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RECEIVED 24 November 2022

ACCEPTED 20 July 2023

PUBLISHED 11 August 2023

CITATION

Rodríguez-Hernández NS, Arango M,
Moreno-Conn LM, Arguello JO,
Bernal-Riobo JH and Pérez-López O
(2023), Grassland management effect on
ecosystem services in the livestock
system in an oxisol from the Eastern high
plains of Colombia.
Front. Environ. Sci. 11:1107466.
doi: 10.3389/fenvs.2023.1107466

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Grassland management effect on ecosystem services in the livestock system in an oxisol from the Eastern high plains of Colombia

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Colombia has a livestock population of approximately 28.2 million heads, of which 20.4% is found in the eastern high plains of the Orinoquia region. The extensive beef cattle system predominates, whose diet is based on native and introduced pastures of the genus *Urochloa* sp. which is deficiently managed, affecting productive and reproductive indexes due to the low adoption of technology. We evaluated potential sustainable intensive systems for cattle production, which contribute to maintaining the provision of environmental services. Between 2011 and 2015 in Agrosavia's Research Center at Carimagua (4° 37'N and 71° 19' W) environmental and animal production variables were monitored in the following systems: a) degraded pastures recovered with tillage and fertilization, b) annual crop rotation with pastures, c) forest arrangements in strips, and d) forest remnants in perimeter areas. Productive and reproductive variables were determined in animals such as weight gain, calving interval, among others while the pasture/crop productive variables included yield and forage quality. Regarding soil ecosystem services (ES) the macrofauna biodiversity, biogeochemical cycles, soil physical and chemical variables were considered. Estimation of indicators was carried out through principal components analysis for soil physical, chemical and macrofauna variables to extract the two main components that explain the variance. For climate regulation of ES, measurement of soil organic carbon (SOC) storage at a depth of 20 cm and the annual accumulated greenhouse gas (GHG) emissions were included. The systems that involved lime amendments and fertilizers increased the value in the year of application. Values in the water regulation indicator did not show significant differences among the options implemented during the years. Edaphic macrofauna biodiversity indicator value was sensitive to changes in management practices, with termites being the group with highest abundance. The indicator related to SOC was higher in the forest area compared to the pasture-crop system. Systems under the agroforestry schemes which integrated various practices such as managed and recovered pastures, crop-pasture rotation, patched areas and tree strips contributed to maintain or improve the ES, although each one of the strategies proposed provided improvement in at least one ES.

KEYWORDS

biodiversity, fertility, water regulation, livestock, Eastern high plains

1 Introduction

In Colombia, since the 18th century, multiple factors such as demographic growth, migration due to violence, and expansion of the agricultural frontier have interacted and even overlapped to forge the transformation of natural ecosystems (Etter et al., 2006). Nearly a third of natural forest areas were changed to productive systems, especially the livestock system, giving way to grazing areas and colonization of lowlands (Etter et al., 2008). Livestock farming is an economic activity that can secure land ownership and helps obtain higher profits in terms of land value, with low investment (Clerici et al., 2020). The transformation of forests to pastureland is one of the main drivers of deforestation (Armenteras et al., 2017; CDP Report, 2018). Extensive cattle raising is a sector that occupies 38% of the Colombian territory. It represents structural barriers to rural development due to the inefficient use of natural resources, limited access to financial resources and technology, low inputs and, therefore, low yields. Reducing the area allocated to livestock farming is a priority in order to contribute to deforestation reduction and forest preservation at the national scale (Correa and Ruiz, 2006; Nepstad et al., 2013). The high plains of the Meta department is located in the Eastern plains of Colombia. There is a cattle population of 5.75 million heads, which represents 20.4% of the national total and covers about 4.6 Mha (Paredes Vega and Hernández Leal, 2013). The high plains of the Meta department is a highly evolved ecosystem. The soils are primarily of the orders oxisols and ultisols, with low cation exchange capacity, high acidity, and Al^{3+} toxicity. From the physical point of view, these soils are fragile and susceptible to compaction, erosion, and surface sealing, with low infiltration and water storage capacity and high bulk density (FAO, 1965). These conditions contrast with the climatic conditions that favor vegetation development, i.e., annual average temperature of 26°C, annual average rainfall of 2,300 mm, a dry season with an average monthly rainfall of less than 100 mm from December to March, and a rainy season with an average monthly rainfall of more than 100 mm from April to November (Decaëns et al., 2003). The current livestock management system is being changed from extensive to intensive livestock farming through a process of sustainable intensification to reduce the impact on the environment. According to Pérez-Lopez et al. (2019), this region is dominated by cattle breeding with birth rates below 55%, mortality in young cattle of 10 %, weaning weight less than 160 kg, and age at first calving greater than 40 months. The cattle feed source is native and introduced grasses, mainly of the *Urochloa* genus; about 70% of the grazing areas have yield less than 700 kg ha⁻¹ of dry matter (DM) with a high neutral detergent fiber (NDF) content of more than 65%, protein concentration of less than 7%, low digestibility (less than 60%), and mineral imbalance, factors that affect the voluntary consumption of cattle and limit production quality.

Unplanned changes in management and limited knowledge of landscape functioning of water, carbon, and nutrient cycling and biodiversity have repercussions on ecosystem services (ES). Evaluation of the ES is required to ensure the viability of

productive systems of the region and thus contribute to the creation of sustainable development opportunities in what is considered the last Colombian agricultural frontier (Rodríguez et al., 2009). The complexity of the interaction between the ecosystem and the productive systems is a challenge for sustainable development.

The objective of this study was to investigate the impact of current and proposed management strategies such as the improvement of degraded grasslands, crop–pasture rotation, and tree establishment on indicators of ES provided by the soil such as fertility, climate control (greenhouse gases (GHGs) and carbon storage), biodiversity (edaphic macrofauna), and water function.

2 Materials and methods

2.1 Location characteristics

The high plain subregion of Colombia is between 100 and 300 m above sea level. Precipitation ranges between 1,500 and 2,500 mm/year, distributed between April and November, with an average temperature of 27°C and maximum and minimum values of 36 °C and 21°C, respectively (Amézquita et al., 2013; Rivera et al., 2013).

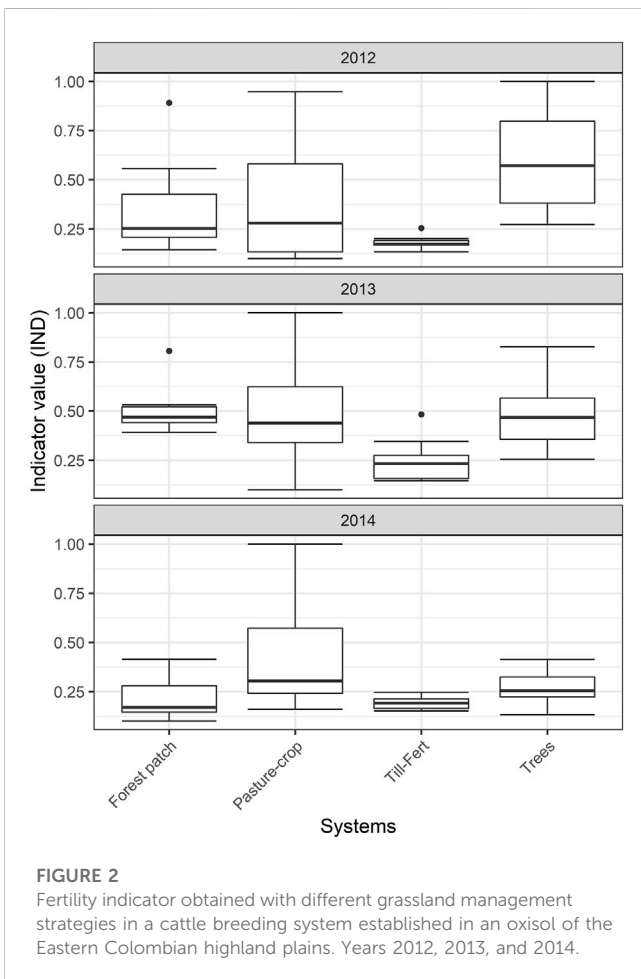
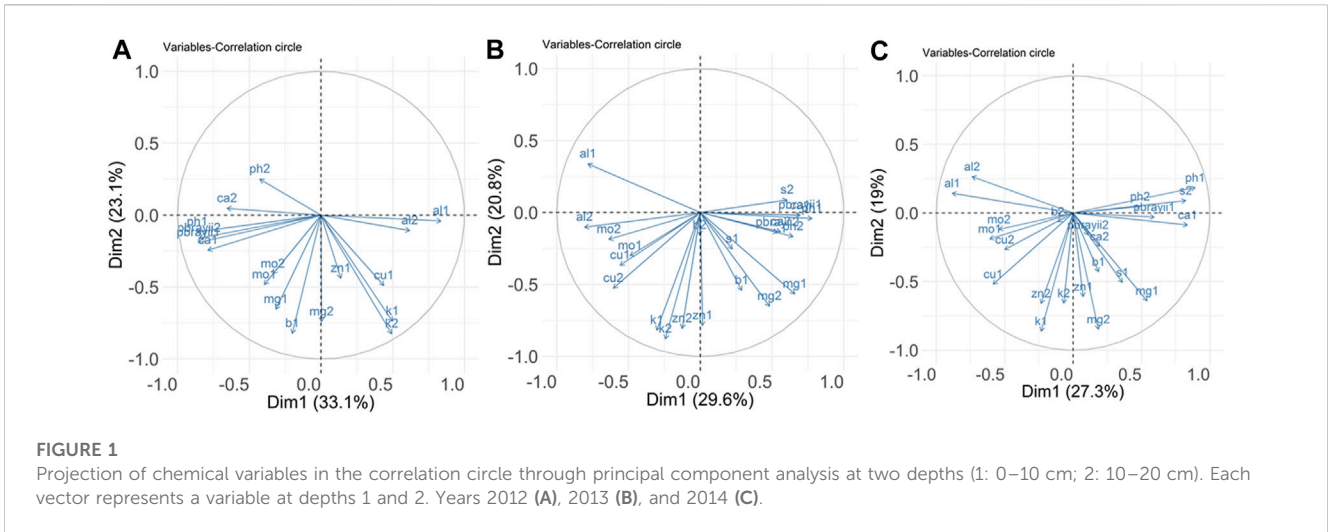
The soils in this area are classified by the US Soil Taxonomy as oxisols and are relatively well-drained, with very low fertility and high aluminum saturation. The soil pH ranges from 4.0 to 4.5, with very low organic matter concentration (1.5%–2.0%), phosphorus (range of 1–2 mg kg⁻¹ Bray II), and exchangeable bases (Ca, Mg, and K), the latter being usually less than 0.1 cmolc kg⁻¹ and exchangeable aluminum saturations between 60% and 90% (Pérez-Lopez et al., 2019). From the physical point of view, the soils are naturally dense, with the bulk density ranging from 1.4 to 1.65 g cm⁻³. The texture ranges from clay loam to sandy textures; they are fragile to mechanization processes, with high susceptibility to compaction, erosion, and surface sealing, and have low infiltration and water storage capacity (Pérez-Lopez et al., 2019).

The study was conducted between 2012 and 2014 at AGROSAVIA's Carimagua Research Center of the Colombian Agricultural Research Corporation in the municipality of Puerto Gaitán (4° 19'01"N; 72° 04'59"W) in the Meta (Colombia). A 70-ha plot with *Urochloa humidicola* was selected with indicators of moderate degradation, such as low forage yield of less than 700 kg DM ha⁻¹year⁻¹, soil compaction, and the presence of termite mounds. A general review of the plot showed the need for pasture renovation and infrastructure adaptation to ensure enhanced management for breeding cattle production. The management strategies evaluated are given as follows:

2.2 Management strategies

2.2.1 Till–fert

Renovation of the *U. humidicola* pastures. Soil cultivation consisted of a harrow pass or a chisel pass. Fertilization and amendments consisted of dolomitic lime (26% Ca and 10% Mg)



at a rate of 50 kg ha⁻¹, phosphate rock (10% P and 25% Ca) at a rate of 250 kg ha⁻¹, Paz del Río[®] Thomas phosphate fertilizer (4% P, 34% Ca, and 1% Mg) at a rate of 250 kg ha⁻¹, and agricultural gypsum (19% Ca and 16% S) at a rate of 150 kg ha⁻¹. Maintenance fertilization varied with the condition of the pasture, with frequencies of 1–2 years for the application of a mixture of urea

at a rate of 46 and 69 kg N ha⁻¹, potassium chloride (KCl) at a rate between 15 and 30 kg K₂O ha⁻¹, DAP (18% N and 20% P) at a rate between 50 and 100 kg ha⁻¹, and Sulcamag (double sulfate of calcium and magnesium, enriched with phosphorus, 18% Ca, 9.6% Mg, and 9% S) at a rate of 50 kg ha⁻¹ (Figure 1; Supplementary Material).

2.2.2 Pasture–crops

Rotation of annual crops (maize, sorghum, and soybean) and grasses (*Urochloa brizantha* cv. Toledo + *U. humidicola*) was applied for silage production and renewal of degraded *U. humidicola* grasslands. Soil tillage consisted of two harrow passes and one rigid chisel pass. A mixture of dolomitic lime (26% Ca and 10% Mg) at a rate of 1,500 kg ha⁻¹, Paz del Río[®] Thomas phosphate fertilizer (4% P, 34% Ca, and 1% Mg) at a rate of 500 kg ha⁻¹, phosphate rock (10% P and 25% Ca) at a rate of 500 kg ha⁻¹, and agricultural gypsum (19% Ca and 16% S) at a rate of 300 kg ha⁻¹ was applied 40–50 days before planting. Two maize hybrids were used: Pioneer 30F35H and P3862. Fertilization was divided into three periods: at planting, 200 kg ha⁻¹ of diammonium phosphate (DAP) (N: P₂O₅ 18: 46) at a rate of 20 kg ha⁻¹ of a source of minor elements B, Zn, Cu, and S (Borozinco[®]), 100 kg ha⁻¹ of a source of Ca and Mg (Sulcamag), and 45 kg K₂O ha⁻¹ as KCl were applied. At 15 days after sowing (DAS), 46 kg N ha⁻¹ as urea and 45 kg K₂O ha⁻¹ as KCl were applied. At 30 DAS, 23 kg N ha⁻¹ as urea was applied. For sorghum crop Corpoica variety JTT18, at planting time, the same doses and inputs were applied as for corn, except for KCl, the dose of which was 50 kg ha⁻¹; at 15 DAS, 46 kg N ha⁻¹ as urea and 50 kg ha⁻¹ of KCl were applied. At 30 DAS, 46 kg N ha⁻¹ as urea was applied. In the soybean crop varieties Panorama 29, Corpoica Taluma 5, and Sabana 7, fertilization consisted of application of a mixture of 75 kg ha⁻¹ of Sulcamag, 200 kg ha⁻¹ of triple superphosphate, 60 kg K₂O ha⁻¹ as KCl, and 20 kg ha⁻¹ of Borozinco[®] at sowing and at 15 DAS, 23 kg N ha⁻¹ as urea and 30 kg K₂O ha⁻¹ as KCl (Figure 2; Supplementary Material).

Harvesting of the maize–soybean or sorghum–soybean crops was carried out simultaneously at 75 DAS with a furrow harvester. A silo press machine was used to preserve the ensiled material in plastic bags to be used for livestock feed in

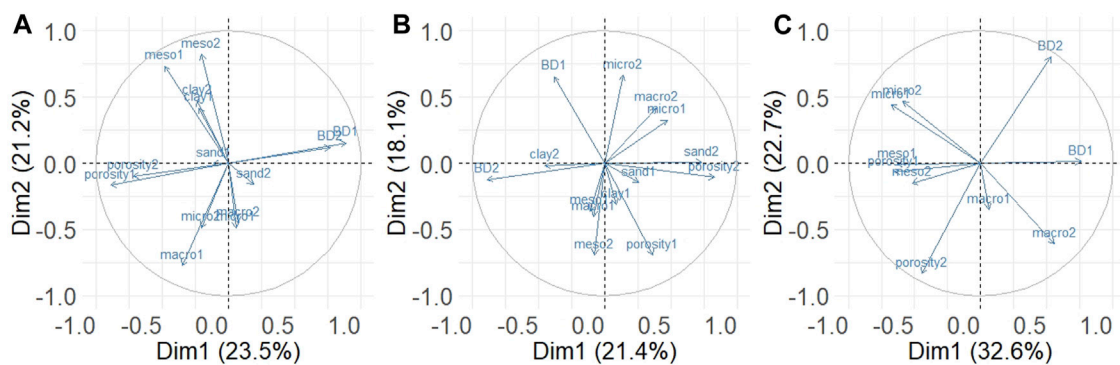


FIGURE 3

Projection of physical variables in the correlation circle through principal component analysis at two depths (1: 0–10 cm; 2: 10–20 cm). Each vector represents a variable at depths 1 and 2. Years 2012 (A), 2013 (B), and 2014 (C).

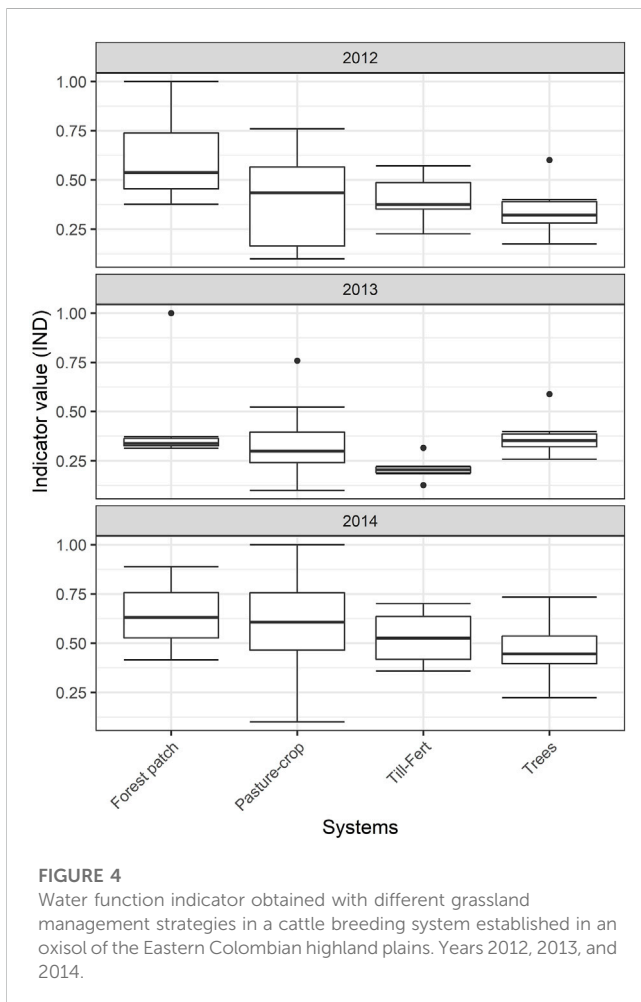


FIGURE 4

Water function indicator obtained with different grassland management strategies in a cattle breeding system established in an oxisol of the Eastern Colombian highland plains. Years 2012, 2013, and 2014.

the dry season. Grazing for cattle was resumed 1–2 months after forage crop harvest, depending on the recovery or pasture development.

2.2.3 Trees

Trees were planted in strips (13 m × 600 m each), with species such as *Piptadenia opacifolia* (yopo), *Acacia mangium*, and

Cassia grandis (Figure 3; Supplementary Material). The seedlings were planted in alternating blocks of 12 trees per species in a staggered arrangement at 7 m between plants and 5 m between rows. Prior to planting, a harrow pass was made, and a mixture of dolomitic lime (26% Ca and 10% Mg) at a rate of 1,000 kg ha⁻¹, phosphate rock (10% P and 25% Ca) at a rate of 250 kg ha⁻¹, Paz del Rio[®] Thomas phosphate fertilizer (4% P, 34% Ca, and 1% Mg) at a rate of 250 kg ha⁻¹, and agricultural gypsum (19% Ca and 16% S) at a rate of 250 kg ha⁻¹ was applied, followed by a chisel pass and a polishing rake pass for amendment. The planting was done with seedlings with a height of 30–45 cm at the definitive site.

2.2.4 Forest patch

Areas of protected and conserved forest relicts through natural regeneration were monitored on the periphery of pastures. They were selected due to their contribution in terms of biodiversity, shade, habitat, and biological corridors for endemic species, among other ES (Figure 4; Supplementary Material).

2.3 Sampling

2.3.1 Soil chemical variables

During the study period, for each management strategy, four disturbed composite soil samples were taken annually at two depths: 0–10 cm and 10–20 cm for quantification of exchangeable bases (Ca, Mg, and K) and exchangeable Al and protons (H), phosphorus (P), sulfur (S), copper (Co), molybdenum (Mo), manganese (Mn), iron (Fe), and boron (B), organic matter, cation exchange capacity, and pH following the methodologies described in Table 1.

2.3.2 Soil physical variables

Soil sampling for physical analysis was performed every year, where four undisturbed soil samples were taken in metal cylinders of fixed volume (100–300 cm³) at two depths (0–10 cm and 10–20 cm). Soil samples were analyzed for texture, bulk density, particle density, total porosity, pore size distribution, volumetric moisture, and retention curves at two points (0.1 and 1.5 bar) following the methodologies cited in Table 2.

TABLE 1 Analytical methods used for the determination of soil chemical properties.

Properties	Analytical methods	References
Organic carbon	Colorimetry	Walkley and Black (1934)
Exchangeable bases: calcium (Ca), magnesium (Mg), potassium (K)—cmol kg ⁻¹	Atomic absorption spectroscopy	Instituto Geográfico Agustín Codazzi – IGAC (2006)
Exchangeable acids: aluminum (Al) and hydrogenum (H)—cmol kg ⁻¹	Volumetry	Instituto Geográfico Agustín Codazzi – IGAC (2006)
pH	Potentiometry	Instituto Geográfico Agustín Codazzi – IGAC (2006)
Phosphorus (P)—mg kg ⁻¹	Colorimetry	
Sulfur (S)—mg kg ⁻¹	Turbidimetry	Instituto Geográfico Agustín Codazzi – IGAC (2006)
Microelements: copper (Co), molybdenum (Mo), manganese (Mn), iron (Fe), and boron (B)—mg kg ⁻¹	Atomic absorption spectroscopy Colorimetry	Instituto Geográfico Agustín Codazzi – IGAC (2006)
	Turbidimetry	

TABLE 2 Analytical methods used for the determination of soil physical properties.

Property	Method	References
Texture	Bouyoucos	Instituto Geográfico Agustín Codazzi – IGAC (2006)
Real density	Pycnometer	Instituto Geográfico Agustín Codazzi – IGAC (2006)
Bulk density	Cylinder of known volume (5 cm diameter and 5 cm height)	Instituto Geográfico Agustín Codazzi – IGAC (2006)
Humidity retention curve	Pressure plates	Instituto Geográfico Agustín Codazzi – IGAC (2006)
Total porosity	Calculation from density	Ditzler et al. (2017)
Size and pore distribution	Calculations from the hydrophysical curve	Forsythe, W. (1985)
Resistance to penetration	Penetrologger	Elaoud et al. (2017)
	Eijkelkamp	
Soil moisture content	Gravimetric	Instituto Geográfico Agustín Codazzi – IGAC (2006)
Available water capacity	Calculations based on moisture retention: water between tensions of 10 kPa and 1,500 kPa	Ditzler et al. (2017)
Structural stability (DPM)	Yoder	Yoder (1936)
		SCCS-Sociedad Colombiana de la Ciencia del Suelo (2013)

2.3.3 Soil macrofauna

Edaphic macroinvertebrates were determined annually using the Tropical Soil Biology and Fertility (TSBF) method (Anderson and Ingram, 1993). A soil section of 25 × 25 × 20 cm was extracted at each depth (0–10 and 10–20 cm). Organisms were separated manually and preserved in 70% alcohol. In the case of earthworms, they were preserved in 4% formaldehyde.

Counting and identification to the order level were made by observation using a stereomicroscope in the soil biology laboratory at the International Center for Tropical Agriculture (CIAT) to determine richness, abundance, Simpson's diversity index, and the dominance index:

$$Ds = 1 - \sum \frac{(ni(ni - 1))}{(N(N - 1))} \quad (1)$$

where ni is the proportion of the number of individuals of species i with respect to the total number of individuals (N).

2.4 GHG fluxes and soil carbon (C) storage

A closed-chamber technique (CCT) was used to determine the net GHG fluxes: methane (CH₄), nitrous oxide (N₂O), and carbon dioxide (CO₂). Sampling was performed monthly and always in the morning hours in each system during 2013 and 2014 for a total of 18 samplings. Three rings were inserted into the soil in each plot and remained throughout the study period. At the time of sampling, the chamber lid was placed, and 20 mL air sample was taken at 0, 20, and 40 min after placing the lid. A 15-mL sample was dispensed in

vacuum tubes and sent to the greenhouse gas laboratory at CIAT. Concentrations of CO₂, N₂O, and CH₄ taken in 2013 were determined in a Shimadzu GC-14A gas chromatograph, equipped with a flame ionization detector for methane analysis and an electron capture detector for nitrous oxide analysis. For the determination of CO₂ concentrations, a Qubit Systems S151 gas analyzer with infrared detection (IRGA) was used. In 2014, a Shimadzu GC-2014 was used that read all three gases simultaneously.

The fluxes were calculated following the GRACEnet methodology (Parkin and Venterea, 2010). Concentration in ppm (v/v) was transformed into mass units according to the ideal gas equation. Once the mass was calculated, the flux was estimated using the HMR function (Pedersen et al., 2010; Pedersen A. R., 2015) by the R 3.2.1 software (R Core Team, 2015). If the concentrations vs. time relationship generated a nonlinear trend, the Hutchinson and Mosier (1981) equation was used. When the trend was linear, the linear regression equation was used. Flows outside the range reported in the literature were discarded from the analysis. The area under the curve was interpolated to estimate the cumulative emission per period. Cumulative emissions of the three gases were converted to CO₂-equivalent values (t ha⁻¹ period⁻¹) using global warming potential values of 25 and 298 for methane and nitrous oxide, respectively.

2.5 Total soil organic carbon (SOC)

SOC was determined from soil chemical and physical sampling, i.e., C concentration (%) and bulk density. Sampling was performed at two depths: 0–10 and 10–20 cm. The C content was determined by applying the following equation:

$$TOC (kg/m^2) = \rho * fC * P_m, \quad (2)$$

where ρ = soil bulk density (kg m⁻³), fC = carbon concentration (kg C kg soil⁻¹) by Walkley & Black, (1934), and P_m = sampling depth (m).

2.6 Forage production and nutritional content

During the dry season (December–March) and rainy season (April–November), the following parameters were evaluated in the pastures.

- Forage yield was quantified by placing a frame (0.5 m × 0.5 m) in 10–20 sites according to the pasture management condition, and the grass was cut with a sickle or scissors. Subsequently, the green weight of the forage was determined using an electronic scale. Forage moisture was quantified after drying at 70°C for 3 days in a forage drying oven and used to calculate forage dry matter content.
- Ground cover (%) was determined by placing a 0.25-m² frame, as for forage yield.
- Forage nutritional quality: plucked samples of forage were taken from the pastures and analyzed for total nitrogen content (AOAC., 1984), neutral detergent fiber (Van Soest.,

1967), acid detergent fiber (Van Soest et al., 1991), and *in vitro* or *in vivo* dry matter digestibility (Tilley and Terry., 1963) at AGROSAVIA's Animal Nutrition Laboratory.

2.7 Statistical analysis

Soil chemical and physical variables were analyzed considering a completely randomized design. The fixed factors were the systems (management strategies), time, and depth, and the random factors were associated with the replicates. For the analysis of the data, an analysis of variance model with repeated measures over time was used when the temporal autocorrelation was significant; in this case, different covariance matrices were evaluated using Proc GLIMMIX of SAS version 9.4 (SAS, 2016). Normality and homoscedasticity assumptions were evaluated based on Q–Q plots and model residual vs. predicted plots.

For the construction of indicators associated with the soil's physical (water regulation), chemical (soil fertility), and biological (biodiversity) variables, a principal component analysis (PCA) was performed using the ADE4 1.4.5 library (Dray and Dufour, 2007) from R software (R Core Team, 2019). Calculations were carried out following the methodology proposed by Velásquez, Lavelle, and Andrade (2007).

An initial PCA allows for identification of the variables that best discriminate among the different management strategies. Variables with significant contribution (>50% of the maximum value) to either of the first two principal component axes were selected, and their contribution to PCA axes 1 and 2 was multiplied by the overall variability explained by each PCA axis to generate a weight factor for each variable. Values for each variable were then multiplied by their corresponding weight factor and summed to generate a raw sub-indicator value using the following formula:

$$I_{i1} = F_1 \times (\alpha I_a + \beta I_b + \gamma I_c \cdot \cdot \cdot) + F_2 \times (\alpha I_{1a} + \beta I_{1b} + \gamma I_{1c} \cdot \cdot \cdot).$$

I_{i1} = value of the ecosystem service indicator at plot 1. F_1 = % of the value of the variance explained by axis 1 of the PCA of the indicator data group I_i . α , β , and γ = respective contributions of variables a, b, and c to factor 1. a, b, and c = values of the variables measured at plot 1.

Finally, the indicators were rescaled between 0.1 and 1 to compare ES such as soil fertility, water functions, and biodiversity (soil macrofauna) associated with management strategies. For the climate regulation indicator, the CO₂-equivalent (t ha⁻¹ period⁻¹) was estimated from GHG fluxes (Rodríguez et al., 2013; Lavelle et al., 2014) and rescaled between 0.1 and 1 for comparison, and stock SOC values were rescaled between 0.1 and 1.

The analysis of variance of the indicators was performed considering a complete randomized design where the fixed factors were the management strategies and time (years) and the interaction strategies × time. The random factor was associated to the replicate.

3 Results

3.1 Soil fertility

Soil Ca, Mg, and P concentrations were affected by management, time (year), and sampling depth. The highest Ca, Mg, and P concentrations were found at 0–10 cm depth, where systems

received corrective applications between 1 and 1.5 t ha⁻¹ of dolomite lime and fertilizers. The concentrations of Ca, Mg, and P ranged from 1.02 to 1.97 cmolc kg⁻¹, 0.24 to 0.48 cmolc kg⁻¹, and 6.1 to 10.9 mg P kg⁻¹ in the pasture–crop system during the time of study. The concentration of Ca, Mg, and P ranged from 1.05 to 2.56 cmolc kg⁻¹, 0.23 to 0.59 cmolc kg⁻¹, and 1.8 to 7 mg P kg⁻¹, respectively, in the tree management system. The concentration of Ca, Mg, and P ranged from 0.46 to 1.8 cmolc kg⁻¹, 0.14 to 0.37 cmolc kg⁻¹, and 2.1 to 2.6 mg P kg⁻¹, respectively, in the till fertilizer system, and the values of Ca, Mg, and P ranged from 0.68 to 0.88 cmolc kg⁻¹, 0.39 to 0.59 cmolc kg⁻¹, and 1.1 to 1.6 mg P kg⁻¹, respectively, in the forest patch management system. According to the PCA, the behavior of soil chemical properties is explained by a total (components 1 and 2) of 56.2%, 50.4%, and 46% for the years 2012, 2013, and 2014, respectively. Trees and pasture–crop rotation management had the highest values of nutrient concentrations.

In general, during the 3 years of evaluation, there was a separation of pH and Al in the two depths sampled and a positive relationship of some elements such as P, Ca, and Mg in the systems with the addition of chemical fertilizers (Figure 1).

For the soil fertility indicator, there were significant differences regarding the time (year) × system effect. In 2012, the pasture–crop rotation system that included fertilizer additions and amendments had the highest indicator values compared to the other management systems (Figure 2). In 2013, the pasture–crop system still had the highest indicator values. The areas with the lowest values corresponded to plots that had lower rates of fertilization such as pasture areas managed only with tillage and fertilization (till–fert). In 2014, the trend was the same as in the previous year; the plots with the highest values corresponded to those that received amendments and fertilizers (pasture–crop and trees).

3.2 Water function

Total porosity was more strongly correlated with mesoporosity and inversely related to microporosity (Figure 3) in all years; however, the strength was not the same across years. Although texture is related to porosity, the contribution to principal components is minimal (Figure 3). Mesoporosity had variations originated by management systems and soil use change in the years of evaluation; its values ranged from 4.98% to 15.6%; the highest values were found for the first year, mainly due to the action of tillage in soil management involving pastures and annual crops (pasture–crop).

The water function indicator was possibly ($\alpha < 0.1$) affected by the management system. The highest values were observed in the pasture–crop and forest patch management systems, which had values above 0.5 for the years 2012 and 2014. In 2013, this indicator decreased in all systems; however, the forest patch and pasture–crop systems still had the highest values (Figure 4).

3.3 Soil macroinvertebrates

In 2012, the Blattodea order, which is composed primarily by termites, had the highest density (individuals m⁻²) at a depth of 0–10 cm, representing 35% (4,613 individuals m⁻²) of the

population, followed by earthworms with 28% (3,632 individuals m⁻²). At the depth of 10–20 cm, this trend was similar, where termites were also the dominant organisms; however, the population decreased with depth, corresponding to 15% with 2032 individuals m⁻², followed by ants (Hymenoptera) and beetles (Coleoptera).

According to the PCA, the first two components of soil macrofauna variables explained 55.1, 56.0, and 54.6% of the variability for the years 2012, 2013, and 2014, respectively (Figure 5). In 2013, a high correlation between termites and total macrofauna abundance (individuals m⁻²) was observed with Blattodea being the order with the highest density, with 17,984 individuals m⁻², representing 72% of the total abundance. For 2014, even though there was a decrease in density, termites were the taxonomic unit with the highest correlation with total abundance and 9,139 individuals m⁻² representing 69% of the population.

In 2012, the till–fert system tended to differentiate from the other management systems with an average value corresponding to 0.32 (Figure 6). In 2013, the forest patch system had a higher value than others, with an indicator value of 0.6. For 2014, no significant differences were found among systems.

3.4 Climate control

3.4.1 GHGs and soil organic carbon content

There was a significant interaction system × year. For 2013, the forest patch had significantly lower fluxes (7.9 t CO₂-eq ha⁻¹ yr⁻¹) than the other management systems. For 2014, the cumulative emissions were similar among management strategies, with values below 10 t CO₂-eq ha⁻¹ yr⁻¹ (Figure 7).

The soil organic carbon content at 10 cm depth was 13% higher in year 2013 than in 2012 and 12% higher in year 2014 than in 2012, with mean values of 31 g C kg⁻¹ and 26.9 g C kg⁻¹, respectively.

3.5 Forage production from grasslands and annual crops

The goal of the systems was to improve zootechnical indicators of the cattle breeding system with financial sustainability and animal welfare. Degraded *U. humidicola* pastures, without intervention during the rainy season (April–November), had a forage yield of 435–839 kg DM ha⁻¹ and yields ranging from 389 to 671 kg DM ha⁻¹ during the dry season. The renovation of *U. humidicola* by means of the crop–grassland strategy increased forage yield from 1,324 to 1,929 kg DM ha⁻¹ in the rainy season and from 869 to 1,152 kg DM ha⁻¹ in the dry season. When pasture recovery was carried out conventionally with soil tillage and fertilization, yields ranged from 1,232 to 1,893 kg DM ha⁻¹ in the rainy season and from 749 to 904 kg DM ha⁻¹ in the dry season (Table 3). The forage quality of the *U. humidicola* pastures showed crude protein (CP) between 3.1% and 6.6% and forage degradability between 51.6% and 63.6% in the dry season. In the rainy season, CP ranged between 5.3% and 11%, and degradability between 60.7% and 77.5%. Regarding corn forage quality, it was found that the CP contents varied between 7.0% and 8.5%, with degradability between 58.5% and 62.1%.

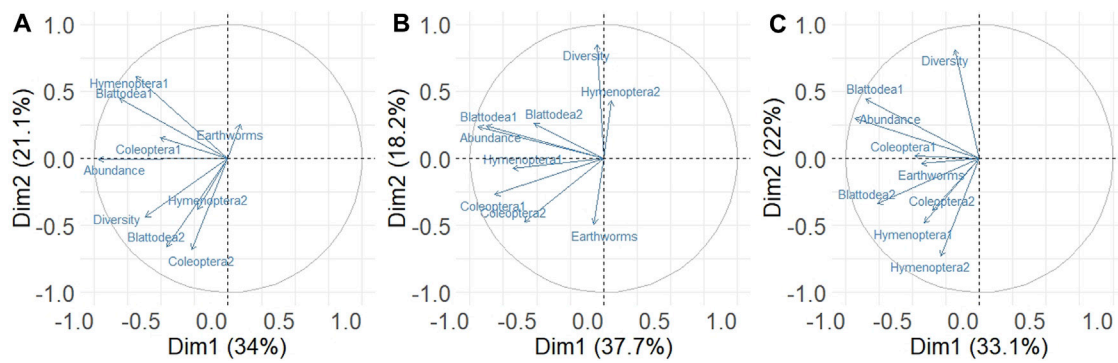


FIGURE 5 Projection of soil macrofauna variables in the correlation circle through principal component analysis at two depths (1: 0–10 cm; 2: 10–20 cm). Each vector represents a variable at depths 1 and 2. Years 2012 (A), 2013 (B), and 2014 (C).

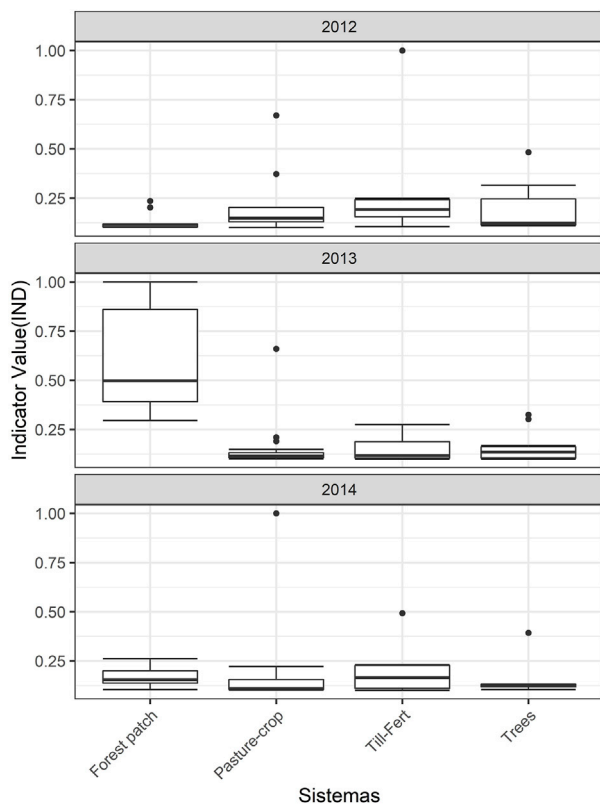


FIGURE 6 Biodiversity indicator, edaphic macrofauna, obtained with different grassland management strategies in a cattle breeding system established in an oxisol of the Eastern Colombian highland plains. Years 2012, 2013, and 2014.

Forage production from annual crops also showed outstanding results. Maize reached green forage yields of 35.1 and 25 t ha⁻¹, with the differences due to the genotypes used. Forage sorghum Corpoica JYT18 had yields that varied between 39.6 and 50 t ha⁻¹. Sorghum also has the possibility of generating a second crop by regrowth, which can represent 36–45% of additional forage to conserve.

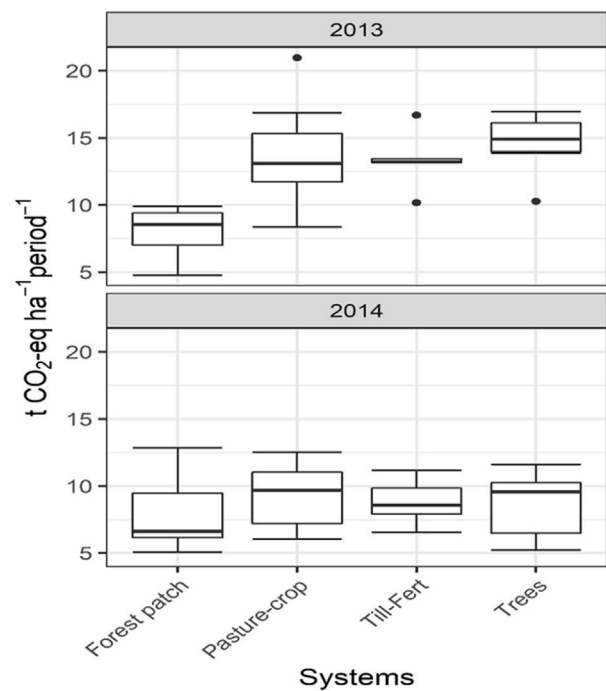


FIGURE 7 GHG emissions (CO₂-eq ha⁻¹ period⁻¹) quantified in different grassland management strategies in a cattle breeding system established in an oxisol of the Eastern Colombian highland plains. Years 2013 and 2014.

TABLE 3 Forage supply (DM kg ha⁻¹) from *U. humidicola* pasture in rainy and dry seasons at the Carimagua Research Center.

Pasture	Forage supply (DM kg ha ⁻¹)	
	Rainy season	Dry season
<i>Urochloa humidicola</i> (degraded)	435–839	389–671
<i>Urochloa humidicola</i> (crops–pasture)	1,324–1929	869–1,152
<i>Urochloa humidicola</i> (conventional)	1,232–1893	749–904

TABLE 4 Production (t ha⁻¹) and nutritional content (%) of green forage of maize, sorghum, and soybean crops in the Carimagua Research Center.

Crop	Green forage (t ha ⁻¹)	CP (%)	Degradability at 48 h (%)
Maize (hybrid)	25.0–35.1	7.0–8.5	58.5–62.1
Sorghum JJT18	39.6–50.0	6.0–7.6	58.1–69.7
Soybean Taluma 5	17.1–27.8	18.4–19.5	56.7–71.1
Grass–legume mixture	-	10.4–12.9	58.4–72.1

Regarding soybean forage production, the Corpoica Taluma 5 variety reached yields between 18.5 and 26.5 t ha⁻¹, the Panorama 29 variety, between 17.1 and 26.4 t ha⁻¹, and the Corpoica Sabana 7 variety, 27.7 t ha⁻¹. Corpoica JJT18 sorghum provided CP values of 6.0–7.6% and degradability between 58.1% and 69.7%. For soybean varieties, CP values ranged from 12% to 24% and degradability from 56.7% to 71.1% (Table 4).

The management and environmental conditions prevailing during the study favored the growth and development of forest species. Diameter at breast height (DBH) recorded in 2015 showed that the highest values were reached by *Acacia mangium* with 58 cm, followed by *Piptadenia opacifolia* with 45 cm, and finally, *Cassia grandis* with 25 cm. With the differential development of the species, ground shade projection related to the growth and architecture of the canopy that generates the microclimate for animal comfort was similar for the first two species and less for the last one. Species growth also plays an important role in C fixation in the wood biomass accumulated, soil and water conservation, biodiversity, and the beauty of the rural landscape.

4 Discussion

The management strategies that were tested in this research were aimed at improving animal productivity through a greater forage supply quality, especially during critical periods. Direct actions into the soil were carried out to improve nutrient supply conditions and soil water storage. Those actions included amendment applications for improving K⁺, Ca²⁺, and Mg²⁺ availability and the use of tillage implements for improving soil porosity.

Soil nutrient availability was improved in the management strategies till–fert, pasture–crop, and trees by application of fertilizers and amendments. Soil chemical properties (Ca²⁺, Mg²⁺, P, pH, and Al³⁺ concentrations) were the main soil fertility indicators. Dolomite lime affects mainly pH and Ca²⁺. Positive pH and calcium relationships, shown in the correlation circles, are associated with the application and incorporation of a mixture of dolomitic lime, Paz del Río[®] fertilizer, phosphate rock, and agricultural gypsum, as well as that of phosphorus by the application of triple superphosphate and DAP. According to Amézquita et al. (2004), the main soil limitation for crop production in highland plains is the high acidity and low nutrient availability; therefore, it is essential to improve soil fertility by improving nutrient availability and reducing Al toxicity to increase productivity.

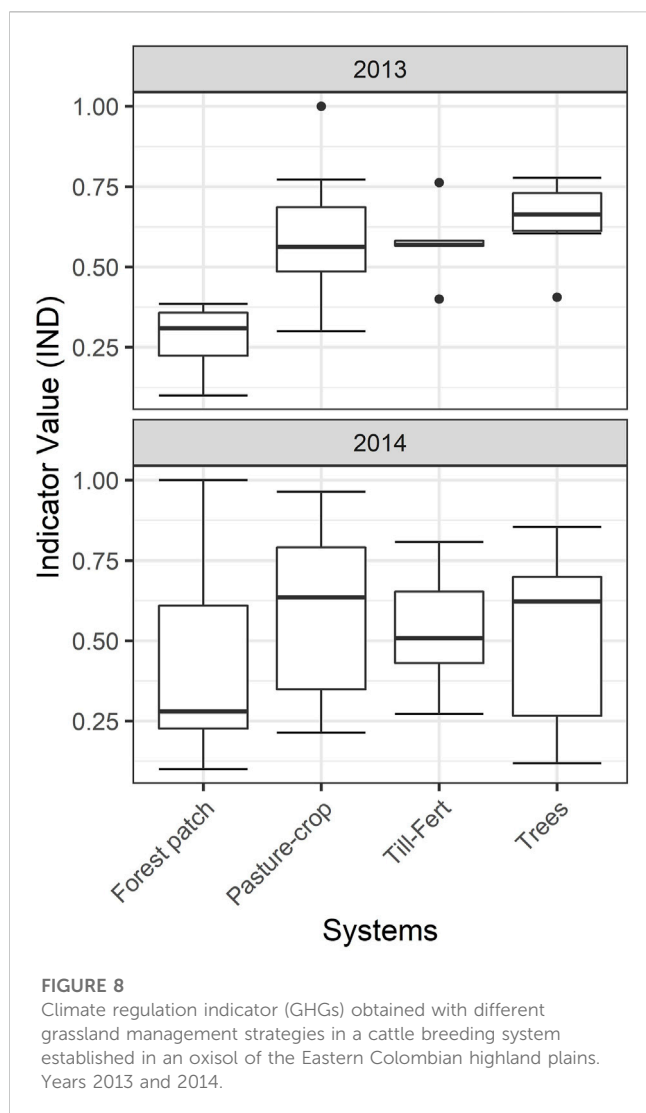
Soil bulk density, macroporosity, and mesoporosity were the main variables associated to the soil water indicator. Tillage practices used in the management strategies mainly impacted macropore and mesopore ratios. Increasing mesoporosity is the key for improving soil water and nutrient availability for established pastures. By

improving soil organic matter through practices such as liming, pasture nutrition, and the use of soil decompaction equipment, root system quantity and quality of established grasses would be increased, thus exerting a positive effect on soil structure, density, and porosity, which in turn contribute to the hydrophysical dynamics that benefit crop development (Torres-Guerrero et al., 2013).

An improvement in chemical and physical conditions was observed in less than 2 years; however, in the medium term (3 years), an increase in soil bulk density was observed, which returned to the initial condition of native savanna, validating the theory of soil resilience (Amézquita et al., 2013). This condition highlighted the importance of including practices such as incorporation of organic matter and deep tillage to maintain better soil conditions for production systems. Amézquita et al. (2013) state that the main challenge facing agricultural soils from the highland plains (Typic Haplustox) is their susceptibility to degradation. Degradation is understood as the loss of some physical, chemical, and biological qualities due to negative human intervention, which will affect agricultural sustainability. In livestock systems, soil compaction is the main physical cause that restricts plant root growth and development.

The improvement of edaphic macrofauna population was observed in forest relict areas (forest patch). It was noted that the greatest variability explained by PCA was related to organisms found at 0–10 cm depth. This may have been influenced by the vegetation cover present in the evaluated systems, which contributed with biomass supply as a food source for organisms and for regulating soil temperature and decreasing light intensity, generating a stable microhabitat for the development of soil epigeal and endogean macroinvertebrate communities. Soil-dwelling biological communities are among the most diverse, comprising a wide range of life forms and functions, and are involved in many ecological processes and provide key ES (Decaëns, 2010). Macrofaunal dynamics can be used as a biological indicator sensitive to impacts on land use, soil management, and soil quality (Rousseau, et al., 2013). These organisms, together with earthworms, are dominant in the edaphic macrofauna of tropical savannas, such as Brazilian Cerrado and the Eastern Plains of Colombia, where they influence soil physical characteristics and soil chemical activity (Decaëns et al., 2003). Biodiversity patterns can be affected not only by the soil physical conditions but also by biotic interactions that are established between species. Therefore, the magnitude and specificity of mechanisms that control a given group of organisms may be different from those of another group (Decaëns, 2010). According to Masters (2004), changes in soil temperature and moisture conditions as a consequence of a lower amount of organic residues affect certain taxonomic entities, such as Oligochaeta or earthworms, which normally require permanent moisture for survival.

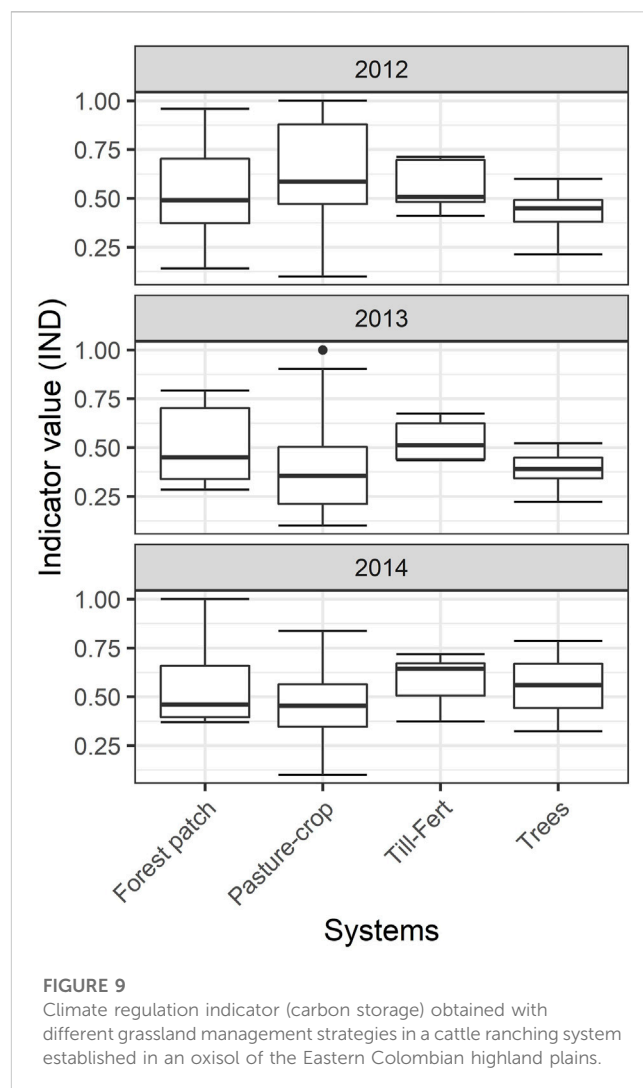
The GHG emissions expressed as accumulated CO₂-equivalent of each gas showed how the system that has no anthropic



intervention (forest patch) registered the lowest emissions and, therefore, was associated with the lowest indicator (Figure 8). Systems where tillage and fertilizer were applied showed higher values (pasture-crop, till-fert, and trees). Year-by-year variation in the accumulated CO₂-equivalent may also be associated with the prevailing climatic conditions; thus, lower values in 2014 may be due to the dry condition in that year for Eastern Colombian highland plains (Vargas-Pineda et al., 2017). Soil moisture is the most important parameter associated with soil GHG dynamics because it controls microbial activity and all related processes.

Soil organic carbon indicator tended to increase in the forest patch and tree management systems, while it was reduced in the pasture-crop system (Figure 9). According to Post and Kwon (2000), there are factors that influence soil carbon storage, which depends on the quality and quantity of inputs and the decomposition rate of the organic matter. Management practices that improve soil physical, chemical, and biological characteristics have positive impacts on the accumulation of above- and below-ground organic carbon.

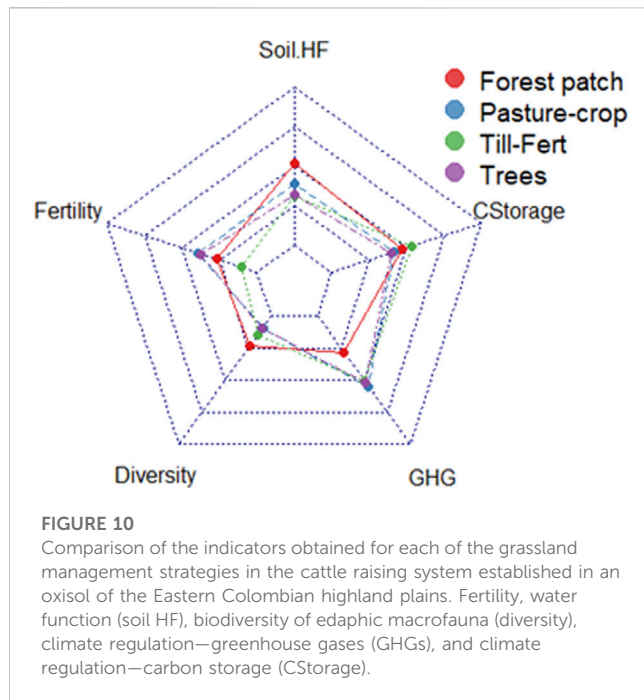
Although it is necessary to implement practical actions in livestock systems to reduce the impacts of climate change, these



should not be in contradiction with the objectives of ending poverty and reducing the hunger rate to 0 by 2030, considering that livestock production provides benefits for human nutrition, health, and welfare. The livestock system is part of the AFOLU sector, which is considered one of the largest emitters, contributing up to 15%, and of the AFOLU sector, 10% is considered as the direct responsibility of the bovine system (FAO, 2019).

The consolidated analysis of the indicators shows how forest relict areas (forest patch) had an expected trend regarding the best performance for the biodiversity of edaphic macrofauna and improved soil physical conditions for better water storage capacity and lower GHG emissions. For fertility indicator and carbon storage, forest patch had an intermediate performance compared to other systems. As expected, the highest emissions were recorded for the pastures and annual crop systems which were intervened with fertilization and/or mechanization (Figure 10).

This study showed that systems under an agroforestry scheme, which in this case translates into a system of recovered pastures, annual crops-pastures (agropastoral), tree strips, and forest-patch areas, can contribute to the maintenance of soil ES, although each one of them provides improvement in at least one soil ecosystem service.



The work developed in soil and climate conditions of the Colombian highlands contributes with evidence that shows that cattle farms can improve their eco-efficiency indicators and, in particular, contribute to carbon sequestration through the implementation of agroforestry systems; remaining forests, areas of secondary vegetation in regeneration, wooded pastures, and other land uses dedicated to agricultural work can act as atmospheric carbon sinks (Post and Kwon, 2000; Ruiz, 2002; Fisher et al., 2004).

5 Conclusion

It is urgent to improve the sustainability of livestock production systems by adopting measures at both local and regional scales, not only for mitigating GHG levels in the atmosphere but also to conserve other ES that are part of the soil dynamics. In this sense, the scheme proposed, which included improving pastures through management, tree incorporation, and preserving forest-patch areas, can be applied in the farms of small and medium producers in the Eastern Colombian highland plains.

Improving management practices in pasture systems and introducing tree cover can enhance zootechnical and environmental indicators due to their positive effect on forage quality and quantity and the potential reduction of carbon footprint.

The sustainable intensification of livestock systems will allow releasing areas not suitable for agricultural production and the use of secondary forests in natural regeneration processes, considering establishment of forest plantations at the farm level.

There are still challenges in agricultural research to evaluate the long-term effect of management strategies on soil properties, water use efficiency, crops, and animal productivity in the high plains of Colombia aimed to contribute to mitigate effects of climate change and variability.

Data availability statement

The raw data supporting the conclusion of this article will be made available by the authors, without undue reservation.

Author contributions

OP-L developed and managed the research project. NR-H, JB-R, JOA, LM-C, and MA participated in the planning of field trials and consolidation of experimental data. MA performed statistical analysis and interpretation of the data. NR-H and OP-L prepared the manuscript with important contributions from all authors. All authors contributed to the article and approved the submitted version.

Funding

This research was supported by the Ministry of Agriculture and Rural Development—AGROSAVIA –CIAT Cooperation Agreement Fund and by AGROSAVIA (formerly Corpoica) Five-Year Agenda Fund implemented between 2011 and 2017.

Acknowledgments

The authors would like to thank the Colombian Ministry of Agriculture and Rural Development for financing this research and the researchers and operators of the International Center for Tropical Agriculture (CIAT) who participated in the execution of activities and field sampling. The authors acknowledge the support of professionals and operators of the Colombian Agricultural Research Corporation AGROSAVIA for their follow-up and support in data collection. Special recognition is due to the technical and administrative staff of CI Carimagua—AGROSAVIA for their cooperation and support in the development of the work.

Conflict of interest

Authors NR-H, MA, LM-C, JA, Jb-R, and OP-L were employed by the Colombian Agricultural Research Corporation—AGROSAVIA.

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Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fenvs.2023.1107466/full#supplementary-material>

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