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Dam failure environmental standards in China based on ecosystem service value

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Dam failure risk standards are the foundation of risk decision-making for dam managers. However, as an important component of dam failure risk standards, there are currently no unified environmental risk standards. Drawing on research ideas of ecological economics on ecosystem service values and equivalent factor methods, this study quantified environmental values and effectively connected environmental standards with existing standards using the ALARP principle and the F-N curve. Considering the differences in environmental and economic conditions in different regions, a risk preference matrix was constructed to determine the risk preference of each region and formulate the dam failure environmental risk standards for China. This study presents a preliminary exploration of the formulation of dam failure environmental risk standards, providing new methods and ideas for subsequent research.

KEYWORDS

dam failure, environmental risk standards, equivalent factor method, risk preference, ecosystem service (ES) values

1 Introduction

Recently, owing to frequent extreme weather events and the increasing occurrence of flood events exceeding standards, and problems such as design defects, poor construction quality, and inadequate operation and management of many reservoir dams in China, dam collapses occur occasionally (Chen Junfei et al., 2020; Xu et al., 2022). With continuous social and economic development, people are paying attention to the quality of their living environments (Tu et al., 2023). Environmental standards for dam collapse form the basis for environmental risk assessments and decision-making. Developing reasonable ecological environment risk standards is conducive to helping decision makers fully understand the risk level of dams to ensure scientific risk decisions, and has important significance for improving the evaluation system for the consequences of dam collapse risks and the theoretical management of dam risks (He et al., 2020).

Currently, there is no unified environmental standard for dam collapses (Li et al., 2019; Ge Wei et al., 2020). This is primarily because the environmental scope of a dam collapse is wide, and there are many influencing factors. Furthermore, there is no unified dimensionality, making it difficult to quantify environmental values. Compared with the loss of life and economic losses caused by dam collapse, little attention has been paid to the environmental impact of dam collapse in previous studies. Therefore, most current research focuses on the construction of dam collapse indicator systems and mathematical method analysis. For example, Wang et al. (2006) believed that common environmental impacts refer to changes in environmental conditions caused by developmental activities or the formation of new environmental conditions. The environmental impacts of dam collapses mainly

include those on river morphology, loss of habitats for organisms and their growth, destruction of human landscapes, and industrial impacts that are susceptible to or cause considerable environmental effects or pollution, and define an impact index to evaluate risk size. He et al. (2008) believed that environmental impact refers to changes in environmental conditions in the area near the reservoir caused by dam collapses, which are specifically reflected in the impact on water, soil, ecology, and human habitats; they applied the Analytic Hierarchy Process and fuzzy mathematics to evaluate risk size. Francisco and Gallardo Izquierdo (2008) believed that environmental risk is primarily related to the erosion of the underlying surface of the dam, seepage, and pollution load, and defined the environmental safety index of the dam to calculate environmental risk. Cheng and Zhou (2013) believed that environmental impact refers to the effects of human activities, including economic and social activities, on the environment, and the resulting changes in the environment. Seven factors—river morphology, water environment, soil environment, vegetation cover, biodiversity, human environment, and pollution industry—were selected for analyzing the environmental impact of dam collapses, and the fuzzy mathematics theory was applied for environmental risk assessment. Xu et al. (2013) analyzed the environmental impact based on factors such as land-carrying capacity, water quality, reservoir siltation, downstream riverbed erosion, and earthquake geological hazards. These studies determined the factors of environmental impact and provide an important research basis for environmental assessment.

The above mentioned methods can be effectively used to rank the severity of environmental impacts caused by dam failures. However, they cannot provide specific representations of the consequences (typically categorized into 3–5 levels of severity), and are therefore semi-quantitative evaluations that may not be universally applicable. To achieve a quantitative representation of the environmental value of dam failures, the energy footprint method was introduced for the risk assessment of the environmental impacts of dam failure. This method is commonly used in ecological impact assessments (Binod et al., 2019; Chen Shurui et al., 2020; Yang et al., 2020). However, the scientific validity and accuracy of the energy conversion rate and energy conversion coefficients still require improvement. Meanwhile, the quantification of Ecosystem Service Value (ESV) based on the equivalent factor method is intuitive, widely used, and requires little data, making it suitable for regional- or global-scale environmental value assessments (Chen et al., 2023; Yang et al., 2023; Zhang et al., 2023).

The essence of the environmental impact is the change in the natural environment and ecological conditions of areas near the reservoir, caused by dam failure. Natural and ecological environments are composed of two parts, biotic and abiotic. Natural ecosystems provide various raw materials or products directly for survival (food, water, oxygen, wood, fibers, etc.) and regulate the climate, purify pollution, retain water sources, maintain soil and water, prevent wind and sand, reduce disasters, and protect biodiversity on a large scale (Enes and Ibrahim, 2022). Through its functions, the natural ecosystem continuously provides environmental conditions and material foundations for humans, creating service value (similar to the GDP created

by human labor) (Langill Jennifer et al., 2022). Therefore, from the human perspective, the concept of environmental impact caused by dam failures can be defined as the destruction of the natural ecosystem service supply, resulting in a reduction in the output of ecosystem service value.

Based on the equivalent factor method for quantifying environmental values, this study explored the preliminary construction of China's regional dam break environmental risk standards using the ALARP principle, F-N curve method, and risk matrix, providing a new approach and method for subsequent research.

2 Methods

2.1 As low as reasonably practicable (ALARP) principle

No system carries risks that cannot be eliminated completely through preventive measures. Although the lower the risk of a system the higher the level of security, it is increasingly difficult and costly to further reduce the risk level. Given China's economic and social development level and the public's willingness to accept risks, a trade-off can be made between the risk level and the cost of a system. Therefore, the ALARP principle was selected to establish risk standards in China.

This principle stands divides risks into three zones: intolerable, ALARP, and broadly acceptable (Langdalen et al., 2020). If the risk falls into an intolerable zone, measures must be taken to reduce the risk. If it falls into the ALARP zone, taking risk control measures depends on whether risk reduction is feasible and whether the benefits outweigh the costs. If a risk falls into the broadly acceptable zone, no risk-control measures are required. This principle is illustrated in Figure 1.

2.2 F-N curve method

In 1967, Farmer proposed a limit curve for the acceptable risk of various accident types using probability theory, known as the famous F-N curve (Pei et al., 2018). The F-N curve is a limiting curve based on probability theory that allows various accidents to occur. It is based on probabilistic analysis and expected benefits, and takes into account the current state of risk in which it is located as well as people's aversion to risk, and it is highly scientific and versatile. It was first used for risk assessment in nuclear power plants and was later widely applied in the construction of social life risk standards for dam engineering. The expression for this curve is as follows:

$$1 - F_N(x) < \frac{C}{x^n} \quad (1)$$

where $F_N(x)$ is the distribution function of the annual mortality rate, representing the probability of x deaths; $1 - F_N(x)$ is the probability of $\leq N$ deaths; C is a constant that determines the position of the risk control line; and n is the slope that represents the degree of aversion to risk, with a larger slope indicating greater ecosystem to risk and less acceptability.

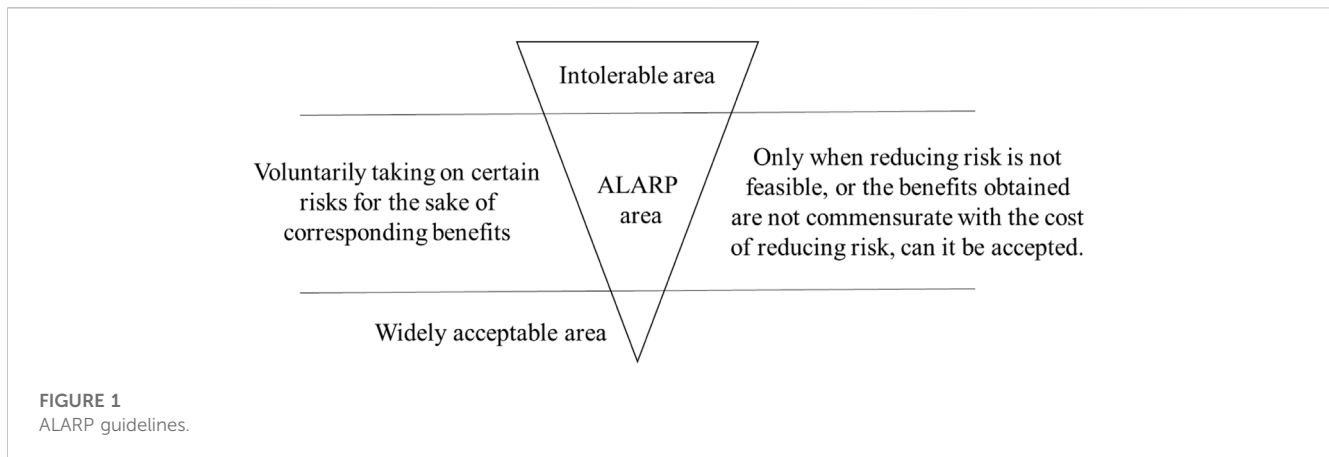


FIGURE 1
ALARP guidelines.

According to the Chinese State Council's "Regulations on Reporting and Investigating Production Safety Accidents," one death is equivalent to a direct economic loss of 3.3–5 million RMB (Li et al., 2015). Therefore, it is suggested to construct China's economic risk standards for reservoir dams based on a ratio of 4 million RMB per person. Although measuring human life values in economic terms is often considered inhumane and has gained strong criticism and opposition, it is reasonable to use life risk standards as a basis for constructing economic risk standards under the premise of controlling life risks at socially acceptable levels.

Constructing the environmental risk criteria for dam failure on the basis of the life and economic risk criteria avoids over-complicating the criteria on the one hand, and on the other hand allows for a strong correlation between the different criteria. Therefore, it is proposed to analyze the economic perspective by simplifying the environmental loss to the money needed to compensate the environmental risk, then the relationship between the loss of environmental value and its exceeding probability is as follows:

$$1 - FN'(x) < \frac{C}{x^n} \quad (2)$$

where $FN'(x)$ is the probability distribution function in which the funds needed to compensate for ecological and environmental risks $<x$, C is a constant, and n is the slope of the standard line

3 Estimation of environmental value

The Equivalent Factor Method is quantitative and based on expert knowledge. It is used to estimate the Ecosystem Service Value (ESV) (Zhang et al., 2022). One standard ecological value equivalent factor is defined as the economic value of natural food production per hectare of farmland with a national average yield (Cao et al., 2021). This reflects the potential contribution of ecosystems to ecological services. Food production reflects the value of the agricultural ecosystem, and the economic value of food production is primarily calculated based on three major grain crops: rice, wheat, and maize. The formula is as follows:

$$D = Sr \times Fr + Sw \times Fw + Sm \times Fm \quad (3)$$

Where D represents the environmental value of one standard equivalent factor (in yuan per hectare); Sr , Sw , and Sm represent the percentage of the planting area of rice, wheat, and maize, respectively, in the total planting area of the three crops; and Fr , Fw , and Fm represent the average net profit per hectare of rice, wheat, and maize, respectively, in China. The basic equivalent factor was determined to be 1482.67 yuan based on the data from China Statistical Yearbook 2017, 2018, and 2019 (National Bureau of statistics of the people's Republic of China, 2019), and the Compilation of National Agricultural Cost and Benefit Data (National Development Planning Commission, 2019).

Based on previous research results and expert experience, we constructed basic equivalent factors for different types of ecosystems and different types of environmental values, as shown in Table 1. Land-use remote sensing data were used to correspond farmland ecosystems to agricultural land types and wetland ecosystems to swamp land types. The average equivalent factors for forest, grassland, water, and desert ecosystems in the second level of classification were used as the calculation factors for the forest, grassland, water, and unused land (excluding swamp land) types in the remote sensing data. Based on Table 1, the ESV for each region in China in 2020 were calculated as shown in Figure 2.

4 Construction of environmental risk standards for dam failure based on ESV

4.1 Principles for constructing risk standards

Owing to substantial differences in politics, economics, culture, and other aspects among different countries (Li et al., 2018; Ge et al., 2020b), foreign risk standards cannot be directly applied to China (Ge et al., 2017; Ge et al., 2020c). Based on relevant domestic and foreign achievements and China's national conditions, the following principles were used to construct risk standards for dams:

- (1) Compliance with the safety status of Chinese dams. Several of China's active reservoir dams were built from the 1950s to the 1970s. They were limited by economic and technological levels at that time, and several have hidden safety hazards. If standards

TABLE 1 Equivalent value of ecosystem services per unit area.

Ecosystem classification		Supply services		Conditioning services				Support services			Cultural services		Total
First-level	Secondary classification	Food production	Raw material production	Water supply	Air regulation	Climate regulation	Environmental purification	Hydrological regulation	Soil conservation	Maintenance of nutrient cycling	Biodiversity	Aesthetic landscape	
Farmland	Dry land	0.85	0.40	0.02	0.67	0.36	0.10	0.27	1.03	0.12	0.13	0.06	4.01
	Paddy field	1.36	0.09	-2.63	1.11	0.57	0.17	2.72	0.01	0.19	0.21	0.09	3.89
Forest	Coniferous	0.22	0.52	0.27	1.70	5.07	1.49	3.34	2.06	0.16	1.88	0.82	17.53
	Coniferous broadleaf hybrid	0.31	0.71	0.37	2.35	7.03	1.99	3.51	2.86	0.22	2.60	1.14	23.09
	Broadleaves	0.29	0.66	0.34	2.17	6.50	1.93	4.74	2.65	0.20	2.41	1.06	22.95
	Shrub	0.19	0.43	0.22	1.41	4.23	1.28	3.35	1.72	0.13	1.57	0.69	15.22
Grassland	Grassland	0.10	0.14	0.08	0.51	1.34	0.44	0.98	0.62	0.05	0.56	0.25	5.07
	Thicket	0.38	0.56	0.31	1.97	5.21	1.72	3.82	2.40	0.18	2.18	0.96	19.69
	Meadow	0.22	0.33	0.18	1.14	3.02	1.00	2.21	1.39	0.11	1.27	0.56	11.43
Wetland	Wetland	0.51	0.50	2.59	1.90	3.60	3.60	24.23	2.31	0.18	7.87	4.73	52.02
Desert	Desert	0.01	0.03	0.02	0.11	0.10	0.31	0.21	0.13	0.01	0.12	0.05	1.1
	Bare ground	0.00	0.00	0.00	0.02	0.00	0.10	0.03	0.02	0.00	0.02	0.01	0.2
Waters	Water system	0.80	0.23	8.29	0.77	2.29	5.55	102.24	0.93	0.07	2.55	1.89	125.61
	Glacial snow	0.00	0.00	2.16	0.18	0.54	0.16	7.13	0.00	0.00	0.01	0.09	10.27

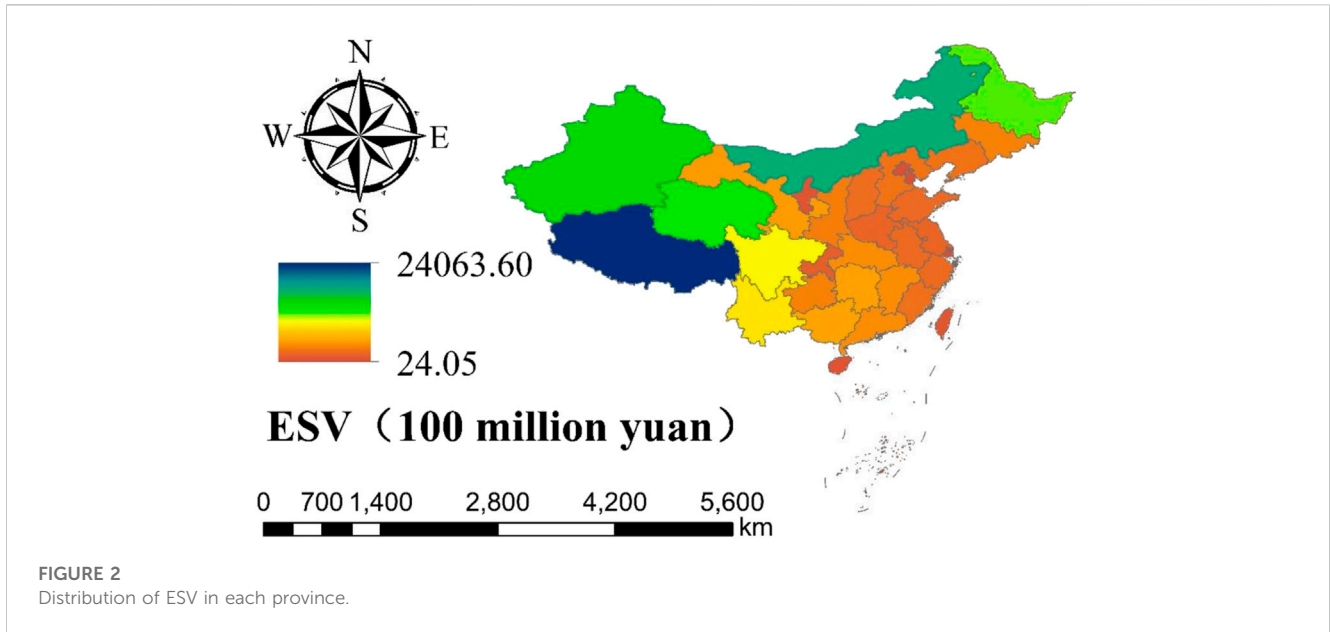


TABLE 2 Values of C adopted by some countries or institutions.

Country/Region/Agency	C value	Country/Region/Agency	C value
Hong Kong, China (GEO)	10^{-3}	Brazil (Rio de Janeiro)	10^{-2}
United Kingdom (HSE)	10^{-2}	Brazil (Sao Paulo)	10^{-3}
Australia (AGS)	Newly built 10^{-4}	Netherlands (VROM)	10^{-3}
	Existing 10^{-3}	United States (Santa Barbara)	10^{-3}
Australia (Victoria)	10^{-2}	Canada (CDA)	10^{-3}
Australia (New South Wales)	3×10^{-3}	Denmark	10^{-2}

are too strict, more reservoirs will be classified as “dangerous or sick.” In the current situation, where funding is relatively tight, several reservoirs will not be able to receive timely reinforcement against risks and will be forced to change operating conditions or stop operating, which will degrade the local economic development environment. It is also not conducive to effectively distinguish the degree of danger between “dangerous and sick” reservoirs.

- (2) Compliance with the latest public acceptance of risks. With the increasing need for a better life for people, risk awareness is continuously increasing; in the event of equally serious accidents, the concern of the public will grow. Therefore, when formulating risk standards, both the engineering and the public’s acceptance of risks should be considered.

4.2 Determination of the C value

C is a constant that determines the starting position of the standard line. The acceptable risk C values adopted by some countries and institutions are listed in Table 2. As seen from table, the value of C is primarily 10^{-3} /year. The values of C in

Table 2 include various application categories such as nuclear power plants, hazardous chemicals, factories, all types of buildings (hospitals, schools, and residential buildings), slopes, and dams. Considering that accidents involving hazardous chemicals, factories, and buildings during new construction, reconstruction, expansion, and in-service production and storage processes are primarily caused by human factors, natural factors beyond human control are also important components of uncertainty during the construction and operation of reservoir dams. Therefore, according to the suggestion of Li et al. (2015), the standard for dam safety can be appropriately lowered to $C = 10^{-2}$. The acceptable risk standard can be one order of magnitude smaller than this; thus $C = 10^{-3}$.

4.3 Determination of the n value

The value of n represents the degree of risk preference. Considering the vast territory of China, there may be considerable differences in economic development, ecological environment quality, and social and cultural backgrounds among the different regions. Therefore, the degree of risk preference varies across the regions; thus, a risk matrix was introduced. This method

TABLE 3 Average ESV *per capita* GDP and unit area of each province in 2020.

Province (district)	Population (residential)/10,000	GDP <i>per capita</i> /10,000 yuan	ESV/100 million	Area (land only)/km ²	Average ESV/(yuan/m ²) per unit area
Anhui	6102.72	6.34	2282.98	140,100	16.30
Macau	68.32	24.63	0.38	33	11.60
Beijing	2189.31	16.49	295.51	16,410	18.01
Chongqing	3205.42	7.80	1451.82	82,402	17.62
Fujian	3973.00	10.57	2742.24	124,000	22.11
Gansu	2501.98	3.60	4797.57	425,800	11.27
Guangdong	12601.25	8.79	4059.66	179,725	22.59
Guangxi	5012.68	4.42	5308.27	237,600	22.34
Guizhou	3622.95	4.62	3678.02	176,167	20.88
Jiangsu	8474.80	12.12	1969.40	107,200	18.37
Hainan	1008.12	5.49	704.38	35,400	19.90
Hebei	7461.02	4.85	2794.48	188,800	14.80
Henan	9936.55	5.53	1974.39	167,000	11.82
Heilongjiang	3185.01	4.30	11614.27	473,000	24.55
Hubei	5775.00	7.52	4057.06	185,900	21.82
Hunan	6644.48	6.29	4997.21	211,800	23.59
Jilin	2690.73	5.11	3670.75	187,400	19.59
Jiangxi	4518.86	5.69	4109.97	166,900	24.63
Liaoning	4259.14	5.90	2874.17	148,600	19.34
Inner Mongolia	2404.92	7.22	17678.74	1,183,000	14.94
Ningxia	720.27	5.44	711.47	66,400	10.71
Qinghai	592.40	5.07	13065.77	722,300	18.09
Shandong	10152.75	7.20	2039.32	157,900	12.92
Shanxi	3491.56	5.06	2554.02	156,700	16.30
Shaanxi	3952.90	6.22	3360.46	205,600	16.34
Shanghai	2487.09	15.56	137.54	6,341	21.69
Sichuan	8367.49	5.81	9464.28	486,000	19.47
Taiwan	2356.12	18.20	946.47	36,014	26.28
Tianjin	1386.60	10.16	246.62	11,966	20.61
Tibet	364.81	5.22	24063.58	1,228,400	19.59
Hong Kong	747.42	32.30	24.05	1,107	21.74
Xinjiang	2585.23	5.34	14638.28	1,660,000	8.82
Yunnan	4858.30	5.19	8584.68	394,100	21.78
Zhejiang	6456.76	10.01	2362.88	105,500	22.40

was proposed by the US Air Force Electronic Systems Center in the mid-to-late 1990s, and has been widely used in US military weapon system development projects as a structural method for identifying the importance of project risks. It can evaluate the potential risks of projects and is currently widely used in research fields, such as

project risk assessment and safety risk rating (Koulinas Georgios et al., 2021).

The higher the *per capita* GDP in a region, the higher the demand for quality of life; and the more emphasis it places on the surrounding ecological environment, the more averse it is to risk

TABLE 4 Risk appetite matrix.

Per capita GDP/10,000 yuan	Average ESV per unit area/(yuan/m ²)		
	(8, 14)	(14, 21)	(21, 28)
(2.58, 5.29)	X	X	Y
(5.29, 8.00)	X	Y	Y
>8.00	Y	Y	Z

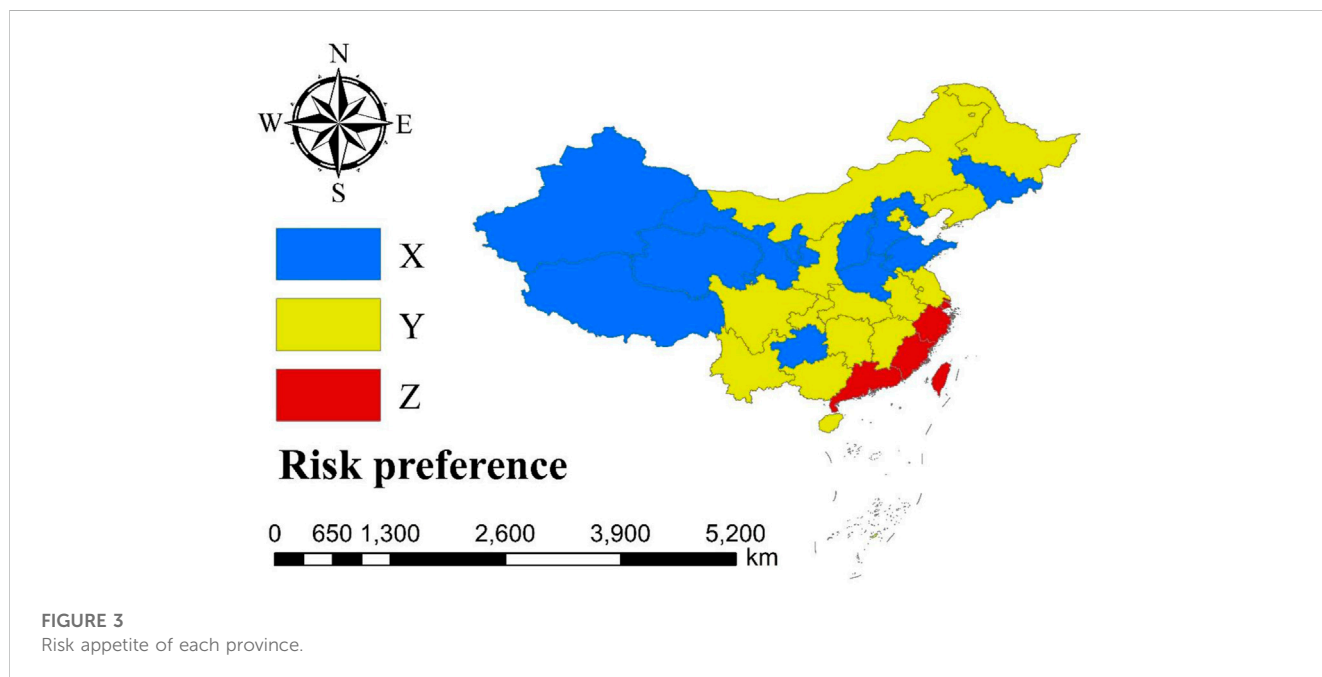


FIGURE 3 Risk appetite of each province.

(Francis et al., 2022). The higher the quality of the ecological environment, the higher the losses caused by disasters, and the more averse they are to risk. The per-unit area average ESV and the per capita ecological environment value reflect the ecological environment quality of a region from different perspectives. The higher the per capita ESV, the stronger sustainable development capability of the region. However, this does not necessarily mean that the ecological damage caused by disasters is more severe. For example, in areas with poor land quality, floods may not cause considerable ecological damage; however, the per capita ESV may still be relatively high because of low population density (Peter et al., 2022). Therefore, the average ESV per unit area is a better indicator of the importance of a region's ecological environment. The higher the average ESV per unit area, the greater are the losses caused by floods. Therefore, per capita GDP and per unit area average ESV were selected to construct a risk preference matrix. Based on statistical data released by the National Bureau of Statistics in 2020 (National Bureau of statistics of the people's Republic of China, 2020), the per capita GDP and per unit area average ESV of each province are shown in Table 3.

According to the 2020 income classification standard of the World Bank, most provinces in China have an annual per capita income in the middle-to high-income range, whereas the remaining provinces have a per capita income in the high-income range (Shah

et al., 2022). To differentiate the per capita GDP income of each province more precisely, the middle-to high-income range was evenly divided into three categories: 2.58–5.29, 5.29–8, and >8 yuan per capita GDP. The average ecological service value per unit area (ESV) for each province was determined by setting the highest and lowest values as the upper and lower limits of the range and equally dividing the range into three categories: 8–14, 14–21, and 21–28 yuan per unit area. The resulting risk-preference matrices are presented in Table 4.

When falling into zone Z, $n = 2$ is assigned, indicating risk aversion; when falling into zone Y, $n = 1.5$ is assigned, indicating slight risk aversion; and when falling into zone X, $n = 1$ is assigned, indicating risk neutrality. Based on the per capita GDP and average ESV per unit area of each province in 2020, the risk preference of each province was determined using a risk matrix, as shown in Figure 3.

As shown in Figure 3, the regions with risk preference X primarily include some western and central provinces of China and Jilin Province. Western regions, such as Xinjiang, Tibet, Qinghai, Gansu, and Ningxia, have relatively underdeveloped economies. People in these areas tend to focus more on economic development than on the ecological environment, resulting in a risk preference of X. In central provinces, such as Hebei, Henan, Shanxi, and Shandong, although the total economic

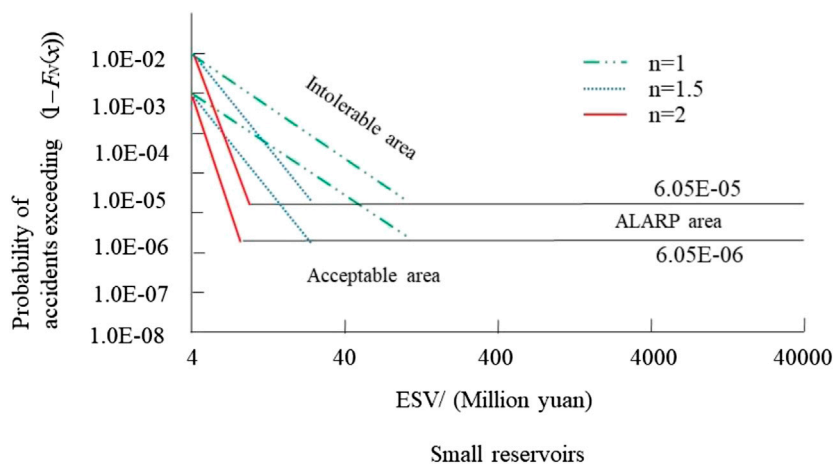


FIGURE 4
Risk standards for environmental losses caused by dam failure in large- and medium-sized reservoirs.

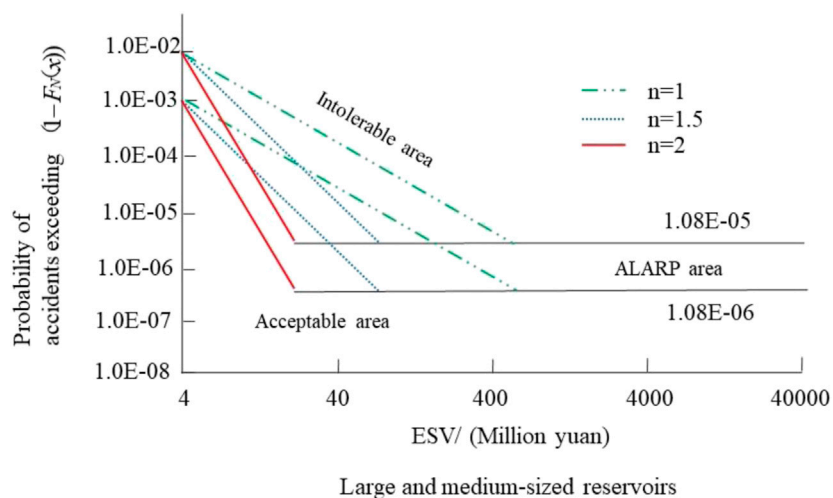


FIGURE 5
Risk criteria for environmental losses caused by dam failure in small reservoirs.

output is large, the population is large, and the *per capita* GDP is relatively small, resulting in a risk preference of X. Among the three northeastern provinces, only Jilin Province has a risk preference of X. Upon comparison, although Jilin's *per capita* GDP is higher than that of Heilongjiang, Heilongjiang has important ecological protection areas, such as the Greater and Lesser Khingan Mountains, with higher ecological value and lower risk tolerance. Therefore, among the three northeastern provinces, only Jilin Province has a risk preference of X.

The risk preference of most provinces in China is Y, where some areas have high ecological environment quality despite being economically underdeveloped, whereas others have moderate ecological environment quality despite being economically developed (i.e., Beijing). Additionally, some regions, such as Hunan, have good overall economic development and ecological environment quality; however, a large population means that their

per capita GDP has not reached the high-income range, resulting in a Y risk preference.

Only Shanghai, Hong Kong, Guangdong, Fujian, Zhejiang, and Taiwan have a Z-risk preference and are concentrated in the southeastern coastal region of China. These areas are economically developed and have a high ecological environment quality, resulting in a lower tolerance for risk.

4.4 Determination of extreme value lines

Most countries or regions have set extreme value lines for the risk standards of the two types. First, the probability extreme value line set by Australia indicates that there is no need to consider the consequences of an accident if its probability is below a certain value (ANCOLD Australia National Committee on Large Dams, 2003).

The other, such as the loss extreme value line set by Hong Kong, indicates that if the accident loss exceeds a certain value, then the risk is intolerable. Considering the safety status of Chinese dams, management levels, and levels of economic and social development, it is currently not feasible to set extreme accident value lines (Li et al., 2015). Therefore, the following two points should be considered when setting the accident probability extreme value line:

- (1) Probability of dam accidents in China. According to (Li et al., 2015; Fan and Jiang, 2005), the annual average probability of dam failure can be considered as 10% for the tolerable risk extreme value line and 1% for the acceptable risk extreme value line. From 1954 to 2018, 3,541 dam failures occurred in China, with an annual failure rate of 6.29×10^{-4} . Therefore, the tolerable and acceptable extreme value lines were 6.29×10^{-5} and 6.29×10^{-6} , respectively.
- (2) Reliability standards in China. Reference (China Planning Press, 2013) stipulates the values of reliability indicators in China: 4.2 for large (I) reservoirs, 3.7 for large (II) and medium-sized reservoirs, and 3.2 for small reservoirs. According to the reliability theory, reliability indicators can evaluate or measure the level of structural reliability, as a quantitative basis for expressed symbolically, its relationship with failure probability P_f is as follows:

$$Pf = 1 - \Phi(\beta) \quad (4)$$

Where $\Phi(\cdot)$ represents the standard normal distribution function.

Assuming that the reliability function is a random variable following a normal distribution (Wilde and Johansson, 2013; Li S. et al., 2015; Ge et al., 2017). The current reliability standards in China are relatively consistent with the dam safety situation, so 10% of P_f can also be used as the tolerable risk extreme line and 1% as the acceptable risk extreme line.

As people's concern for the ecological losses caused by dam failures is significantly lower than that for the loss of life and economic losses, these requirements can be appropriately reduced. Moreover, the funding guarantees and management levels of large reservoirs were better than those of medium and small reservoirs. Among the 84 dam failures from 2000 to 2018, there were no failures in large reservoirs, 7 in medium-sized reservoirs, and the remainder in small reservoirs (Li H. et al., 2021). Therefore, to avoid excessive complexity in standards, large (I) reservoirs can adopt the same standards as large (II) and medium reservoirs. Thus, for large- and medium-sized reservoirs, the tolerable and acceptable risk probability extreme value lines are 1.08×10^{-5} and 1.08×10^{-6} , respectively, whereas for small reservoirs, the tolerable and acceptable risk probability extreme value lines are 7×10^{-5} and 6.29×10^{-6} , respectively.

To simultaneously meet the extreme value lines determined according to the safety level of the dam and current reliability standards, a smaller value was considered as the risk standard. Finally, the tolerable and acceptable risk probability extreme value lines for medium-sized reservoirs are 1.08×10^{-5} and 1.08×10^{-6} , respectively, whereas the tolerable and acceptable risk probability extreme value lines for small reservoirs are 6.29×10^{-5} and 6.29×10^{-6} , respectively.

5 Results

Based on the determination of the C value, n value, and limit line, environmental risk standards for Chinese reservoir dams were established, as shown in Figures 4, 5.

6 Conclusion

Reasonable dam-failure risk standards form the foundation for accurate risk assessment and management. There are difficult to define issues regarding the environment and its impacts, this study introduced ESV as a practical and operational method for determining environmental standards. The different provinces in China were divided into regions with distinctive characteristics, and the risk matrix method was used to determine the risk standards for these regions and established more targeted risk standards. This study determined the tolerable and acceptable standard values C of China's environmental impact risk standards, respectively. According to the GDP and environmental conditions of different provinces in China, a risk matrix is introduced to determine the risk appetite of each province. Finally, after the extreme value line has been determined, environmental loss risk criteria for reservoir dams of different sizes have been developed. These results were consistent with reality. The results can effectively help decision makers fully understand the level of dam risk management, promote further development of environmental risk research, and provide references and guidance for subsequent studies on environmental risks of dam failure.

Data availability statement

The datasets presented in this study can be found in online repositories. The names of the repository/repositories and accession number(s) can be found in the article/Supplementary Material.

Author contributions

Conceptualization, WL and JY; methodology, JL and JY; validation WL and HZ; formal analysis, WL and HZ; investigation, WL and JL; writing—original draft preparation, WL; writing—review and editing, WL and YJ; supervision, WL; funding acquisition, WL. All authors contributed to the article and approved the submitted version.

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Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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