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Changes in land use and ecosystem service value in desert areas of China after reform and opening up

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Desert areas of China have important ecological functions, so analyzing changing characteristics of ecosystem service values in this region is important for sustainable development. Using land use change data for desert areas from 1978 to 2022, this paper combined the equivalent factor method and the actual situation of the study area to revise value coefficients of ecosystem services and study characteristics of land use and ecosystem service value changes after reform and opening up. The results showed that after reform and opening up, the area of plow, water, and other lands in the study area decreased, while grassland, forest, construction land, and sandy land increased. The conversion of a land use type mainly occurred as conversion of grassland and other lands to other land use types. The value of ecosystem services increased, increasing by 19.63×10^8 CNY. From the perspective of land use type in 2022, the ecosystem service value of grassland was the highest, reaching 12.19×10^8 CNY, an increase of 11.73×10^8 CNY compared with the early stage of reform and opening up. From the perspective of ecosystem service types in 2022, the value of maintaining soil ecological services was the highest, reaching 3.07×10^8 CNY, an increase of 2.97×10^8 CNY compared with the beginning of reform and opening up. From the perspective of the ecological sensitivity index, the sensitivity index results for the ecosystem service value in the study area were all <1 , and the research results were credible. From the perspective of the land ecological coordination degree, the overall land ecological coordination degree in the study area was at a moderate coordination level, indicating that land use change did not lead to environmental deterioration, but there was a crisis. Therefore, increasing and maintaining ecological land is the key to improving ESV in the study area.

KEYWORDS

land use, land cover change, ecosystem service value, value equivalent factor method, desert areas of China, reform and opening up

1 Introduction

Ecosystem services refer to the natural environmental conditions and utility that human beings depend on for survival. They are formed and maintained by ecosystems and ecological processes (Aryal et al., 2022). The main indicators of ecosystem services include supply, adjustment, support, and cultural services. With the in-depth study of global change, people find that land use change is an important reason affecting the global environmental change

(Chen et al., 2018). Land use change not only affects the climate change but also further affects the spatial distribution of the ecosystem service value and the ecological process of landscape pattern change (Zheng et al., 2008; Chen et al., 2019; Xie et al., 2020). Land use change plays a decisive role in maintaining the services of an ecosystem (Fang et al., 2022).

With the deepening of social research on the ecosystem service function, “environmental service function” has gradually evolved into the ecological service value (Takacs and O’Brien, 2023). The method of assessing the value of ecological services was systematically elaborated in the 1997 book “The Services of Nature: Society’s Dependence on Natural Ecosystems.” In 1999, Costanza estimated the economic value of 17 types of global ecosystems, which clarified the development of ecological service function value assessments. The principles and methods of this research provided new ideas for many global scholars to carry out further research. Subsequently, Camacho studied the impact of land use in coastal cities on the value of ecological services (Camacho-Valdez et al., 2014). Based on the experimental method, Harpinder calculated the ecological service value of organic cultivated land (Harpinder et al., 2007). Costanza’s research results have played a significant role in promoting the research progress on ecological service value in China. In 2003, Xie Gaodi et al. carried out a study evaluating the ecological service value of various grasslands on the Qinghai–Tibet Plateau (Xie et al., 2017). In 2008, Xie Gaodi et al. constructed an “ecosystem service value (ESV) equivalent factor table” applicable to China’s regional situation (Xie et al., 2017). The ESV equivalent method, which is based on the current scale developed by experts, has some credibility. Based on this method, China has studied ESV from the scale of river basins (Liu et al., 2023), urban agglomerations (Gong et al., 2023), provinces and cities (Normyle et al., 2023), the country (Sharma et al., 2023), etc. At present, the use of an economic value assessment mainly follows the ESV assessment model established by Costanza et al. According to the actual situation in China, Xie Gaodi et al. developed a table of ecological service value per unit area based on this model, which has been widely used (Jia et al., 2019; Yang et al., 2020).

Desert areas are unique natural ecosystems in northern China (Xin et al., 2023) and play an important role in national ecological security. Since the reform and opening up, with the rapid transformation of China’s social economy and rapid urbanization, rapid change in land circulation and land use types has been triggered (Liu et al., 2023). At the same time, ecological and environmental problems such as soil degradation and soil erosion in the area have become increasingly prominent. Land use will not only cause major changes to the earth’s surface structure and reshape landscapes but also affect the supply capacity of regional ecosystem services by changing regional climate, hydrology, soil, etc. (Hu et al., 2023) and driving the value of ecosystem services to respond positively or negatively (Tiandraza et al., 2023). However, since 2012, China has placed the construction of ecological civilization in a prominent position overall, and the environment in the area has significantly improved. Therefore, it was important to carry out a long-term quantitative evaluation and provide a scientific theoretical basis for regional sustainable utilization and improvement of ecosystem services. In summary, this study selected China’s desert region as the study area, based on the land use data on the study area for many years and supported by RS and GIS technologies (Jia et al., 2018), adopted the service value measurement method of China’s terrestrial

ecosystem to analyze and process the land use data on the study area, and studied the change characteristics and rules of land use and ecosystem service value in the desert region since the reform and opening up. The results are of great significance for the policy formulation of ecosystem protection and restoration, and the coordination of the relationship between regional economic development and environmental protection.

2 Study area

Desert areas of China (30°–50° N, 75°–130° E) cover an area of 826,072.70 km² (Dong et al., 2013; Ma et al., 2022), specifically distributed in China’s inland arid and semi-arid climate regions, including Heilongjiang, Jilin, Liaoning, Inner Mongolia, Shanxi, Ningxia, Gansu, Qinghai, and Xinjiang (Wu et al., 2023) (Figure 1). Influenced by the topography, such as the Tibetan Plateau and Mongolian Plateau, climatic conditions, and atmospheric circulation, annual precipitation gradually decreases from southeast to northwest. The average annual precipitation in the northwest arid region was less than 200 mm, and the average annual precipitation in the eastern semi-arid region was less than 500 mm. The wind force in the wind season was greater than level 5, with 20–100 sandstorm days and annual average temperatures of –5–20°C. There were 2,600–3,400 h of sunshine, and the frost-free period is 150–260 days. The underlying strata in the area provide abundant sand sources for the formation of deserts, including river alluvial deposits, alluvial lacustrine deposits, alluvial deposits, and weathered residual deposits of bedrock.

3 Materials and methods

3.1 Data sources and pre-processing

Using the Google Earth Engine (GEE) platform, Landsat sequence images of the study area were obtained for 1978, 1990, 2000, 2010, and 2022. The spatial reference is the Krasovsky coordinate Albers projection, and the spatial resolution of all the calculated indicators was resampled to 1 km × 1 km raster data obtained from the data source (Jia et al., 2012). According to the conditions of the study area, land use types were divided into eight categories: plow, forest, grassland, water, construction land, sandy land, and other lands. The deep learning method was adopted for interpretation, and the accuracy of the interpretation results was evaluated using existing data products (Yang and Huang, 2021). The random point method was used to evaluate the results, and the precision was 97.23%.

3.2 Research method

3.2.1 Revision of the ESV model

The biomass of each land use type in the study area has the following relationship with the net primary productivity of the ecosystem (Fang et al., 2007; Guo et al., 2021; Wang et al., 2022; Zhou et al., 2023).

$$h(\text{plough, grassland}) = NPP/0.45, \quad (1)$$

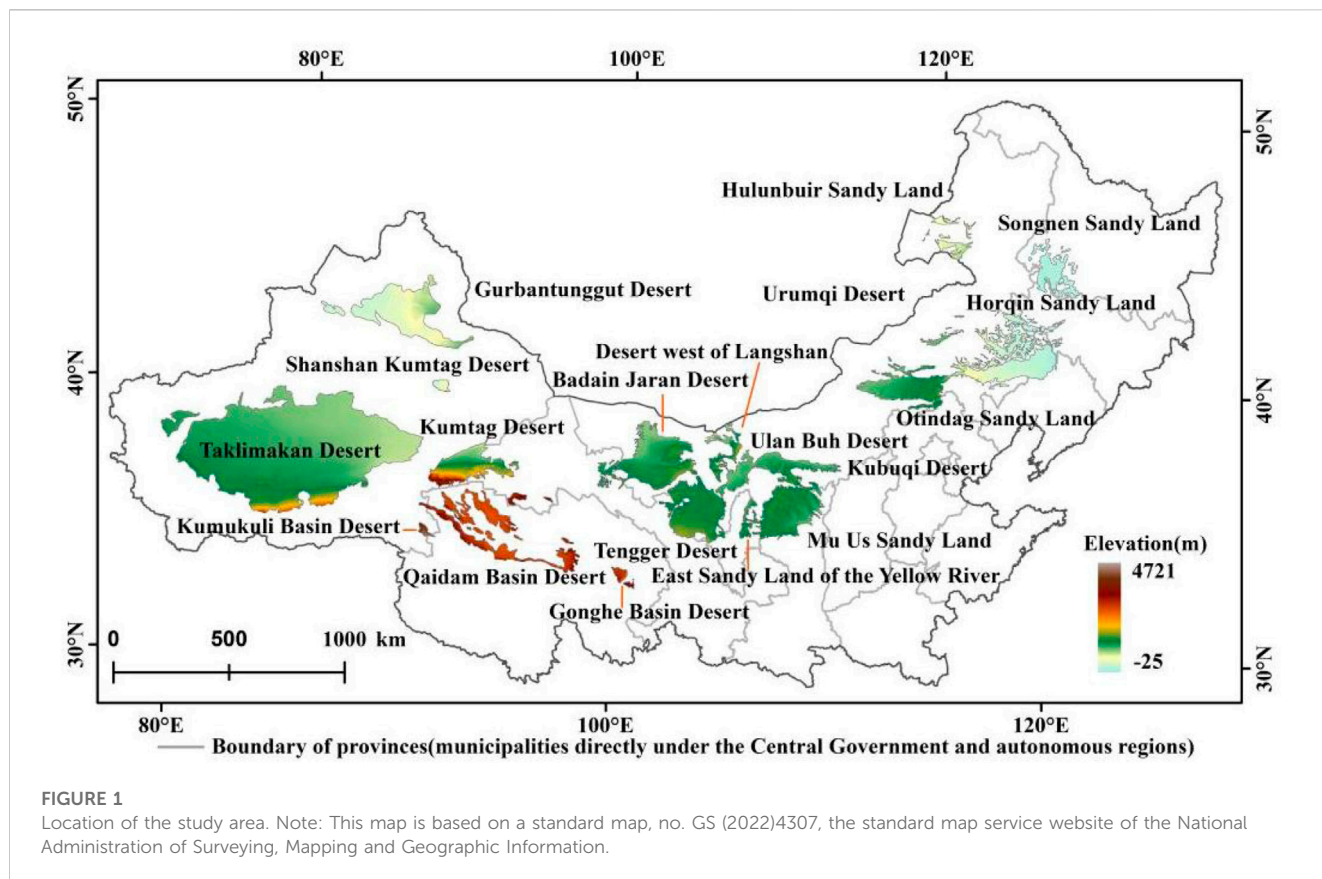


FIGURE 1

Location of the study area. Note: This map is based on a standard map, no. GS (2022)4307, the standard map service website of the National Administration of Surveying, Mapping and Geographic Information.

TABLE 1 Study area biomass (g·m⁻²).

Year	Forest	Grassland	Plow	Water	Sandy land	Total
1978	30.26	461.67	552.09	175.49	123.92	1572.13
1990	29.66	422.38	486.36	158.68	119.80	1428.88
2000	30.53	300.24	485.69	98.98	99.07	1194.16
2010	30.67	320.60	320.60	133.81	99.07	1111.22
2022	31.68	344.18	575.58	133.81	100.66	1407.74

$$h(water) = NPP, \tag{2}$$

$$h(sandland) = 0.3982 \cdot NPP + 84.0181, \tag{3}$$

$$h(forest) = 0.045 \cdot NPP + 17.53, \tag{4}$$

where h is the biomass of a given land use type (g·m⁻²) and NPP is the net primary productivity of the ecosystem (t·km⁻²). Finally, the biomass of each ecosystem in each year of the study area was obtained (Table 1).

The ESV equivalent factor is defined as the value of the annual natural grain yield of cultivated land with an average yield of 1 hm². After comparative analysis, the value of an ESV equivalent factor is equal to 1/7 of the market value of grain yield per unit area of the study area in the current year (Xie et al., 2015; Wu et al., 2023), and it is calculated as follows:

$$M = (m \times n)/7, \tag{5}$$

where M is the value of an ESV equivalent factor in the study area (CNY·km⁻²), m is the average unit price of grain in the study area (CNY·kg), and n is the average grain yield of 1 km² in the study area (kg·km⁻²). The value of an ESV equivalent factor was 3 512.11 CNY based on the market economic value of the average grain yield per unit area in the study area (CNY).

3.2.2 Calculation method of ESV

The benchmark unit price of ecosystem service functions was calculated as follows:

$$A_i = M \cdot a_i, \tag{6}$$

where A_i is the benchmark unit price of ecosystem service functions in the study area (CNY), M is the value of an ecosystem service function equivalent factor in the study area

TABLE 2 Type division of the land ecological coordination degree.

Indication range	$LEC \geq 1$	$0.5 \leq LEC < 1$	$0 \leq LEC < 0.5$	$-0.5 \leq LEC < 0$	$-1 \leq LEC < -0.5$	$LEC < -1$
Type	High coordination	Moderate coordination	Low coordination	Low-level conflict	Middle-level conflict	High-level conflict

(CNY·km²), a_i is the equivalent factor of the service function value for different ecosystems; $i = 1, 2, 3, \dots, 9$, respectively, represent ecosystem service functions, such as food production, raw material production, gas regulation, climate regulation, hydrological regulation, waste disposal, soil conservation, biodiversity maintenance, and esthetic landscape provision.

Based on the revised benchmark unit price of ecosystem service functions, the ESV per unit area of the study area was constructed as follows:

$$A_{ij} = (h_j/H) \cdot A_i, \quad (7)$$

where A_{ij} is the ESV per unit area of the study area (CNY), h_j is the biomass of a j -type ecosystem (g·m²), H is the average biomass per unit area of the study area (g·m²·km²), A_i is the benchmark unit price of ecosystem service functions in the study area (CNY); $j = 1, 2, 3, \dots, 6$, respectively, represent plow, forest, grassland, water, sandy land, and other ecosystem types. The total value of ecosystem services was calculated as follows:

$$D_{ij} = A_{ij} \cdot A_j, \quad (8)$$

where D_{ij} is the total service value of different ecosystems in the study area (CNY), A_{ij} is the ESV per unit area of the study area (CNY), and A_j is the area of different ecosystems in the study area (km²).

3.2.3 Sensitivity analysis

Sensitivity mainly reflects the degree of dependence of ESV change over time on the change of the ESV coefficient or the ESV change that was caused by the change in the ESV coefficient (Hatamkhani et al., 2023). If the sensitivity index was greater than 1, it indicates that ESV was elastic to the value coefficient. If the sensitivity index was less than 1, it indicates that ESV in the study area was not elastic to the value coefficient (Yang et al., 2022a), and the research results were credible. The sensitivity index was calculated as follows (Millennium, 2005):

$$CS = \frac{(ESV_j - ESV_i) / (ESV_i)}{(VC_{jk} - VC_{ik}) / (VC_{ik})}, \quad (9)$$

where CS represents the ESV sensitivity index of the study area, ESV_i represents the total value of the original ecosystem services (CNY), ESV_j represents the adjusted ESV (CNY), VC_{jk} represents the original value coefficient, VC_{ik} represents the adjusted value coefficient, and k represents the land use type.

3.2.4 Land ecological coordination degree

The degree of coordination refers to the harmonious balance between systems or system elements in the development process (Hatamkhani et al., 2023). In this paper, the land ecological coordination degree (LEC) was used to represent the coordination between land use change and ESV in the study area (Table 2). The coordination degree was calculated as follows:

$$LEC = \frac{V_r}{L_r}, \quad (10)$$

$$V_r = \frac{V_j - V_i}{V_i}, \quad (11)$$

$$L_r = \frac{\sum_{i=1}^n \Delta L_{ui-j}}{2 \sum_{i=1}^n L_{ui}}, \quad (12)$$

where V_r is the change rate of ESV during the study period, L_r is the change rate of the land use type area during the study period, V_i and V_j represent ESV at the beginning and end of the study period (CNY), respectively, L_{ui} represents the area of the i land use type at the beginning of the study period (km²), and ΔL_{ui-j} represents the total area converted from the i land use type to other land use types during the study period (km²).

4 Analysis results

4.1 Temporal and spatial characteristics of land use change in desert areas

4.1.1 Status quo of land use in desert areas

In 2022, sandy land and grassland were the main types of land use in the study area, with an area of 552,771.66 km², accounting for 64.12% of the total study area. The area of grassland was 156,982.86 km², accounting for 19.00% of the total study area. The sizes of other land use types were as follows: the area of other lands was 41,219.69 km² (4.99% of the total study area), plow was 40,425.77 km² (4.89%), forest was 15,017.72 km² (1.82%), construction land was 4,339.79 km² (0.52%), and water was 3,568.51 km² (0.43%). In the past 44 years, the spatial distribution pattern of land use in the study area showed that grassland and sandy land were concentrated in the core area of the study area, forest land and urban and rural construction lands were concentrated in the edge area of the study area, and rivers and wetlands were scattered.

4.1.2 Temporal changes of land use in desert areas

Since the reform and opening up, there has been a clear land use change in this region, with the grassland area changing the most (46,016.33 km²), followed by other lands (30,402.17 km²), water area (826.15 km²), and construction land (1,487.06 km²). The total area of plow, other lands, and water decreased, while the total area of grassland, forest, construction land, and sandy land increased.

From the perspective of land use type transformation, from 1978 to 2000, land use transformations mainly occurred when grassland and sandy land were transformed into plow land (Figure 2). After the implementation of the Western development policy, with accelerated urban agglomeration and the increase in population, the demand for food became more and more urgent, and there were more projects to destroy grass and open fields,

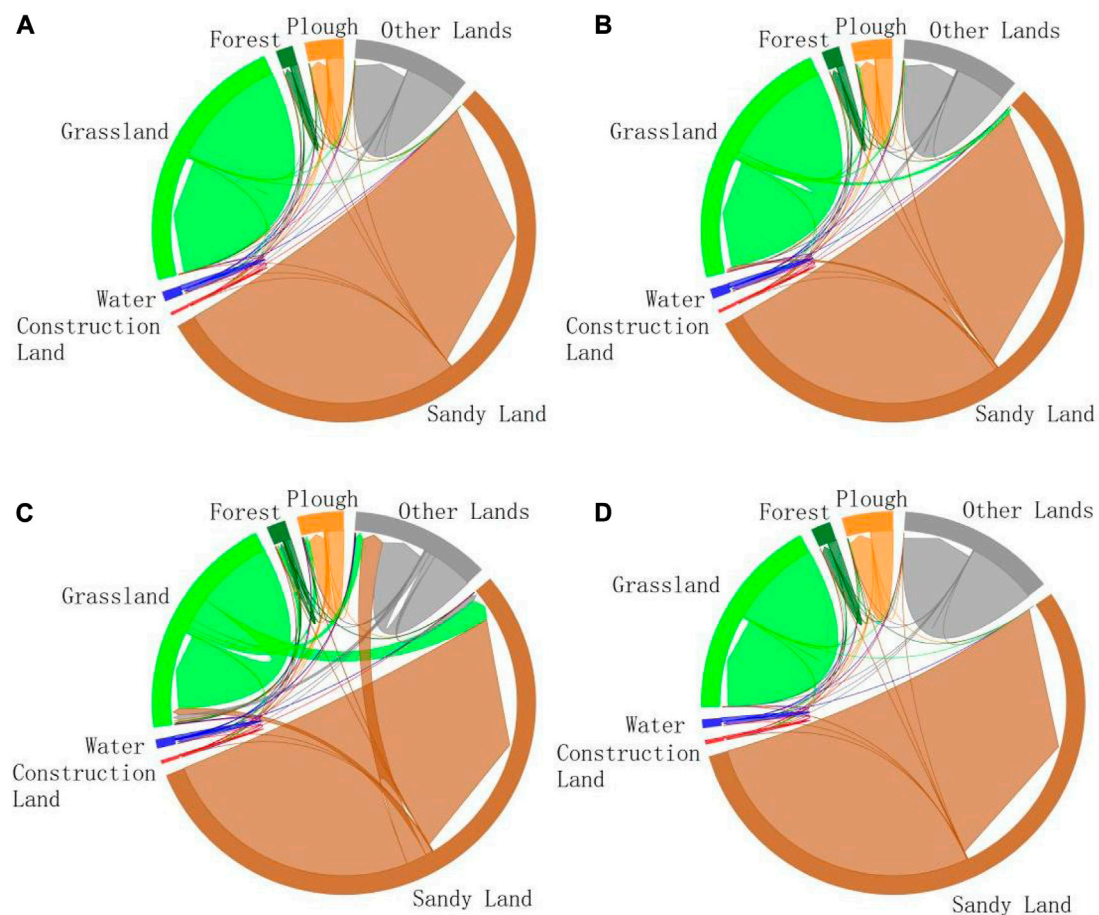


FIGURE 2
Dynamic change of land use in desert areas. (A) 1978–1990; (B) 1990–2000; (C) 2000–2010; and (D) 2010–2022.

leading to the expansion of the plow area. From 2000 to 2022, land use conversions in the study area mainly occurred as the conversion of plow, sandy land, and other lands to grassland and forest. Among them, the transfer of grassland was more significant because the state proposed the construction of ecological civilization. This promoted the “Three-North shelter belt” project in the north and implemented a policy of returning farmland to forest, making the transfer of plow to grassland obvious. In addition, social and economic development is needed to develop other lands and occupy more plow and grassland to meet the needs of construction land, such as the construction of public facility land, industrial and mining land, and urban and rural residential areas. This was another reason for the increase in construction land after reform and opening up.

4.1.3 Spatial differentiation of land use in desert areas

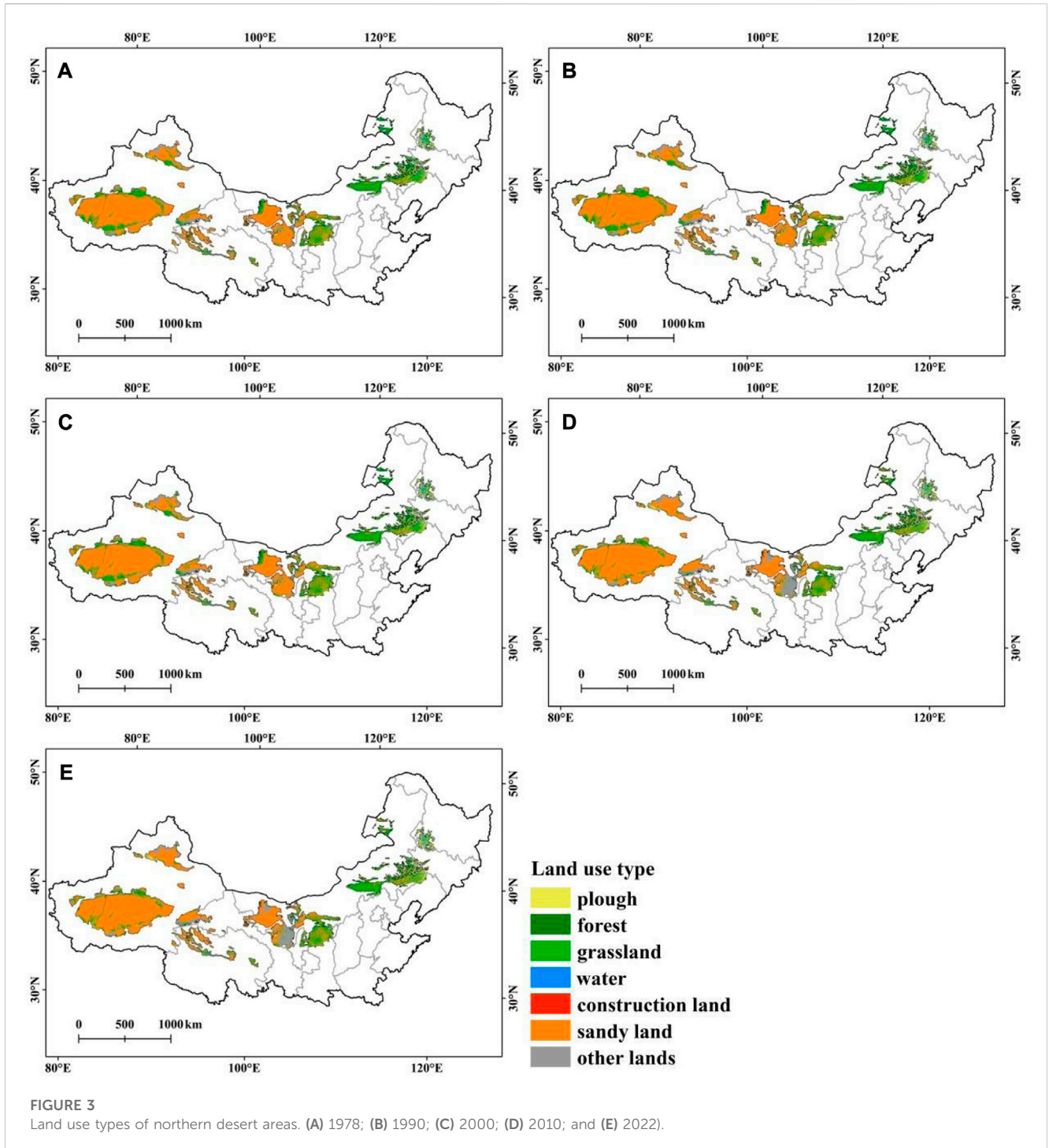
Sandy land and other lands were mainly concentrated in the central and western parts of the study area, distributed in the Taklimakan Desert, Gurbantunggut Desert, Badain Jaran Desert, Kumukuri Desert, and Kumtag Desert (Figure 3). Grassland and forest, as the main types of green land cover in the study area, were mainly distributed in the Songnen Sandy Land, Horqin Sandy Land, Hunshandak Sandy Land, and Mu Us Sandy Land in

the eastern part of the study area. From the perspective of land use spatial differentiation, grassland changes from 1978 to 2022 were mainly distributed in the Badain Jaran Desert, Gonghe Basin Desert, Hedong Sandy Land, Hulun Buir Sandy Land, Horqin Sandy Land, and Songnen Sandy Land. Changes to plow were mainly distributed in Hunshandake Sandy Land and Songnen Sandy Land in the eastern part of the study area. The variation in the water area was distributed in the Gonghe Basin Desert in the central and southern parts of the study area. Changes to sandy land were mainly distributed in the eastern Badain Jaran Desert, the southern Qaidam Desert, the eastern Gonghe Basin Desert, and the eastern Gurbantunggut Desert. Changes to construction land were mainly distributed in Ulan Buh and the northern part of the desert.

4.2 ESV variation characteristics in desert areas

4.2.1 Changes in the total ecosystem service value in desert areas

Since the reform and opening up, ESV in desert areas has increased overall (Table 3) by a total of 19.64×10^8 CNY,



specifically due to the large-scale reclamation of plow. In the order of ESV magnitude, the highest was grassland, followed by plow, sandy land, water, and forest. After reform and opening up, grassland ESV was always in the first place, which demonstrated that grassland played a leading role in maintaining ESV in the study area. ESV of all land types increased. Among them, ESV of plow and grassland increased faster, and ESV of grassland increased by 11.73×10^8 CNY, but ESV of grassland decreased from 72.14% in 1978 to 60.15% in 2022. ESV of plow increased by

3.52×10^8 CNY, and the percentage of ESV for plow increased from 9.03% in 1978 to 17.65% in 2022. ESV of forest, water, and sandy land increased slowly, and ESV of forest increased by 0.19×10^8 CNY. The ESV percentage increased from 0.61% in 1978 to 0.97% in 2022. ESV of water increased by 0.46×10^8 CNY, but the percentage of ESV in water decreased from 3.30% in 1978 to 2.37% in 2022. ESV of sandy land increased by 1.51×10^8 CNY, and ESV percentage of sandy land increased from 6.11% in 1978 to 7.62% in 2022.

TABLE 3 Total value of ecosystem services in northern China, 1978–2022.

Year	Project	Plow	Forest	Grassland	Water	Sandy land	Total
1978	Magnitude of value	0.06	0.01	0.45	0.02	0.03	0.63
	Percent/%	9.03	0.61	72.14	3.30	6.11	100.00
1990	Magnitude of value	0.19	0.02	1.48	0.08	0.13	2.13
	Percent/%	9.33	0.71	69.86	3.85	6.27	100.00
2000	Magnitude of value	0.49	0.03	2.42	0.08	0.30	3.71
	Percent/%	13.26	0.96	65.22	2.35	8.14	100.00
2010	Magnitude of value	1.08	0.08	5.39	0.19	0.97	8.74
	Percent/%	12.35	0.96	61.64	2.22	11.14	100.00
2022	Magnitude of value	3.57	0.19	12.19	0.47	1.54	20.26
	Percent/%	17.65	0.97	60.15	2.37	7.62	100.00

TABLE 4 Single values of ecosystem services in northern China desert areas from 1978 to 2022.

Type of ecosystem services	Year 1978	Year 1990	Year 2000	Year 2010	Year 2022
Food production	0.02	0.07	0.14	0.31	0.82
Raw material production	0.01	0.06	0.10	0.24	0.57
Gas regulation	0.06	0.22	0.37	0.84	1.94
Climate regulation	0.08	0.27	0.46	1.11	2.51
Hydrological regulation	0.08	0.27	0.45	1.07	2.43
Waste treatment	0.08	0.28	0.48	1.17	2.65
Soil conservation	0.10	0.34	0.58	1.32	3.07
Maintenance of biodiversity	0.09	0.32	0.55	1.31	2.87
Esthetic landscape provision	0.04	0.16	0.27	0.66	1.39
Total	0.61	2.04	3.45	8.07	18.29

4.2.2 Changes in the value of individual ecosystem services in desert areas

After reform and opening up, ESV in the study area has been mainly focused on soil conservation and biodiversity maintenance (Table 4). Single ESV has increased, among which the fastest increase occurred for soil function, from 0.11×10^8 CNY in 1978 to 3.07×10^8 CNY in 2022. The slowest increase was for the production of raw materials, from 0.02×10^8 CNY in 1978 to 0.85×10^8 CNY in 2022.

4.3 ESV sensitivity analysis

The ESV sensitivity index of different periods ranged from 0 to 1, indicating that the ESV coefficient was inelastic and the research results are credible (Table 5). The sensitivity index values for land use types in descending order were grassland, forest, plow, water, other lands, sandy land, and construction land. During the study period, the sensitivity index of plow, forest, and other lands all

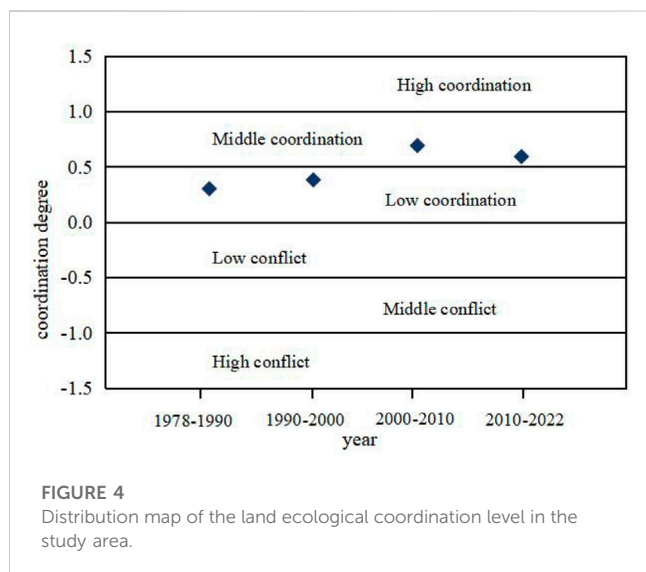
increased, while the sensitivity index of other land use types declined.

4.4 Coordination degree between land use and ESV

The overall land ecological coordination degree in the study area was at a moderate level (Figure 4), indicating that land use change did not lead to environmental deterioration, but there was a crisis (Table 2). From 1978 to 2000, the land ecological coordination degree was low but increasing, indicating that land use change did not have a negative impact on the environment and was developing in a good direction. From 2000 to 2010, the ecological coordination degree of land use in the study area was at a moderate level, indicating that land use change did not cause deterioration of the environment, but there was a potential crisis. From 2010 to 2022, the land ecological coordination degree was 0.51, which was still at a moderate level,

TABLE 5 Sensitivity index of the ESV for each land use type in the study area.

Year	Plow	Forest	Grassland	Water	Construction land	Sand	Other lands
1978	0.1126	0.1832	0.6301	0.0185	0.0009	0.0059	0.0062
1990	0.1251	0.1958	0.6258	0.0180	0.0009	0.0061	0.0063
2000	0.1387	0.1859	0.6247	0.0172	0.0008	0.0064	0.0064
2010	0.1459	0.1847	0.6123	0.0189	0.0007	0.0063	0.0064
2022	0.1599	0.1866	0.6051	0.0177	0.0006	0.0065	0.0065



but it was decreasing compared to 2000–2010. This showed that the development of urban agglomerations did not pay enough attention to the irrational use of land, so the level of conflict between the environment and land use was more obvious in the later period of the study.

5 Discussion

Land use change is the main factor influencing ESV change, and land use change can lead to a series of ecological processes, thus affecting the supply of ecosystem services (Millennium, 2005; Hasan et al., 2020). For example, the regulation and support services of forest are high, while the food supply services are relatively low (Lei et al., 2021; Gascoigne et al., 2011; Zhao et al., 2022). Agricultural product supply services of plow are relatively high, while regulation, support, and cultural services are relatively low (Yang et al., 2022b). The land use type in the study area is mainly grassland, and the construction land area is small. During 1978–2022, the forest land, grassland, water area, and construction land in the study area increased, while the unused land area decreased, which is consistent with the research results of Shao et al. (2022). The contribution of the grassland ecosystem service value was the highest, which was consistent with the research results of Zhang et al. (2023). The degree of land use is negatively correlated with ecosystem services. Contrary to the research results of Zhao et al. (2022), the main reason is that the changing trend of the land use degree is

different due to the difference in land use change. In China, sandy land and grassland were the main land use types, and the areas of grassland, forest, construction land, and sandy land were all increasing. The areas of plow, water, and other lands decreased. In particular, from 1978 to 2010, the area of construction land increased significantly, especially because economic development and the construction of small towns have led to urban expansion since the reform and opening up (Wang et al., 2020). This also showed that human activities are an important driving force for short-term regional landscape changes (Li et al., 2019). In recent years, with policy support for poverty alleviation and development and rural revitalization, China's land use in desert areas has shown new characteristics of change (Liu et al., 2014) and has changed from "sand advancing and people retreating" (Moharram and Sundaram, 2023) into a new situation of "harmony between man and sand" (Parker-Shames et al., 2023).

After reform and opening up, there was an overall ESV increase of 19.64×108 CNY, but the increase in raw material production was the slowest. Because the contribution rates of grassland and plow to ESV were as high as 80%, the two had a significant impact on ESV change, which was consistent with the research results of Sun Menghua (Yang et al., 2022a). In terms of ESV, the values of food production and raw material production were relatively low, specifically because the productivity of desertification grassland resources was insufficient and grassland resources were overexploited and utilized (Liu et al., 2017). Therefore, it is necessary to strengthen the protection, restoration, and management of desertification grassland. In the future, consideration should be given not only to land classes with a high ESV but also to land use types that play an important role in raw materials and food production to maintain ecosystem service functions and stability.

ESV of the study area should be scientifically assessed, according to the principles of ecological economics (Shao et al., 2022), and intuitively reflected in monetary form, which is an important basis for formulating regional ecological compensation policies (Yang and Huang, 2021). However, the longitude span of the deserts in China is large, and the geographical location, climatic characteristics, and human environment of the deserts vary greatly. On the whole, the land ecological coordination degree in the study area is at a moderate coordination level, and the ecological environment management has a long way to go. Ecological and environmental management projects should be implemented in the study area to strengthen environmental protection, reduce the damage of human activities to the environment, and prevent further deterioration of the environment. For ecologically fragile areas, the local governance model should be maintained. Therefore, the next step will be to revise the ESV coefficient, according to the characteristics of each desert region to obtain more accurate calculation results.

6 Conclusion

From the perspective of land use structure, the main land use types in the study area were sandy land and grassland, accounting for more than 80% of the total area. From the perspective of land use change from 1978 to 2020, the areas of plow, water, and other lands all decreased. Meanwhile, grassland, forest, construction land, and sandy land increased, with the fastest change occurring in grassland. Over the whole study period, the period of change rate was the largest from 2000 to 2010 and the smallest from 1978 to 1990. From the perspective of land use transfer, the plow in the study area was mainly converted into grassland, forest, and construction land. Forest was mainly converted into grassland, grassland was mainly converted into sandy land, construction land was mainly converted into plow, and the area of other lands was mainly converted into grassland. Therefore, the main reason behind land use change in the study area was the implementation of returning plow to forest.

From 1978 to 2022, ESV in the study area increased from 0.63×10^8 CNY to 20.26×10^8 CNY, an increase of 19.63×10^8 CNY. In terms of time period, the growth rate was the fastest from 2010 to 2022 and the slowest from 1978 to 1990. From the perspective of land use type in 2022, ESV of grassland was the highest, reaching 12.19×10^8 CNY, while ESV of water was the lowest, at only 0.47×10^8 CNY. From the perspective of the ecosystem service type in 2022, soil conservation was the highest, reaching 3.07×10^8 CNY, and raw material production was the lowest, at only 0.82×10^8 CNY. From the perspective of the ecological sensitivity index, the ESV sensitivity index results for the study area were all less than 1. Grassland had the highest sensitivity index, while construction land had the lowest. The ESV coefficient was inelastic, so the research results were credible. From the perspective of the land ecological coordination degree, the land ecological coordination degree in the study area was between low and medium coordination degrees. Therefore, land management and planning should be strengthened to promote coordinated development between land use and economic development.

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Data availability statement

The original contributions presented in the study are included in the article/Supplementary Material; further inquiries can be directed to the corresponding author.

Author contributions

ZB, HD, and EH contributed to the conception and design of the study, and organized the database. XL performed the statistical analysis. All authors contributed to the article and approved the submitted version.

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Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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