



Effective Conservation of Desert Riverscapes Requires Protection and Rehabilitation of In-Stream Flows With Rehabilitation Approaches Tailored to Water Availability

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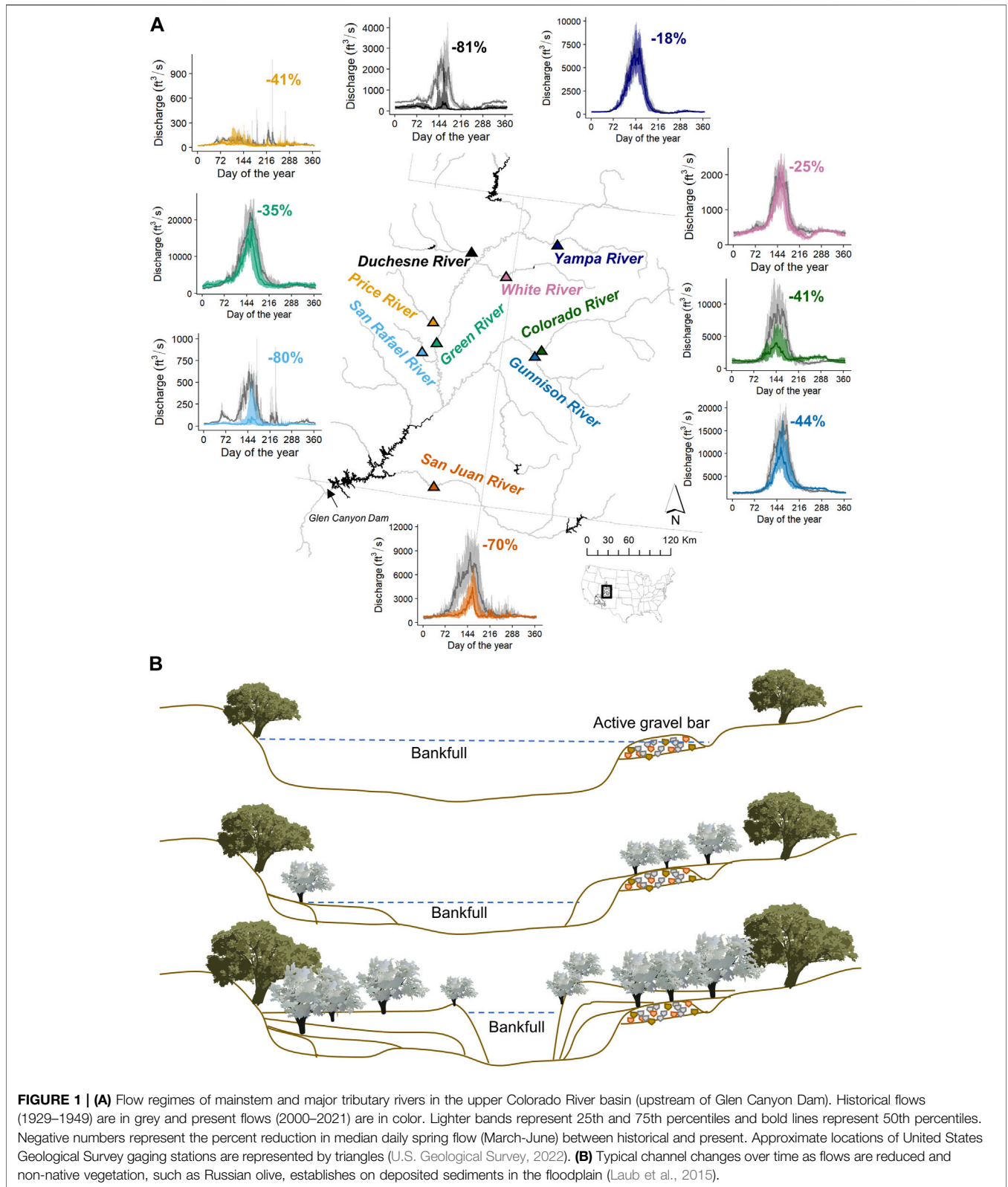
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Desert riverscape rehabilitation practitioners must contend with compounding effects of increasing human water demand, persistent drought, non-native species establishment, and climate change, which further stress desert riverine ecosystems such as rivers in the Colorado River basin, United States. Herein, we provide our perspective on the importance of natural flows, large floods in particular, for successful conservation and rehabilitation of riverscapes. We present ideas developed from our experience with rehabilitation projects across multiple desert tributary rivers with varying levels of habitat degradation and water abstraction. We propose spatially extensive measures such as protection of in-stream flows, tailoring rehabilitation efforts to available annual water availability, and working with nature using low-tech process-based techniques to more completely address the mechanisms of habitat degradation, such as flow reduction and vegetation-induced channel narrowing. Traditionally, rehabilitation efforts in the Colorado River basin take place at relatively small spatial extents, at convenient locations and, largely focus on reducing non-native plant and fish species. We suggest that we need to think more broadly and creatively, and that conservation or recovery of natural flow regimes is crucial to long-term success of almost all management efforts for both in-stream and riparian communities.

Keywords: rivers, arid lands, climate change, flow regime, native species, environmental flow (e-flow)

THE ISSUE PART I

River-dwelling organisms, such as fishes, need adequate flow to complete their life histories and thrive. Whereas at least some minimum amount of water is required for fish to simply persist, flows of large enough magnitude and long enough duration to create and maintain physical habitat and provide lateral connection to productive floodplains are vital to thriving fish assemblages (Matthews 1998; Marchetti and Moyle 2001; Bunn and Arthington 2002). This simple fact is well established (Poff et al., 1997; reviewed by Palmer and Ruhi 2019), but management of river flows rarely follows this basic principle (e.g., Pennock et al., 2022a). Several challenges exist to the management of natural stream flows, particularly in arid-lands such as the Colorado River basin. Water in the Colorado River basin is overallocated (Kuhn and Fleck 2019); consequently, natural flow regimes are altered



across the upper Colorado River basin (Figure 1A). Overall allocation is fed by implicit societal support, by some, of western water law used to dictate all river flows for upstream

consumption (e.g., *The Law of the River*; Kuhn and Fleck 2019). Base flows fed by groundwater discharge contribute an average of 56% of surface flows in the upper Colorado River basin (Miller

et al., 2016), which are threatened by groundwater mining, particularly during dryer periods (Castle et al., 2014). Moreover, human-induced climate change is causing basin-wide aridification leading to declining run-off basin-wide (Udall and Overpeck 2017; Overpeck and Udall 2020).

Reduction of large spring floods, and run-off in general, in desert rivers of the Colorado River basin (and elsewhere) has contributed to a positive feedback loop such that river channels have become aggraded and narrowed as stream power to erode sediment is lost (Allred and Schmidt 1999; Grams and Schmidt 2002; Dean and Schmidt 2011). Concomitantly, this process is often aided by encroachment and establishment of riparian vegetation on previously active alluvial surfaces, primarily by introduced, non-native species (Figure 1B; Manners et al., 2014; Laub et al., 2015; Scott et al., 2018). This feedback loop between declining flow and vegetation encroachment has resulted in the simplification and loss of habitats important for many native fishes and wildlife, such as floodplain wetlands and riparian gallery forests (Pennock et al., 2022a).

THE ISSUE PART II

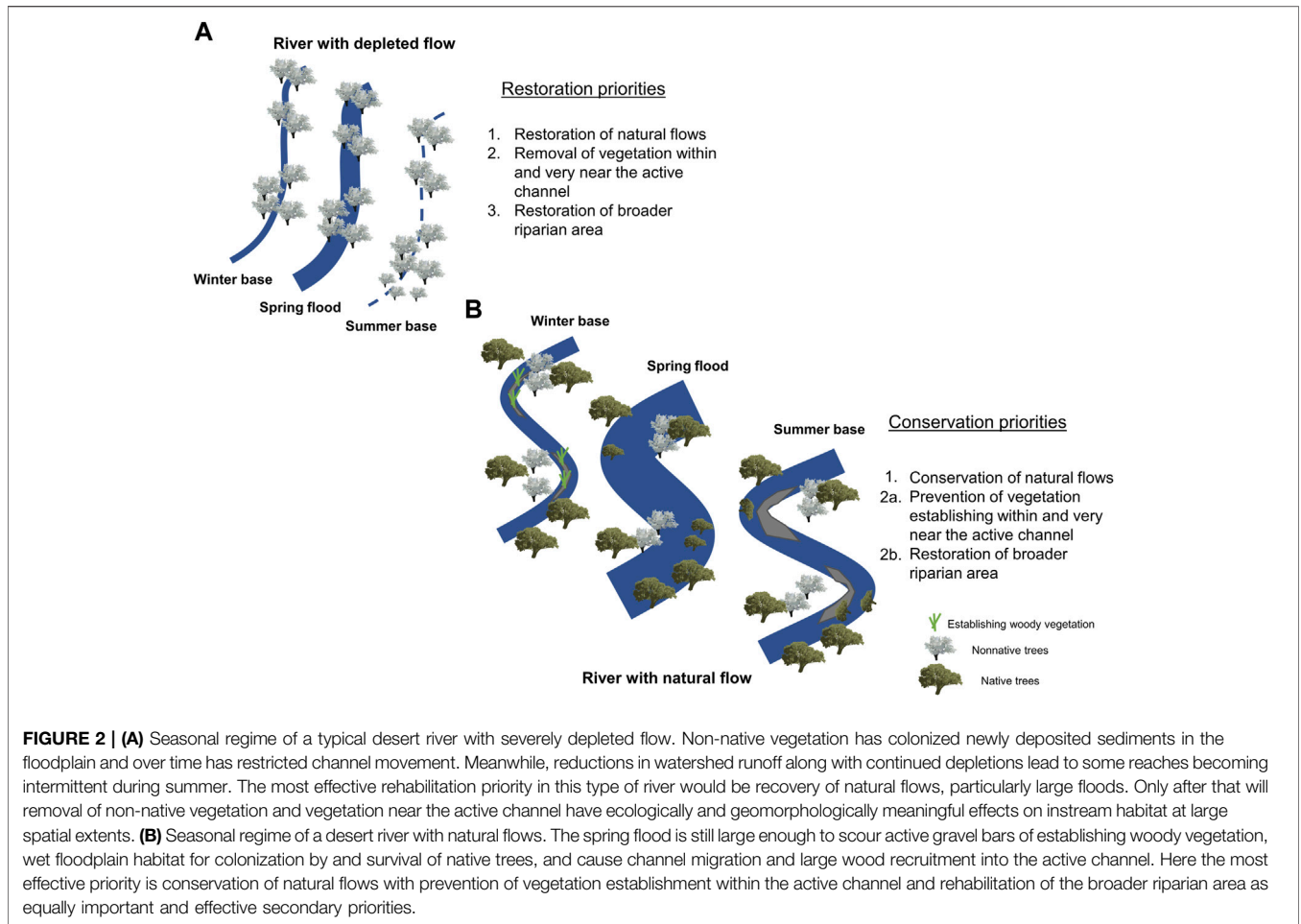
Managers and researchers have attempted to remedy the loss of complex habitat by performing active rehabilitation and, in some instances, attempting to provide environmental flows (e.g., Gido and Propst 2012; Yarnell et al., 2015). Despite good intentions, many of these conservation and rehabilitation efforts take place over relatively small spatial and temporal extents, while the greater ecological degradation takes place over much broader extents, an issue alluded to more than a century ago (e.g., Marsh 1864; Fausch et al., 2002). Additionally, desert riverine fishes evolved in highly dynamic and connected riverscapes; within contemporary riverscapes, fishes are often highly mobile, moving amongst complementary habitat types (e.g., Cathcart et al., 2018; Hooley-Underwood et al., 2019; Pennock et al., 2020) to reproduce, feed and grow, and overwinter (Schlosser 1995; Fausch et al., 2002). Climate change induced flow reductions, water attenuation, increased human water demand, insufficient water releases from dams, groundwater mining, and failure to prioritize environmental flows over consumptive use have all stymied attempts to provide flows of adequate magnitude and duration to maintain habitat and inundate floodplains in the Colorado River basin.

Many of the challenges to the goal of conserving and recovering riverscapes and native fishes are exemplified by a few specific examples from desert tributary systems where we are actively working. First, on two tributary rivers, the Price and San Rafael rivers, our own previous rehabilitation efforts have taken place on only several kilometers of river out of nearly 100 km per watershed. Although those smaller scale efforts have been critical to rehabilitating local habitats, they have not contributed to resolving the mechanism behind habitat degradation, a lack of water. Second, in the San Juan River, flow releases from an upstream dam are managed to mimic the natural flow regime, aided by major downstream tributary inputs (e.g., Gido and Propst 2012). However, because of infrastructure limitations,

climate change, and increased consumptive water use, these managed flows have failed to truly mimic historical natural flows in terms of all attributes of the natural flow regime, as is seen by the reduction in magnitude and duration of the spring flood from the early 20th century to present day (Figure 1A; Pennock et al., 2022c). Third, current mitigation and management actions by two federal recovery programs in the upper Colorado River basin include stocking native fishes, habitat rehabilitation, non-native fish removal, and, in some limited cases, management of flow releases from dams. However, actions available to mitigate widespread habitat loss and fragmentation from dam construction and reservoir filling and consumptive water use are limited. The stated goal of these programs is recovering endangered fishes, while not hindering further water development (<https://coloradoriverrecovery.org>). The existence of the recovery programs suggests consistent compliance of current and future federal water projects with the Endangered Species Act (Benson 2010). Compliance is achieved through the above-mentioned mitigation and management actions, but achieving their stated goal of recovery is becoming more and more unlikely as water availability continues to decline (Williams et al., 2022). Further, the programs generally perform management actions in designated critical habitats, which for most listed species includes very little habitat in smaller tributaries widely used by many fishes (Bottcher et al., 2013; Webber et al., 2013; Laub et al., 2018), and so protections for tributary habitats is limited. Due in part to our demonstrated inability to manage desert rivers for natural flow regimes required for native fishes to thrive, we are witnessing widespread declines in distribution and abundance of native, and often endemic, fishes, including extirpation (Budy et al., 2015; Laub et al., 2018; Pennock et al., 2022a).

PERSPECTIVES ON POTENTIAL SOLUTIONS

We believe there are several promising ways to overcome some of the challenges described above. First, there is a need to scientifically evaluate ecological processes, such as fish and riparian vegetation population and community dynamics across large segments of riverscapes and prioritize conservation and rehabilitation actions across larger spatial extents, where these important ecological processes play out (Schlosser 1995; Palmquist et al., 2018). Moreover, the underlying causes of degradation must be remedied for long-term conservation to be successful, which for arid land rivers in the southwest United States requires at least some level of protection or recovery of natural flow regimes (Stromberg 2001; Pennock et al., 2022a). This approach would require working with states to modify laws to include water for environmental flows as a beneficial use, as a start. Establishment of incentive-based programs that pay water rights holders to not divert water (e.g., Van Kirk et al., 2020), or other promising approaches that leverage existing water law could also increase in-stream flows (e.g., Van Kirk et al., 2019; Van Kirk et al., 2020). For instance, using managed aquifer



recharge to bolster local groundwater resources along with joint management of groundwater and surface water could help maintain baseflows in severely depleted rivers (Castle et al., 2014; Van Kirk et al., 2020). We also suggest shifting to a more holistic approach, by leveraging annual flow variability, and partnering with nature using low-tech process-based restoration practices (Wheaton et al., 2019).

On some tributary rivers with severely depleted flow, prior rehabilitation efforts have focused primarily on removing localized monocultures of non-native riparian vegetation, such as Russian olive (*Elaeagnus angustifolia*) and tamarisk (*Tamarix* spp.), an important component of riverscape rehabilitation because of the impacts of non-native vegetation encroachment on in-stream and riparian habitat (Keller et al., 2014; Laub et al., 2015; Scott et al., 2018). We suggest the most effective approach in these depleted rivers will first require recovery of some level of natural flows (volume and timing), then accompanied removal of established vegetation within and near the active channel, followed by, rehabilitation of the broader riparian area by removal of non-native vegetation throughout the riverscape (Figure 2A). Recovery of natural flows would allow for dynamic geomorphic and biotic processes to return to the riverscape once non-native vegetation is removed, potentially

even with spring floods of lower magnitude than observed historically. For instance, removal of vegetation near the active channel could allow for scour and lateral channel migration in years of modest to larger spring floods (e.g., Keller et al., 2014). Finally, recovery of in-stream flows and the broader riparian area at the riverscape-scale has a high probability to increase recruitment success of native vegetation, such as Fremont’s cottonwood (*Populus fremontii*) and reduce propagule pressure of non-native vegetation (Schlatter et al., 2017; Scott et al., 2018; West et al., 2020).

In the White River in Utah, a river with a relatively natural flow regime, our approach is similar, but focuses more on conserving the natural ecological and geomorphological processes that still exist (Figure 2B; Pennock et al., 2022b). We predict the most effective strategy in rivers, such as the White River, is to first conserve natural flows (Pennock et al., 2022a; Pennock et al., 2022b). The simple fact here is that rivers with natural flow regimes still have a spring flood of large enough magnitude and long enough duration to scour bar features of emerging vegetation (non-native and native) and to provide lateral connection to floodplains. However, with climate change-induced declines in watershed runoff in the Colorado River basin, all rivers have experienced some degree of flow

reduction from the 20th to 21st century and the predictions for the future are no better (Udall and Overpeck 2017; Williams et al., 2022). The second priority is rehabilitation of the riparian plant community across the riverscape by removing established non-native vegetation, as described above.

In light of basin-wide reductions in flow, even in rivers with natural flow, we are working with spring floods of lower magnitude, and as such, we propose conservation actions tailored to annual flows. We can do this by pairing annual conservation and rehabilitation actions with current annual flows, and in some cases, annual flows in prior years. For example, in years with back-to-back below average floods, we are focusing our efforts on preventing vegetation encroachment on the active channel, which would force the river towards a more aggraded and narrowed state (Grams and Schmidt 2002; Dean and Schmidt 2011; Laub et al., 2015). Conversely, in a year with an average or above average flood, we will focus efforts on recruiting large wood into the channel and regaining channel width in localized areas where aggradation has occurred.

Furthermore, we propose to partner with nature by using low-tech process-based restoration techniques that have recently become popular and demonstrated to be effective at rehabilitating physical, chemical, and biological processes in riverscapes (e.g., Beechie et al., 2010; Wheaton et al., 2019; <https://lowtechpbr.restoration.usu.edu/>). We are adapting some of the more popular of these techniques to larger rivers, but the long-term success of this strategy also likely hinges on the ability to conserve some degree of large spring floods. For example, recruitment of large wood into the active channel is an important process on the White River aided by its relatively natural flow regime, recruiting cottonwoods, and active beaver tree-felling, resulting in densities of large woody debris 2-6 × higher relative to other tributaries to the Green River (Pennock et al., 2022a). This large woody debris maintains and creates habitat complexity including scour pools, low-velocity areas, and colonization areas for macroinvertebrates, habitat benefitting fish and other wildlife. Thus, more specifically, one of our strategies is to mimic beaver activity on meander bends by selectively felling trees to promote hydraulic, hydrologic, and geomorphic effects similar to post-assisted-log structures (PALS) in wadeable streams (Wheaton et al., 2019).

In tributary systems across the upper Colorado River basin, we are incorporating natural flow conservation, tailoring of conservation and rehabilitation actions with annual water availability, and partnering with nature by implementing dynamic and low-tech, process-based techniques (versus hard engineering) in three different ways. A) In the White River, as discussed above, we are working to conserve the natural hydrograph and tailoring annual conservation and rehabilitation activities to the predicted water year and anticipated discharge and allowing the river to collect and use removed non-native trees to create habitat. B) On the San Juan River, where non-native Russian olive have established dense monocultures throughout the riverscape, we are experimentally adding semi-anchored Russian olive branches to the river channel to enhance habitat for native fishes. C) On the Price and San Rafael rivers, we are pairing the translocation of nuisance beavers

with installation of high densities of PALS and non-native vegetation removal to increase habitat complexity (Doden et al., 2022), promote lateral channel movement, and increase water retention in the riverscape. In 2022, Utah's first ever water banking application was approved for the Price River, which allows for The Nature Conservancy, Trout Unlimited, and Utah Division of Wildlife Resources to lease water for environmental benefit. These entities are currently working to lease water for instream purposes including maintaining minimum flows and providing some semblance of a spring flood. For all of these examples, we combine our rehabilitation activities with structured monitoring programs such that we can learn from our successes and mistakes and manage future efforts based on lessons learned (Palmer et al., 2005). Although these rivers differ greatly in their level of degradation, and thus, the specific techniques we are implementing also vary widely, the long-term success of our efforts in each river still hinges on the ability to implement flow conservation and recovery, particularly moderate to large spring snowmelt floods.

DATA AVAILABILITY STATEMENT

The original contributions presented in the study are included in the article/Supplementary Material, further inquiries can be directed to the corresponding author.

AUTHOR CONTRIBUTIONS

CP and PB led writing of the manuscript with input from WM. All authors edited and approved the final version of the manuscript and developed the original ideas jointly.

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