



# Biomarkers in Ringed Seals Reveal Recent Onset of Borealization in the High- Compared to the Mid-Latitude Canadian Arctic

Camille de la Vega<sup>1\*†</sup>, Claire Mahaffey<sup>1</sup>, David J. Yurkowski<sup>2</sup>, Louisa Norman<sup>1</sup>, Elysia Simpson<sup>1</sup>, Sophie Smout<sup>3</sup>, Steven H. Ferguson<sup>2</sup> and Rachel M. Jeffreys<sup>1</sup>

<sup>1</sup> School of Environmental Sciences, University of Liverpool, Liverpool, United Kingdom, <sup>2</sup> Freshwater Institute, Fisheries and Oceans Canada, Winnipeg, MB, Canada, <sup>3</sup> Sea Mammal Research Unit, Gatty Marine Laboratory, St. Andrews, United Kingdom

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### \*Correspondence:

Camille de la Vega  
camille.de-la-vega@  
liverpool.ac.uk

### † Present address:

Camille de la Vega,  
Leibniz Institute for Baltic Sea  
Research, Warnemünde, Germany

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Warming of the Arctic has resulted in environmental and ecological changes, termed borealization, leading to the northward shift of temperate species. Borealization has occurred across all trophic levels, altering the structure of the food web. The onset and rate of borealization likely varies with latitude, depending on local warming and advection of warmer water into the Arctic. In order to assess latitudinal trends in food web structure in the Arctic, we analyzed stable nitrogen isotopes of specific amino acids alongside bulk stable carbon isotopes in ringed seal muscle tissue from the Canadian Arctic Archipelago (high-Arctic) and Southern Baffin Bay (mid-Arctic) from 1990 to 2016. Our results indicate a shift in food web structure in the high-Arctic that has occurred more recently when compared with the mid-Arctic. Specifically, over the past 25 years, the trophic position of ringed seals from the mid-Arctic was largely constant, whereas the trophic position of ringed seals decreased in the high-Arctic, reaching similar values observed in the mid-Arctic in 2015–2016. This suggests a potential shortening of the food chain length in the high-Arctic, possibly driven by changes in zooplankton communities feeding complexity in association with sea ice decline. This study identifies a temporal offset in the timing of borealization in the Canadian Arctic, resulting in different response of food webs to ecological changes, depending on latitude.

**Keywords:** borealization, Canadian Arctic, ringed seals, stable isotopes, amino acids, latitudes, trophic position, food web structure

## INTRODUCTION

The Arctic is changing rapidly, warming twice as fast as the global average causing sea ice extent and thickness to decline (Meredith et al., 2019). Increased sea surface temperature has led to northward shifts of subarctic and temperate phytoplankton, zooplankton and fish species, a process described as “borealization” (Fossheim et al., 2015; Möller and Nielsen, 2020). These changes in community structure have resulted in changes to predator-prey interactions, with implications for food web stability and resilience (Kortsch et al., 2015; Yurkowski et al., 2018; Pecuchet et al., 2020). Monitoring changes in the marine environment, especially in polar regions, is logistically

challenging due to strong seasonality and remoteness. Archives of marine predator tissues offer a unique resource in overcoming the limited environmental sampling opportunities in the Arctic.

Trophic position is a fundamental property of ecological communities that can reflect integrated changes in ecosystems. Food web structure and the functional roles of predators can be assessed using trophic position as an expression of food chain length using top and near-top trophic level species (Post, 2002a). Here, we used the trophic position of ringed seals (*Pusa hispida*), a generalist ice-dependent Arctic predator, to assess latitudinal differences in multi-decadal trends in food chain length in the Canadian Arctic. Ringed seals are circumpolar, and have an extensive distribution in northern Canada, occurring continuously across latitudes from Newfoundland and Labrador to the Arctic basin (Ferguson et al., 2020). They are predators and display high plasticity in their diet with dietary items consisting of a wide variety of pelagic invertebrates, e.g., mysids and amphipods species, and forage fish, e.g., Arctic cod (*Boreogadus saida*) and capelin (*Mallotus villosus*; Ogloff et al., 2019; Bengtsson et al., 2020). Despite being generalists, ringed seals display a preference for Arctic prey, especially the lipid-rich Arctic cod (Hamilton et al., 2019; Bengtsson et al., 2020), which makes them excellent indicators of biotic changes in Arctic food webs.

The ratios between heavy ( $^{15}\text{N}$  and  $^{13}\text{C}$ ) and light ( $^{14}\text{N}$  and  $^{12}\text{C}$ ) stable nitrogen and carbon isotopes, referred to as  $\delta^{15}\text{N}_{\text{bulk}}$  and  $\delta^{13}\text{C}_{\text{bulk}}$  hereafter, are a common tool for studying food webs.  $\delta^{13}\text{C}_{\text{bulk}}$  typically undergoes minimal fractionation ( $0.4 \pm 1.3\text{‰}$ ; Post, 2002b) between trophic steps and has been used to trace the origin of resources, e.g., phytoplankton versus sea ice algae (Hobson et al., 1995).  $\delta^{15}\text{N}_{\text{bulk}}$  is a commonly used tracer for estimating trophic position, as it generally undergoes greater fractionation between each trophic level ( $3.4 \pm 1.0\text{‰}$ ; Post, 2002b), but is heavily dependent on knowledge of  $\delta^{15}\text{N}$  at the base of the food web, or “baseline.”  $\delta^{15}\text{N}$  of nitrate, taken up by phytoplankton, represents the  $\delta^{15}\text{N}$  at the baseline (Mariotti et al., 1981). The  $\delta^{15}\text{N}$ -nitrate differs between Atlantic and Pacific waters (Somes et al., 2010), thus supplying nitrate with distinct water mass properties alongside  $\delta^{15}\text{N}$ -nitrate values. Spatial trends in water mass influence alongside temporal changes in circulation leads to a heterogenous and evolving  $\delta^{15}\text{N}$  baseline across the Arctic, that needs to be accounted for in order to reliably detect changes in trophic position of predators (de la Vega et al., 2020).

Compound-specific stable nitrogen isotopes of amino acids ( $\delta^{15}\text{N}_{\text{AA}}$ ) can overcome this issue.  $\delta^{15}\text{N}_{\text{AA}}$  is a powerful biomarker approach that disentangles baseline and fractionating trophic effects when using  $\delta^{15}\text{N}$  values to estimate trophic position. The  $\delta^{15}\text{N}$  of the “source” amino acid (AA) phenylalanine ( $\delta^{15}\text{N}_{\text{Phe}}$ ), can conservatively trace the  $\delta^{15}\text{N}$  at the baseline as it experiences minimal fractionation ( $<1\text{‰}$ ) during trophic transfer (McMahon and McCarthy, 2016). The  $\delta^{15}\text{N}$  of “trophic” AAs ( $\delta^{15}\text{N}_{\text{trophic}}$ ), e.g., glutamic acid, undergoes significant fractionation ( $>3\text{‰}$ ) resulting in  $^{15}\text{N}$  enrichment between each trophic step (McMahon and McCarthy, 2016). This approach simultaneously fingerprints both the  $\delta^{15}\text{N}$  baseline and trophic information in a predator from the analysis of predator

tissue alone. This allows accurate estimation of changes in relative trophic position while accounting for variation in  $\delta^{15}\text{N}$  baseline, using baseline-corrected  $\delta^{15}\text{N}_{\text{trophic}}$  ( $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$ ; de la Vega et al., 2020). This is particularly powerful when studying multi-decadal trends in predators that integrated temporally evolving baselines across latitudes.

Mechanisms involved in the northward expansion of boreal and temperate species are directly associated with the local rate of warming as well as the advection of warmer water into the Arctic (Wassmann, 2015). Hence, the rate and timing of borealization likely varies with latitude (Kortsch et al., 2015). Here, we used temporal trends in mean sea ice concentration and length of the open water period (OWP) as an indication of the rate of warming across latitudes. The Canadian Arctic Archipelago (CAA) is one of the two Arctic outflow shelves, and receives water exiting the Arctic, mainly consisting of Pacific waters flowing southward into the Baffin Bay (Figure 1; Carmack and Wassmann, 2006). In addition, Southern Baffin Bay receives warmer Atlantic water advected via the Irminger current and flowing northward along the West coast of Greenland (Figure 1; Tang et al., 2004), potentially enhancing the northward spread of boreal and temperate species (Wassmann, 2015), and providing new potential prey species for opportunistic predators, such as ringed seals.

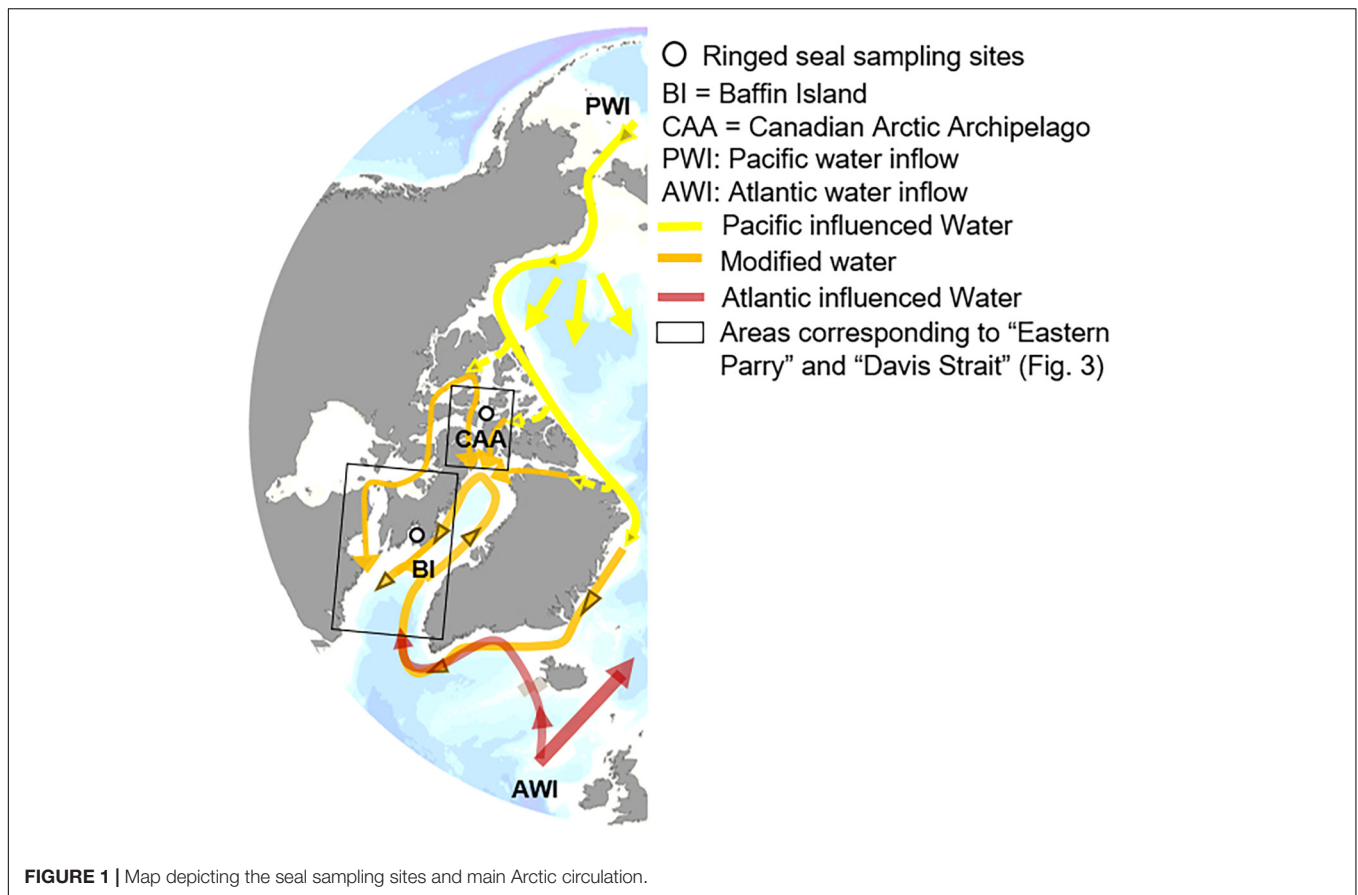
The main objective of this study was to assess the latitudinal difference in the response of food webs to environmental changes in the Arctic. We used  $\delta^{15}\text{N}_{\text{AA}}$  and  $\delta^{13}\text{C}_{\text{bulk}}$  in ringed seal muscle from 1990 to 2016, from the CAA and Baffin Island to monitor the trophic structure of both systems, respectively at high- and mid-Arctic latitudes, influenced by different water masses (Figure 1). We predicted that changes in the mid-Arctic will be more prominent than in the high-Arctic in relation to the changing environment, which in turn influences the rate of borealization in these systems. We hypothesized that: (1)  $\delta^{15}\text{N}_{\text{Phe}}$  and  $\delta^{13}\text{C}_{\text{bulk}}$ , representing isotopic baselines, will vary temporally in association with environmental changes; (2)  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$ , representing the relative trophic position of ringed seals will vary temporally and reflect changes in trophic structure; and (3) the temporal trends in  $\delta^{13}\text{C}_{\text{bulk}}$ ,  $\delta^{15}\text{N}_{\text{Phe}}$ , and  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  will differ between the high- and mid-Arctic, reflecting differences in the onset and rate of environmental change, and in turn borealization between latitudes.

## MATERIALS AND METHODS

### Environmental Data

Sea ice concentrations for the CAA and the Baffin Island region were obtained using Canadian Ice Service’s IceGraph 2.0 tool<sup>1</sup> which disseminates sea ice data on regional and sub-regional spatial scales across the Canadian Arctic. We queried databases for each of the pre-defined sub-regions of “Eastern Parry” and “Davis Strait,” which encompassed Resolute and Pangnirtung where ringed seal muscle were collected, for total accumulated ice concentrations from early February to end of September,

<sup>1</sup><http://iceweb1.cis.ec.gc.ca/IceGraph>, accessed July 2021



representing the period integrated in ringed seal muscles, per year from 1978 to 2021. OWP was estimated as the number of days on which the weekly sea ice concentration reached and remained below 50% in “Eastern Parry” and “Davis Strait” sub-regions.

## Seal Sampling Design

Muscle tissue from adult ringed seals were opportunistically sampled in the CAA and Baffin Island as part of Inuit subsistence harvests. Samples from the CAA were collected in Resolute, Nunavut (latitude 75°N) from 1992 to 2016 ( $n = 66$ ) and samples from Baffin Island were collected in Pangnirtung, Nunavut (latitude 66°N) from 1990 to 2016 ( $n = 39$ , **Table 1**). All samples were stored at  $-20^{\circ}\text{C}$  prior to analysis.  $\delta^{15}\text{N}$  and  $\delta^{13}\text{C}$  values of marine mammals do not change through long-term storage in freezers (Yurkowski et al., 2017).

The majority of the samples were collected between early summer (June) and early fall (September). The turnover time

of muscle tissues is poorly known for marine mammals. Vander Zanden et al. (2015) estimated the stable isotope half-life of muscle tissue for a mammal to be 29.4 days for 50 g of body mass. Assuming an average weight for ringed seals of 60 kg, ringed seal's muscle tissue integrates  $\delta^{15}\text{N}$  and  $\delta^{13}\text{C}$  over  $\sim 5$  months, therefore mainly reflecting the seal foraging from spring (March) to summer (July). In the Canadian Arctic, ringed seals tend to be resident during the ice-season and can undertake extensive movements during the ice-free season (Yurkowski et al., 2016b). Here, muscle tissue from ringed seals from both the CAA and Baffin Island mainly represented the diet during the ice-season and break-up, likely reflecting the stable isotope signal from the CAA and Southern Baffin Bay, respectively.

## Stable Isotope Analyses of Seal Samples

Stable isotope ( $\delta^{13}\text{C}_{\text{bulk}}$ ,  $\delta^{15}\text{N}_{\text{bulk}}$ , and  $\delta^{15}\text{N}_{\text{AA}}$ ) analysis of seal muscle tissue was carried out at the Liverpool Isotopes for Environmental Research Laboratory, University of Liverpool and results are reported in standard  $\delta$ -notation (‰).

## Stable Isotope Analyses of Bulk Tissue

Each muscle sample was freeze-dried and homogenized. Approximately 0.5 mg of sample was precisely weighed ( $\pm 1 \mu\text{g}$ ) and sealed in a tin capsule. Samples were run for  $\delta^{15}\text{N}_{\text{bulk}}$  and  $\delta^{13}\text{C}_{\text{bulk}}$  separately. Lipids were removed from the samples using

**TABLE 1** | Sampling years, age class and number of muscle samples ( $n$ ) of ringed seals from the Canadian Arctic Archipelago (CAA) and the Baffin Island (BI).

Site	Sampling years	Age	$n$
CAA	1992–2016	>5	66
BI	1990–2016	>5	39

five repeated rinses with 2:1 dichloromethane:methanol to avoid the bias due to the depletion in  $^{13}\text{C}$  in lipids relative to the diet, before  $\delta^{13}\text{C}_{\text{bulk}}$  analyses.

Samples were analyzed using an elemental analyzer (Costech) coupled to Delta V isotope ratio mass spectrometer (IRMS; Thermo Fisher Scientific). Isotope values are reported in standard  $\delta$ -notation (‰) relative to Vienna PeeDee Belemnite and atmospheric  $\text{N}_2$  for  $\delta^{13}\text{C}_{\text{bulk}}$  and  $\delta^{15}\text{N}_{\text{bulk}}$ , respectively. We used an internal standard of ground prawn *Penaeus vannamei* with well characterized  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  values ( $-22.6$  and  $6.8$ ‰, respectively) to monitor precision. We analyzed this internal standard every 10 samples, and precision was  $< 0.2$ ‰ for both  $\delta^{13}\text{C}_{\text{bulk}}$  and  $\delta^{15}\text{N}_{\text{bulk}}$ . To determine accuracy, we used international reference standards with known  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  values, USGS40 ( $\delta^{13}\text{C} = -26.24$ ‰ and  $\delta^{15}\text{N} = -4.52$ ‰) and USGS41a ( $\delta^{13}\text{C} = 36.55$ ‰ and  $\delta^{15}\text{N} = 47.55$ ‰). Both standards were analyzed at the beginning, middle and end of each run and accuracies for all the EA-IRMS runs combined here. For USGS40, mean accuracy was  $0.13 \pm 0.09$ ‰ for  $\delta^{13}\text{C}$  and  $0.03 \pm 0.11$ ‰ for  $\delta^{15}\text{N}$ . For USGS41a, mean accuracy was  $0.01 \pm 0.13$ ‰ for  $\delta^{13}\text{C}$  and  $0.01 \pm 0.15$ ‰ for  $\delta^{15}\text{N}$  ( $1\sigma$ ,  $n = 19$ ).

### $\delta^{15}\text{N}$ Analyses on Amino Acids

Approximately 7 mg of sample was hydrolyzed and propylated. Amino acid methyl esters were then acetylated, dissolved in DCM and stored at  $-20^\circ\text{C}$  until analysis. For a full description of the laboratory protocol see the **Supplementary Material**.

$\delta^{15}\text{N}_{\text{AA}}$  values were determined using a Trace Ultra gas chromatograph (GC) coupled to a Delta V Advantage IRMS with a ConFlo IV interface (Thermo Fisher Scientific). A liquid nitrogen trap was added after the reduction oven to remove  $\text{CO}_2$  from the sample stream. The separation of AA was achieved using an HP INNOWax capillary column (Agilent). Each sample was analyzed in triplicate (measurement error  $< 1.0$ ‰). Precision and accuracy were determined using a mixed AA standard prepared from eight AA with known  $\delta^{15}\text{N}$  values (University of Indiana, United States and SI Science, Japan), analyzed every four injections. The mean precision of all standards was  $\pm 0.9$ ‰ and ranged from  $\pm 0.7$ ‰ for phenylalanine and  $\pm 1.2$ ‰ for valine ( $n = 35$ ; **Supplementary Table 1**). The mean accuracy of all standards was  $\pm 0.7$ ‰ and ranged from  $\pm 0.1$ ‰ for leucine and  $\pm 1.4$ ‰ for glutamic acid ( $n = 35$ ; **Supplementary Table 1**). Details of the method, typical precisions and accuracies, and  $\delta^{15}\text{N}$  of all identified AA are shown in **Supplementary Table 1**.

### Seal Trophic Position Estimation

We used  $\delta^{15}\text{N}_{\text{Phe}}$ , to track the  $\delta^{15}\text{N}$  baseline in ringed seals. We used the mean  $\delta^{15}\text{N}$  value of three main trophic amino acids (glutamic acid, aspartic acid, and leucine; de la Vega et al., 2020), to estimate the relative trophic position of ringed seals. The uncertainty regarding trophic fractionation factors between “source” and “trophic” amino acids across taxa prevents accurate estimation of an organism’s absolute trophic position (Nielsen et al., 2015). To assess the decadal trends in relative trophic position, we averaged the  $\delta^{15}\text{N}$  values of the three trophic amino

acids that we baseline-corrected using  $\delta^{15}\text{N}_{\text{Phe}}$  (Eq. 1), providing baseline-corrected  $\delta^{15}\text{N}_{\text{trophic}}$  values ( $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$ ).

$$\text{Cor-}\delta^{15}\text{N}_{\text{trophic}} = \delta^{15}\text{N}_{\text{trophic}} - \delta^{15}\text{N}_{\text{Phe}} \quad (1)$$

### Statistical Analyses

Statistical analyses were performed in R version 3.5.1 (R Core Team, 2018). We assessed the relationship between  $\delta^{15}\text{N}_{\text{bulk}}$ ,  $\delta^{15}\text{N}_{\text{Phe}}$ , and  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  using linear models. To analyze temporal variation in sea ice concentration and OWP, and  $\delta^{13}\text{C}_{\text{bulk}}$ ,  $\delta^{15}\text{N}_{\text{Phe}}$ , and  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  in ringed seals, linear models were fitted with sea ice concentration, OWP,  $\delta^{13}\text{C}_{\text{bulk}}$ ,  $\delta^{15}\text{N}_{\text{Phe}}$ , and  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  as a function of year for each site separately. We used the Gaussian family with no transformation of the data, assuming that measurement errors were normally distributed. Model fit was checked by residual analyses with visual inspection of quantile–quantile plots, and residuals and standardized residuals versus fitted values plots (**Supplementary Material**). To check for potential influential data points on the temporal trends in stable isotopes, we used leverage (i.e., hat values), studentized residuals and the Cook’s distance (Zuur et al., 2009), using the “Car package” in R (Fox et al., 2007; **Supplementary Material**).  $p$ -Values,  $R^2$ ,  $F$ -statistics, and degrees of freedom are reported for each model (**Table 2**). We compared the  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  values between Southern Baffin Bay and the CAA for years when data was available from both regions, i.e., 1993–1994, 2006 and 2015–2016, using Student  $t$ -tests. Significance was set at  $\alpha = 0.01$ .

## RESULTS

### Temporal Trends in Sea Ice Concentration

In the CAA, mean sea ice concentration decreased by  $0.7 \pm 0.2\%$  year $^{-1}$  (linear model:  $p < 0.01$ ,  $R^2 = 33.7\%$ ) and did not vary in Southern Baffin Bay from 1990 to 2016 (**Figure 2A** and **Table 2**). The OWP became longer in the CAA by  $2.0 \pm 0.6$  days year $^{-1}$  (linear model:  $p < 0.01$ ,  $R^2 = 33.7\%$ ) and did not vary in Southern Baffin Bay from 1990 to 2016 (**Figure 2B** and **Table 2**). Generally, the CAA had higher sea ice concentration ( $58.0 \pm 9.2\%$ ) and a much shorter OWP ( $96 \pm 27$  days) than Southern Baffin Bay ( $38.5 \pm 7.3\%$  and  $224 \pm 32$  days, respectively; **Figures 2A,B**).

### Temporal Trends in Stable Isotopes Values in Ringed Seals

$\delta^{13}\text{C}_{\text{bulk}}$  became depleted in  $^{13}\text{C}$  (hereafter decreased) by  $-0.029 \pm 0.008$ ‰ y $^{-1}$  in ringed seals from the CAA from 1992 ( $-17.7 \pm 0.7$ ‰) to 2016 ( $-18.6 \pm 0.8$ ‰; linear model:  $p < 0.01$ ,  $R^2 = 9.4\%$ ) and by  $-0.021 \pm 0.007$ ‰ y $^{-1}$  in ringed seals from Baffin Island from 1990 ( $-18.1 \pm 0.5$ ‰) to 2015–2016 ( $-18.9 \pm 0.2$ ‰; linear model:  $p < 0.01$ ,  $R^2 = 17.0\%$ ; **Figures 3A,B** and **Table 2**).

$\delta^{15}\text{N}_{\text{Phe}}$  became enriched in  $^{15}\text{N}$  (hereafter increased) by  $0.060 \pm 0.013$ ‰ year $^{-1}$  in ringed seals from the CAA (linear

**TABLE 2** | Statistical parameters of the linear models for stable isotopes of ringed seal muscle, sea ice concentration, and open water period from each Arctic regions.

Site	Model	Explained variable	Explanatory variable	<i>n</i>	Intercept (±SD)	<i>p</i> -value	Slope (±SD)	<i>p</i> -value	<i>R</i> <sup>2</sup> (%)	<i>F</i> -value (df)
CAA	Linear	Sea ice	Year	44	7.69 ± 1.80	<0.01	−0.004 ± 0.001	<0.01	25.1	15.42 (42)
BI	Linear	Sea ice	Year	44	3.03 ± 1.90	0.119	−0.001 ± 0.001	0.172	2.1	1.93 (42)
CAA	Linear	OWP	Year	44	−2707 ± 588.6	<0.01	1.397 ± 0.294	<0.01	33.4	22.52 (42)
BI	Linear	OWP	Year	44	1105.7 ± 1072.5	0.308	−0.443 ± 0.537	0.413	0.0	0.68 (43)
CAA	Linear	δ <sup>13</sup> C <sub>bulk</sub>	Year	66	23.31 ± 15.33	0.025	−0.029 ± 0.008	<0.01	19.4	10.40 (38)
BI	Linear	δ <sup>13</sup> C <sub>bulk</sub>	Year	39	24.02 ± 14.91	0.116	−0.021 ± 0.007	<0.01	17.0	8.18 (34)
CAA	Linear	δ <sup>15</sup> N <sub>Phe</sub>	Year	66	−109.91 ± 25.33	<0.01	0.060 ± 0.013	<0.01	25.2	22.95 (64)
BI	Linear	δ <sup>15</sup> N <sub>Phe</sub>	Year	39	−52.95 ± 21.88	0.021	0.032 ± 0.011	<0.01	16.8	8.47 (36)
CAA	Linear	Cor-δ <sup>15</sup> N <sub>trophic</sub>	Year	66	167.45 ± 48.01	<0.01	−0.077 ± 0.024	<0.01	12.6	10.39 (64)
BI	Linear	Cor-δ <sup>15</sup> N <sub>trophic</sub>	Year	39	48.50 ± 29.82	0.113	0.019 ± 0.015	0.216	1.6	1.59 (36)
CAA	Linear	δ <sup>15</sup> N <sub>bulk</sub>	δ <sup>15</sup> N <sub>Phe</sub>	66	10.77 ± 0.66	<0.01	0.564 ± 0.058	<0.01	59.2	95.32 (64)
BI	Linear	δ <sup>15</sup> N <sub>bulk</sub>	δ <sup>15</sup> N <sub>Phe</sub>	39	6.33 ± 1.02	<0.01	0.880 ± 0.095	<0.01	69.6	85.83 (36)
CAA	Linear	δ <sup>15</sup> N <sub>bulk</sub>	Cor-δ <sup>15</sup> N <sub>trophic</sub>	66	17.99 ± 0.65	<0.01	−0.062 ± 0.051	0.234	0.7	1.45 (64)
BI	Linear	δ <sup>15</sup> N <sub>bulk</sub>	Cor-δ <sup>15</sup> N <sub>trophic</sub>	39	15.49 ± 1.53	<0.01	0.025 ± 0.139	0.858	0.0	0.03 (36)

df, degree of freedom; *n*, number of samples; SD, standard deviation; CAA, Canadian Arctic Archipelago; BI, Baffin Island. Lines in bold are considered significant.

model:  $p < 0.01$ ,  $R^2 = 25.2\%$ ) and by  $0.032 \pm 0.011\text{‰ year}^{-1}$  in ringed seals from Baffin Island (linear model:  $p < 0.01$ ,  $R^2 = 16.8\%$ ; **Figures 3C,D** and **Table 2**).  $\delta^{15}\text{N}_{\text{Phe}}$  increased from the early 1990s to the mid-2010s from  $10.4 \pm 0.7\text{‰}$  (1992–1993) to  $11.9 \pm 0.6\text{‰}$  (2015–2016) in ringed seals from the CAA and from  $10.7 \pm 0.5\text{‰}$  (1990–1992) to  $11.1 \pm 0.6\text{‰}$  (2015–2016) in ringed seals from Baffin Island.

Cor- $\delta^{15}\text{N}_{\text{trophic}}$  of ringed seals from the CAA decreased by  $-0.077 \pm 0.024\text{‰ year}^{-1}$  from 1992 to 2016 (linear model:  $p < 0.01$ ,  $R^2 = 12.6\%$ ; **Figure 4A** and **Table 2**). Cor- $\delta^{15}\text{N}_{\text{trophic}}$  of ringed seals from Baffin Island did not vary with time (**Figure 4A** and **Table 2**). Cor- $\delta^{15}\text{N}_{\text{trophic}}$  values were enriched in  $^{15}\text{N}$  in ringed seals from the CAA compared to Baffin Island in 1992–1993 ( $12.6 \pm 1.6$  and  $10.9 \pm 0.9\text{‰}$ , respectively; Student test:  $p < 0.01$ ,  $n = 18$ ) and in 2006 ( $12.5 \pm 0.7$  and  $10.7 \pm 0.9\text{‰}$ , respectively; Student test:  $p < 0.01$ ,  $n = 10$ ; **Figure 4A**). Cor- $\delta^{15}\text{N}_{\text{trophic}}$  values were similar in ringed seals from the CAA ( $11.4 \pm 1.4\text{‰}$ ) and Baffin Island ( $10.8 \pm 0.5\text{‰}$ ) in 2015–2016 (Student test:  $p = 0.159$ ,  $n = 17$ ; **Figure 4A**).

## Relationship Between $\delta^{15}\text{N}_{\text{bulk}}$ and $\delta^{15}\text{N}_{\text{AA}}$

There was a strong positive correlation between  $\delta^{15}\text{N}_{\text{bulk}}$  and  $\delta^{15}\text{N}_{\text{Phe}}$  in ringed seals muscle from the CAA (linear model:  $p < 0.01$ ,  $R^2 = 59.2\%$ ; **Figure 5A**) and Baffin Island (linear model:  $p < 0.01$ ,  $R^2 = 69.6\%$ ; **Figure 5B** and **Table 2**).  $\delta^{15}\text{N}_{\text{bulk}}$  was not correlated with Cor- $\delta^{15}\text{N}_{\text{trophic}}$  at either location (**Figures 5C,D** and **Table 2**).

## DISCUSSION

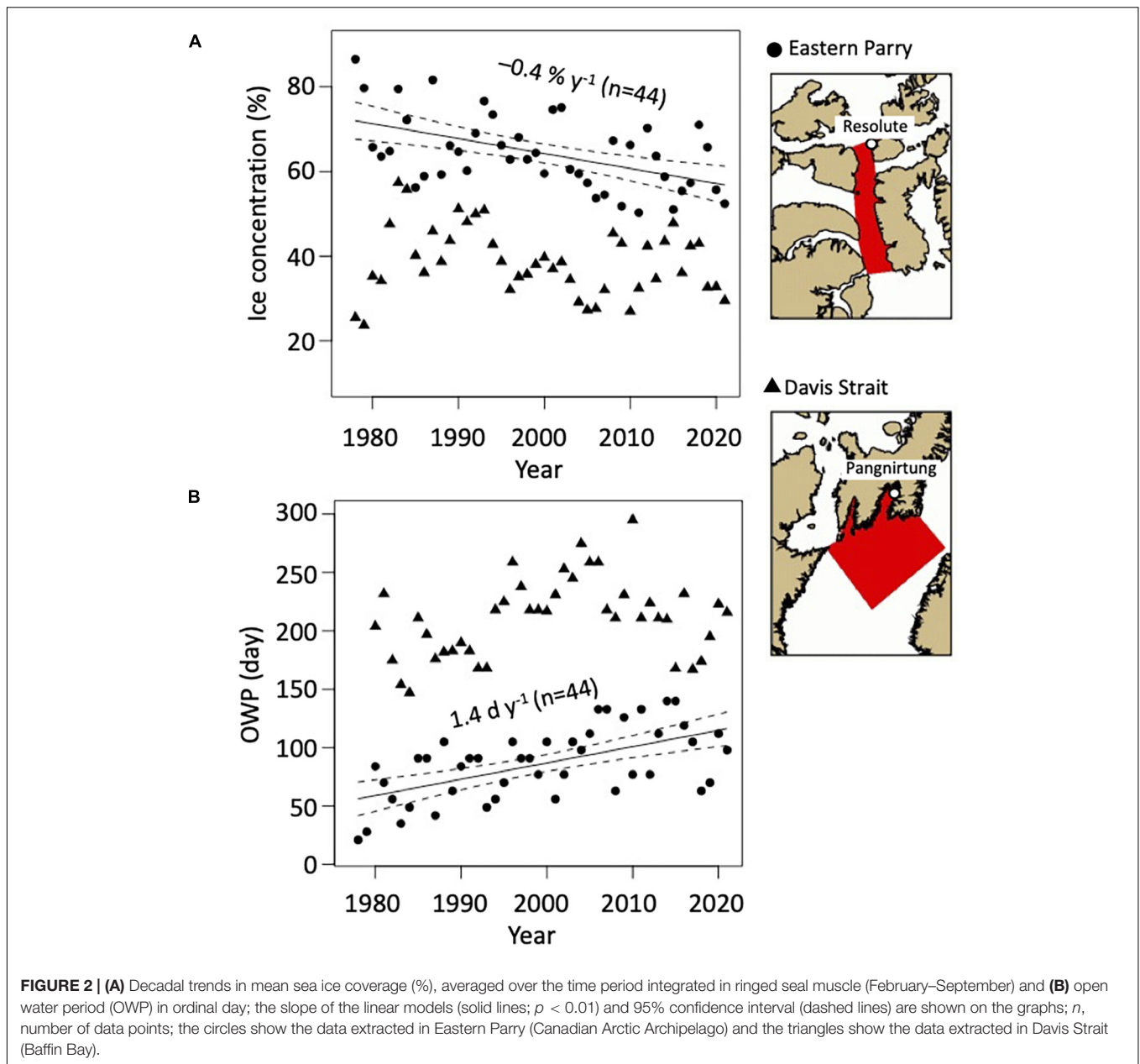
### Temporal Trends in $\delta^{13}\text{C}_{\text{bulk}}$

The annual decline in  $\delta^{13}\text{C}_{\text{bulk}}$  in ringed seals of  $-0.029\text{‰ year}^{-1}$  in the CAA and  $-0.021\text{‰ year}^{-1}$  in the Baffin Island region (**Figures 3A,B**) is in accordance with the temporal decline in  $\delta^{13}\text{C}_{\text{bulk}}$  of  $-0.020 \pm 0.003\text{‰ year}^{-1}$  in Northern

fur seals from the Bering Sea (Newsome et al., 2007) and of  $-0.021 \pm 0.006\text{‰ year}^{-1}$  in beluga whales from Baffin Bay (Matthews and Ferguson, 2014). However, these trends need to be interpreted with caution, as a few points were driving the trends in both the CAA (two points in 1992 and two points in 2015) and the Baffin Island region (two points in 1990; **Supplementary Material**). This decline is likely due to a combination of several factors. First, enhanced emission of anthropogenic carbon dioxide ( $\text{CO}_2$ ), which is  $^{13}\text{C}$ -deplete relative to natural  $\text{CO}_2$  (Quay et al., 2003), has led to the increase in oceanic  $\text{CO}_2$  (Sabine et al., 2004) and a decline in  $\delta^{13}\text{C}$  of dissolved inorganic carbon ( $\delta^{13}\text{C}_{\text{DIC}}$ ). This is known as the Suess effect.  $\delta^{13}\text{C}_{\text{DIC}}$  is fractionated by phytoplankton during photosynthesis and represents the  $\delta^{13}\text{C}$  baseline in marine food webs (Young et al., 2013). In the Arctic, the decline in  $\delta^{13}\text{C}_{\text{DIC}}$  is estimated to be in average  $-0.011\text{‰ year}^{-1}$  and is transferred through the food web to predators (de la Vega et al., 2019). Second, sea ice derived carbon is generally  $^{13}\text{C}$ -enriched compared to phytoplankton derived carbon (Hobson et al., 1995; de la Vega et al., 2019). The strong sea ice decline alongside increased OWP in the CAA (**Figures 2A,B**) could result in a decreased contribution of  $^{13}\text{C}$ -enriched sea-ice sympatric production versus  $^{13}\text{C}$ -depleted pelagic phytoplankton, leading to an overall  $^{13}\text{C}$ -depleted carbon pool entering the food web in the high-Arctic (Yurkowski et al., 2020). Third, climate change has led to increased coastal erosion, river flow and melting of permafrost in the Canadian Arctic (MacDonald and Birchall, 2020; Terhaar et al., 2021). This could result in an increased influence of  $^{13}\text{C}$ -depleted terrestrial organic matter (Boutton, 1991) into the surrounding marine environment, contributing to the decline in  $\delta^{13}\text{C}_{\text{bulk}}$  in ringed seals from both the CAA and the Baffin Island region.

### Temporal Trends in the $\delta^{15}\text{N}$ Baseline

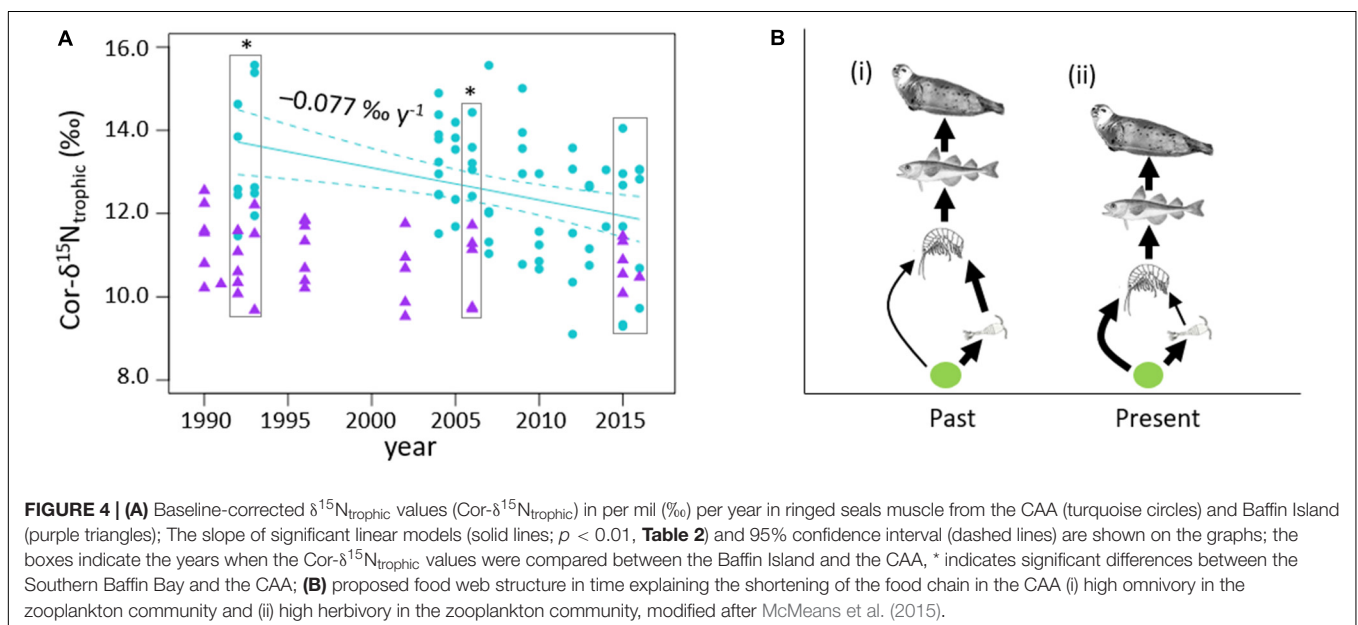
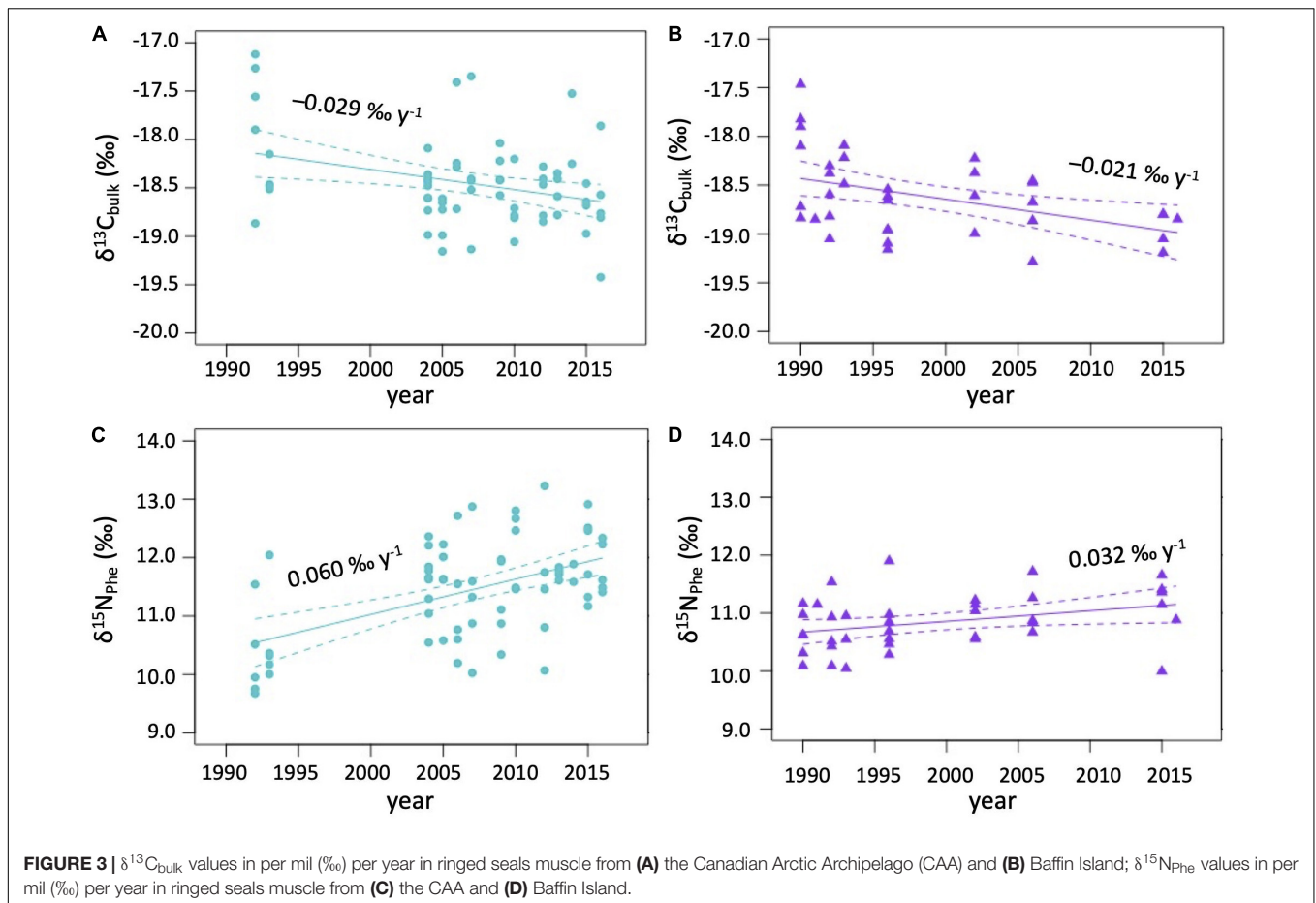
The strong positive correlation between  $\delta^{15}\text{N}_{\text{bulk}}$  and  $\delta^{15}\text{N}_{\text{Phe}}$ , and the lack of correlation between  $\delta^{15}\text{N}_{\text{bulk}}$  and Cor- $\delta^{15}\text{N}_{\text{trophic}}$  (**Figure 5**), demonstrates that temporal changes in  $\delta^{15}\text{N}_{\text{bulk}}$  of



ringed seals from both the CAA and Baffin Island was mainly driven by change in the  $\delta^{15}\text{N}$  at the baseline. The  $\delta^{15}\text{N}$  baseline in the marine environment is represented by the  $\delta^{15}\text{N}$  of nitrate, an essential nutrient taken up by phytoplankton (Mariotti et al., 1981). Phytoplankton take on the isotopic signature of nitrate, to create the particulate organic matter that forms the base of marine food webs. Nitrate is supplied to the Arctic mainly by Pacific water, entering *via* the Bering Strait, and Atlantic water, entering *via* the Barents Sea and Fram Strait (Torres-Valdés et al., 2013), and to a lesser extent *via* rivers (McClelland et al., 2007; Fouest et al., 2013). Water column denitrification in the Pacific Ocean leaves an enriched imprint on  $\delta^{15}\text{N}$  of nitrate, resulting in a 2–3‰  $^{15}\text{N}$ -enrichment end-member in

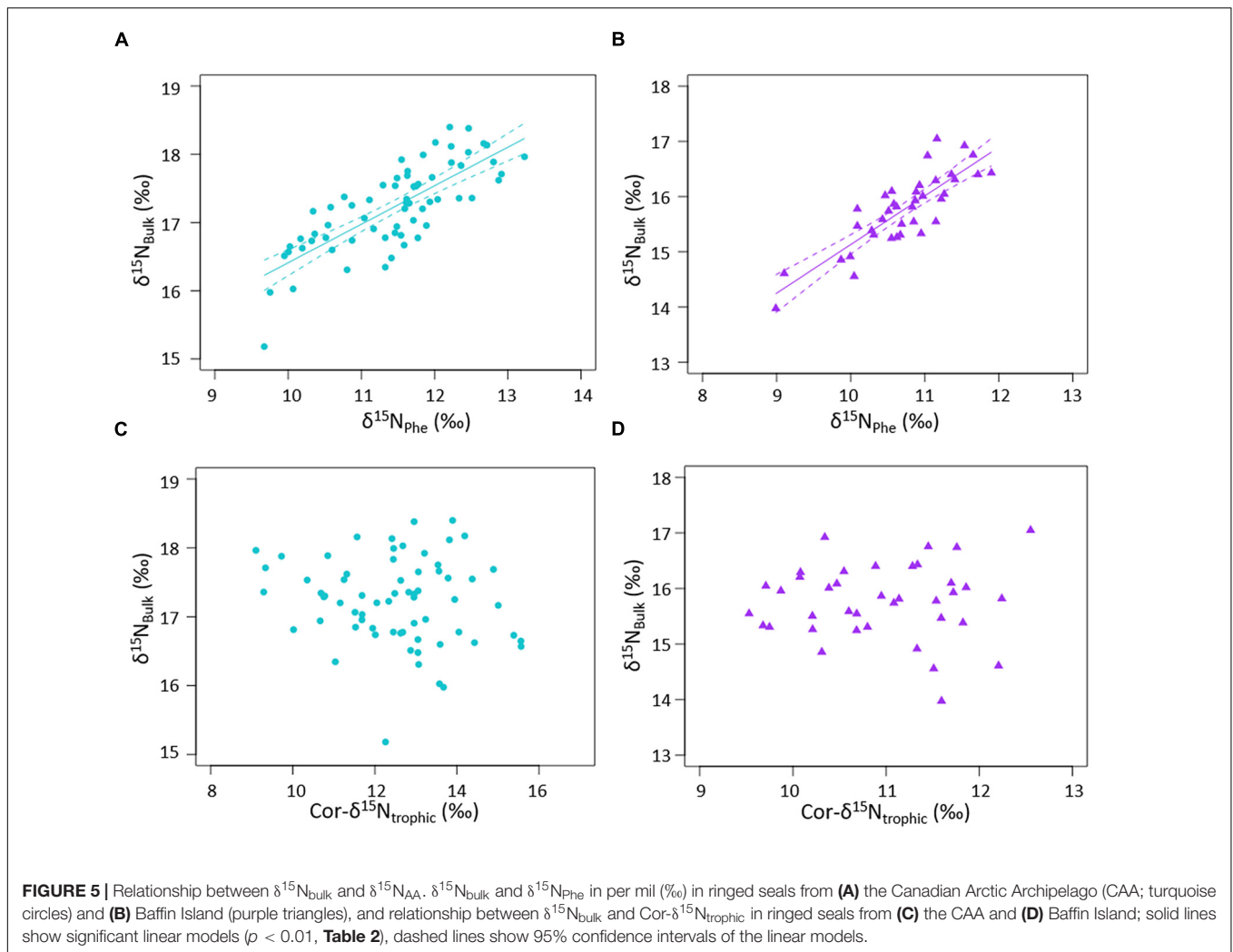
the Pacific Arctic inflow compared to the Atlantic Arctic inflow (Somes et al., 2010). The  $\delta^{15}\text{N}$  of particulate organic matter depends on the  $\delta^{15}\text{N}$  of nitrate, and is also influenced by fractionation during primary production (Sigman et al., 2009) and bacterial activity (Morata et al., 2008).

The increasing trend in  $\delta^{15}\text{N}_{\text{Phe}}$  in ringed seals from both the CAA and Baffin Island (Figures 3C,D) could be explained by several processes. First, the Pacific inflow through the Bering Strait has increased by more than 50% as a result of warming and increased winds (Woodgate, 2018). In conjunction, the Beaufort gyre has expanded in areal extent and sea surface height (Regan et al., 2019), increasing the surface slope between the Canadian Basin and Baffin Bay (Michel et al., 2006). Although the



downstream effects of these changes are not well known, model predictions suggest that the export of water from the Beaufort Gyre through the CAA is increasing and will continue to increase

into the coming decades, influencing the surface waters as far South as the Labrador Sea (McGeehan and Maslowski, 2012). The increasing influence of  $^{15}\text{N}$ -enriched Pacific derived water



in the CAA and Baffin Bay might explain the increasing trend in  $\delta^{15}\text{N}_{\text{Phe}}$  in ringed seals from the CAA and the similar but less pronounced increasing trend in ringed seals from Baffin Island, which could be a result of a dilution effect of the Pacific water influence when reaching Southern Baffin Bay.

Second, benthic denitrification in the western Arctic and Bering Sea has also been shown to impart a  $^{15}\text{N}$ -enriched signature in the overlying water column, known as the “sedimentary isotope effect” (Granger et al., 2011; Fripiat et al., 2018). Previous work has suggested that increasing net primary production on Arctic shelves would increase organic matter supply to sediments, leading to increased benthic denitrification rates (Arrigo and van Dijken, 2015), resulting in an enrichment in  $^{15}\text{N}$  of water column nitrate. The increasing trend in  $\delta^{15}\text{N}_{\text{Phe}}$  in ringed seals from the CAA and Baffin Island could reflect the effect of increasing primary production and in turn benthic denitrification, either occurring locally (i.e., within the CAA and the Baffin Island shelf), or occurring in the Beaufort Sea and the Canadian Basin (Lewis et al., 2020), the resulting  $^{15}\text{N}$ -enriched water column nitrate being subsequently transported *via* the Arctic outflow through the CAA. Finally, increasing bacterial

degradation activity with increasing temperature in the Arctic (Vaqué et al., 2019) may also increase the  $\delta^{15}\text{N}$  values of organic matter entering the food web (Morata et al., 2008).

The difference in the  $\delta^{15}\text{N}$  at the baseline between the early 1990s and the mid-2010 was  $\sim 1.5\text{‰}$  in ringed seals from the CAA. This is close to representing an increase of half a trophic position, as trophic discrimination factors for trophic amino acids in marine tertiary and higher consumers range from 2.5 to 4.3‰ (McMahon and McCarthy, 2016). Not accounting for this change in the  $\delta^{15}\text{N}$  at the baseline would lead to misestimation of ringed seal trophic position using  $\delta^{15}\text{N}_{\text{bulk}}$  alone (de la Vega et al., 2020). This highlights the power of using  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  values to assess temporal trend in trophic position of ringed seals from the Canadian Arctic, especially when the  $\delta^{15}\text{N}$  at the baseline is increasing at varying rates by latitude.

### Temporal Trend in Food Web Structure

In the CAA, the decrease in  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  in ringed seals of  $\sim 1.5\text{‰}$  (Figure 4A) represents a change of approximately half a trophic level (McMahon and McCarthy, 2016). This



could be explained by a change in diet, for example from a more piscivorous to an omnivorous diet consisting of more invertebrates. Conversely, the decrease in the difference between  $\delta^{15}\text{N}_{\text{Glu}} - \delta^{15}\text{N}_{\text{Phe}}$  ( $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$ ) could also suggest a change in the ringed seals diet from a “low quality diet” (e.g., omnivorous diet consisting of more invertebrates) to a “high quality diet” (e.g., more piscivorous diet), as fractionation of individual amino acids is influenced by the degree of similarity in amino acid composition between the prey and consumers (McMahon and McCarthy, 2016). Currently there is no direct documented evidence to support this diet shift as ringed seals collected from the CAA primarily consume Arctic cod (Yurkowski et al., 2016a). A change in the diet of their prey items, resulting in change in amino acid composition of these prey, could also influence the  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  values of ringed seals. For example, the decrease in Arctic sea ice might affect the feeding patterns and prey composition of Arctic cod (Matley et al., 2013) potentially modifying their amino acid composition, but such change is not currently documented.

Alternatively, the decreasing  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  could reflect a change near the base of the food chain which indirectly influences the trophic position of the ringed seals. The copepods *Calanus glacialis* and *Calanus finmarchicus*, and the amphipod, *Themisto libellula*, are key components of the Arctic and sub-Arctic food webs, constituting the main trophic links to Arctic cod, and predators such as seabirds and marine mammals (Dalpadado et al., 2016; Møller and Nielsen, 2020). In the Barents Sea and West Greenland, increasing sea water temperature and decreased sea ice extent, has led to an increase in the relative abundance of the smaller herbivorous Atlantic crustacean species, *C. finmarchicus* compared to the large omnivorous and carnivorous Arctic crustacean species, e.g., *C. glacialis* and *T. libellula* (Dalpadado et al., 2016; Møller and Nielsen, 2020). Warming in the CAA, reflected in the strong decrease in sea ice concentration and increase of OWP (Figure 2), might have led to a decrease in omnivory versus herbivory in the zooplankton communities, resulting in a shortened food chain length (Figure 4B; McMeans et al., 2015) in the high-Arctic, reflected in the gradual decrease in  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  values of ringed seals from the CAA since 1992.

It is likely that such changes in zooplankton dynamics also occurred at lower latitudes. The CAA is typically more ice-covered throughout the year and has exhibited stronger environmental change compared to Southern Baffin Bay (Figure 2), and Hudson Bay since 1990 (Ferguson et al., 2020). The consistency in  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  values of ringed seals from Baffin Island in the last 25 years suggests that the food web structure in Southern Baffin Bay has remained stable since the 1990s. The similar  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  values in ringed seals from the two regions in 2015–2016, in contrast with the  $^{15}\text{N}$ -enriched  $\text{Cor-}\delta^{15}\text{N}_{\text{trophic}}$  values in ringed seals from the CAA compared to those from the Baffin Island in 1992–1993 (Figure 4A) suggests that the food chain length in the CAA became similar to the food chain length in the Baffin Bay in recent years in response to environmental change. This observation supports latitudinal differences in the onset and rate of environmental changes, and

in turn borealization between mid- and high-Arctic, and suggests that the high-Arctic responded more recently than the mid-Arctic. We have not been able to determine the onset and rate of borealization of the food chain in the mid-Arctic in this study, because access to sample archives or data from earlier than the 1990s would be required.

As the Arctic continues to warm, borealization of the Arctic will likely continue, with consequences for ringed seal population dynamics. Firstly, dietary analyses of Arctic predators such as sea birds, beluga whales, and harp seals, have suggested a recent increase of capelin in Southern Baffin Bay, coinciding with a decreased availability of Arctic cod (Gaston et al., 2012; Yurkowski et al., 2018). A shift in prey fish composition from Arctic cod to capelin is unlikely to reduce the food quality of the prey available to marine predators (Pedro et al., 2019). However, ringed seals from the Canadian Arctic have continued to feed primarily on Arctic cod despite high availability of capelin (Ogloff et al., 2019). In the Northern Barents Sea, ringed seals have also maintained their dietary preferences for arctic prey species in a restricted foraging habitat at an increased energetic expense (Hamilton et al., 2019; Bengtsson et al., 2020). Considering that Arctic prey might become less abundant, the selective pressure on ringed seals to modify or adapt their foraging strategies to consume more boreal species might increase, leading to higher energetical stress. Second, species shift in *Calanus* communities, which have recently become dominated by less lipid-rich smaller individuals, decreases the availability of energy rich lipids for higher trophic level species leading to negative impacts on their body condition (Møller and Nielsen, 2020), with consequences for reproduction (Ferguson et al., 2017). This is in accordance with the decreasing blubber thickness of ringed seals from the CAA since 1990 (Harwood et al., 2020), and coincides with our finding of a decreasing trophic position over time. Ferguson et al. (2020) forecasted reproductive difficulties and higher mortality for ringed seals in the low Arctic, which showed greater seasonal fluctuation in body condition compared to ringed seals from the high-Arctic. This could be amplified by the longer exposure of ringed seals to the less lipid-rich food web in the mid-Arctic compared to ringed seals in the CAA. Management and implementation of protection measures in the Arctic should take into account the stage and state of adaptation of each ecosystem, as we observe and suggest here, that the high-Arctic has become increasingly “borealized” during the last two decades, whilst the food chain in the mid-Arctic remained stable for the last 25 years of monitoring.

## DATA AVAILABILITY STATEMENT

The datasets presented in this study can be found in online repositories. The names of the repository/repositories and accession number(s) can be found below: De la Vega, C., Kershaw, J., Yurkowski, D. J., Ferguson, S. H., Stenson, G. B., Haug, T., Biuw, M., Frie, A. K., Smout, S. C., Jeffreys, R. M., and Mahaffey, C. (2021). ARISE project–Work package 3: stable nitrogen isotopes of bulk tissue and amino-acids of ringed seals muscle and teeth’s growth layer groups of harp seals from the

Arctic and sub-Arctic (Version 1.0) (Data set). NERC EDS UK Polar Data Centre. <https://doi.org/10.5285/6AAA53E8-3D0A-48FE-838E-31C5B5F24CE7>.

## ETHICS STATEMENT

Ethical review and approval was not required for the animal study because all tissue samples were acquired through Fisheries and Oceans Canada Licences to Fish for Scientific Purposes.

## AUTHOR CONTRIBUTIONS

CV, CM, and RJ designed the study. DY and SF collected the samples. CV, LN, and ES analyzed the samples. CV performed the data and statistical analyses and wrote the first draft of the manuscript. CM, DY, SS, SF, and RJ contributed substantially to the revisions. All authors contributed to the article and approved the submitted version.

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## SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2021.700687/full#supplementary-material>

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