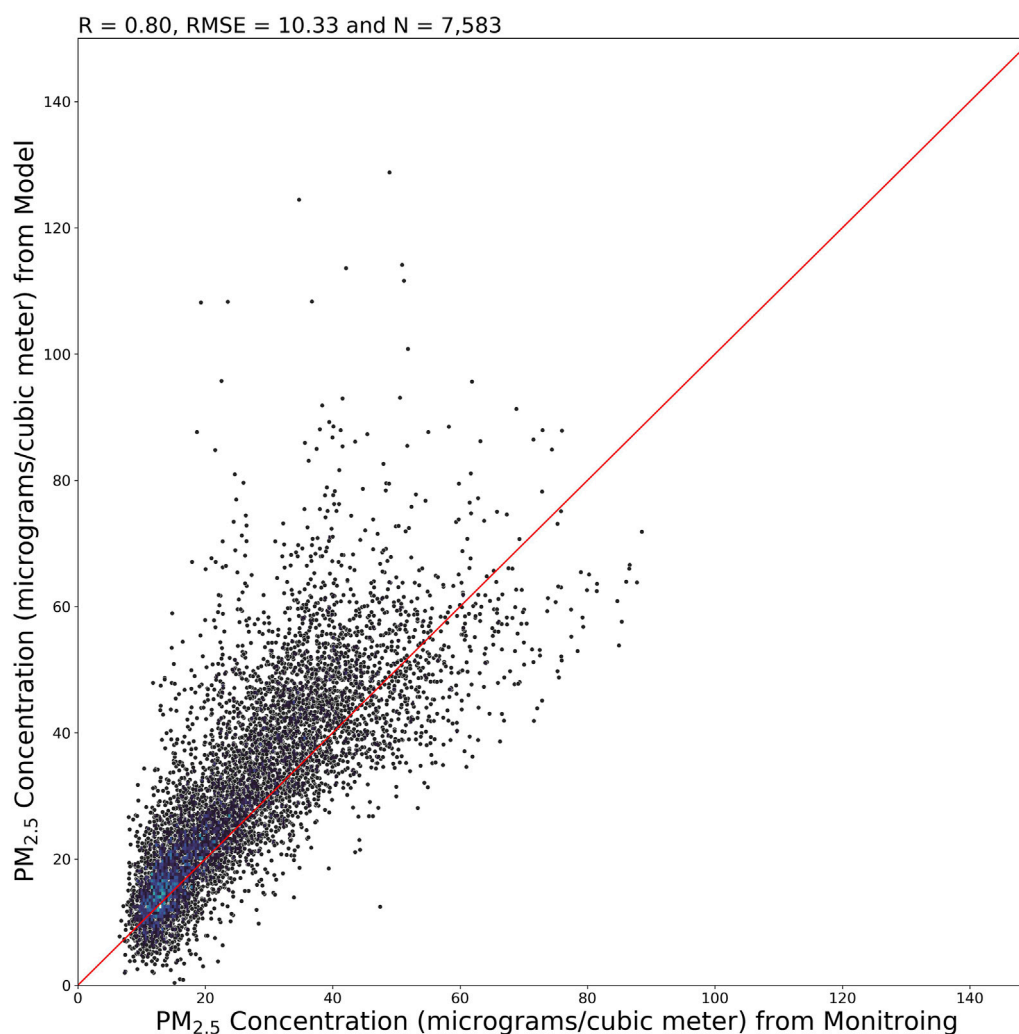


FIGURE 7  
Scatter Plot of Hourly Average  $PM_{2.5}$  Concentrations: Model Estimation vs. Monitoring (2018–2021).

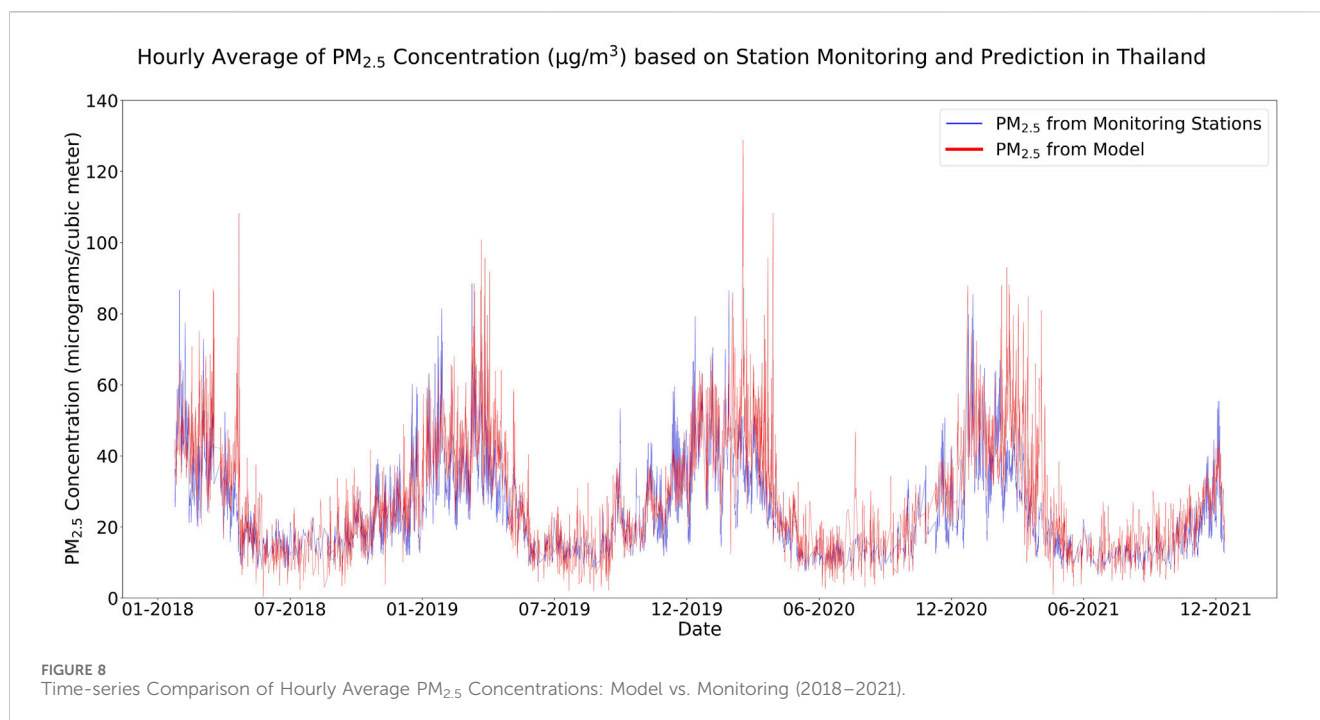
sources to dry-season  $PM_{2.5}$  concentration, it is important to note that traffic and other emissions are consistent year-round sources of  $PM_{2.5}$ . Despite biomass burning accounting for 25%–79% of  $PM_{2.5}$  during this period (Hu et al., 2017), traffic emissions also play a significant role in contributing to haze pollution (Fu et al., 2023).

Figure 10D yearly average of  $PM_{2.5}$  concentration: The northern and western regions of Thailand exhibited higher  $PM_{2.5}$  levels compared to other areas. This observation is consistent with the spatial distribution of active fires in Figure 9, indicating a correlation between high  $PM_{2.5}$  levels and the occurrence of active fire hotspots in the northern and western regions.

## 4 Discussion and conclusion

In this study, relationship between AOD of Himawari-8 and MAIAC, WRF model parameters (wind speed, precipitation,

pressure, and temperature) and  $PM_{2.5}$  concentrations in Thailand during 2018–2021 has been investigated. The hourly  $5 \times 5 \text{ km}^2$  AOD from Himawari-8 was downscaled to  $1 \times 1 \text{ km}^2$  using the AOD distribution pattern of the MAIAC product on the same day. More than 250 million data points for training and testing were employed during the downscaling process to enhance the model's accuracy by XGBoost model. However, our findings revealed that the R-square value of the training model was relatively low during rainy season in Thailand (May to August). This can be attributed to the time gap between Himawari-8 and MAIAC-product passing over the study area. Additionally, the presence of cloud interference during the rainy season contributed to the availability of the data during this period. Nonetheless, it is important to note that our study utilized training and testing data encompassing all regions of Thailand. One potential factor contributing to the observed lower accuracy may be the presence of anthropogenic aerosols concentrated in urban and agricultural areas, exerting a substantial influence on the scattering and absorption of solar radiation (Luo et al., 2019). Nevertheless, our study has developed a downscaling AOD methodology that

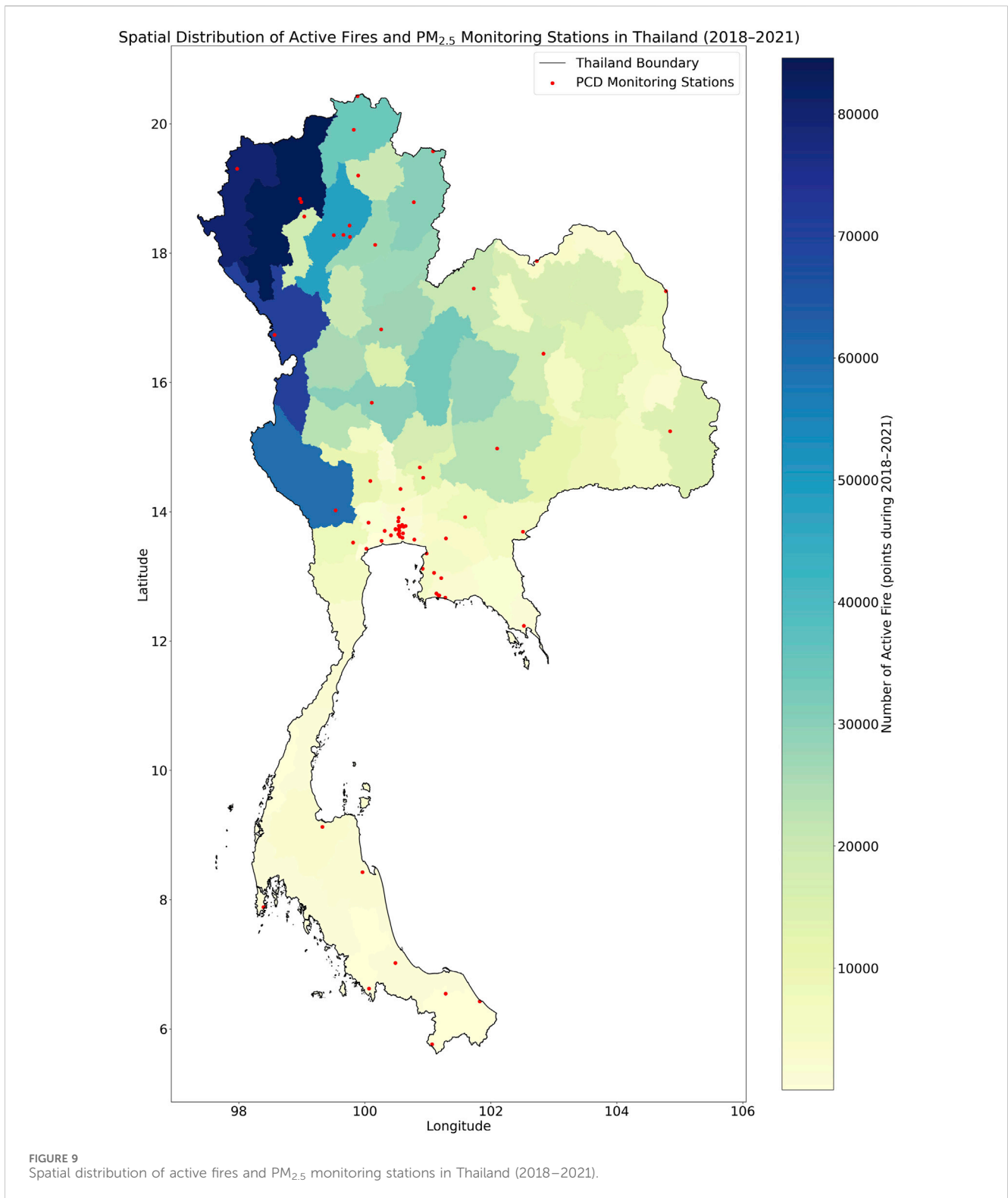


optimizes the impact of anthropogenic factors and significantly improves spatial and temporal resolution. Furthermore, AOD accuracies are influenced by various factors, including topography, seasonally changing surface characterization (e.g., vegetation cover, surface reflectance), aerosol type, size distribution, and vertical distribution of aerosols in the atmospheric column, along with sensor and solar viewing geometry (Gupta et al., 2021). Despite the multifaceted nature of these influences, this study specifically examined AOD as the most critical factor in predicting PM<sub>2.5</sub> concentrations. Consequently, AOD emerges as a primary source of error in PM<sub>2.5</sub> concentration predictions, particularly as we aimed to estimate concentrations across all regions in Thailand. Many previous studies have revealed that regression model with the MODIS AOD measurement can be used as a predictor variable to estimate the spatial ground-level PM<sub>2.5</sub> concentrations, taking into account many potential confounders (Hu et al., 2014; Zheng et al., 2016; Wang et al., 2019; Guo et al., 2021). This method would benefit for exposure assessment for epidemiological research, especially in the areas with no monitoring station network that is commonly used to explore the association of PM<sub>2.5</sub> with morbidity and mortality (Peng-in et al., 2022). The prediction of PM<sub>2.5</sub> concentrations commenced by identifying relevant variables through the Pearson's correlation and recursive feature elimination based on the XGBoost model. Following the testing phase, the downscaled AOD, wind speed, temperature, and pressure were selected as the predicting variables for hourly PM<sub>2.5</sub> concentration estimation. Utilizing the downscaled AOD with meteorological data, this study predicted PM<sub>2.5</sub> concentrations over Thailand, encompassing over 600,000 grids at a 1 × 1 km<sup>2</sup> resolution.

The PM<sub>2.5</sub> estimation models (developed for each month in Thailand) yielded a diverse range of R<sup>2</sup> results for the training dataset, but, in overall, exhibited consistently high accuracy. The R<sup>2</sup> values for the testing dataset varied from 0.20 to 0.91, with

corresponding root mean square errors ranging from 7.07 to 36.93 µg/m<sup>3</sup>. Then, the PM<sub>2.5</sub> concentration from the model estimation were compared with those from ambient air monitoring stations at the corresponding grid and time. The model's performance shows R<sup>2</sup> value of 0.64 and an RMSE of 10.33 µg/m<sup>3</sup>. The accuracy result of estimating PM<sub>2.5</sub> concentration was similar to the previous studies in Thailand and other countries (Ma et al., 2014; Zheng et al., 2016; Guo et al., 2021; Xu et al., 2021; Buya et al., 2023). However, the estimated PM<sub>2.5</sub> concentration from March to April was higher than the monitoring data due to the significant of open burning of agricultural residues in the study area. These effects contribute to higher AOD than other periods and are related to PM<sub>2.5</sub>, AOD and other pollutants (Amnuaylojaroen, 2022). Thus, the developed models can be used to estimate PM<sub>2.5</sub> concentration on an hourly basis with a resolution of 1 × 1 km<sup>2</sup> covering Thailand.

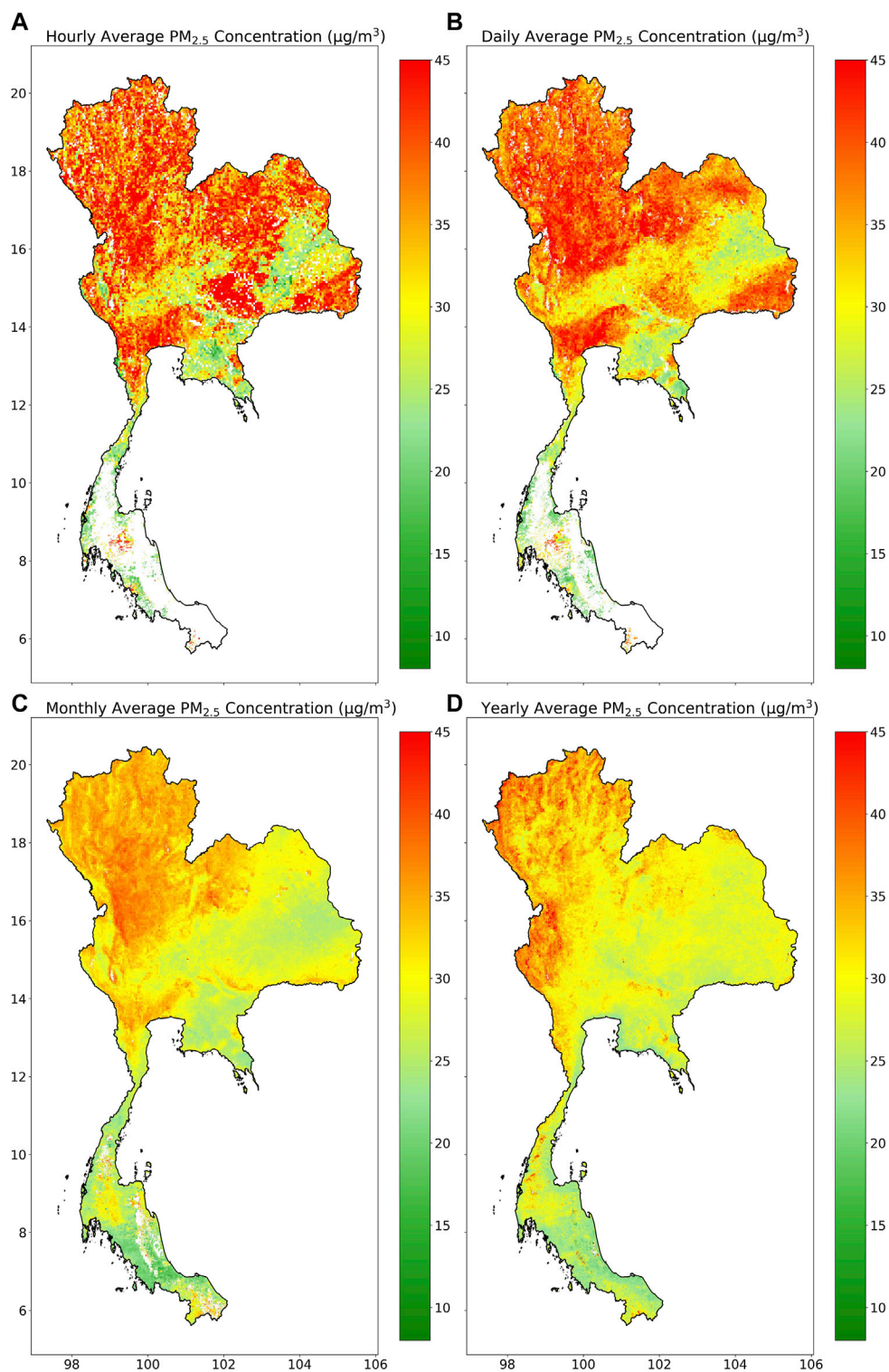
The accuracy of our PM<sub>2.5</sub> estimations was subject to the influence of several pivotal factors. The AOD-PM<sub>2.5</sub> relationship proved sensitive to variables, such as aerosol concentrations, relative humidity, cloud cover, boundary layer height (Gupta and Christopher, 2009; Chitranshi et al., 2015). However, our analysis encountered limitations, notably the restricted scope of meteorological data from the available WRF model output from TMD. Additionally, our reliance on an AOD inversion algorithm introduced limitations that affected our ability to estimate PM<sub>2.5</sub> concentrations across all regions. Notably, the current AOD inversion algorithm struggles with cloud recognition, often mistaking haze for clouds, potentially resulting in the absence of aerosol products under heavy pollution conditions (Bilal et al., 2017). As a potential solution to these limitations, we suggest exploring multisource data inversion algorithms to enhance spatial coverage (Shi et al., 2018). The complexity of PM<sub>2.5</sub> prediction is further compounded by the influence of various variables, including weather conditions and environmental seasonality (Amnuaylojaroen, 2022). In particular,



Buya et al. (2023) enhanced the accuracy of PM<sub>2.5</sub> estimation by incorporating additional factors, such as Normalized Difference Vegetation Index (NDVI), Elevation, Week of the Year, and year in Thailand. The resulting model, based on the XGBoost algorithm, achieved an R-squared value of 0.45 and an RMSE of 12.12 µg/m<sup>3</sup>. Furthermore, Luo et al. (2019) conducted a study that explored the

influence of surface solar radiation on AOD. This relationship arises from solar radiation affecting the scattering and absorption of aerosols, which, in turn, impacts AOD. Higher solar radiation levels can intensify photochemical reactions that result in the formation of secondary aerosols, subsequently elevating AOD (Luo et al., 2019).

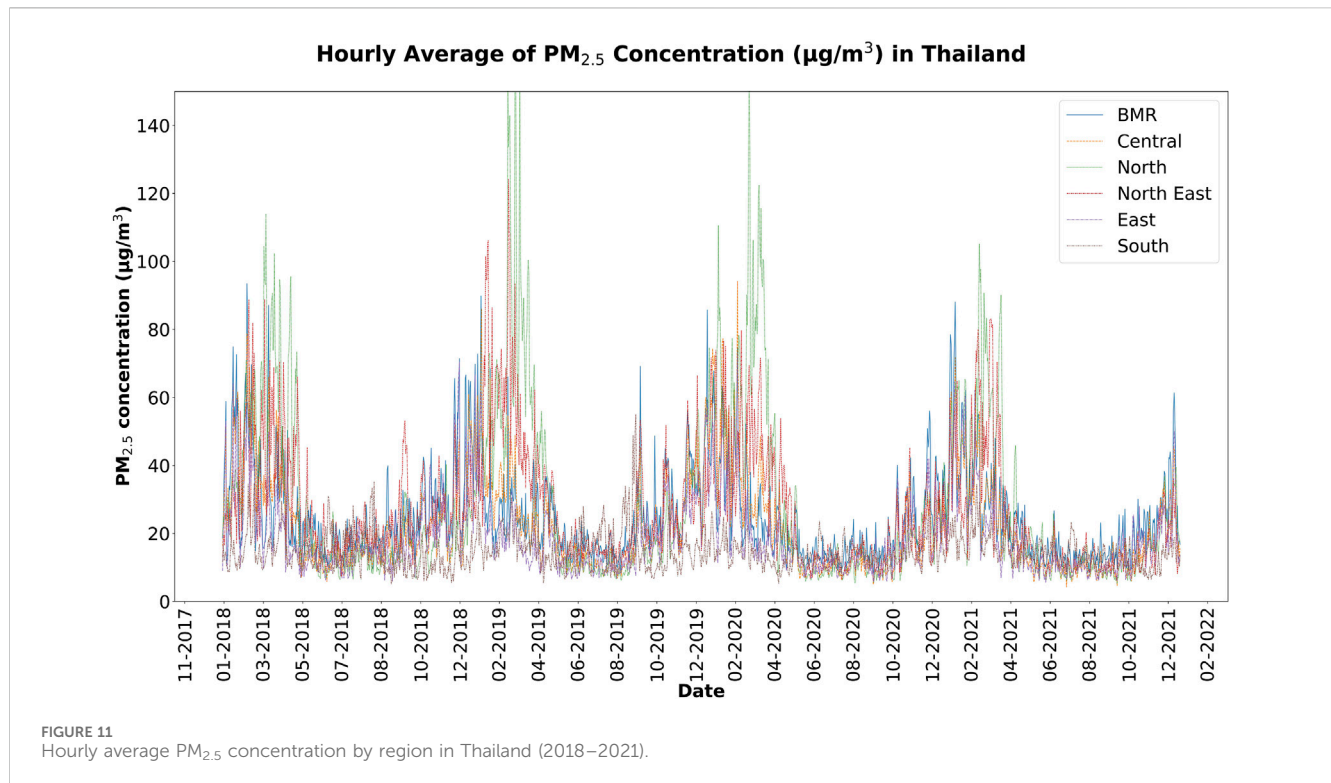




**FIGURE 10** Spatial distribution ( $1 \times 1 \text{ km}^2$ ) of  $\text{PM}_{2.5}$  concentration ( $\mu\text{g}/\text{m}^3$ ) in Thailand from model estimation: (A) hourly average (11 December 2019 at 10:00am), (B) daily average (11 December 2019), (C) monthly average (December 2019) and (D) yearly average (year 2019).

Despite the valuable insights gained from our study, it is crucial to acknowledge that we primarily relied on AOD data for  $\text{PM}_{2.5}$  estimation, introducing variability across different regions and potentially affecting the overall accuracy of our findings, given

the utilization of two AOD products from Himawari-8 and MAIAC. In future research, it is crucial to consider incorporating additional factors into AOD downscaling, such as land use data, solar radiation, week of the year.



## Data availability statement

The original contributions presented in the study are included in the article/Supplementary material, further inquiries can be directed to the corresponding author.

## Author contributions

PP: Conceptualization, Methodology, Visualization, Formal Analysis, Writing–original draft. EW: Conceptualization, Methodology, Visualization, Project administration, Writing–review and editing. PK: Conceptualization, Writing–review and editing. SV: Writing–review and editing. WX: Writing–review and editing. TN: Writing–review and editing.

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## Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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