



Simulating Effects of Variable Stoichiometry and Temperature on Mixotrophy in the Harmful Dinoflagellate *Karlodinium veneficum*

Chih-Hsien Lin¹, Kevin J. Flynn^{2*}, Aditee Mitra² and Patricia M. Glibert¹

¹ Horn Point Laboratory, University of Maryland Center for Environmental Science, Cambridge, MD, United States,

² Department of Biosciences, Wallace Building, Swansea University, Swansea, United Kingdom

OPEN ACCESS

Edited by:

Matthew D. Johnson,
Woods Hole Oceanographic
Institution, United States

Reviewed by:

Lasse Tor Nielsen,
Technical University of Denmark,
Denmark
Shovonlal Roy,
University of Reading,
United Kingdom

*Correspondence:

Kevin J. Flynn
k.j.flynn@swansea.ac.uk

Specialty section:

This article was submitted to
Marine Ecosystem Ecology,
a section of the journal
Frontiers in Marine Science

Received: 28 March 2018

Accepted: 22 August 2018

Published: 10 September 2018

Citation:

Lin C-H, Flynn KJ, Mitra A and
Glibert PM (2018) Simulating Effects
of Variable Stoichiometry
and Temperature on Mixotrophy
in the Harmful Dinoflagellate
Karlodinium veneficum.
Front. Mar. Sci. 5:320.
doi: 10.3389/fmars.2018.00320

Results from a dynamic mathematical model are presented simulating the growth of the harmful algal bloom (HAB) mixotrophic dinoflagellate *Karlodinium veneficum* and its algal prey, *Rhodomonas salina*. The model describes carbon-nitrogen-phosphorus-based interactions within the mixotroph, interlinking autotrophic and phagotrophic nutrition. The model was tuned to experimental data from these species grown under autotrophic conditions and in mixed batch cultures in which nitrogen:phosphorus stoichiometry (input molar N:P of 4, 16, and 32) of both predator and prey varied. A good fit was attained to all experimentally derived carbon biomass data. The potential effects of temperature and nutrient changes on promoting growth of prey and thus *K. veneficum* bloom formation were explored using this simulation platform. The simulated biomass of *K. veneficum* was highest when they were functioning as mixotrophs and when they consumed prey under elevated N:P conditions. The scenarios under low N:P responded differently, with simulations showing larger deviation between mixotrophic and autotrophic growth, depending on temperature. When inorganic nutrients were in balanced proportions, lower biomass of the mixotroph was attained at all temperatures in the simulations, suggesting that natural systems might be more resilient against *Karlodinium* HAB development in warming conditions if nutrients were available in balanced proportions. These simulations underscore the need for models of HAB dynamics to include consideration of prey; modeling HAB as autotrophs is insufficient. The simulations also imply that warmer, wetter springs that may bring more N with lower N:P, such as predicted under climate change scenarios for Chesapeake Bay, may be more conducive to development of these HABs. Prey availability may also increase with temperature due to differential growth temperature responses of *K. veneficum* and its prey.

Keywords: mixotrophy, *Karlodinium veneficum*, *Rhodomonas salina*, harmful algal blooms, mathematical model, stoichiometry, temperature responses

INTRODUCTION

In conjunction with the growing recognition that harmful algal blooms (HABs) are promoted by increasing nutrient loads to marine and freshwaters (e.g., Anderson et al., 2002; Glibert et al., 2005; Heisler et al., 2008; Glibert and Burford, 2017), there is also an enhanced appreciation for the importance of mixotrophy in the nutrition of many HAB taxa (Jeong et al., 2005a,b; Burkholder et al., 2008; Flynn et al., 2013; Stoecker et al., 2017). The complexities of understanding the dynamics of HABs dominated by mixotrophs compound the already difficult study of system regulation by simple autotrophic physiology (Flynn, 2009; Mitra and Flynn, 2010; Ghyoot et al., 2017; Flynn and McGillicuddy, 2018; Glibert et al., 2018). For instance, some mixotroph species are primarily autotrophic but ingest prey under light limitation or conditions of nutrient limitation or imbalance, while some others appear to be primarily heterotrophic but photosynthesize under certain conditions by retaining chloroplasts from their prey (Stoecker, 1998). The challenge in understanding HAB dynamics where mixotrophy is an important nutritional mode is thus further complicated by the need to not only understand the ecophysiology of the HAB species, but also that of its prey species. Additionally, mixotrophy may also be related to toxin production in some HABs (e.g., Blossom et al., 2012).

Mixotrophy in protists is not a simple additive process of autotrophy plus phagotrophy, but rather a complex integration of physiological interactions (Flynn and Mitra, 2009; Mitra and Flynn, 2010). Only a few physiological experiments (Lundgren et al., 2016; Lin et al., 2017)—and even fewer model constructs—consider the feedback function of rates of change during mixotrophic feeding, the nutritional status of both predator and prey, and linkages to nutrient physiological interactions (Mitra and Flynn, 2006; Flynn, 2010). At present, predictive capabilities that include the role of mixotrophy in bloom formation are just beginning to be developed (Flynn, 2005, 2010; Glibert et al., 2010; Mitra et al., 2014, 2016; Flynn and McGillicuddy, 2018; Flynn et al., 2018).

There is also a growing appreciation that changes in climate may expand the potential niches for some harmful or toxic algal blooms (Hallegraeff, 2010; Fu et al., 2012; Wells et al., 2015; Glibert and Burkholder, 2018). The most direct effects of climate change are those associated with rising temperatures. The frequency and severity of blooms may be exacerbated due to temperature-driven competitive advantages for HAB species over non-HAB species (Hallegraeff, 2010) and other HAB-favorable conditions may expand, such as increased stratification or altered precipitation patterns that affect the timing of freshwater and associated nutrient delivery (e.g., Heisler et al., 2008; Moore et al., 2015; Glibert and Burkholder, 2018). The mixotrophic chrysophyte *Ochromonas* sp., for example, has been found to become more heterotrophic with increased temperature (Wilken et al., 2013). However, other climate changes events such as ocean acidification with (de)eutrophication may also affect the consortium of organisms that co-occur with the HAB species, and which can be food sources for these mixotrophs (e.g., Fu et al., 2012; Flynn et al., 2015; Wells et al., 2015; Glibert et al., 2018). There is evidence that ingestion, growth rates and

cell volume of the heterotrophic dinoflagellate *Oxyrrhis marina* respond differently to temperature-prey interactions, indicating complex and non-linear predator-prey dynamics with increasing temperatures (Montagnes et al., 2003; Kimmance et al., 2006). Many important questions related to the interactive effects of temperature changes and mixotrophy on HAB dynamics clearly remain.

The mixotrophic dinoflagellate *Karlodinium veneficum* (formerly *Gymnodinium galatheanum* and *K. micrum*) is a common toxigenic species that can produce a suite of unique polyketide compounds, karlotoxins (Van Dolah, 2000; Kempton et al., 2002). This species is a constitutive mixotroph (Mitra et al., 2016), possessing the ability to make its own chloroplasts, and is capable of forming blooms of up to 10^7 – 10^8 cells L^{-1} (e.g., Adolf et al., 2008) that have been associated with fish and shellfish mortality, both in natural waters and aquaculture farms worldwide (Braarud, 1957; Nielsen, 1993; Glibert and Terlizzi, 1999; Deeds et al., 2002; Stoecker et al., 2008; Zhou et al., 2011). Blooms of *K. veneficum* appear to be increasing in size and frequency of occurrence in estuaries such as Chesapeake Bay, MD, United States (Li et al., 2015) and elsewhere worldwide (Place et al., 2012 and references therein; Dai et al., 2013; Adolf et al., 2015). This mixotroph is also a species for which there are considerable physiological data related to mixotrophy (e.g., Li et al., 1999, 2000; Place et al., 2012; Lin et al., 2017). Given its prevalence and potential threats to natural resources around the world, improved forecasting and predictive ability of this HAB taxon would be an aid to managers.

Here, applying both previously published and newly acquired experimental data, a model was developed to improve our understanding of the interactions of the growth of this mixotroph and its common prey, *Rhodomonas salina*, under varying nutrient and temperature conditions. The resultant simulations were used to address the hypothesis that growth of this mixotroph may increase due to the combination of increased nutrient concentrations, altered nutrient ratios, and raised temperature, combinations of conditions that may be expected under future climate conditions in eutrophic estuaries. While such responses may seem intuitive, the extent to which changes in nutrient (in form and proportion) together with temperature alter growth of the mixotroph and its prey are not well resolved and therefore simulations enable such interactions to be explored for multiple temperature and nutrient conditions. Results from using such models may help inform nutrient management plans under future climate conditions.

MATERIALS AND METHODS

Overall Approach

A mechanistic model was developed based on the framework of an existing variable stoichiometric, photo-acclimative mixotrophy model, namely the “perfect beast” construct of Flynn and Mitra (2009). This describes carbon-nitrogen-phosphorus (C-N-P)-based interactions within a mixotroph cell, and builds upon the variable stoichiometric zooplankton model of Mitra (2006) and the phytoplankton model of Flynn (2001). To

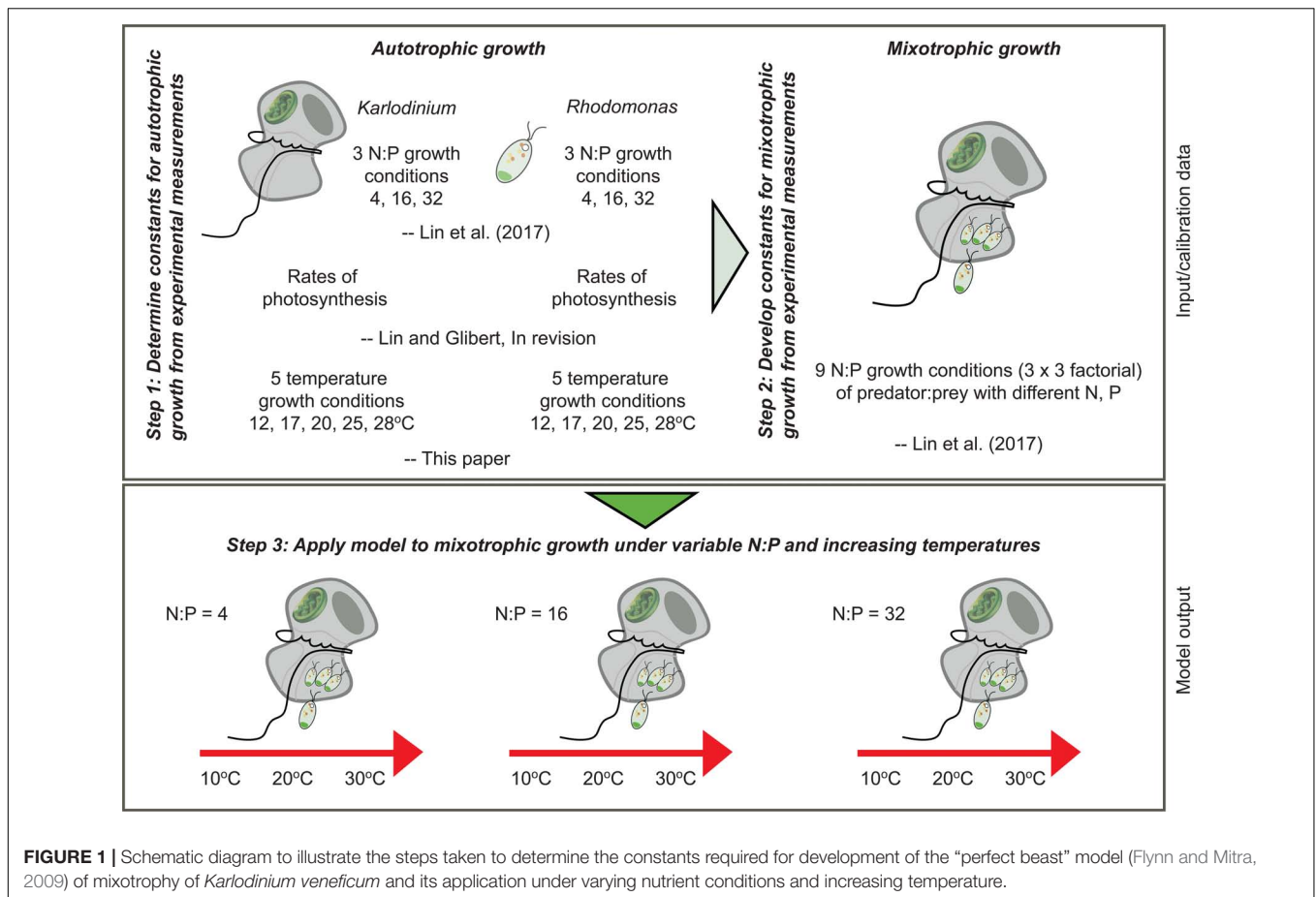


FIGURE 1 | Schematic diagram to illustrate the steps taken to determine the constants required for development of the “perfect beast” model (Flynn and Mitra, 2009) of mixotrophy of *Karlodinium veneficum* and its application under varying nutrient conditions and increasing temperature.

parameterize the model, previously published data as well as new data on temperature responses of both the mixotroph and prey were applied (Figure 1).

In brief, the “perfect beast” model has eight state variables (Figure 2) describing C, N, and P and chlorophyll (Chl) associated with the core mixotroph (m) biomass (mC, NC, PC, and ChlC) and also the same constituents associated with the contents of the food (F) vacuole (namely, FC, FNC, FPC, and FChlC) after the m have fed on algal prey. The amount of material associated with the F vacuole is relative to the core mC biomass. Thus, the total C associated with the m is $mC \cdot (1 + FC)$ with the unit of $gC L^{-1}$. Here the m model was configured to be consistent with the status of *K. veneficum* as a constitutive m, with its own photoacclimative description of Chl:C. Prey were described using the variable stoichiometric photoacclimative phytoplankton model of Flynn (2001), as deployed in Flynn and Mitra (2009). The full model accounts for predator stoichiometry, prey stoichiometry (i.e., F quality) and their feedback interactions (Figure 2). The model operates using ordinary differential equations (ODEs) using an Euler integration routine with a timestep of 0.0039 d (5.625 min). Equations are provided in the associated **Supplementary Table S1**; these are given in a linear form to aid reconstruction in the modeling platform of choice.

The model was built, and simulations run, using the Powersim Constructor platform, with tuning (calibration) to experimental

data performed using the evolutionary algorithm supported by Powersim Solver v2 (Isdalstø, Norway). This algorithm maximizes the likelihood of resolving a global, rather than a local, minimum, and produces the fit closest to the presented data (Haefner, 2005; Flynn, 2018). Most of the constants within the model are not tuned (Tables 1, 2); they are used to modulate physiological feedback processes and the model is not sensitive to their precise value (see source papers for further details).

To configure the full model describing both the mixotroph and its prey, the constants that constrain the autotrophic physiology of predator and prey were first determined from experimental data (Figure 1). Once rates of photosynthesis and inorganic nutrient uptake for these species were calculated for varying nutrients and temperatures, the parameters that control mixotrophic performance of the predator were then ascertained (again through reference to experimental data), for conditions in which predator and prey were both grown under varying nutrient stoichiometry. Finally, the tuned model was run to simulate (predict) growth of *K. veneficum* and its prey under variable N:P and temperature conditions.

Data Sources, Experimental Conditions and Model Parameters

Previously available experimental data were first exploited. These included rates of autotrophic growth of both *K. veneficum* and

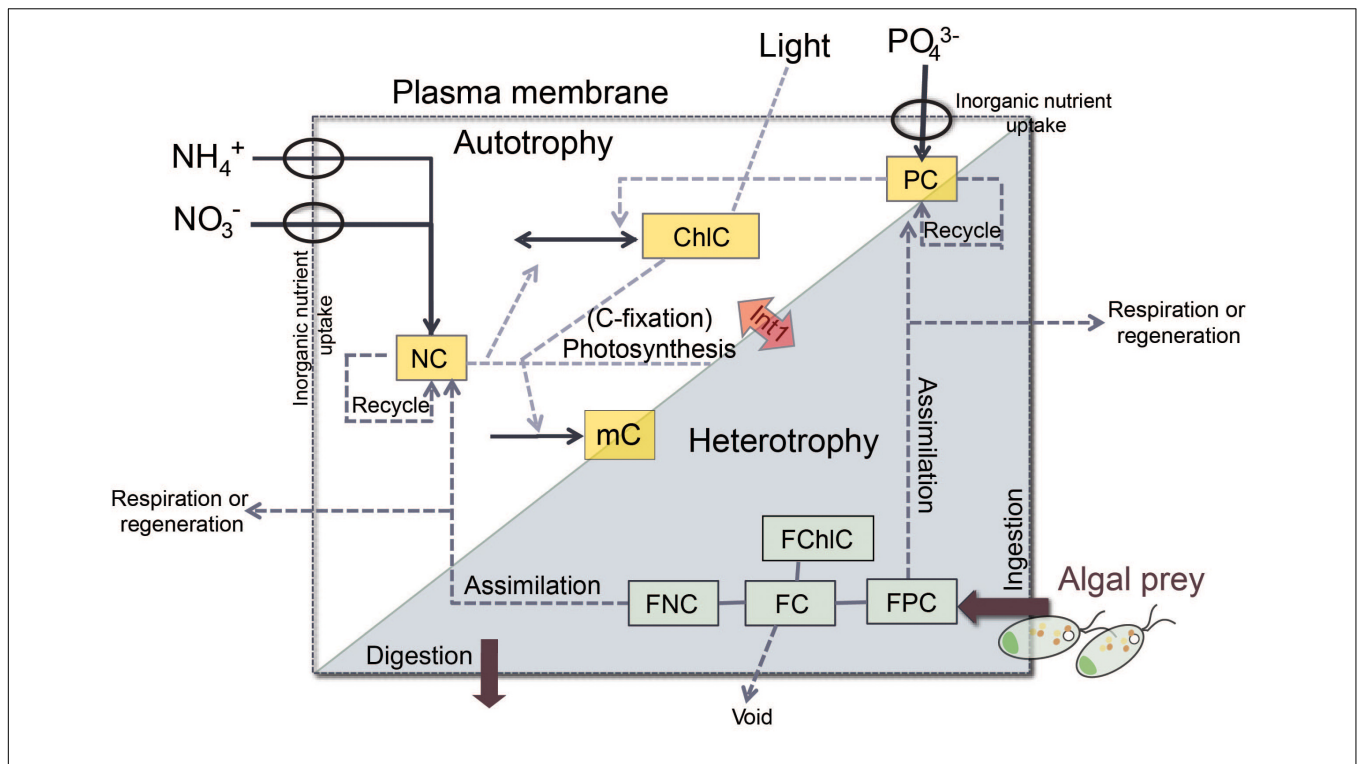


FIGURE 2 | Schematic of the structure of the “perfect beast” model, showing major flows in and out of state variables (solid arrows and boxes) from the external parameters (NO_3^- , PO_4^{3-} , and Light), and the major feedback processes (dashed arrows). Autotrophic growth uses inorganic nutrients and light via the photosystems of the mixotroph (phototrophy; white part). A proportion of activity leading to growth is required to support synthesis of those photosystems. Predation brings algal prey into the food vacuole within the confines of the mixotroph cell (heterotrophy; gray part). Interactions between phototrophic and heterotrophic nutrition (Int1) influence the growth of the mixotrophy (Flynn and Mitra, 2009). The state variables (yellow boxes) that describe carbon (C), nitrogen (N), phosphorus (P), and chlorophyll (Chl) associated with core mixotroph biomass are mC (C-biomass of the mixotroph), ChlC (chlorophyll C quota), NC (cellular NC quota), and PC (cellular PC quota), while the same constituents (green boxes) associated with the content of food vacuole are FC (food vacuole C content relative to mC), FChlC (food vacuole Chl content relative to mC), FNC (food vacuole N content relative to mC), and FPC (food vacuole P content relative to mC).

R. salina under varied N:P stoichiometry (molar N:P of 4, 16, and 32) in exponential growth phase (Lin et al., 2017; **Figure 1**); these experimental data are provided in the **Supplementary Table S2**). In order to establish the variable N:P conditions in those experiments, NO_3^- concentrations were held constant at $\sim 1232 \mu\text{g N L}^{-1}$, while initial PO_4^{3-} concentrations were varied. The constants that describe the autotrophic physiology of prey

(e.g., half saturation constants for N and P uptake, aK_{Ni} , aK_{P} , and of growth, $a\mu_{\text{max}}^{\text{phot}}$) and m (e.g., mK_{Ni} , and mK_{P}) were obtained from model tuning based on the change in residual NO_3^- and PO_4^{3-} concentrations (nutrient kinetics), and Chl and C biomass (see below) during culture growth (**Table 1**). From a second set of experiments, based on monocultures of *K. veneficum* and *R. salina*, growing autotrophically, initial

TABLE 1 | Autotrophic state constants that were calculated and gained from tuning against changes in experimental monoculture cultures of *Rhodomonas salina* and *Karlodinium veneficum*.

Parameters	Units	<i>Rhodomonas salina</i>		<i>Karlodinium veneficum</i>		Sources
		Abbr.	Values	Abbr.	Values	
Half saturation for NO_3^- transport	$\mu\text{g N L}^{-1}$	aK_{Ni}	57.437	mK_{Ni}	14.628	Tuned herein
Half saturation for PO_4^{3-} transport	$\mu\text{g P L}^{-1}$	aK_{P}	1.550	mK_{P}	118.490	Tuned herein
Chl-specific initial slope to PI curve (α)	$(\text{m}^2\text{g}^{-1} \text{Chl } a) (\text{mgC micromol photon}^{-1})$	$a\alpha^{\text{Chl}}$	0.192	$m\alpha^{\text{Chl}}$	0.011	Lin and Gilbert, unpublished
Maximum N:C	gN gC^{-1}	$a\text{NC}_{\text{max}}$	0.200	$m\text{NC}_{\text{max}}$	0.200	Calculated herein
Minimum N:C	gN gC^{-1}	$a\text{NC}_{\text{min}}$	0.050	$m\text{NC}_{\text{min}}$	0.050	Calculated herein
Maximum P:C	gP gC^{-1}	$a\text{PC}_{\text{max}}$	0.020	$m\text{PC}_{\text{max}}$	0.020	Calculated herein
Minimum P:C	gP gC^{-1}	$a\text{PC}_{\text{min}}$	0.005	$m\text{PC}_{\text{min}}$	0.001	Calculated herein
Maximum rate of phototrophic growth	d^{-1}	$a\mu_{\text{max}}^{\text{phot}}$	1.280	$m\mu_{\text{max}}^{\text{phot}}$	ND	Tuned herein

Autotrophic growth rate of K. veneficum was not tuned here (ND; not determined, but see Table 2).

TABLE 2 | Constants obtained from tuning the “perfect beast” model of Flynn and Mitra (2009) against experimentally derived changes in carbon biomass in mixed cultures of *Karlodinium veneficum* (mixotroph) with *Rhodomonas salina* (prey) when each was grown separately in different N:P condition (low NP = 4, Redfield = 16, and high N:P = 32 on a molar basis) and also combined in a total of 9 combinations.

Parameters	Scenario		
	Low-N:P prey	Redfield-N:P prey	High-N:P prey
Low-N:P <i>K. veneficum</i>			
<i>mAE_{min}</i>	0.658	0.840	0.814
<i>mcap_a</i>	0.050	0.050	0.050
<i>mK_{as}</i>	0.999	0.835	0.647
<i>mK_{Ing}</i>	0.490	0.295	0.204
<i>mμ_{max}^{phot}</i>	0.200	0.400	0.400
<i>mμ_{max}^{het}</i>	0.450	0.477	0.504
Redfield-N:P <i>K. veneficum</i>			
<i>mAE_{min}</i>	0.743	0.370	0.736
<i>mcap_a</i>	0.050	0.050	0.050
<i>mK_{as}</i>	0.010	0.840	0.997
<i>mK_{Ing}</i>	0.086	0.551	0.464
<i>mμ_{max}^{phot}</i>	0.200	0.400	0.400
<i>mμ_{max}^{het}</i>	0.451	0.887	0.842
High-N:P <i>K. veneficum</i>			
<i>mAE_{min}</i>	0.537	0.821	0.832
<i>mcap_a</i>	0.163	0.050	0.050
<i>mK_{as}</i>	0.363	0.450	0.720
<i>mK_{Ing}</i>	0.010	0.296	0.309
<i>mμ_{max}^{phot}</i>	0.200	0.400	0.400
<i>mμ_{max}^{het}</i>	0.545	0.501	0.759

mAE_{min}, minimum assimilation efficiency; *mcap_a*, the likelihood of ingestion following encounter; *mK_{as}*, mixotroph half saturation for digestion rate; *mK_{Ing}*, mixotroph half saturation for ingestion; *mμ_{max}^{phot}*: mixotroph maximum rate of phototrophic growth; *mμ_{max}^{het}*, mixotroph maximum rate of heterotrophic growth.

slopes of photosynthesis-irradiance (PI) curves were calculated (Lin and Glibert, unpublished). Additional physiological data (Supplementary Table S1) for parameterizing the autotrophic component of the model were obtained from Flynn and Mitra (2009). The C biomass of *K. veneficum* was estimated based on a cellular C-volume relationship for dinoflagellates (Menden-Deuer and Lessard, 2000) and a conversion factor of 0.2 pg C μm⁻³ from volume to C for *R. salina* was applied (Jakobsen and Hansen, 1997). Cell size of predator and prey were recorded in parallel with the cell densities in the study of Lin et al. (2017); cellular volume (CV) was estimated from these data using the following equation:

$$CV = 0.1875WL^2$$

where W and L are the width and length of cells.

To parameterize temperature responses, new experimental data were also obtained on rates of autotrophic growth of *K. veneficum* and *R. salina* across a temperature gradient (Figure 1). The same strains of *K. veneficum* and *R. salina* used in Lin et al. (2017) were inoculated separately into *f*/2 media (Guillard, 1975) and maintained at 12, 17, 20, 25,

28°C under irradiance of 430 micromol photons m⁻² s⁻¹ in a 12 h light:12 h dark cycle in batch cultures. The strains were acclimated for 2 weeks to the experimental conditions, after which growth was monitored over 96 h. Aliquots (2 mL) were collected for cell enumeration at 0, 24, 48, 72, and 96 h from each flask and were preserved in paraformaldehyde (final concentrations of 1% v/v) at 4°C for later cell enumeration. The cells were identified and gated based on size, shapes, and auto-fluorescence using a BD Accuri C6 flow cytometry. Then, cell-specific growth rates of predator and prey were determined separately based on the rates of changes in the slopes of the regression of natural log-transformed cell-densities change over 96 h.

After the mixotroph model was calibrated, and after autotrophic temperature responses of predator and prey were experimentally determined (and assumed to remain the same under mixotrophic conditions), the model was used to simulate 10-day growth responses of the mixotroph and its prey under 3 nutrient conditions (N:P = 4, N:P = 16, and N:P = 32 for both predator and prey) under varying temperatures (Figure 1). Rates of maximum growth of *K. veneficum* are reported in autotrophic and in mixotrophic nutritional modes.

Statistical Analyses

All statistical analyses were performed with R. The Shapiro-Wilk test was used to verify normality of the experimental data while the Levenes' test was used to assess the homogeneity of variance. Cell-specific growth rates of *K. veneficum* and *R. salina* were compared for statistical differences in slopes of regression of natural log-transformed data under each temperature conditions (ANCOVA test). Two-way analysis of variance was applied to test for the interactive effects between temperature and species. Regressions were considered significant at $p < 0.05$ with the adjusted r^2 value.

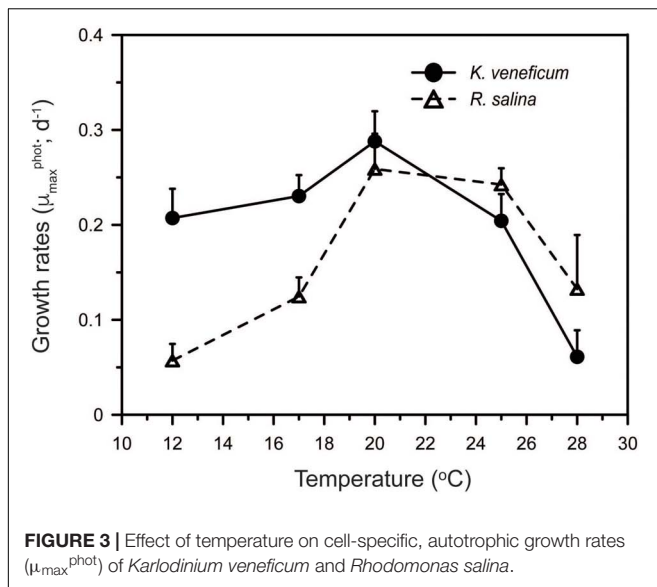
RESULTS

Temperature Effects on Growth

The responses to temperature of the mixotroph and its prey differed. Autotrophic growth rates of *K. veneficum* ranged between 0.06 and 0.29 d⁻¹ and increased with increasing temperature up to 20°C above which growth rates fell sharply (Figure 3). The growth rates of *R. salina* had a similar range as those in the predator, from 0.06 to 0.26 d⁻¹, but the prey grew significantly faster than its predator at temperatures > 20°C (ANCOVA, $p < 0.001$; Figure 3). Variations between temperature responses of the maximal growth rates of *K. veneficum* and *R. salina* were statistically significant (two-way ANOVA: F -value = 2.88, $p = 0.011$; Figure 3).

Model Tuning to Experimental Data Sets

Model tuning was undertaken in 2 steps. Half saturation values (i.e., mK_{Ni} , mK_P , aK_{Ni} , and aK_P) were calculated based on nutrient depletion to determine transport of nutrients in relation to the cell quotas, and photosynthetic rates were used to determine cell C when cells were in autotrophic growth.



Maximum and minimum ratios of N:C and P:C define the nutrient status of the cells (Table 1). These parameters were then applied to mixotrophic growth.

The mixotroph model was successfully calibrated against previously available data on the growth of the mixotroph and its prey as a function of variable nutrient stoichiometry (Figure 4). Between 67 and 97% of the variations in the nine experimental data sets could be explained by the simulations for biomass of the mixotroph, with no significant difference between the observed and predicted data ($p < 0.05$).

The parameters that control mixotrophic growth showed variability with the nutritional status (C:N:P) of *K. veneficum* and its prey (Table 2). The most sensitive parameters were assimilation efficiency (mAE_{min}) and half saturation constant for ingestion (mK_{Ing}) and maximum growth rate of heterotrophic growth ($m\mu_{\max}^{\text{het}}$; Table 2). For example, the minimum assimilation efficiency (mAE_{min}) of C in the mixotroph from the ingested prey ranged from 0.370 under conditions in which both the mixotroph and prey were grown with nutrients supplied in Redfield conditions (N:P = 16), to 0.840 when the prey was under Redfield growth conditions but *K. veneficum* was grown under low N:P conditions. Thus, there was significantly higher mAE_{min} when the mixotroph was initially under low N:P conditions than when it was under Redfield N:P conditions and given the same quality prey. In addition, the half saturation for ingestion (mK_{Ing}) for *K. veneficum* grown under low N:P conditions was significantly lower when they were mixed with the high N:P prey than when *K. veneficum* in the same nutrient state was given prey grown under Redfield N:P and low N:P conditions (0.204 vs. 0.295 and 0.490, respectively). For those *K. veneficum* grown under high N:P conditions, their mK_{Ing} ranged from 0.010 to 0.309 with the lowest value corresponding to prey grown under low N:P. Maximum growth rate of *K. veneficum* as a photo-heterotroph ($m\mu_{\max}^{\text{het}}$) was consistently higher than maximum growth rate of *K. veneficum* as an autotroph (μ_{\max}^{phot}) by factors of 1.19 (low N:P *K. veneficum* with

Redfield N:P prey) to 2.72 (high N:P *K. veneficum* with low N:P prey; Table 2).

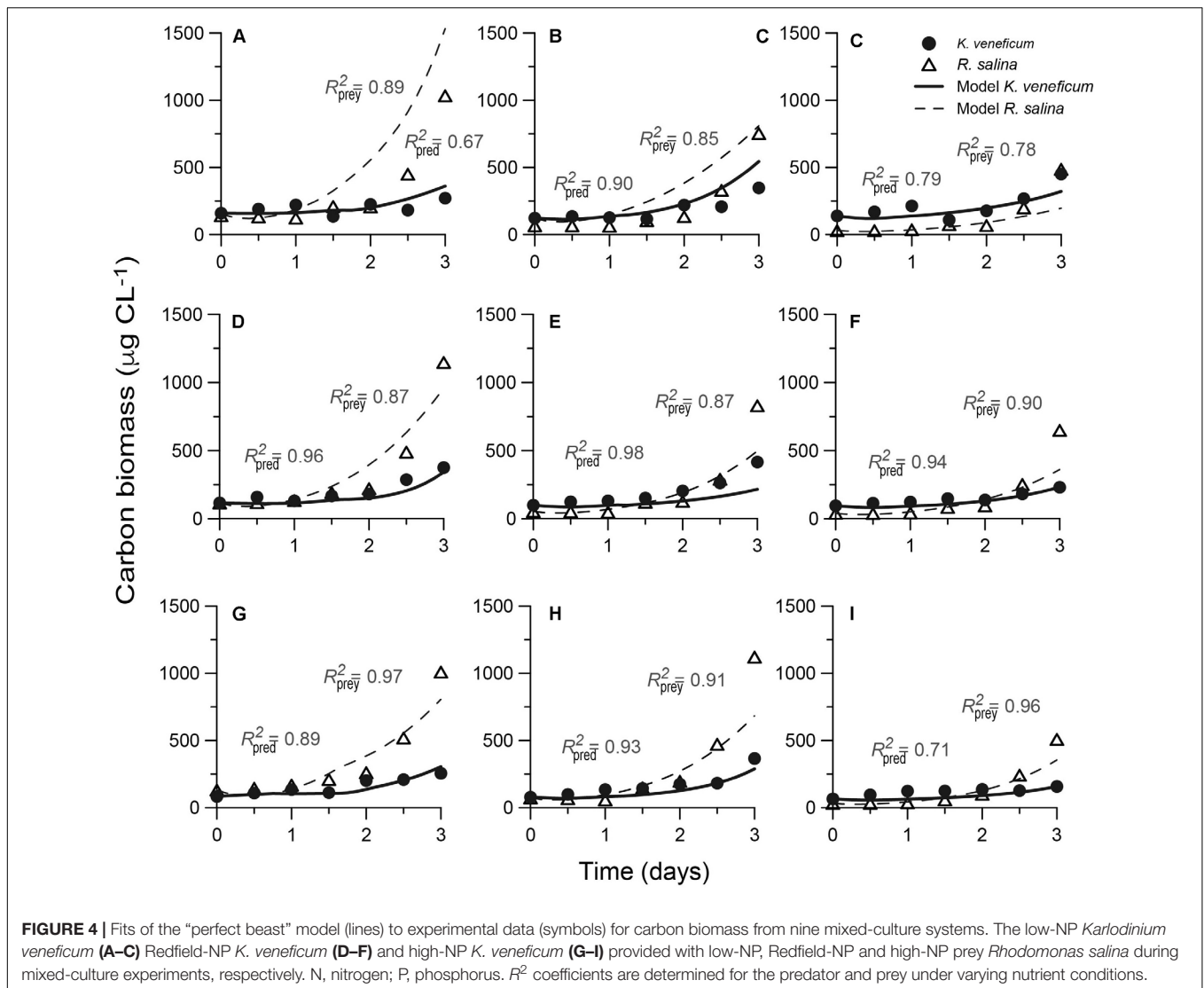
Simulating Growth Under Variable Stoichiometry and Temperature

Using the tuned mixotroph-prey models, and having established individual temperature responses of *K. veneficum* and *R. salina*, scenarios were developed to estimate growth of both species under variable stoichiometry and temperature conditions (Figure 1). For *K. veneficum*, growth as an autotroph and as a mixotroph were compared, and for *R. salina*, growth with and without the predator were estimated. Three stoichiometric conditions were simulated, holding both species in the same N:P condition for each scenario (N:P = 4, 16, 32).

In the low N:P scenario, a significantly higher C biomass of *K. veneficum* in mixotrophic growth was attained in simulations with increased temperature ($\geq 25^\circ\text{C}$) compared with that under comparable autotrophic conditions (ANCOVA, $p < 0.001$; Figures 5A,D). The highest mixotrophic growth rate, 0.30 d^{-1} , was attained at 25°C , which was 1.5-fold higher than the simulation without prey at this temperature. In the Redfield N:P scenarios, there were no significant differences between autotrophic and mixotrophic growth rates of *K. veneficum* (ANCOVA, $p = 0.161$) and relatively low overall C biomass of the mixotroph was attained in the 10-day simulation (Figures 5B,E). Differences between mixotrophic and autotrophic growth of *K. veneficum* under high N:P conditions considered a distinct trend toward statistical significance (ANCOVA, $p = 0.072$). The growth patterns of *K. veneficum* in the two nutritional modes were very similar, showing increases in biomass reaching the maximum value of $\sim 2200\ \mu\text{gC L}^{-1}$ at 20°C under mixotrophic conditions, but a lower growth rate and corresponding lower C biomass accumulation at the highest temperature (Figures 5C,F).

The accumulation of C biomass and growth rates of *R. salina* in the presence and absence of the mixotroph were also estimated under variable nutrient and temperature conditions (Figure 6). In the presence of the mixotroph, prey biomass in all N:P conditions declined, but the patterns of decline varied depending on the nutrient conditions (Figures 6A-C). Under low N:P conditions, prey biomass gradually declined to zero within the 10-day simulation. Under Redfield conditions, prey remained detectable, but low, throughout this period. In the highest N:P simulation, prey biomass declined most quickly, to a near-zero biomass within 4 days. The patterns of the changes in prey biomass without predator were comparable among the three nutrient conditions, but higher biomass values were usually attained in the N-rich conditions at the near-highest temperatures (Figure 6F).

Changes in cellular N:P of *K. veneficum* with time was also explored in the model output in autotrophic and mixotrophic growth to determine the extent to which the mixotroph was using inorganic nutrients under the different stoichiometric and temperature conditions. The cellular elemental ratios of *K. veneficum* under low N:P and Redfield N:P growth conditions varied considerably in the first 2 days of simulated growth, then converged at a value of ~ 7 , but those of *K. veneficum* grown

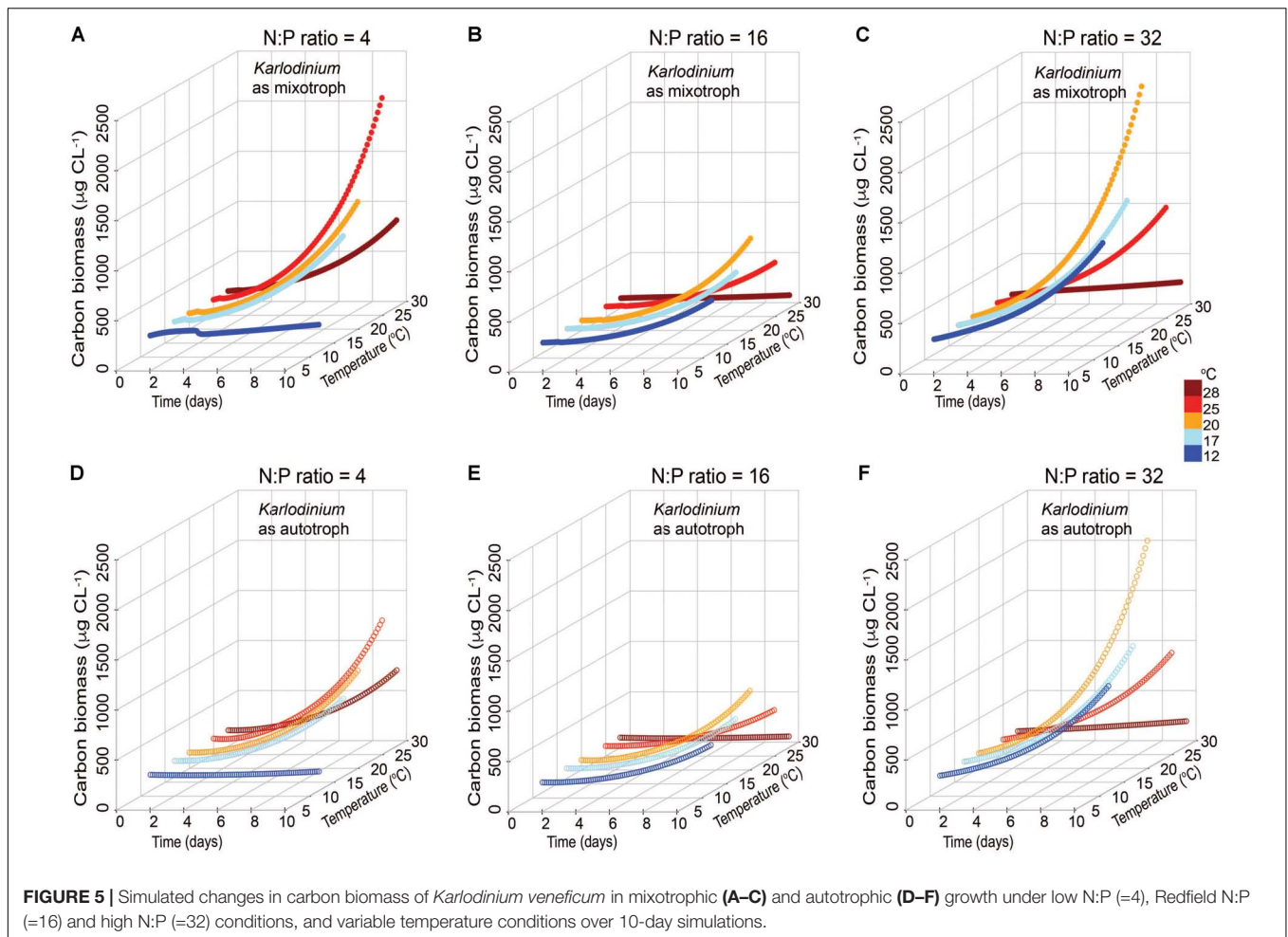


in high N:P conditions converged on a value of ~ 5 within 2 days under all temperatures conditions (Figure 7). Thus, in the simulations, under the condition of excess N (high N:P), *K. veneficum* appeared to become increasingly enriched with internal P through mixotrophy.

DISCUSSION

Despite the increasing recognition of the importance of mixotrophy in planktonic communities, especially HABs (Jeong et al., 2005a,b; Burkholder et al., 2008; Flynn et al., 2013; Stoecker et al., 2017), modeling of plankton dynamics that incorporates mixotrophy is in its infancy (but see Thingstad et al., 1996; Stickney et al., 2000; Ward et al., 2011; Våge et al., 2013; Mitra et al., 2014; Berge et al., 2017; Ghyyot et al., 2017). For simplicity, most model approaches have assumed independence between phototrophic and phagotrophic regulations (e.g., Thingstad et al., 1996; Baretta-Bekker et al., 1998; Jost et al., 2004; Våge et al.,

2013; Ward and Follows, 2016). The “perfect beast” model (Flynn and Mitra, 2009) integrates phototrophy vs. phagotrophy with feedback functions to better represent a nearly true of mixotrophic behaviors, especially for predicting rates of ingestion (Mitra and Flynn, 2010). The “perfect beast” is a construct that can be configured to represent different types of constitutive and non-constitutive mixotrophs, and is consistent with our understanding of the different physiologies associated with these different types of mixotrophs (Mitra et al., 2016). Although the model has been previously configured to represent different generic mixotroph types, and used to explore the implications of different types of mixotrophy in oligotrophic through to eutrophic conditions (Flynn and Mitra, 2009; Mitra and Flynn, 2010; Flynn and Hansen, 2013; Mitra et al., 2014), this is the first time that this, or indeed any, multi-stoichiometric model of protist mixotrophy has been specifically tuned to simulate experimental data of the complexity explored here. In large measure, this reflects the paucity of such data not only for mixotroph activity but also for the prey. Indeed, very

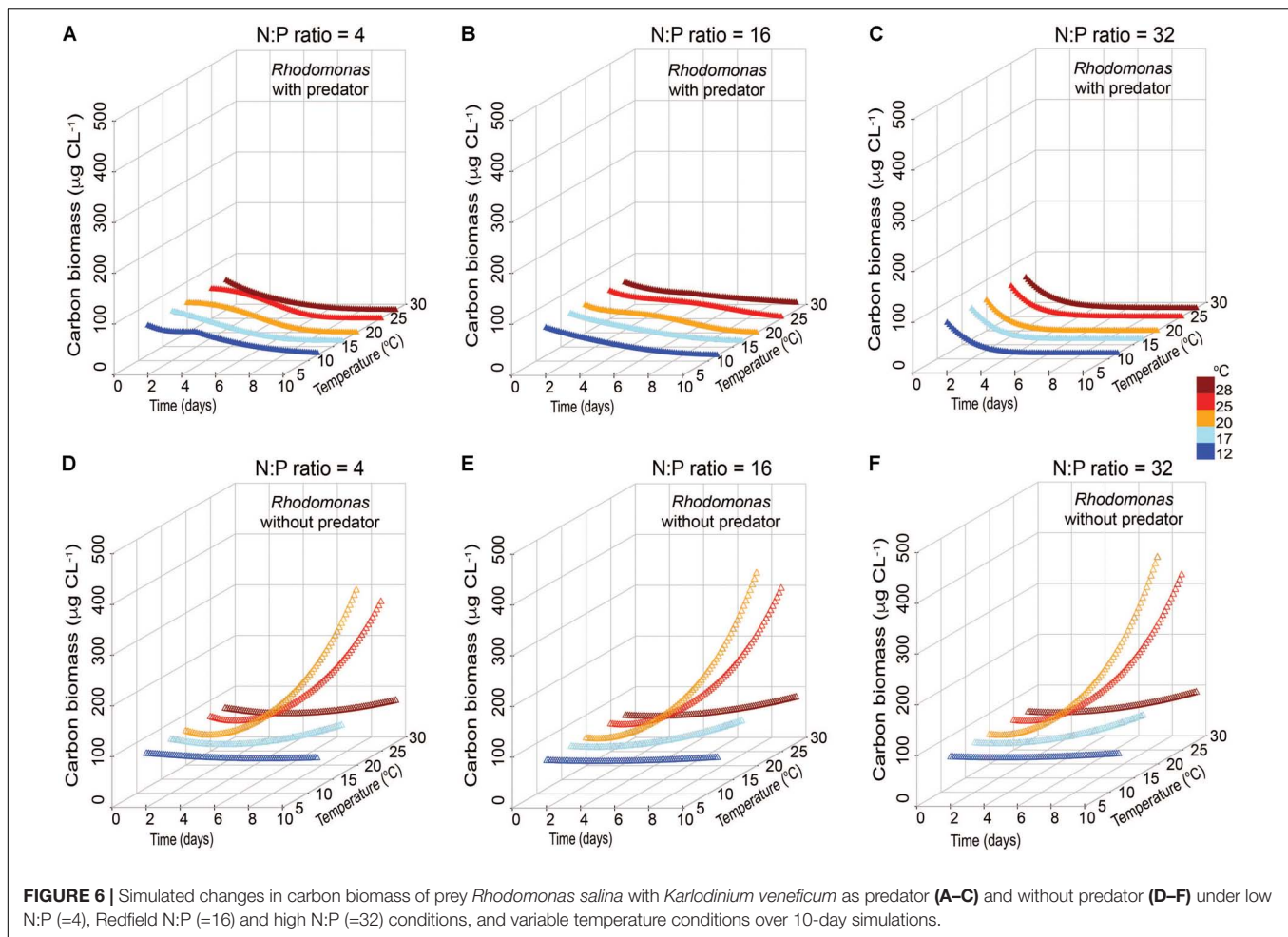


few empirical or modeling studies of phytoplankton describe multiple stoichiometries (i.e., C:N:P), despite increasing evidence that in nature such multiple nutrient states are important features structuring ecology and that nutrient stoichiometry is changing in many systems with anthropogenic nutrient loads (e.g., Peñuelas et al., 2012; Sutton et al., 2013; Glibert, 2017).

This study has successfully tuned the “perfect beast” to experimental data sets of the harmful dinoflagellate, *K. veneficum* and its prey, *Rhodomonas* sp., under varying nutrient conditions. Temperature growth responses were added to the original ‘perfect beast’ construct, thus allowing scenarios of mixotrophic growth under both varying nutrient and temperature conditions to be explored. The modeled scenarios highlighted several distinct differences in responses of *K. veneficum* as an autotroph and as a mixotroph in different nutrient and temperature conditions. Both autotrophic and mixotrophic *K. veneficum* attained much higher biomass in non-Redfieldian nutrient conditions compared to balanced nutrient growth. The modeled results in autotrophic *K. veneficum* biomass were quite consistent with the experimental data from Lin et al. (2017). Thus, simple stoichiometric relationships for predicting HAB developments need careful reconsideration, particularly in eutrophic systems

where the corresponding changes in nutrient ratios and forms as well as prey stoichiometry may ultimately affect cellular functions (such as C:N:P ratio) within the harmful algal mixotrophs (Mitra and Flynn, 2005, 2010). Indeed, it has been increasingly recognized that the benefits of mixotrophy to cells are synergistic, not additive (e.g., Mitra and Flynn, 2010).

In both low N:P and high N:P simulations, *K. veneficum* appeared to be more mixotrophic and attained higher biomass with increasing temperature compared with growth under Redfield conditions or growth as an autotroph (Figure 5). Under low N:P condition, temperature effects on mixotrophic growth rates of *K. veneficum* were also significant, with mixotrophic growth rates increasing faster than autotrophic growth rates at the highest temperature (i.e., 25°C; Figure 5A). This growth stimulation may imply a higher demand for C from prey. These growth patterns support the notion that mixotrophy is likely to be greater under nutrient imbalanced conditions, that is, that mixotrophy is not just a mechanism to acquire C, but also a mechanism by which nutrients are acquired (e.g., Glibert and Burkholder, 2011). For example, the assimilation efficiency (mA_{Emin}) in the mixotrophs was the highest for those *K. veneficum* grown under low N:P conditions and mixed with prey in Redfield N:P conditions, indicating nutrient sources from

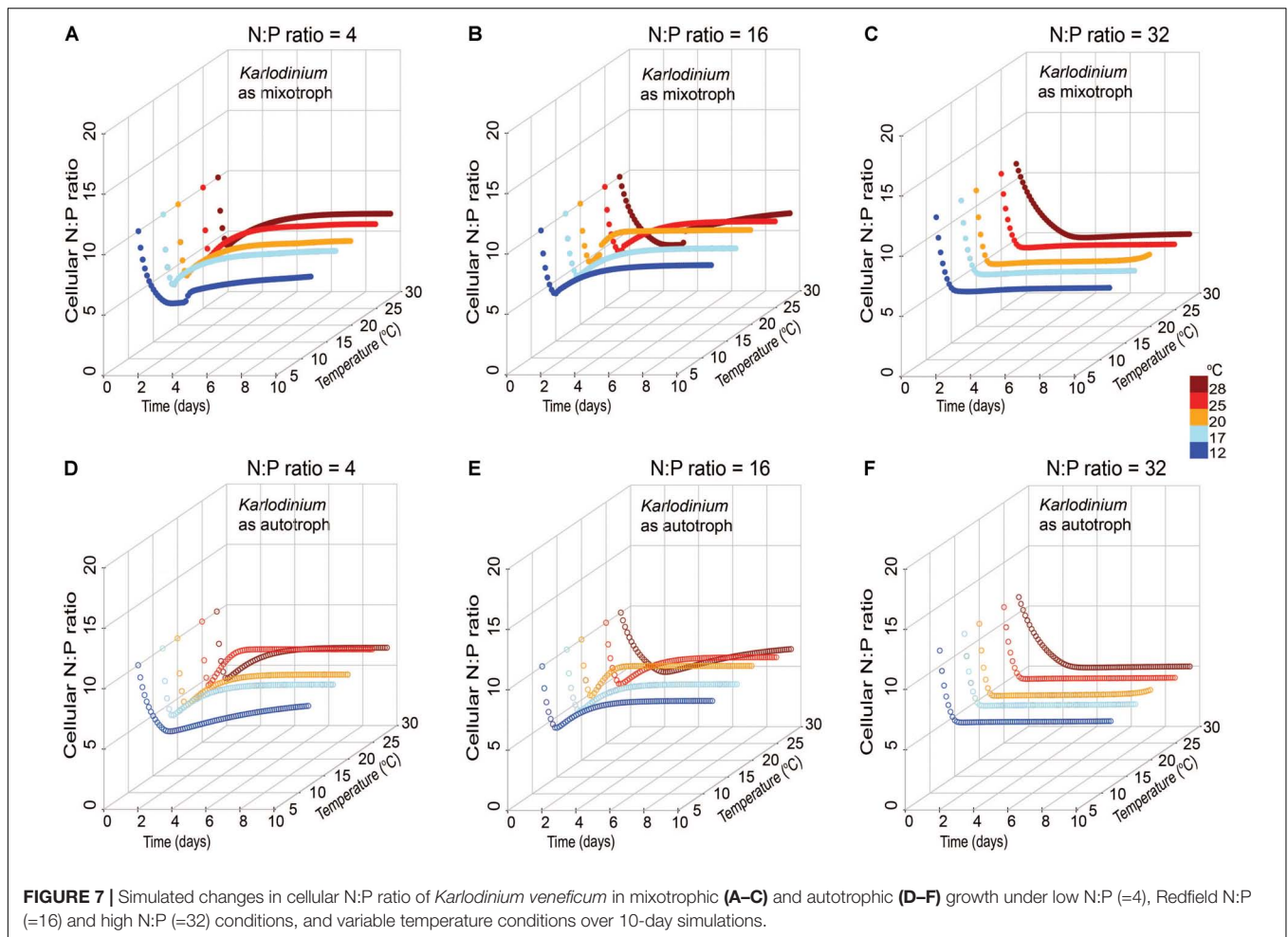


ingested prey were required (Table 2). On the other hand, the cellular N:P of *K. veneficum* in simulations under high N:P was low compared to the other nutrient conditions for this mixotroph (Figure 7C). The present simulations agreed with laboratory data: the incidence of *K. veneficum* feeding are enhanced when ambient nutrient ratios deviated from Redfield ratio with either N and/or P deficiency depending on the cellular status of the mixotrophic dinoflagellate (Lin et al., 2017). These results also suggest that under the warmest temperatures simulated, increased growth rates of prey could contribute to an increased growth of the mixotroph (Figure 6). The growth rates of *R. salina* were higher than those of *K. veneficum* > 20°C, and thus prey availability increases faster at these temperatures. This situation may be enhanced in the environments that deviate from balanced nutrient proportions.

Outputs from these simulated scenarios have implications for growth of this HAB in eutrophic conditions in warming environments. In eutrophic estuaries such as Chesapeake Bay, there are large seasonal variations in nutrient loads and in their stoichiometry (e.g., Kemp et al., 2005; Li et al., 2015). A conceptual model was previously developed of summer blooms of *K. veneficum* in Chesapeake Bay that incorporates the role of prey with a high N:P ratio originating from river inputs and a

source inocula of *K. veneficum* from southern Bay waters with a lower N:P content (Lin et al., 2018). Nutrient inputs through tributaries are greatest during the high-flow period, typically starting through March to May. During this period, prey can accumulate in the tributaries and are typically characterized by high N:P ratios due to disproportionate high N loading (Fisher et al., 1992; Kemp et al., 2005). The peak in summer *K. veneficum* blooms generally occurs 1–3 months later relative to these inputs, in June through September (Li et al., 2015). In this regard, according to previous laboratory experiments, growth performance of low NP-*K. veneficum* had a two-fold increase when fed upon prey with N-rich conditions compared to same nutrient conditions of prey (Lin et al., 2017). Thus, the enhanced growth of *K. veneficum* derived from the oceanic end member of the Bay may be enhanced if they encountered prey originating from the tributaries with different patterns of nutrient loading.

With accelerating climate change, mean temperatures may rise by 2–6°C by the end of the century in all seasons for Chesapeake Bay (Muhling et al., 2018). This may expand the window for *K. veneficum* growth in several ways. Prey availability may increase due to growth stimulation at higher temperatures. Also, recent Mid-Atlantic climate projections show that warming will likely increase current interannual variability, and that



winter/spring increases in precipitation are likely (e.g., Najjar et al., 2010), bringing increased N and high N:P conditions with these flows. These wetter spring conditions, with more nutrients may lead to more N-rich and/or P-deplete prey that may further support the development of these HABs. As mixotrophs may be more temperature sensitive than their autotrophic prey, the increased temperatures could enhance their ingestion capabilities and effectively control the growth of autotrophic prey (e.g., Yang et al., 2016). The modeled biomass of *K. veneficum* as a mixotroph was found to achieve the highest biomass when they consumed prey under high N:P conditions (Figure 5). Interestingly, the model suggests that while the highest biomass for *K. veneficum* is attained at 20°C under high N:P, and falls off rapidly above 20°C, under low N:P conditions, highest biomass is attained at 25°C. These differing temperature responses raise important questions that warrant further exploration experimentally.

Using models of mixotrophy, based on food uptake and photosynthesis measurements of *K. veneficum* and its congener, *K. armiger*, and assuming constant Redfield ratios, Berge et al. (2017) predicted succession of these species and their relative investments in autotrophy and phagotrophy. Their model suggested that nutrient uptake and high investments in photosynthesis would yield high autotrophic growth rates in

spring, but increased phagotrophy in summer. In another recent model, Ghyoot et al. (2017) developed a flexible model in which a distinction was made between constitutive mixotrophs, those that synthesize and maintain their chloroplasts, and non-constitutive mixotrophs, those that acquire chloroplasts. The next important step in mixotroph modeling will be to incorporate variable nutrient stoichiometry in a model of seasonal succession of both constitutive and non-constitutive mixotrophs.

CONCLUSION

In conclusion, the current study expanded modeling of mixotrophic growth to conditions of variable stoichiometry and temperature. These simulations have highlighted the consideration of particulate prey in modeling HAB dynamics under future warming; it is insufficient to only consider dissolved nutrients. The complexities shown here in consideration of differential impacts of temperature upon the growth of mixotroph predator and prey, parallel those expected under (de)eutrophication scenarios with ocean acidification (Flynn et al., 2015). Such changes in species competitive advantage under multi-stressor environments (light, temperature, pH, nutrients)

will require a concerted effort in physiology-modeling research to conceptualize adequately to aid ecosystem management. Even though the current models are based only on bottom-up, nutrient conditions, and are focused on only one typical prey species without modeling the role of the toxic contents of *K. veneficum* in predation purpose (Sheng et al., 2010), they have provided some insight into the potential trend in HABs under future eutrophication and warming conditions.

AUTHOR CONTRIBUTIONS

C-HL and PG conceived and designed the experiments and analyzed the simulation model outputs. KF and AM developed the model structures. C-HL and KF performed the model tuning. PG contributed the experimental materials, reagents, and tools. CL wrote the paper with editorial contributions from all authors.

FUNDING

C-HL was supported by funding from the Taiwanese Government, the Bay and Rivers Fellowship from the

REFERENCES

- Adolf, J. E., Bachvaroff, T., and Place, A. R. (2008). Can cryptophyte abundance trigger toxic *Karlodinium veneficum* blooms in eutrophic estuaries? *Harmful Algae* 8, 119–128. doi: 10.1016/j.hal.2008.08.003
- Adolf, J. E., Bachvaroff, T. R., Deeds, J. R., and Place, A. R. (2015). Ichthyotoxic *Karlodinium veneficum* (Ballantine) J Larsen in the upper Swan river estuary (Western Australia): ecological conditions leading to a fish kill. *Harmful Algae* 48, 83–93. doi: 10.1016/j.hal.2015.07.006
- Anderson, D. M., Glibert, P. M., and Burkholder, J. M. (2002). Harmful algal blooms and eutrophication: nutrient sources, composition, and consequences. *Estuaries* 25, 704–726. doi: 10.1007/BF02804901
- Baretta-Bekker, J. G., Baretta, J. W., Hansen, A. S., and Riemann, B. (1998). An improved model of carbon and nutrient dynamics in the microbial food web in marine enclosures. *Aquat. Microb. Ecol.* 14, 91–108. doi: 10.3354/ame014091
- Berge, T., Chakraborty, S., Hansen, P. J., and Andersen, K. H. (2017). Modeling succession of key resource-harvesting traits of mixotrophic plankton. *ISME J.* 11, 212–223. doi: 10.1038/ismej.2016.92
- Blossom, H. E., Daugbjerg, N., and Hansen, P. J. (2012). Toxic mucus traps: a novel mechanism that mediates prey uptake in the mixotrophic dinoflagellate *Alexandrium pseudogonyaulax*. *Harmful Algae* 17, 40–53. doi: 10.1016/j.hal.2012.02.010
- Braarud, T. (1957). A red water organism from Walvis Bay (*Gymnodinium galatheanum* n. sp.). *Galathea Deep Sea Exped.* 1, 137–138.
- Burkholder, J. M., Glibert, P. M., and Skelton, H. M. (2008). Mixotrophy, a major mode of nutrition for harmful algal species in eutrophic waters. *Harmful Algae* 8, 77–93. doi: 10.1016/j.hal.2008.08.010
- Dai, X., Lu, D., Guan, W., Wang, H., He, P., Xia, P., et al. (2013). Newly recorded *Karlodinium veneficum* dinoflagellate blooms in stratified water of the East China Sea. *Deep Sea Res. II* 101, 237–243. doi: 10.1016/j.dsr2.2013.01.015
- Deeds, J. R., Terlizzi, D. E., Adolf, J. E., Stoecker, D. K., and Place, A. R. (2002). Toxic activity from cultures of *Karlodinium micrum* (= *Gyrodinium galatheanum*) (Dinophyceae)—a dinoflagellate associated with fish mortalities in an estuarine aquaculture facility. *Harmful Algae* 1, 169–189. doi: 10.1016/S1568-9883(02)00027-6
- Fisher, T. R., Peele, E. R., Ammerman, J. W., and Harding Jr, L. W. (1992). Nutrient limitation of phytoplankton in Chesapeake Bay. *Mar. Ecol. Prog. Ser.* 82, 51–63. doi: 10.3354/meps082051
- Flynn, K. J., Mitra, A., Glibert, P. M., and Burkholder, J. M. (2018). “Mixotrophy in HABs: by whom, on whom, when, why and what next,” in *Global Ecology and*
- Horn Point Laboratory, and travel support from the Ryan Saba Memorial Fellowship. Additional support for this work was provided by the National Oceanic and Atmospheric Administration National Centers for Coastal Ocean Science Competitive Research program under award no. NA17NOS4780180. This is contribution number 5519 from the University of Maryland Center for Environmental Science.

ACKNOWLEDGMENTS

This work honors the memory of a friend of C-HL, Ryan Saba.

SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2018.00320/full#supplementary-material>

- Oceanography of Harmful Algal Blooms*, eds P. M. Glibert, E. Berdalet, M. A. Burford, G. C. Pitcher, and M. Zhou (Cham: Springer Press), 113–132.
- Flynn, K. J. (2001). A mechanistic model for describing dynamic multi-nutrient, light, temperature interactions in phytoplankton. *J. Plankt. Res.* 23, 977–997. doi: 10.1093/plankt/23.9.977
- Flynn, K. J. (2005). Castles built on sand: dysfunctionality in plankton models and the inadequacy of dialogue between biologists and modellers. *J. Plankt. Res.* 27, 1205–1210. doi: 10.1093/plankt/fbi099
- Flynn, K. J. (2009). Going for the slow burn: why should possession of a low maximum growth rate be advantageous for microalgae? *Plant Ecol. Divers.* 2, 179–189. doi: 10.1080/17550870903207268
- Flynn, K. J. (2010). Do external resource ratios matter?: implications for modelling eutrophication events and controlling harmful algal blooms. *J. Mar. Syst.* 83, 170–180. doi: 10.1016/j.jmarsys.2010.04.007
- Flynn, K. J. (2018). *Dynamic Ecology - An Introduction to the Art of Simulating Trophic Dynamics*. Swansea: Swansea University.
- Flynn, K. J., Clark, D. R., Mitra, A., Fabian, H., Hansen, P. J., Glibert, P. M., et al. (2015). Ocean acidification with (de)eutrophication will alter future phytoplankton growth and succession. *Proc. R. Soc. Lond. B Biol. Sci.* 282:20142604. doi: 10.1098/rspb.2014.2604
- Flynn, K. J., and Hansen, P. J. (2013). Cutting the canopy to defeat the “selfish gene”; conflicting selection pressures for the integration of phototrophy in mixotrophic protists. *Protist* 164, 811–823. doi: 10.1016/j.protis.2013.09.002
- Flynn, K. J. and McGillicuddy, D. J. Jr. (2018). “Modeling marine harmful algal blooms; current status and future prospects,” in *Harmful Algal Blooms: A Compendium Desk Reference*, eds S. E. Shumway, J.-A.M. Burkholder, and S. Morton (New York, NY: Wiley Science Publishers).
- Flynn, K. J., and Mitra, A. (2009). Building the “perfect beast”: modelling mixotrophic plankton. *J. Plankt. Res.* 31, 965–992. doi: 10.1093/plankt/fbp044
- Flynn, K. J., Stoecker, D. K., Mitra, A., Raven, J. A., Glibert, P. M., Hansen, P. J., et al. (2013). Misuse of the phytoplankton–zooplankton dichotomy: the need to assign organisms as mixotrophs within plankton functional types. *J. Plankt. Res.* 35, 3–11. doi: 10.1093/plankt/fbs062
- Fu, F. X., Tatters, A. O., and Hutchins, D. A. (2012). Global change and the future of harmful algal blooms in the ocean. *Mar. Ecol. Prog. Ser.* 470, 207–233. doi: 10.3354/meps10047
- Glycoot, C., Flynn, K. J., Mitra, A., Lancelot, C., and Gypens, N. (2017). Modeling plankton mixotrophy: a mechanistic model consistent with the Shuter-type biochemical approach. *Front. Ecol. Evol.* 5:78. doi: 10.3389/fevo.2017.00078

- Glibert, P. M. (2017). Eutrophication, harmful algae and biodiversity—challenging paradigms in a world of complex nutrient changes. *Marine Poll. Bull.* 124, 591–606. doi: 10.1016/j.marpolbul.2017.04.027
- Glibert, P. M., Allen, J. I., Bouwman, A. F., Brown, C. W., Flynn, K. J., Lewitus, A. J., et al. (2010). Modeling of HABs and eutrophication: status, advances, challenges. *J. Mar. Syst.* 83, 262–275. doi: 10.1016/j.jmarsys.2010.05.004
- Glibert, P. M., and Burford, M. A. (2017). Globally changing nutrient loads and harmful algal blooms: recent advances, new paradigms, and continuing challenges. *Oceanography* 30, 58–69. doi: 10.5670/oceanog.2017.110
- Glibert, P. M., and Burkholder, J. M. (2011). Harmful algal blooms and eutrophication: “strategies” for nutrient uptake and growth outside the Redfield comfort zone. *Chin. J. Oceanogr. Limnol.* 29, 724–738. doi: 10.1007/s00343-011-0502-z
- Glibert, P. M., and Burkholder, J. M. (2018). “Causes of harmful algal blooms,” in *Harmful Algal Blooms: A Compendium Desk Reference*, eds S. Shumway, J. M. Burkholder, and S. L. Morton (West Sussex: Wiley Press), 1–21.
- Glibert, P. M., Heil, C. A., Wilkerson, F. P., and Dugdale, R. C. (2018). “Nutrients and harmful algal blooms: dynamic kinetics and flexible nutrition,” in *Global Ecology and Oceanography of Harmful Algal Blooms*, eds P. M. Glibert, E. Berdalet, M. A. Burford, G. C. Pitcher, and M. Zhou (Cham: Springer Press), 93–112.
- Glibert, P. M., Seitzinger, S., Heil, C. A., Burkholder, J. M., Parrow, M. W., Codispoti, L. A., et al. (2005). The role of eutrophication in the global proliferation of harmful algal blooms. *Oceanography* 18, 198–209. doi: 10.5670/oceanog.2005.54
- Glibert, P. M., and Terlizzi, D. E. (1999). Cooccurrence of elevated urea levels and dinoflagellate blooms in temperate estuarine aquaculture ponds. *Appl. Environ. Microbiol.* 65, 5594–5596.
- Guillard, R. R. (1975). “Culture of phytoplankton for feeding marine invertebrates,” in *Culture of Marine Invertebrate Animals*, eds W. L. Smith and M. H. Chanley (New York, NY: Plenum Press), 29–60.
- Haefner, J. W. (2005). *Modeling Biological Systems: Principles and Applications*. Berlin: Springer Science & Business Media.
- Hallegraeff, G. M. (2010). Ocean climate change, phytoplankton community responses, and harmful algal blooms: a formidable predictive challenge. *J. Phycol.* 46, 220–235. doi: 10.1111/j.1529-8817.2010.00815.x
- Heisler, J., Glibert, P. M., Burkholder, J. M., Anderson, D. M., Cochlan, W., Dennison, W. C., et al. (2008). Eutrophication and harmful algal blooms: a scientific consensus. *Harmful Algae* 8, 3–13. doi: 10.1016/j.hal.2008.08.006
- Jakobsen, H. H., and Hansen, P. J. (1997). Prey size selection, grazing and growth response of the small heterotrophic dinoflagellate *Gymnodinium* sp. and the ciliate *Balanion comatum*—a comparative study. *Mar. Ecol. Prog. Ser.* 158, 75–86. doi: 10.3354/meps158075
- Jeong, H. J., Park, J. Y., Nho, J. H., Park, M. O., Ha, J. H., Seong, K. A., et al. (2005a). Feeding by red-tide dinoflagellates on the cyanobacterium *Synechococcus*. *Aquat. Microb. Ecol.* 41, 131–143. doi: 10.3354/ame041131
- Jeong, H. J., Yoo, Y. D., Park, J. Y., Song, J. Y., Kim, S. T., Lee, S. H., et al. (2005b). Feeding by phototrophic red-tide dinoflagellates: five species newly revealed and six species previously known to be mixotrophic. *Aquat. Microb. Ecol.* 40, 133–150. doi: 10.3354/ame040133
- Jost, C., Lawrence, C. A., Campolongo, F., Van de Bund, W., Hill, S., and DeAngelis, D. L. (2004). The effects of mixotrophy on the stability and dynamics of a simple planktonic food web model. *Theor. Popul. Biol.* 66, 37–51. doi: 10.1016/j.tpb.2004.02.001
- Kemp, W., Boynton, W., Adolf, J., Boesch, D., Boicourt, W., Brush, G., et al. (2005). Eutrophication of Chesapeake Bay: historical trends and ecological interactions. *Mar. Ecol. Prog. Ser.* 303, 1–29. doi: 10.3354/meps303001
- Kempton, J. W., Lewitus, A. J., Deeds, J. R., Law, J. M., and Place, A. R. (2002). Toxicity of *Karlodinium micrum* (Dinophyceae) associated with a fish kill in a South Carolina brackish retention pond. *Harmful Algae* 1, 233–241. doi: 10.1016/S1568-9883(02)00015-X
- Kimmance, S. A., Atkinson, D., and Montagnes, D. J. S. (2006). Do temperature–food interactions matter? Responses of production and its components in the model heterotrophic flagellate *Oxyrrhis marina*. *Aquat. Microb. Ecol.* 42, 63–73. doi: 10.3354/ame042063
- Li, A., Stoecker, D. K., and Adolf, J. E. (1999). Feeding, pigmentation, photosynthesis and growth of the mixotrophic dinoflagellate *Gyrodinium galatheanum*. *Aquat. Microb. Ecol.* 19, 163–176. doi: 10.3354/ame019163
- Li, A., Stoecker, D. K., and Coats, D. W. (2000). Mixotrophy in *Gyrodinium galatheanum* (Dinophyceae): grazing responses to light intensity and inorganic nutrients. *J. Phycol.* 36, 33–45. doi: 10.1046/j.1529-8817.2000.98076.x
- Li, J., Glibert, P. M., and Gao, Y. (2015). Temporal and spatial changes in Chesapeake Bay water quality and relationships to *Prorocentrum* minimum, *Karlodinium veneficum*, and CyanoHAB events, 1991–2008. *Harmful Algae* 42, 1–14. doi: 10.1016/j.hal.2014.11.003
- Lin, C.-H., Accoroni, S., and Glibert, P. M. (2017). *Karlodinium veneficum* feeding responses and effects on larvae of the eastern oyster *Crassostrea virginica* under variable nitrogen: phosphorus stoichiometry. *Aquat. Microb. Ecol.* 79, 101–114. doi: 10.3354/ame01823
- Lin, C.-H., Lyubchich, V., and Glibert, P. M. (2018). Time series models of decadal trends in the harmful algal species *Karlodinium veneficum* in Chesapeake Bay. *Harmful Algae* 73, 110–118. doi: 10.1016/j.hal.2018.02.002
- Lundgren, V. M., Glibert, P. M., Granéli, E., Vidyarthna, N. K., Fiori, E., Ou, L., et al. (2016). Metabolic and physiological changes in *Prymnesium parvum* when grown under, and grazing on prey of, variable nitrogen: phosphorus stoichiometry. *Harmful Algae* 55, 1–12. doi: 10.1016/j.hal.2016.01.002
- Menden-Deuer, S., and Lessard, E. J. (2000). Carbon to volume relationships for dinoflagellates, diatoms, and other protist plankton. *Limnol. Oceanogr.* 45, 569–579. doi: 10.4319/lo.2000.45.3.0569
- Mitra, A. (2006). A multi-nutrient model for the description of stoichiometric modulation of predation in micro- and mesozooplankton. *J. Plankt. Res.* 28, 597–611. doi: 10.1093/plankt/fbi144
- Mitra, A., and Flynn, K. J. (2005). Predator–prey interactions: is ‘ecological stoichiometry’ sufficient when good food goes bad? *J. Plankt. Res.* 27, 393–399. doi: 10.1093/plankt/fbi022
- Mitra, A., and Flynn, K. J. (2006). Accounting for variation in prey selectivity by zooplankton. *Ecol. Model.* 199, 82–92. doi: 10.1016/j.ecolmodel.2006.06.013
- Mitra, A., and Flynn, K. J. (2010). Modelling mixotrophy in harmful algal blooms: more or less the sum of the parts? *J. Mar. Syst.* 83, 158–169. doi: 10.1016/j.jmarsys.2010.04.006
- Mitra, A., Flynn, K. J., Burkholder, J., Berge, T., Calbet, A., Raven, J. A., et al. (2014). The role of mixotrophic protists in the biological carbon pump. *Biogeosciences* 10, 13535–13562. doi: 10.5194/bg-11-995-2014
- Mitra, A., Flynn, K. J., Tillmann, U., Raven, J. A., Caron, D., Stoecker, D. K., et al. (2016). Defining planktonic protist functional groups on mechanisms for energy and nutrient acquisition: incorporation of diverse mixotrophic strategies. *Protist* 167, 106–120. doi: 10.1016/j.protis.2016.01.003
- Montagnes, D. J. S., Kimmance, S. A., and Atkinson, D. (2003). Using Q10: can growth rates increase linearly with temperature? *Aquat. Microb. Ecol.* 32, 307–313. doi: 10.3354/ame032307
- Moore, S. K., Johnstone, J. A., Banas, N. S., and Salathe, E. P. Jr. (2015). Present-day and future climate pathways affecting *Alexandrium* blooms in Puget Sound, WA, USA. *Harmful Algae* 48, 1–11. doi: 10.1016/j.hal.2015.06.008
- Muhling, B. A., Gaitán, C. F., Stock, C. A., Saba, V. S., Tommasi, D., and Dixon, K. W. (2018). Potential salinity and temperature futures for the Chesapeake Bay using a statistical downscaling spatial disaggregation framework. *Estuar. Coast.* 41, 349–372. doi: 10.1007/s12237-017-0280-8
- Najjar, R. G., Pyke, C. R., Adams, M. B., Breitbart, D., Hershner, C., Kemp, M., et al. (2010). Potential climate-change impacts on the Chesapeake Bay. *Estuar. Coast. Shelf. Sci.* 86, 1–20. doi: 10.1016/j.eccs.2009.09.026
- Nielsen, M. V. (1993). Toxic effect of the marine dinoflagellate *Gymnodinium galatheanum* on juvenile cod *Gadus morhua*. *Mar. Ecol. Prog. Ser.* 93, 273–277. doi: 10.3354/meps095273
- Peñuelas, J., Sardans, J., Rivas-ubach, A., and Janssens, I. A. (2012). The human-induced imbalance between C, N and P in Earth’s life system. *Glob. Change Biol.* 18, 3–6. doi: 10.1111/j.1365-2486.2011.02568.x
- Place, A. R., Bowers, H. A., Bachvaroff, T. R., Adolf, J. E., Deeds, J. R., and Sheng, J. (2012). *Karlodinium veneficum* —the little dinoflagellate with a big bite. *Harmful Algae* 14, 179–195. doi: 10.1016/j.hal.2011.10.021
- Sheng, J., Malkiel, E., Katz, J., Adolf, J. E., and Place, A. R. (2010). A dinoflagellate exploits toxins to immobilize prey prior to ingestion. *Proc. Natl. Acad. Sci. U.S.A.* 107, 2082–2087. doi: 10.1073/pnas.0912254107

- Stickney, H. L., Hood, R. R., and Stoecker, D. K. (2000). The impact of mixotrophy on planktonic marine ecosystems. *Ecol. Model.* 125, 203–230. doi: 10.1146/annurev-marine-010816-060617
- Stoecker, D. K. (1998). Conceptual models of mixotrophy in planktonic protists and some ecological and evolutionary implications. *Eur. J. Protistol.* 34, 281–290. doi: 10.1016/S0932-4739(98)80055-2
- Stoecker, D. K., Adolf, J. E., Place, A. R., Glibert, P. M., and Meritt, D. W. (2008). Effects of the dinoflagellates *Karlodinium veneficum* and *Prorocentrum minimum* on early life history stages of the eastern oyster (*Crassostrea virginica*). *Mar. Biol.* 154, 81–90. doi: 10.1007/s00227-007-0901-z
- Stoecker, D. K., Hansen, P. J., Caron, D. A., and Mitra, A. (2017). Mixotrophy in the marine plankton. *Annu. Rev. Mar. Sci.* 9, 311–335. doi: 10.1146/annurev-marine-010816-060617
- Sutton, M. A., Bleeker, A., Howard, C. M., Erisman, J. W., Abrol, Y. P., Bekunda, M., et al. (2013). *Our Nutrient World. The Challenge to Produce More Food and Energy with Less Pollution*. Edinburgh: Centre for Ecology & Hydrology.
- Thingstad, T. F., Havskum, H., Garde, K., and Riemann, B. (1996). On the strategy of “eating your competitor”: a mathematical analysis of algal mixotrophy. *Ecology* 77, 2108–2118. doi: 10.2307/2265705
- Våge, S., Castellani, M., Giske, J., and Thingstad, T. F. (2013). Successful strategies in size structured mixotrophic food webs. *Aquat. Ecol.* 47, 329–347. doi: 10.1007/s10452-013-9447-y
- Van Dolah, F. M. (2000). Marine algal toxins: origins, health effects, and their increased occurrence. *Environ. Health Perspect.* 108, 133–141.
- Ward, B. A., Dutkiewicz, S., Barton, A. D., and Follows, M. J. (2011). Biophysical aspects of resource acquisition and competition in algal mixotrophs. *Am. Nat.* 178, 98–112. doi: 10.1086/660284
- Ward, B. A., and Follows, M. J. (2016). Marine mixotrophy increases trophic transfer efficiency, mean organism size, and vertical carbon flux. *Proc. Natl. Acad. Sci. U.S.A.* 113, 2958–2963. doi: 10.1073/pnas.1517118113
- Wells, M. L., Trainer, V. L., Smayda, T. J., Karlson, B. S. O., Trick, C. G., Kudela, R. M., et al. (2015). Harmful algal blooms and climate change: learning from the past and present to forecast the future. *Harmful Algae* 49, 68–93. doi: 10.1016/j.hal.2015.07.009
- Wilken, S., Huisman, J., Naus-Wiezer, S., and Donk, E. (2013). Mixotrophic organisms become more heterotrophic with rising temperature. *Ecol. Lett.* 16, 225–233. doi: 10.1111/ele.12033
- Yang, Z., Zhang, L., Zhu, X., Wang, J., and Montagnes, D. J. S. (2016). An evidence-based framework for predicting the impact of differing autotroph-heterotroph thermal sensitivities on consumer–prey dynamics. *ISME J.* 10, 1767–1778. doi: 10.1038/ismej.2015.225
- Zhou, C., Fernández, N., Chen, H., You, Y., and Yan, X. (2011). Toxicological studies of *Karlodinium micrum* (Dinophyceae) isolated from East China Sea. *Toxicon* 57, 9–18. doi: 10.1016/j.toxicon.2010.08.014

Conflict of Interest Statement: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

The handling Editor declared a past co-authorship with the authors KF and PG.

Copyright © 2018 Lin, Flynn, Mitra and Glibert. This is an open-access article distributed under the terms of the Creative Commons Attribution License (CC BY). The use, distribution or reproduction in other forums is permitted, provided the original author(s) and the copyright owner(s) are credited and that the original publication in this journal is cited, in accordance with accepted academic practice. No use, distribution or reproduction is permitted which does not comply with these terms.