



# Limnochemistry and Plankton Diversity in Some High Altitude Lakes of Kashmir Himalaya

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High altitude lakes (HALs) of Kashmir Himalaya are the important ecosystems in the mountain ecology of the broader Hindukush Himalayan region. This article provides a comprehensive information about the plankton (phytoplankton and periphyton) assemblages, water quality (WQ), bathymetry, morphometry, and land use land cover (LULC) of some select high altitude mountain lakes of Kashmir Himalaya. LULC analysis revealed that the catchment of the lakes spread over an area of about 16179 ha, is covered by different land cover types dominated by pastures (50.8%), followed by barren rocky (32.6%), snow and glaciers (11.9%), lakes (2.5%), forest (2%), and streams (0.2%). Bathymetric and morphometric analysis revealed that the Gangbal Lake is the deepest (84 m) and largest (162.4 ha) among the investigated lakes. The water quality index revealed that all the HALs have the excellent water quality category. Statistical analysis (Wilk's  $\lambda$ ) depicted that nitrate-nitrogen ( $\text{NO}_3^-$ -N), nitrite nitrogen ( $\text{NO}_2^-$ -N), ammoniacal nitrogen ( $\text{NH}_3$ -N), total phosphorus (TP), and magnesium hardness (Mg-H) are responsible for major variability between all HALs sites. The cations followed the order of  $\text{Ca}^{2+} > \text{Mg}^{2+} > \text{Na}^+ > \text{K}^+$  while as anions followed the order as  $\text{HCO}_3^{3-} > \text{Cl}^- > \text{SO}_4^{2-}$ . Algal composition (phytoplankton and periphyton) assessment revealed the presence of 61 taxa belonging to Bacillariophyceae (45), Chlorophyceae (14), Cyanophyceae (1), and Xanthophyceae (1). The higher dominance of Bacillariophyceae indicates oligotrophic nature of the lakes. Canonical correspondence analysis (CCA) highlighted the role of various water quality parameters like pH, EC, and TDS on the composition of phytoplankton and periphyton species among the lakes. The present study therefore generated a baseline database for some of the HALs of Kashmir Himalaya that can act as a precursor for more research on future changes in the lake ecosystems of the region.

**Keywords:** high altitude lakes, biodiversity, bathymetry, morphometry, Kashmir Himalaya

## INTRODUCTION

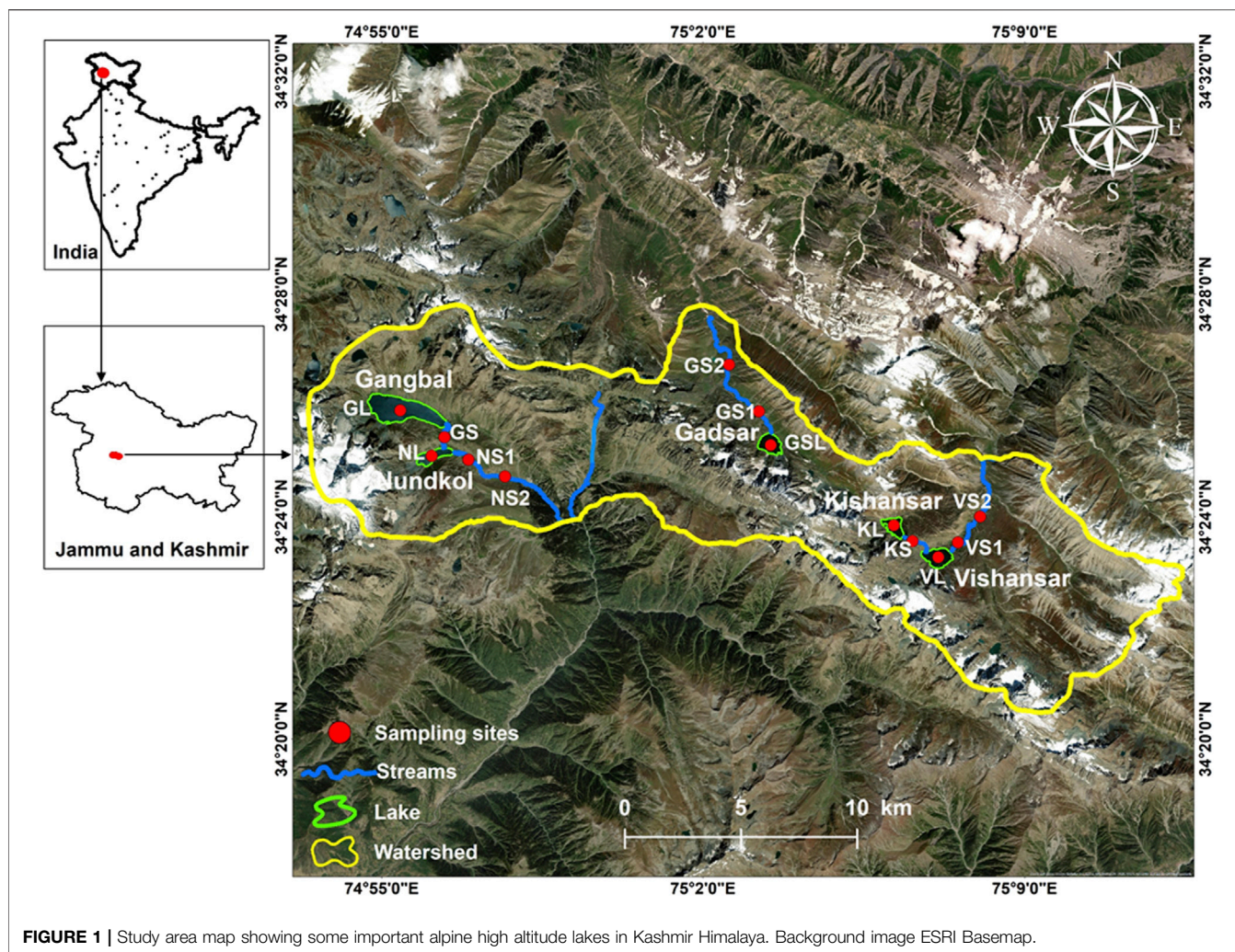
High Altitude Lakes (HALs) are the most important, pristine, and ecologically sensitive ecosystems which lie usually above 3,000 m altitude (Chatterjee et al., 2010; Pastorino et al., 2020; Dar et al., 2021a). They are considered as one of the most significant natural resources for the survival of living beings (Walsh and Milon, 2016). HALs provide a variety of ecological services and functions, apart from being a major source of the transboundary river ecosystems, and play an important role in

hydropower generation, irrigation facilities, besides they help in satisfying the drinking water demand to a noteworthy percentage of the world population (Chatterjee et al., 2010; UNCSO, 2012). HALs are known as “natural laboratories” wherein global biological diversity and water quality changes could be examined (Rogora et al., 2018; Pastorino and Prearo, 2020) to identify the important anthropogenic pressures (Battarbee et al., 2009). HALs are categorized by a distinctive water sources, species, habitats, and communities, and have particular spiritual, and cultural importance (WWF, 2009; Gurung et al., 2018). Although apparent to be in a pristine state, they are potentially at risk from overgrazing (Pastorino et al., 2018), climate change (Pastorino and Prearo, 2020), long-range transport of atmospheric pollutants (Schindler and Smol, 2006), water abstraction and exploitation (Tiberti et al., 2019), alien species introduction (Pastorino et al., 2020b), mountain tourism activities (Rupakheti et al., 2017) and mountain farming (Tiberti et al., 2014). Remote HALs are highly vulnerable to pollution from human dominated mid and low altitude lake systems due to combined effect of high ultra-violet radiation exposure and cold distillery phenomenon (Elser et al., 2020; Dar et al., 2021a). The concentration of precipitation borne pollutants increases in ice and snow because of sublimation and lower temperature slows down contaminant degradation (Daly and Wania 2005; Elser et al., 2020). HALs function as a natural storeroom for ecological evidence linked to changing climate and other environmental parameters (Raut et al., 2012; Kaphle et al., 2021). Lake/river water chemistry along with sediment chemistry plays a significant part in explaining the underlying forces of earth surface (Sheikh et al., 2014). As lakes are relatively less dynamic, they are prone to exogenic influences and deposit pollutants gradually (Saini et al., 2008). The hydro-chemical profile of these lake systems is mostly influenced by atmospheric deposition and climatic parameters; therefore, they act as indicators of climate change, atmospheric deposition, and air pollution (Rogora et al., 2008). Generally, the hydrochemistry of the lakes explains the temporal changes in the behavior of ions and watershed features (Anshumali and Ramanathan, 2007). LULC change is most effective factor responsible for changing landscapes (Dar et al., 2021c; 2021d). Land cover maps represent inclusive information about the status of specific landmarks under study. Bathymetric and morphometric attributes are important factors for better understanding of lacustrine system structure, physical topography, and functioning (Raut et al., 2012). GIS based quantitative morphometric, bathymetric measurements and LULC maps provides a pivotal information in support of concluding results to proper lake management and protection (Dar et al., 2021c; 2021d).

The interest of the scientific community towards the phytoplankton and periphytic assemblages increased massively after Hutchinson's (1961) seminal publication on the paradox of the plankton. Hutchinson contrast Hardin's competitive exclusion theory (Hardin, 1960) by providing a plausible explanation for the co-existence of the species in a

homogeneous environment. Ever since the resurgence in phytoplankton research might be elucidated by development in the theoretical and methodological characteristics. Lakes at high altitudes are unique ecosystems that are often described by lower temperatures, oligotrophy, low productivity, simple food web structure, and strong ultra-violet radiation in surface waters (Xing et al., 2009; Kuefner et al., 2021). Biological assemblages at higher elevations are modified to severe environmental conditions of unfavorable boreo-alpine environment (Kuefner et al., 2021). Plankton display a significant role in the productivity and transmission of nutrients in the aquatic ecosystems and therefore acts as a vital bioindicator for monitoring and evaluating the degree of contamination in the aquatic systems (Gao et al., 2018). Increasing anthropogenic pressures and global warming induced latitudinal shifts in climate zones, lead to the modification in the hydrological regime with harmful inference for aquatic systems inclusive of planktons. This contemporary confronts may influence the near future advancements in phytoplankton related studies (Borics et al., 2020). The hydro-biological status of HALs acts as a reference system for complex biophysical and chemical processes occurring within their catchment areas (Steingruber and Colombo, 2006; Singh et al., 2008).

Previously, the studies on the phytoplankton community of the HALs were limited to large species like dinoflagellates and diatoms, as large mesh-size plankton nets were used (Tolotti et al., 2006). The findings guided to the intention that HALs are very low in productivity, and thus aren't capable of supporting a well-established phytoplankton community (Naumann, 1924; Monti and Stella, 1934). This hypothesis was not annulled till the late 1950s when the autotrophic nature of HALs was identified (Rodhe, 1962). This realization stirred a chain of intensive studies which considerably enhanced the acquaintance of plankton distribution and their ecological status in HALs. In Kashmir Himalaya, the limnological investigation on HALs began with the works of Hutchinson, (1937), Löffler, (1969), and (Zutshi et al., 1972). The restricted research that exists is limited in diffuse and at times in unpublished sources. Although, several studies have been conducted to understand biological diversity and hydro-chemical profile of the valley lakes in Kashmir Himalaya (Romshoo and Muslim, 2011; Parvez and Bhat, 2014; Nissa and Bhat, 2016; Rashid et al., 2017a; Khanday et al., 2018; Bashir et al., 2020; Dar et al., 2020a, b; Dar et al., 2021b), very few studies have focused on hydrobiological assessments of the HALs in Kashmir Himalaya (Zutshi, 1980; Zutshi and Vass, 1982; Zutshi, 1991; Shah et al., 2014). Thus, to fulfill this research gap the five HALs in the Kashmir Himalaya were selected to provide a comprehensive information on the LULC, bathymetry, hydro-chemical profile including the plankton diversity. The study will contribute to the further understanding of the hydro-chemical and biological (Phytoplankton and Periphyton) spectrum and physical status of alpine lake ecosystems and will also serve as a reference for assessing future alterations in these alpine systems of the Kashmir Himalayas especially in the climate change impact scenario.



## MATERIALS AND METHODS

### Study Area

In the Indian Himalayan region, a total of 4,703 HALs have been mapped, which are situated above 3,000 m, covering an area of 126,249 ha (SAC, 2011; Singh et al., 2016). In the Jammu and Kashmir state alone, there are about 2014 HALs. The present study is focused on the five HALs viz., Gangbal, Nundkul, Gadsar, Kishansar, and Vishansar (also known as the Great Lakes of Kashmir) and their associated streams in Kashmir Himalaya (Figure 1). Since these HALs are located in remote, tough mountain terrains, these are assessed only through marathon trekking. The scale of landscape makes the study area a moderately-difficult trek spanning over 10 days. The trek introduces many challenging situations and great skills are needed to overcome them. The trekking route passes through many picturesque flower-stern meadows, stupefied by the sight of Himalayan wildlife and striking flora. The study area is a conglomerate of wild landscapes, tough rugged mountains, and turquoise HALs. The lakes under investigation are fed by

glacier melt, precipitation, and springs. The Gangbal and Nundkol Lakes are situated at the foothills of Mount Haramukh in the Sindh basin of the Ganderbal district. Both the lakes are alpine high altitude oligotrophic lakes and home to many species of fish including the brown trout. Gadsar Lake also known as Yemsar Lake is a picturesque, alpine high altitude oligotrophic lake, the floating icebergs could be easily seen on its surface in the month of July-August. The Gadsar Lake outflows through a stream, north-westwards, and joins Neelum River at Tulail which flows to Pakistan. The Kishansar and Vishansar Lake are situated 10 kms away from the Sonamarg hill station. Kishansar Lake outflows through a small channel to the Vishansar Lake. The Vishansar Lake outflows in a north-westward direction up to Badoab and then westwards through Gurez along the Line of Control to Pakistan. The watershed as dominated by exposed barren and rocky areas, the vegetation is comprised of grasses, herbs, and a patch of forest on one side. The catchment area of all the lakes is heavily used as a grazing land by the herders and the area is also used for the collection of medicinal plants. Threats of pollution from both mountain

**TABLE 1** | Study area site description.

| S. No | Site description    | Site code | Altitude (m) | Coordinates                       |
|-------|---------------------|-----------|--------------|-----------------------------------|
| 1     | Gangbal lake        | GL        | 3582         | 34° 25' 56.76" N–74° 55' 19.02" E |
| 2     | Gangbal stream      | GS        | 3580         | 34° 25' 24.23" N–74° 56' 22.41" E |
| 3     | Nundkul lake        | NL        | 3505         | 34° 25' 01.93" N–74° 56' 02.97" E |
| 4     | Nundkul stream I    | NS1       | 3488         | 34° 25' 08.29" N–74° 56' 44.78" E |
| 5     | Nundkul stream II   | NS2       | 3467         | 34° 24' 40.39" N–74° 57' 50.55" E |
| 6     | Gadsar lake         | GSL       | 3584         | 34° 25' 18.76" N–75° 03' 27.05" E |
| 7     | Gadsar stream I     | GS1       | 3540         | 34° 25' 53.57" N–75° 03' 12.76" E |
| 8     | Gadsar stream II    | GS2       | 3424         | 34° 26' 51.26" N–75° 02' 34.65" E |
| 9     | Kishansar lake      | KL        | 3815         | 34° 23' 47.87" N–75° 06' 10.26" E |
| 10    | Kishansar stream    | KS        | 3832         | 34° 23' 32.68" N–75° 06' 36.95" E |
| 11    | Vishansar lake      | VL        | 3683         | 34° 23' 15.46" N–75° 07' 05.28" E |
| 12    | Vishansar stream I  | VS1       | 3677         | 34° 23' 32.24" N–75° 07' 41.01" E |
| 13    | Vishansar stream II | VS2       | 3657         | 34° 24' 17.04" N–75° 08' 11.34" E |

tourism and herders exist in the area. Animals such as bears, Snow leopard, Musk deer, and Yak have been reported in the area.

## Water Quality

The sampling for collection of water and biological samples was carried out in ice free period (July–August) during 2015 and 2017. A total of thirteen sites comprising five sampling sites from lakes and eight sampling sites from associated streams were selected for collection of water and biological samples (Figure 1). A detailed description of the selected sites is provided in Table 1. As the study sites are located in deep mountainous terrains, the sampling was carried out by trekking which extended for a period of 10 days each year. Collection, preservation, and transfer of samples to the laboratory was carried out as per the standard methods (APHA, 2017). A portable global positioning system (GPS) was employed to assess the geographical coordinates of the sampling stations. Water temperature (WT), electrical conductivity (EC), pH, and total dissolved solids (TDS) were measured with the assistance of the Digital multi-parameter probe (Eutech PCSTEST35-01x441506). Dissolved oxygen (DO) was determined by using Hannah's digital DO meter (Hanna HI 9146). Nitrate nitrogen ( $\text{NO}_3^-$ -N), nitrite nitrogen ( $\text{NO}_2^-$ -N), ammoniacal nitrogen ( $\text{NH}_3$ -N), orthophosphate phosphorus ( $\text{PO}_4$ ), total phosphorus (TP), sulphate ( $\text{SO}_4^{2-}$ ), and dissolved silica (DS) were estimated using the UV-VIS Spectrophotometer. Total hardness (TH), calcium hardness (Ca-H), magnesium hardness (Mg-H), total alkalinity (TA), and chloride ion ( $\text{Cl}^-$ ) were determined by titrimetric method, following the standard protocols (APHA, 2017).

## Morphometry and Bathymetric Analysis

Bathymetric maps for most of the HALs worldwide are still lacking, except for a small number of lakes that were measured mainly by sonar (Khattab et al., 2017). Kashmir Himalaya has a rich heritage of HALs. These lakes have high ecological, economic, and cultural importance but most of these are yet to be studied due to remote location, tough terrain, and lack of accessibility. A bathymetric survey was performed in August 2017 in the lakes under investigation to generate baseline information. Depth measurements were obtained from different locations in a boat using a graduated rope. The

lake bathymetry surfaces were created using natural neighbor interpolation (NNI)/Sibson's interpolant in ArcMap 10.1 (Ledoux and Gold, 2005; Dar et al., 2020b). The NNI uses measured values of points and their associated areas to calculate weights of the neighboring unsampled points. If we consider that each data point in S (a set of n points) has an attribute  $a_i$  (a scalar value), the NNI function is given as

$$f(x) = \sum_{i=1}^k w_i(x)a_i \quad (1)$$

where  $f(x)$  is the interpolated function value at the location  $x$  and  $w_i$  is the value or weight of a sample.

For the morphometric characterization of the lakes, the following morphometric parameters were obtained viz: lake area, maximum length ( $L_{\max}$ ), and maximum width ( $B_{\max}$ ). Values for these parameters were obtained from measurements performed with the ArcGIS 10.1 software (ESRI Inc.). The boundary of the lakes was digitized manually using high-resolution Google base map satellite imagery as a data source in ArcGIS 10.1 (Dar et al., 2021c). Area of the lakes was calculated using Calculate Geometry tool in ArcGIS 10.1. Scale tool was used to measure the  $L_{\max}$  along North-South axis and  $B_{\max}$  along East-West axis.

## LULC Analysis

Since, catchment characteristics determine the water quality of lakes, the main rationale behind the LULC analysis was to look over the percent land area available for grazing. This was pursued keeping in view the fact that the catchment area of the lakes under investigation undergoes heavy grazing and browsing and it transports all the nutrients and fecal matter of cattle and livestock into the lakes. High resolution Google base map imagery was used for generating a LULC map. The imagery was co-registered for removal of geometric incongruity. Owing to the high resolution of the satellite data, six LULC classes including forest, lakes, pastures, barren rocky, snow/glaciers and streams were delineated using digital interpretation method at 1:2000 scale in ArcMap 10.1 (Rashid et al., 2016). The methodological sequence consisted of the following steps: image selection, digital processing, interpretation, ground

truthing, quantification of land use classes, and qualitative analysis of the results.

## Phytoplankton and Periphyton Collection, Identification, and Enumeration

Phytoplankton samples for qualitative and quantitative analysis were collected from different lakes. Phytoplankton net with mesh size of 64  $\mu\text{m}$  was employed for filtration and a volume of 50 L of water were filtered from each site. The Periphytic algal samples were collected by scraping the 3  $\text{cm}^2$  surface area of stones, boulders etc. using blade, scale, and brushes from stream sites. The material collected was then stored in vials having capacity of 25 ml. Samples were preserved in 4% formalin and algal identification and enumeration was carried out with the help of standard taxonomic works (Prescott, 1939; Edmondson, 1992; Cox, 1996; Jumppanen, 1976; Biggs and Kilroy, 2000; APHA, 2017). 1 ml of preserved phytoplankton sample was placed in Sedgwick-Rafter (SR) cell, it was allowed to settle for about 15 min and counted. Counting and identification of phytoplankton community was carried out using Binocular microscope (Magnus MLX-DX Olympus (India) pvt. ltd.) at 10x and 40x magnification. Samples were discarded and replaced by diluted ones if it was too dense to count, 1 cm of filamentous organisms was taken as one individual (Ahmed et al., 2017). Phytoplankton diversity indices were estimated using Paleontological statistics software for education and data Analysis (PAST 3.21 software) (Hammer et al., 2001).

## Water Quality Index

Water quality index (WQI) is appropriate mathematical approach for assessing the quality of water (Sanchez et al., 2007; Sener et al., 2017; Khanday et al., 2021). To each WQ parameter, weights were given ranging from 1 (least impact on WQ) to 5 (highest impact on WQ).

WQI is estimated by the following equations:

$$Rw_i = \frac{w_i}{\sum_{i=1}^n w_i} \quad (2)$$

where,  $Rw_i$  is relative weight,  $w_i$  is WQ parameter weight and  $n$  is the total number of WQ parameters. After that to every WQ parameter quality rating is assigned as below:

$$Qr_i = \frac{C_i}{S_i} \times 100 \quad (3)$$

where  $Qr_i$  is the quality rating,  $C_i$  is WQ parameter concentration in each sample and  $S_i$  is drinking water standard for every parameter as per the guidelines of Bureau of Indian Standards (BIS) and World Health Organization (WHO).

Then, first sub-index ( $SI_i$ ) is calculated as:

$$SI = Rw_i \times Qr_i \quad (4)$$

Where in,  $SI$  is the sub-index of  $i$ th parameter, relative weight and quality rating of  $i$ th parameter are given by  $Rw_i$  and  $Qr_i$  respectively.

Then water quality index is calculated as

$$WQI = \sum_{i=1}^n SI \quad (5)$$

The calculated WQI values are classified into five classes as: <50 as excellent water quality category; 50–100 highlights good water quality category; 100–200 denotes poor water quality category; 200–300 symbolizes very poor water quality class and >300 specifies the water quality category which is unsafe for drinking (Sanchez et al., 2007; Khanday et al., 2021).

## Statistical Treatment

Cluster analysis (CA) a multivariate statistical tool was applied to data sets to find the resemblance and dissimilarities among the investigated sampling sites on the basis of physico-chemical parameters using Euclidean distance (Dar et al., 2021e). In order to evaluate whether the dataset contains meaningful cluster as highlighted in the cluster plot. “Hopkins’s statistic” was used to assess the clustering tendency in the dataset using 0.5 as threshold. Cluster validation was carried out using Silhouette coefficient measures, which indicates how similar or different an object is in its own cluster or neighboring cluster. Wilk’s  $\lambda$  test was performed to check which physico-chemical parameters were mainly responsible for the clustering of the sites. Correlation plots were plotted for determining the relation between the physicochemical parameters ( $p < 0.05$ ). Canonical correspondence analysis (CCA) is a linear depiction learning technique that seeks maximally correlated parameters in multi-view data and this method was employed to establish the association among algal assemblages and various environmental variables (Devi et al., 2016). CA, clustering tendency, cluster validation, Wilk’s  $\lambda$  and correlation, was performed using R statistical package with “NbClust”, “cluster”, “clustertend”, “factoextra”. “fpc”, and “cVvalid” packages and SPSS-16. CCA was performed using package “lab dsv” (Roberts, 2016).

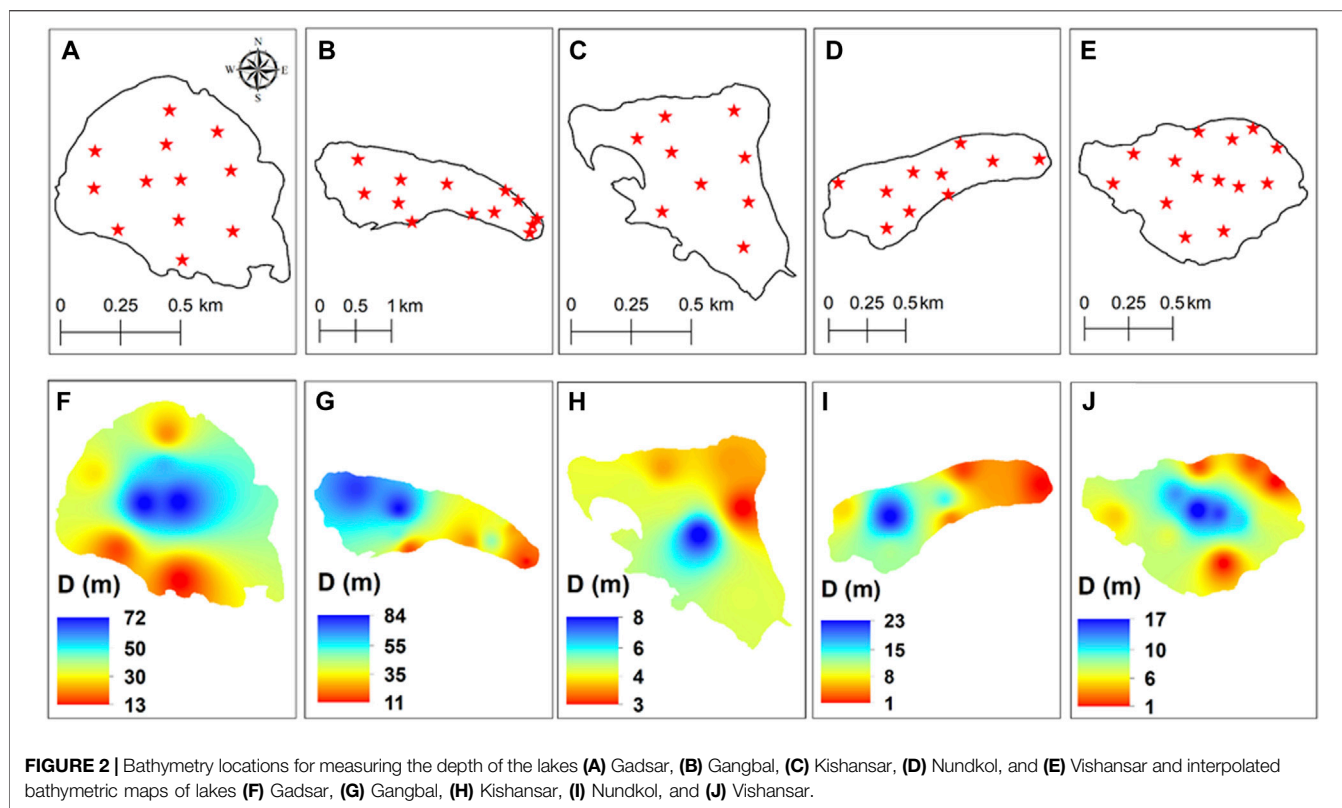
## RESULTS

### Water Quality

The results of the different physico-chemical parameters of the five HALs measured from 13 sampling sites are presented in Table 2. The WT varied from 9–14°C. The lakes and streams were having good concentration of dissolved oxygen ranging from 7.6–8.4  $\text{mg L}^{-1}$ . The pH of the lakes highlighted neutral to alkaline pH with values ranging from 7.3 to 8.1. The lakes showed the lowest concentration of dissolved ions with EC values ranging from 15–53  $\mu\text{S cm}^{-1}$ , and TDS values ranging from 10–37  $\text{mg L}^{-1}$ . Total alkalinity values ranging from 36–70  $\text{mg L}^{-1}$  is indicative of the good buffering capacity of the lakes. The order of equivalency of cations followed the order of  $\text{Ca}^{2+} > \text{Mg}^{2+} > \text{Na}^+ > \text{K}^+$ , while as anions followed the order of  $\text{HCO}_3^{2-} > \text{Cl}^- > \text{SO}_4^{2-}$ . The concentration of chloride content ranging from 8–20  $\text{mg L}^{-1}$  in the lakes clearly shows the role of catchment characteristics being

**TABLE 2 |** Showing the variation in physicochemical parameters in the five HALs and associated streams.

| Sites                            | GL  |      | GS  |      | NL  |      | NS1  |      | NS2  |      | GSL |     | GS1  |      | GS2 |     | KL   |      | KS   |      | VL   |      | VS1  |     | VS2  |       | Range    | BIS 2012 standards | WHO 2017 standards |
|----------------------------------|-----|------|-----|------|-----|------|------|------|------|------|-----|-----|------|------|-----|-----|------|------|------|------|------|------|------|-----|------|-------|----------|--------------------|--------------------|
|                                  | 1   | 2    | 1   | 2    | 1   | 2    | 1    | 2    | 1    | 2    | 1   | 2   | 1    | 2    | 1   | 2   | 1    | 2    | 1    | 2    | 1    | 2    | 1    | 2   | 1    | 2     |          |                    |                    |
| pH                               | 7.5 | 7.33 | 7.4 | 7.46 | 7.6 | 7.5  | 7.8  | 7.8  | 7.9  | 8    | 7.8 | 7.4 | 8.1  | 7.8  | 8   | 7.9 | 7.6  | 7.31 | 7.9  | 7.32 | 7.6  | 7.51 | 7.9  | 7.6 | 8    | 7.9   | 8.1–7.31 | 6.5–8.5            | 6.5–8.5            |
| EC (µS/cm)                       | 22  | 20   | 15  | 40   | 16  | 24   | 16   | 28   | 16   | 30   | 18  | 30  | 25   | 34   | 53  | 38  | 28   | 30   | 28   | 37   | 34   | 40   | 35   | 38  | 37   | 42    | 53–15    | –                  | –                  |
| DO (mg L <sup>-1</sup> )         | 7.8 | 7.8  | 8   | 8.2  | 8.4 | 8.4  | 7.8  | 7.6  | 8    | 8    | 8   | 7.8 | 8.2  | 8.2  | 8.5 | 8.4 | 7.8  | 7.8  | 8    | 8    | 8    | 7.8  | 8    | 8   | 8.2  | 8.2   | 8.5–7.6  | –                  | –                  |
| TDS (mg L <sup>-1</sup> )        | 15  | 14   | 11  | 28   | 10  | 16   | 11   | 17   | 12   | 18   | 13  | 18  | 28   | 19   | 37  | 29  | 19   | 18   | 19   | 22   | 23   | 25   | 24   | 23  | 22   | 24    | 37–10    | 500                | –                  |
| T A (mg L <sup>-1</sup> )        | 70  | 68   | 56  | 64   | 36  | 48   | 50   | 58   | 52   | 60   | 46  | 59  | 48   | 56   | 50  | 60  | 52   | 58   | 48   | 60   | 44   | 60   | 46   | 54  | 52   | 58    | 70–36    | 200                | 500                |
| Cl (mg L <sup>-1</sup> )         | 12  | 11   | 14  | 10   | 17  | 12   | 15   | 14   | 15   | 14   | 20  | 18  | 17   | 18   | 16  | 14  | 17   | 14   | 13   | 12   | 14   | 13   | 19   | 11  | 18   | 8     | 20–8     | 250                | 250                |
| TH (mg L <sup>-1</sup> )         | 120 | 122  | 124 | 124  | 120 | 122  | 122  | 118  | 124  | 116  | 118 | 120 | 112  | 122  | 114 | 114 | 108  | 120  | 94   | 128  | 106  | 124  | 110  | 128 | 104  | 122   | 128–94   | 300                | 200                |
| Ca C (mg L <sup>-1</sup> )       | 34  | 44   | 38  | 37   | 36  | 37   | 37   | 33   | 35   | 34   | 34  | 41  | 36   | 38   | 34  | 37  | 37   | 39   | 29   | 40   | 32   | 41   | 34   | 40  | 29   | 41    | 44–29    | 75                 | –                  |
| Mg C (mg L <sup>-1</sup> )       | 24  | 11   | 20  | 21   | 20  | 20   | 24   | 24   | 22   | 22   | 13  | 16  | 19   | 19   | 16  | 12  | 17   | 16   | 19   | 18   | 16   | 18   | 19   | 20  | 14   | 24–11 | 30       | –                  |                    |
| Nitrate (µg L <sup>-1</sup> )    | 142 | 152  | 130 | 134  | 124 | 110  | 88   | 74   | 68   | 64   | 208 | 192 | 176  | 178  | 160 | 162 | 184  | 164  | 188  | 182  | 160  | 142  | 140  | 138 | 138  | 126   | 208–64   | 45000              | –                  |
| Nitrite (µg L <sup>-1</sup> )    | 60  | 55   | 54  | 48   | 38  | 42   | 34   | 36   | 30   | 34   | 68  | 82  | 60   | 58   | 46  | 64  | 44   | 52   | 58   | 54   | 60   | 64   | 52   | 60  | 48   | 56    | 82–30    | –                  | –                  |
| A Nitrogen (µg L <sup>-1</sup> ) | 84  | 82   | 70  | 64   | 40  | 50   | 37   | 44   | 35   | 38   | 74  | 108 | 70   | 76   | 70  | 82  | 90   | 92   | 84   | 76   | 72   | 68   | 60   | 64  | 58   | 55    | 108–35   | –                  | –                  |
| TP (mg L <sup>-1</sup> )         | 80  | 76   | 64  | 112  | 56  | 48   | 38   | 52   | 44   | 46   | 168 | 158 | 142  | 130  | 140 | 124 | 106  | 128  | 63   | 138  | 152  | 140  | 124  | 134 | 60   | 94    | 168–38   | –                  | –                  |
| OP (mg L <sup>-1</sup> )         | 58  | 56   | 48  | 72   | 39  | 28   | 26   | 32   | 31   | 28   | 60  | 62  | 39   | 32   | 43  | 40  | 27   | 42   | 46   | 52   | 58   | 58   | 48   | 46  | 43   | 56    | 72–26    | –                  | –                  |
| S (mg L <sup>-1</sup> )          | 4   | 4    | 2   | 4    | 3   | 4    | 4    | 3    | 5    | 5    | 3   | 4   | 5    | 6    | 2   | 4   | 4    | 5    | 2    | 5    | 5    | 4    | 3    | 4   | 4    | 4     | 6–2      | 200                | 250                |
| Si (mg L <sup>-1</sup> )         | 6   | 6    | 4   | 8    | 4   | 7    | 5    | 5    | 5    | 4    | 3   | 4   | 4    | 5    | 5   | 4   | 3    | 4    | 6    | 12   | 2    | 4    | 3    | 4   | 6    | 5     | 12–2     | –                  | –                  |
| Velocity (m s <sup>-1</sup> )    | 0   | 0    | 0.7 | 0.7  | 0   | 0    | 1.1  | 1.2  | 0.9  | 1    | 0   | 0   | 0.71 | 0.68 | 1.1 | 0.9 | 0    | 0    | 0.5  | 0.4  | 0    | 0    | 1    | 0   | 1    | 1.2   | 1.2–0    | –                  | –                  |
| Discharge (m s <sup>-1</sup> )   | 0   | 0    | 6   | 5.22 | 0   | 0    | 10.1 | 12.4 | 10.8 | 14.6 | 0   | 0   | 5.2  | 4    | 5.7 | 6.1 | 0    | 0    | 0.7  | 0.8  | 0    | 0    | 3.27 | 0   | 3.27 | 4.44  | 14.6–0   | –                  | –                  |
| WT (C)                           | 12  | 12   | 13  | 13   | 13  | 12.8 | 13.2 | 13   | 13.2 | 13.4 | 10  | 9   | 10   | 10.4 | 9   | 9   | 13.2 | 13.4 | 13.6 | 13.4 | 13.8 | 13   | 13   | 13  | 14   | 14    | 14–9     | –                  | –                  |



**TABLE 3 |** Morphometric measurements of the five high altitude lakes in Kashmir Himalaya.

| Parameters                  | Gadsar | Gangbal | Kishansar | Nundkol | Vishansar |
|-----------------------------|--------|---------|-----------|---------|-----------|
| Direction of major axis     | E-W    | NW-SE   | NW-SE     | NE-SW   | E-W       |
| Altitude (m)                | 3264   | 3612    | 3819      | 3505    | 3949      |
| Area (ha)                   | 39.8   | 162.4   | 31.9      | 37.4    | 47.2      |
| Maximum depth $Z_{max}$ (m) | 72     | 84      | 8         | 23      | 17        |
| Minimum depth $Z_{min}$ (m) | 13     | 11      | 3         | 1       | 1         |
| $L_{max}$ (m)               | 830    | 1135    | 875       | 561     | 780       |
| $B_{max}$ (m)               | 944    | 2605    | 825       | 1317    | 1266      |

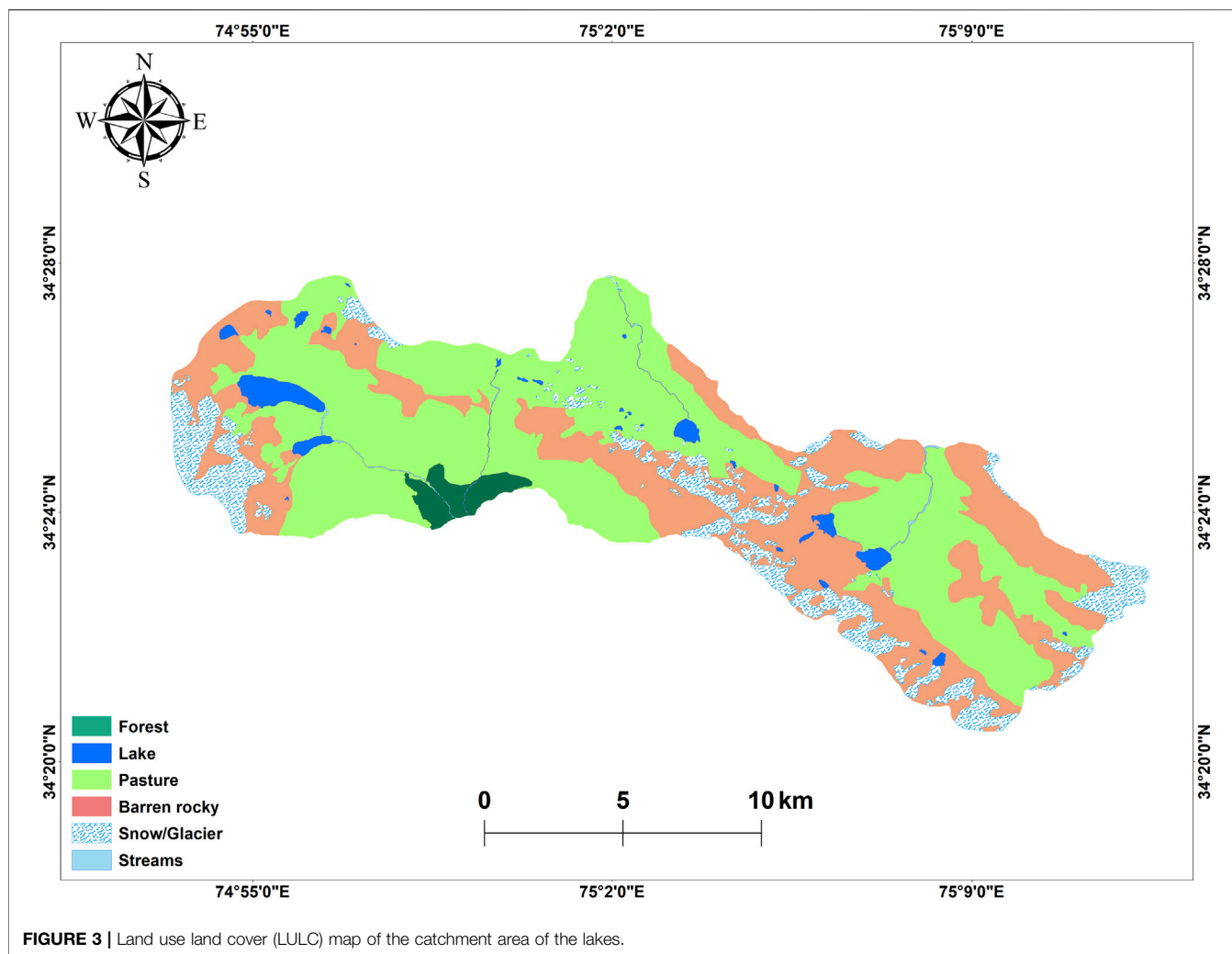
largely exposed to extensive grazing activities by cattle and livestock (Gurung et al., 2018). All the nutrients recorded low concentration in the lakes with  $\text{NO}_2^-$ -N values ranging from 82–30  $\mu\text{g L}^{-1}$ ,  $\text{NO}_3^-$ -N between 64 to 208  $\mu\text{g L}^{-1}$ , and  $\text{NH}_3$ -N ranging from 35 to 108  $\mu\text{g L}^{-1}$ . During the present study, concentration of  $\text{PO}_4$  ranged from 21–72  $\mu\text{g L}^{-1}$  and that of total phosphorus ranged from 32–168  $\mu\text{g L}^{-1}$ . Measurements of the flow data revealed that Nundkul stream has the highest discharge (14.6  $\text{m}^3 \text{S}^{-1}$ ) and velocity (1.2  $\text{m S}^{-1}$ ) among all the investigated streams. The measurements of the WQI revealed that all the 13 sites fall within the excellent category with the WQI values ranging from 23.3–39.9 (Supplementary Table S1).

It is evident from the Wilk's  $\lambda$  quotient Mg-H, TDS,  $\text{NO}_3^-$ -N,  $\text{NO}_2^-$ -N,  $\text{NH}_3$ -N, TP and ortho-P varied significantly ( $p < 0.05$ ) throughout the study period (Supplementary Table S2). Highest values of TDS (37  $\text{mg L}^{-1}$ ) and  $\text{Mg}^+$  (20  $\text{mg L}^{-1}$ ), were recorded at VS2, GS2. While as highest values of  $\text{NO}_3^-$ -N (208  $\mu\text{g L}^{-1}$ ),  $\text{NO}_2^-$ -N

(68  $\mu\text{g L}^{-1}$ ),  $\text{NH}_3$ -N (108  $\mu\text{g L}^{-1}$ ), TP (168  $\mu\text{g L}^{-1}$ ) and Ortho-P (60  $\mu\text{g L}^{-1}$ ) were recorded at GSL site.

## Bathymetry and Morphometry

The Bathymetric analysis of the lakes revealed that the Gangbal lake is the deepest lake with a maximum depth of 84 m (Figure 2). This lake has a crevice as the deepest zone near the Mount Harmukh in the western part of the lake. The altitudinal variation of the lakes ranges from 3,505 m to 3,832 m. The morphometric features of the lakes are provided in Table 3. All the lakes are situated in an asymmetrical basin. Gangbal and Nundkol lakes are crescent shaped and are fed by the snow and glacial melt originating from the Harmukh glacier located on the western side of both the lakes. Besides snow and glacial melt, Nundkol lake is also getting its inflow through an outflow channel from the Gangbal lake. The outlet from the Nundkol lake forms Wangath nallah which after traversing a distance of 22 kms



**TABLE 4 |** LULC characteristics of the catchment area of the five high altitude lakes during 2020.

| S. No | LULC feature      | Area (ha) | % Land cover |
|-------|-------------------|-----------|--------------|
| 1.    | Forest            | 324.94    | 2            |
| 2.    | Lakes             | 401.74    | 2.5          |
| 3.    | Pasture           | 8216.67   | 50.8         |
| 4.    | Rocky             | 5273.86   | 32.6         |
| 5.    | Snow and Glaciers | 1922.31   | 11.9         |
| 6.    | Streams           | 40        | 0.2          |
|       | Total             | 16179.5   |              |

joins the Sindh River at Kangan. Gadsar Lake is more or less circular in shape and receives the snow melt water from the glaciers on the southern side. An outflow channel from the Gadsar Lake joins Neelum river at Tulail which flows through the Pakistan. Kishansar and Vishansar lakes are asymmetrical in shape. Kishansar Lake is fed by the melting of snow and glaciers and drains out through a small stream which falls into the Vishansar Lake. There is one outflow from the Vishansar Lake forming a tributary of Neelum river flowing through Pakistan.

## LULC

Using on-screen digitization of High resolution Basemap imagery, six LULC classes which include Forest, Lakes, Pasture, Barren rocky, Snow/Glacier, and Streams were delineated. The LULC map is shown in **Figure 3**, and details are provided in **Table 4**. From the analysis of the 2018 satellite data, the catchment of the lakes comprises of an area of 16179.5 ha. The predominant LULC feature in the catchment of the lakes is covered by pastures covering an area of 8,216.7 ha (50.8%). These pastures are mostly used for grazing by cattle and livestock. Second predominant LULC feature in the catchment comprises of barren rocks covering an area of 5,273.9 ha (32.6%). Snow and glaciers covering 1922.3 ha (11.9%) of the catchment area, are the main contributing sources to lakes and streams. Lakes cover an area of 401.7 ha (2.5%), followed by forest 324.9 ha (2%), and least area was covered by streams 40 ha (0.2%).

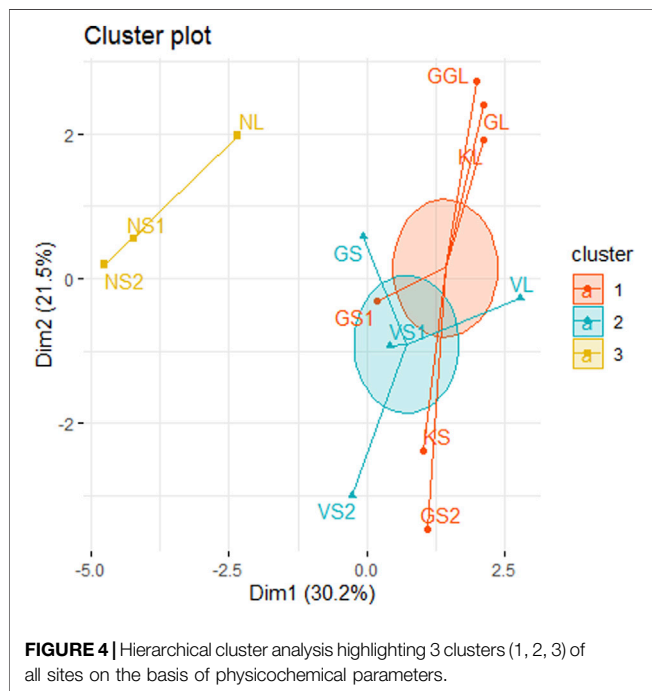
## Algal Composition

The analysis of the data collected from the 13 sites revealed the presence of 63 genera of planktonic algae wherein the class Bacillariophyceae was dominant with 46 taxa followed by

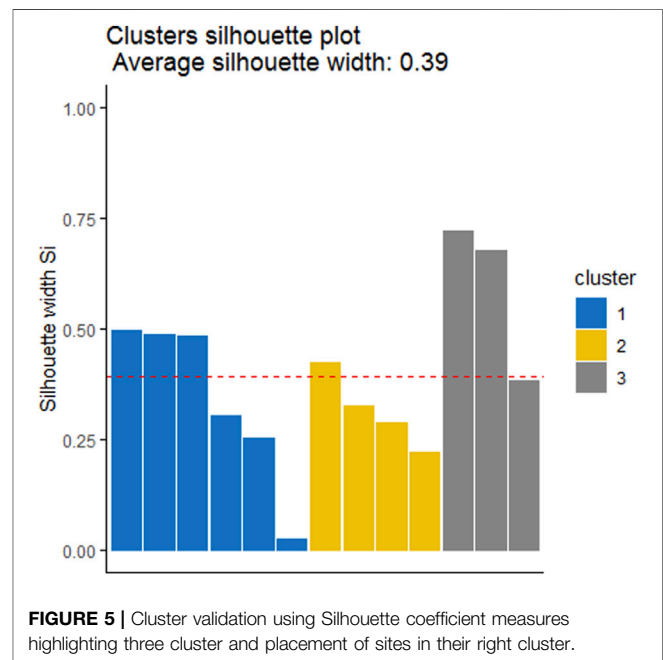


**TABLE 5** | Biodiversity indices of planktonic species collected at different sites during the course of study.

|               | GL   | GS   | NL   | NS1  | NS2  | GSL  | GS1  | GS2  | KL    | KS   | VL    | VS1  | VS2  |
|---------------|------|------|------|------|------|------|------|------|-------|------|-------|------|------|
| Taxa_S        | 35   | 27   | 35   | 33   | 32   | 40   | 32   | 33   | 35    | 38   | 38    | 34   | 36   |
| Individuals   | 9750 | 475  | 9170 | 788  | 867  | 7820 | 811  | 931  | 10340 | 802  | 10670 | 789  | 876  |
| Simpson_1-D   | 0.94 | 0.94 | 0.94 | 0.93 | 0.93 | 0.95 | 0.95 | 0.95 | 0.94  | 0.94 | 0.93  | 0.95 | 0.94 |
| Shannon_H     | 3.25 | 2.99 | 3.22 | 3.05 | 3.05 | 3.31 | 3.2  | 3.24 | 3.1   | 3.23 | 3.15  | 3.23 | 3.19 |
| Evenness_eH/S | 0.73 | 0.73 | 0.71 | 0.64 | 0.65 | 0.69 | 0.76 | 0.77 | 0.64  | 0.67 | 0.61  | 0.74 | 0.67 |

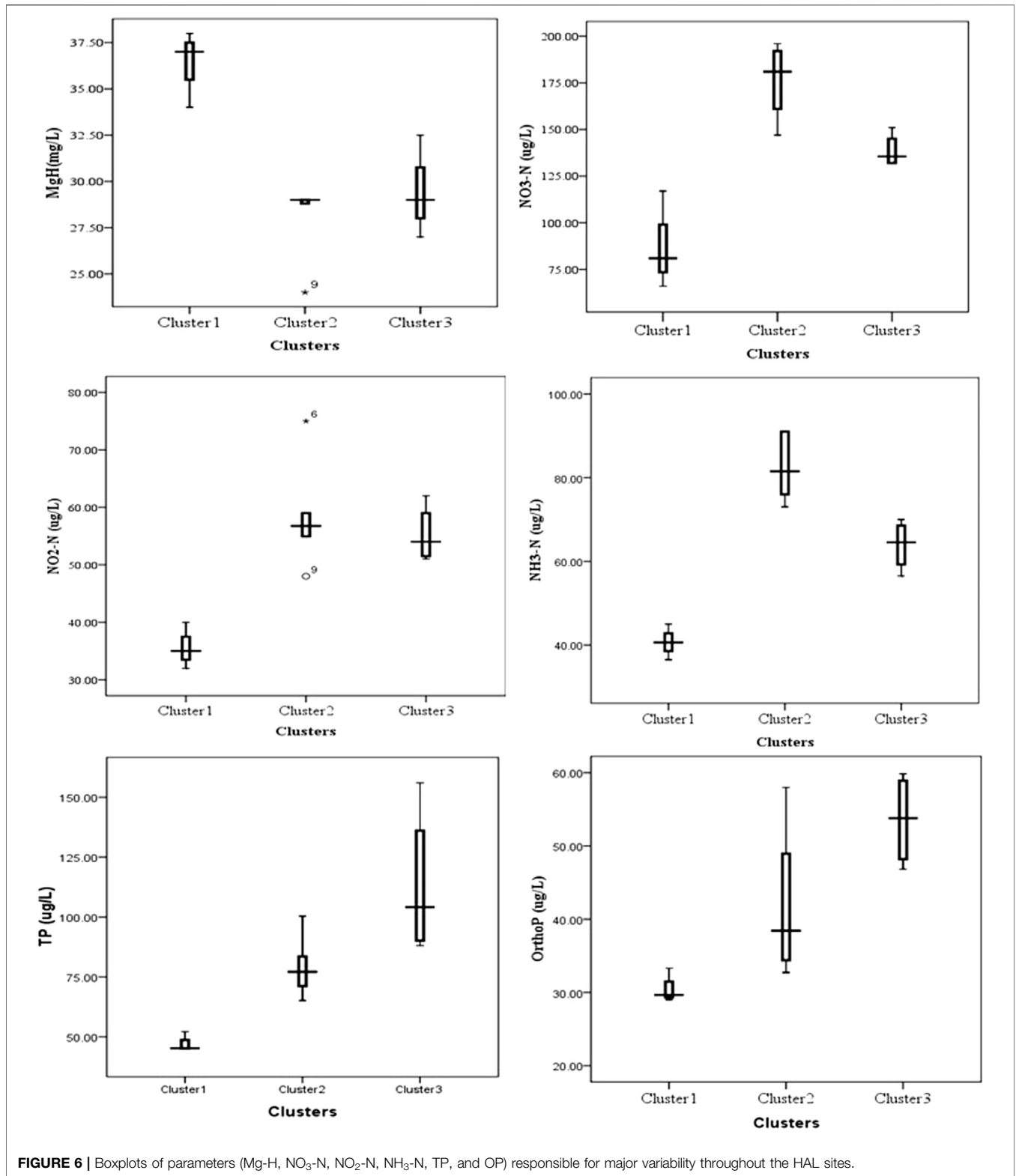


Chlorophyceae having 16 taxa, and 1 taxon of Xanthophyceae (**Supplementary Table S3, S4**). Floristically, Bacillariophyceae and Chlorophyceae were dominant in both the lake and the stream sites. Among Bacillariophyceae, *Amphora* sp., *Cymbella* sp., *Diatoma* sp., *Didymosphenia* sp., *Pinnularia* sp., *Navicula* sp., *Navicularadiosa*., *Fragilaria* sp., *Fragilariforma* sp., *Frustulia* sp. and *Gomphonema* sp., were dominant. In case of stream periphytic algae, *Cymbella aspera*, *Cymbella kappi*, *Cymbella* sp., *Diatoma* sp., *Diatomella* sp., *Gomphonema* sp., *Gomphonema minutum*, *Gomphonies* sp., *Navicula radiosa*, *Navicula* sp., were found to dominate the overall species diversity. The highest number of phytoplankton individuals were found in Kishansar Lake (12987 ind/ L), followed by Vishansar Lake (11334 ind/ L), and Gangbal Lake (9,750 ind/ L), whereas in case of High-altitude streams, Gadsar stream reported highest number of periphytic algal density (1,020 Ind/ cm<sup>2</sup>), followed by Vishansar stream (954 ind/ cm<sup>2</sup>). The Shannon-Weiner index didn't show much variation throughout the lakes and varied between 3.330 and 2.098 (**Table 5**). The Simpson and Evenness coefficients in both lakes and stream sites ranged between 0.9675 to 0.7754 and 0.7890 to 0.5178 respectively.



## Statistical Treatment

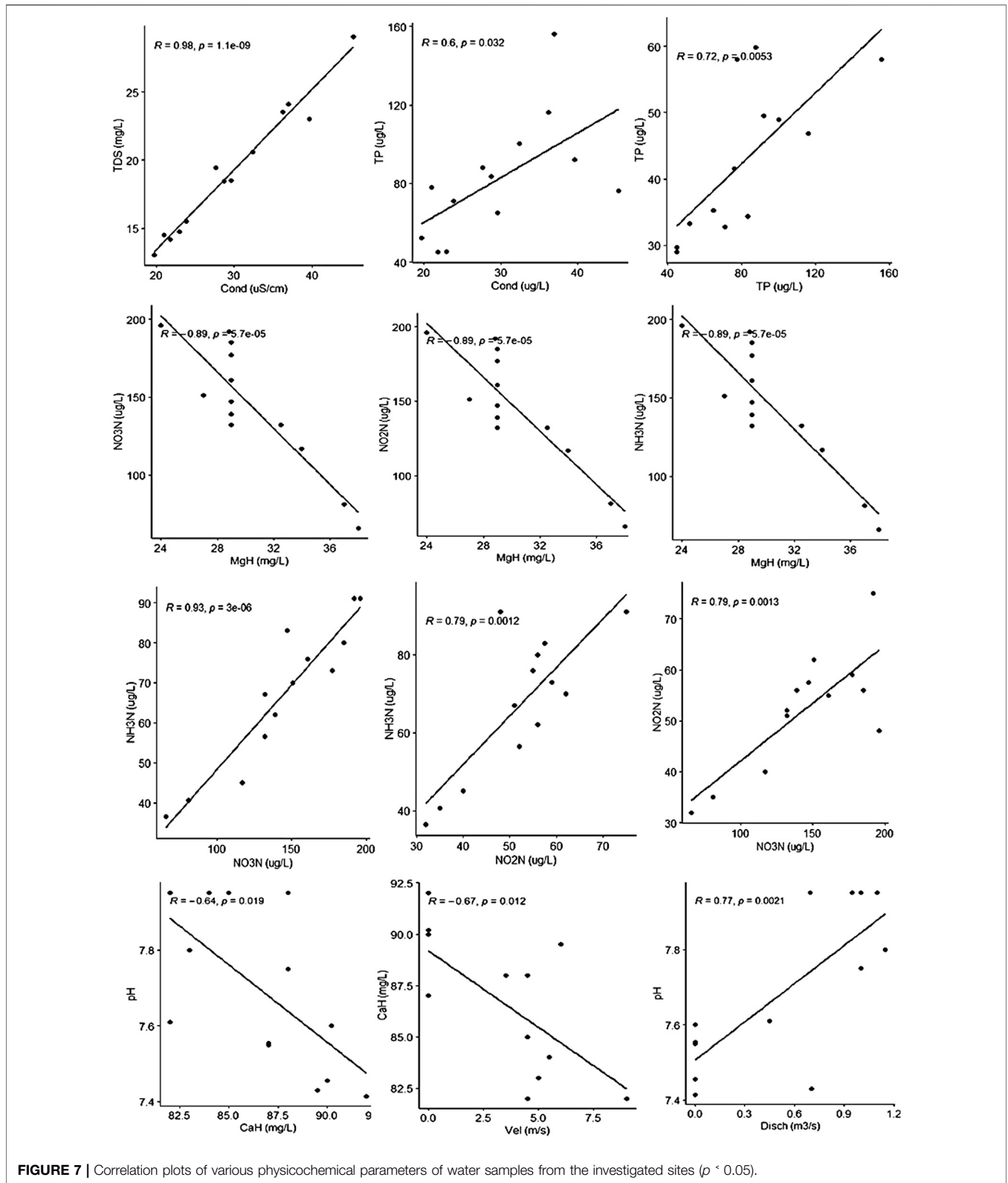
Cluster analysis on the basis of Hydro-chemical variables and algal assemblages categorized all the sites into three clusters 1, 2, and 3 (**Figure 4**). The cluster 1 comprises of sites Nundkul lake and Nundkul stream sites. Cluster two consists of Vishansar Lake and stream sites and Gangbal stream sites. Cluster three consists of Krishansar Lake and stream sites, Gadsar Lake and stream sites and Gangbal Lake site. The results of the Hopkins test are 0.60, which is greater than the threshold limit indicating that dataset is significantly clusterable. Three cluster solution gives us an average silhouette coefficient of 0.39 (**Figure 5**) Furthermore, no negative average silhouette coefficients were observed in three obtained clusters, indicating sites are present in the right cluster. Canonical discriminant functions depicted clear separation of sites into three clusters that were categorized by cluster analysis (**Figure 5**). Tests of Equality of Group Means (Wilk's  $\lambda$ ) depicted that nutrient concentration (TP, PO<sub>4</sub>, NH<sub>3</sub>-N, NO<sub>2</sub><sup>-</sup>-N, NO<sub>3</sub><sup>-</sup>-N) and Mg-H were major variables that are responsible for the variability among all selected sites (**Supplementary Table S2**). Boxplots depict that Mg-H was highest in cluster C1 comprising of Nundkul Lake and stream sites, while as nutrient concentration (TP, OP, NO<sub>2</sub><sup>-</sup>-N, NO<sub>3</sub><sup>-</sup>-N, and NH<sub>3</sub>-N) was highest at all remaining sites (**Figure 6**). Correlation plots depict that



**FIGURE 6 |** Boxplots of parameters (Mg-H, NO<sub>3</sub>-N, NO<sub>2</sub>-N, NH<sub>3</sub>-N, TP, and OP) responsible for major variability throughout the HAL sites.

parameter like TDS and TP were positively correlated with EC (Figure 7). Variables such as NO<sub>2</sub>-N, NO<sub>3</sub>-N, and NH<sub>3</sub>-N were found negatively correlated with Mg-H whereas, Ca-H was negatively correlated with velocity and pH. Discharge was

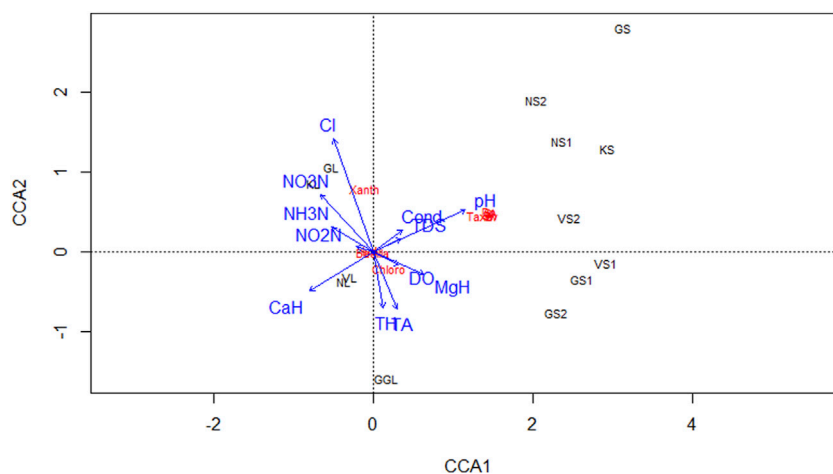
found to be in positive correlation with pH. CCA was conducted between density of the dominant genera belonging to algal assemblages and different physico-chemical parameters. This special set of scalars explained 61.55 and 27.86% of



**FIGURE 7 |** Correlation plots of various physicochemical parameters of water samples from the investigated sites ( $p < 0.05$ ).

correlation along axis 1 and 2 between physico-chemical parameters of water and algal groups respectively (**Figure 8**). The variables such as pH, conductivity, and TDS showed positive

correlation with the number of Taxa at the sampling stations VS2, NS1, NS2, KS, and GS whereas they were negatively correlated with Ca-H. During the study, it was observed that Xanthophyceae



**FIGURE 8** | Canonical correspondence analysis (CCA) biplot, between physicochemical parameters and algal composition at HAL sites.

were positively correlated with nitrate, nitrite, and ammoniacal nitrogen at sampling sites GL and KL. However, Chlorophyceae and Bacillariophyceae reflected the negative relationships with DO, Mg-H, TH, and TA at sampling sites GS1, GS2, GGL, and VS.

## DISCUSSION

Water quality of the lakes shows the variability in the behavior of ionic composition, nutrient dynamics, and catchment characteristics. pH values in all the HALs indicates the alkaline character, the results are in agreement with the work carried out on some other HALs of Kashmir valley Zutshi (1991) and Nepal region (Gurung et al., 2018). pH regulates the weathering processes and dissolved solids concentration in lake waters (Anshumali and Ramanathan, 2007). The variation in pH between different sites depends mainly on glacier melt water quality, and terrain (Deka et al., 2015), and biological processes such as photosynthesis and respiration (Hamid et al., 2020). EC is the assessment of the ionic potential and mainly depends on the magnitude and rate of movement of ionic species (Das and Kaur, 2001; Deka et al., 2015). Furthermore, EC indicates biotic and abiotic contamination in the water (Upadhyay et al., 2012; Sharma and Tiwari, 2018). EC varied from 15 to 53  $\mu\text{S}/\text{cm}$ , and was found to decrease with increase in the altitude, the results are in consonance with the findings of Zutshi et al. (1980), Sharma and Tiwari (2018) and Gurung et al. (2018). Lower values of EC of HALs could be credited to the lower values of all major ions.  $\text{Ca}^{2+}$  and  $\text{Mg}^{2+}$  ions were found to be the major dominant cations, while  $\text{HCO}_3^{2-}$  was the major dominant anion in all the HALs.  $\text{Ca}^{2+}$  and  $\text{Mg}^{2+}$  ions contribute 88% of total cationic composition while as  $\text{HCO}_3^{2-}$  contribute 70% of total anionic composition. The cations in decreasing order of concentration were  $\text{Ca}^{2+} >$

$\text{Mg}^{2+} > \text{Na}^+ > \text{K}^+$ , while as anions in decreasing order of concentration were  $\text{HCO}_3^{2-} > \text{Cl}^- > \text{SO}_4^{2-}$ . The higher concentration of  $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ , and  $\text{HCO}_3^{2-}$  suggests that weathering of carbonate rocks in the catchment areas is the major source of ions in these systems (Singh et al., 2016; Lone et al., 2019; Kaphle et al., 2021). Most of the Indian mountain lakes and glaciers also highlight that carbonate weathering is dominant process in these areas (Das and Kaur, 2001; Anshumali and Ramanathan, 2007; Singh et al., 2016). Ionic composition in freshwaters is mainly controlled by weathering process, atmospheric precipitation, and other existing atmospheric processes (Lacoul and Freedman, 2005; Anshumali and Ramanathan, 2007; Kaphle et al., 2021; Khadka and Ramanathan, 2021).  $\text{Ca}^{2+}$  is globally dominant cation in fresh water bodies (Wetzel, 2001; Gurung et al., 2018), but  $\text{Mg}^{2+}$  has also a significant role in defining water chemistry of lakes. The main source of these cations is the erosion of rocks (limestone and dolomite) and minerals (calcite and magnesite) (Hamid et al., 2020). The current study depicted that Mg-H is a dominant factor for clustering and categorizing the sites like Nundkul Lake and stream sites due to its higher values. Inorganic nitrogen forms ( $\text{NO}_3^-$ -N,  $\text{NO}_2^-$ -N, and  $\text{NH}_3$ -N), TP and  $\text{PO}_4$  throughout the lakes were determining factors responsible for clustering and variability among all the sites. The values obtained during the study for  $\text{NO}_3^-$  and ammonium ( $\text{NH}_4^+$ ) may be attributed to the hydrology of area which is snow-melt dominated and lead to seasonal spikes in the concentration (Campbell et al., 2000). Himalayan high altitude lake ecosystems are mostly phosphorus limited as the concentration of nitrogen is higher than that of phosphorus (Gurung et al., 2020). The concentrations of phosphorus in our study highlights that these lake ecosystems are phosphorus limited. The excellent WQI measured at all the sites may be attributed to their proximity to the water source, and less anthropogenic pressure (Yaseen et al., 2015). TDS is measurement of all

the dissolved particles including organic as well as inorganic substances, TDS was positively correlated with EC and TP. This highlights that the concentration of TP increases with the concentration of TDS and EC. Concentration of TDS, EC, and TP mainly depends on surface runoff and sediment deposition in these lakes from the catchment area (Sharma and Tiwari 2018). All the inorganic forms of nitrogen ( $\text{NO}_2^-$ -N,  $\text{NO}_3^-$ -N, and  $\text{NH}_3$ -N) were negatively correlated with the Mg-H. Ca-H was negatively correlated with stream velocity and pH at all the studied sites. Alpine stream outlets of all these lakes were having higher pH as evident from the correlation plot which shows discharge has a significant positive correlation with pH.

Land use changes directly influences the quality of freshwater ecosystems, particularly by enhancing the concentration of nutrient materials (Shi et al., 2017). Any alteration in the land cover of the watershed is considered as explanatory parameter behind the alterations of the hydrological unit, which ultimately changes runoff and water (Huang et al., 2013). The concentration of nutrients in the lake ecosystems are attributed to natural processes like heavy rainfall, soil erosion, and the runoff from the catchment areas which is laden with the fecal matter livestock (Kumar et al., 2019; Gurung et al., 2020). Same observations were recorded in the Sheshnag, a HAL in Kashmir Himalaya (Zutshi, 1991).

Bathymetric and morphometric analysis revealed that all the lakes are dammed glacier lakes, positioned in an asymmetrical basin. The lakes are deep because the lakes are formed in deep gorge like structures covered on all sides by the mountains and the water levels in these alpine systems are increasing presently, because of high melting of snow and glaciers in the catchment areas. The shoreline is gentle on the southern and eastern sides, but steep on the western and northern sides. Most of the shoreline portion in the lakes has no vegetation. A visual observation of the water markings in the riparian region of the lakes designates a seasonal water level fluctuation of 1–2 m as witnessed during sampling/field survey. The increase and changes in lake levels are in consonance with the results of Shekhar et al. (2010) and Srivastava et al. (2013) they attributed lake level fluctuations to seasonal snow and glacier melting due to increased air temperature in the northwestern Himalaya.

The land cover in the catchment of the lakes is dominated with meadows and pastures, glaciers, boulders/rocks, and eroding debris. Steep and eroding mountain slopes form the boundary of the lakes. Pastures are mainly covered by grasses and herbs are extensively used for grazing and browsing activities. There are bare land areas categorized as rocky in the catchment area. The Lakes are mainly fed by snow and glaciers which cover an area of 1922.31 ha. Forest area mainly comprises of Pinus, deodar, other herbs, and shrubby vegetation. There is also heavy grazing from July–September in the catchment area of the lakes.

Alpine lakes are remote and extremely sensitive ecosystems inhabited by scarce but well adapted species (Pastorino et al., 2020a). These lake ecosystems are characterized by small growing seasons, lower temperatures, longer ice periods, extended snow cover, higher elevations, and strong ultra-violet radiations. Small

variations in these characteristics results in strong reactions of biodiversity, community structure, and productivity of alpine lake phototrophs, therefore forming planktonic biota sensitive senitals of minute environmental variations (Moser et al., 2019; Kuefner et al., 2021). Some studies on diversity and assemblage pattern of phytoplankton in HALs of Kashmir had been carried out by (Hutchinson, 1937); (Löffler, 1969); (Zutshi et al., 1972). As pollution is not the dominant factor in our study, altitude is one of the important parameters which structures the planktonic composition. The altitude determines its surface freezing duration (Pourriot et al., 1995), which may impact lake biota and thus diatom species composition. As only few diatomic species are able to survive in these extreme habitats, may explain to some extent the low diversity reported in this study (Lotter and Bigler, 2000). The low diversity is normally attributed to low productivity of the lakes and harsh environmental conditions Cartuche et al. (2019) but equally the number of samplings conducted may also have role to play in underestimation of the diversity. The overall phytoplankton composition and abundance suggest oligotrophic status of these lakes (Kumar et al., 2012). Low temperatures together with the low conversion efficiency of solar energy by phytoplankton have been observed to be responsible for low productivity elsewhere (Talling, 1965; Khan, 1986). One of the prominent characteristics of phytoplankton and periphyton assemblages of the HALs is the dominance of Bacillariophyceae (Shah et al., 2014; Bashir et al., 2020) in lakes and wetlands. This may be attributed to their ability to get adapted to lower temperatures than other algae (Lüring et al., 2013). Nutrient enrichment significantly alters the structure and functions of primary producers. Anthropogenic activities have been observed to influence on algal composition (diversity and density pattern) and biomass trends (Dubey and Dutta, 2020). The fallout of the grazing and tourist activity which currently is very low needs to be taken care of on long-term basis to protect and safeguard the environmental quality of these important lakes.

## CONCLUSION

The range of physico-chemical parameters and water quality index calculated after accumulating various hydro-biological parameters corroborate the excellent water quality of the investigated HALs. Bathymetric analysis revealed that all the lakes are deep with Gangbal Lake attaining a maximum depth of 84 m. The catchment of the lakes is dominated by pastures and the area undergoes grazing during summers which to some extent affects the nutrient concentration of the lakes. Phytoplankton being dominated by Bacillariophyceae would require more in-depth studies to reach some clear conclusion on diversity and productivity of such lakes. We anticipate the increasing grazing activity and the renewed tourist attraction towards these lakes can add to water quality degradation which needs attention of the concerned agencies at an early stage.

## DATA AVAILABILITY STATEMENT

The original contributions presented in the study are included in the article/**Supplementary Materials**, further inquiries can be directed to the corresponding author.

## AUTHOR CONTRIBUTIONS

SI, SD, MS, and SUB conceived, designed, and executed this research work. SI, SD, MS, SUB, AJ, AB carried out the field sampling while SI, SD, MS, IS, and AH carried out the analysis of the samples in the laboratory. SI, SD, and MS wrote the original manuscript draft reviewed and edited by SUB. All the authors at the end read the manuscript and approved the same for its submission for publication.

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## SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fenvs.2021.681965/full#supplementary-material>

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