



Variation in Coarse Particulate Matter (PM₁₀) and Its Characterization at Multiple Locations in the Semiarid Region

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Introduction: The elemental composition and morphological study of particulate matter are very important to understand the nature of particles influencing the environment, climate, soil, and health.

Methods: The PM₁₀ samples were collected during the winter season (2018) in Nowshera city, KPK, Pakistan, in three locations, namely, urban, industrial, and suburban. Scanning electron microscopy (SEM) and electron-dispersive X-ray (EDX) spectroscopy were used to examine the PM samples for morphological examination and elemental composition.

Results: The average mass concentrations of particulate matter (PM₁₀) at the urban, industrial, and suburban locations were 238.5, 505.1, and 255.0 μg m⁻³, respectively. The average PM₁₀ mass concentration was higher than the WHO and National Ambient Air Quality Standards (NAAQS). The results of EDX showed that samples contained variable amounts of thirteen elements, such as oxygen, carbon, silicon, magnesium, sodium, calcium, iron, aluminum, potassium, sulfur, titanium, gold, and chlorine. The probable sources of PM were biogenic like plant debris, pollen, and diatoms; geogenic like road dust and resuspended soil dust; and anthropogenic like carbonaceous particles and fly ash, as confirmed by SEM-EDX. The carbonaceous species, that is, OC and EC, had average values of 55.8 ± 13.1 and 4.6 ± 0.6, 5.2 ± 3.2, and 36.4 ± 10.4, 40.0 ± 2.6 and, 6.3 ± 0.2 in industrial, urban, and suburban locations, respectively. Similarly, OC/EC had average values of 12.0 ± 1.2, 8.0 ± 3.0, and 6.3 ± 0.2 in industrial, urban, and suburban locations, respectively. Highly significant correlations among water-soluble ions (K⁺), OC, and EC were found in each location.

Conclusions: The examined PM₁₀ mass concentration in Nowshera city was above the thresholds of National Ambient Air Quality Standards (NAAQS) set by the U.S. Environmental Protection Agency (EPA). In addition, the concentration of pollutants

was the highest at the industrial site compared to the other sites. The HYSPLIT model showed that the air mass originated from local sources like cement industries, brick kiln industries, and others.

Keywords: particulate matter, Nowshera, SEM-EDX, OC, EC, OC/EC ratio

INTRODUCTION

Ambient air pollution is one of the biggest threats all around the world. In various countries, governments are developing technologies to decrease pollutants in the air. A different pollutant source emits pollutants in the form of gases and particulates into the earth's atmosphere. Particulate matter (PM) is a suspended combination of solid particles and liquid droplets. Suspended PM contains dust and other particulates of size $<< 1\text{--}200\ \mu\text{m}$ (Chandrasekaran et al., 1997). The most important physical properties of PM are its size and shape. Thus, the PM based on size is divided into two main groups or types. The aerodynamic diameter of the coarse particles spans from 2.5 to 10 μm . Mechanical disruption, that is, crushing, grinding, dust from roads, uncovered sand, industry, and construction fly ash from fossil fuel burning, as well as from non-combustible material, such as pollen grain and mold spores, causes them to form. The life span of coarse particles (PM_{10}) ranges from minutes to hours, and their transit distance varies from 1 to 10 km. The aerodynamic diameter of fine particles ($\text{PM}_{2.5}$) is less than 2.5 μm and comprises multiple sulfates, nitrates, and carbon compounds, as well as ammonium, hydrogen ions, and metals (Pb, Cu, Zn, Mn, and Fe). Fossil fuel combustion, biomass burning, smelting, and metal processing are the main producers of $\text{PM}_{2.5}$. Their life span ranges from days to weeks, and they may travel between 100 and 1,000 km (Pipal and Satsangi, 2015).

PM has a key role in changing the Earth's climate (Fowler et al., 2009). Particulate matter with an aerodynamic diameter of 2.5–10 μm ($\text{PM}_{10-2.5}$) can readily pass through and settle in a thoracic area of airways (Berube et al., 2007). On the other hand, PM with an aerodynamic diameter between 1 and 2.5 μm ($\text{PM}_{2.5-1}$) can enter deep into the distal lung, which eventually intensifies never-ending respiratory and cardiovascular diseases (Stevanovic et al., 2013). The number of particles and their surface area are also important because of the ultrafine proportion of coarse particles (Levin et al., 2016). There are very limited studies about the characterization of the size, chemical composition, and morphology of PM, which can provide substantial insights into their origin (Khan et al., 2010; Deshmukh et al., 2012; Kumar et al., 2013). The chemical composition of PM plays a significant function in its characterization and identification of various emission sources (Wu et al., 2013). For the determination of the chemical composition of PM, various techniques are used like XRF, ICP-MS, PIXE, INAA, and XPS. However, most of these techniques do not reveal chemical composition based on the shape and size of the object (Ram et al., 2012). For that reason, SEM-EDS is used to find sources based on morphological properties such as size, shape, equivalent spherical diameter, aspect ratio, circularity, and perimeter (Kulkarni et al., 2011; Satsangi and Yadav, 2014; Mishra et al., 2015).

Because of the climatic implications, atmospheric carbonaceous particles have a key role (Frazer, 2002; Rajput et al., 2014). The types of carbonaceous PM including elemental carbon (EC) are generated from the combustion process and are called primary pollutants. Similarly, organic carbon (OC) can be emitted directly from anthropogenic and biogenic sources, that is, primary organic carbon, or it can be generated in the atmosphere by photochemical reactions, that is, secondary organic compounds *via* volatile organic carbon gas-to-particle conversion (Cao et al., 2003; Satsangi et al., 2012). Elemental carbon is strongly absorbed *via* solar radiation and has a significant role in climate change after CO_2 (Hansen et al., 2000; Rajput et al., 2013). Similarly, OC scatters solar radiation and causes negative radiative forcing (Li and Bai, 2009).

Most cities in the world are facing challenges associated with air pollution. Air pollution has become a severe problem in densely populated nations like Pakistan, owing to industrialization and the uncontrolled release of air-polluting agents (Liu and Diamond, 2005; Bradsher and Barboza, 2006; Kahn and Yardley, 2007; Kerchich et al., 2011). Air pollution is caused by particulate matter (PM_{10} , $\text{PM}_{2.5}$) and other organic agents emitted by congested road traffic systems, as well as other reasons (Chow et al., 1994; Eldred et al., 1997). Dust, particles, and aerosols all contribute to air pollution, which disrupts the dynamic atmospheric balance (Dhakal, 2003; Rajput et al., 2017). PM, one of the most frequent ambient air pollutants, is currently the subject of epidemiological and toxicological research.

Previous research revealed that environmental epidemiologists focused mostly on PM with an aerodynamic diameter of 10 μm (PM_{10}) (Goldberg et al., 2001; Janssen et al., 2002). According to a recent report by the World Health Organization (WHO) on ambient air pollution, annual mean PM_{10} concentrations increased by more than 5% in 720 cities throughout the world between 2008 and 2013 (Goel and Kumar, 2014). There are very limited studies on the characterization of PM in the Khyber Pakhtunkhwa Province of Pakistan (Alam et al., 2015; Zeb et al., 2018), and there is no study in Nowshera city. As per the National Disaster Management Plan 2012-22, Nowshera has been regarded as the most "at-risk district" in Khyber Pakhtunkhwa, Pakistan. Various types of hazards have been proposed for Nowshera, such as riverine floods, flash floods, earthquakes, land sliding, soil erosion, epidemics, drought, pest attacks, waterborne disease, hailstorms, as well as industrial fires, sectarian violence, terrorism, and refuges (District Disaster Management Unit, 2014). Therefore, it is the need of the hour to find the extent of different types of pollutants, especially air pollutants, including particulate matter. In the current study, we attempted to study the concentration of PM_{10} , its chemical composition, and morphological analysis collected from three locations in

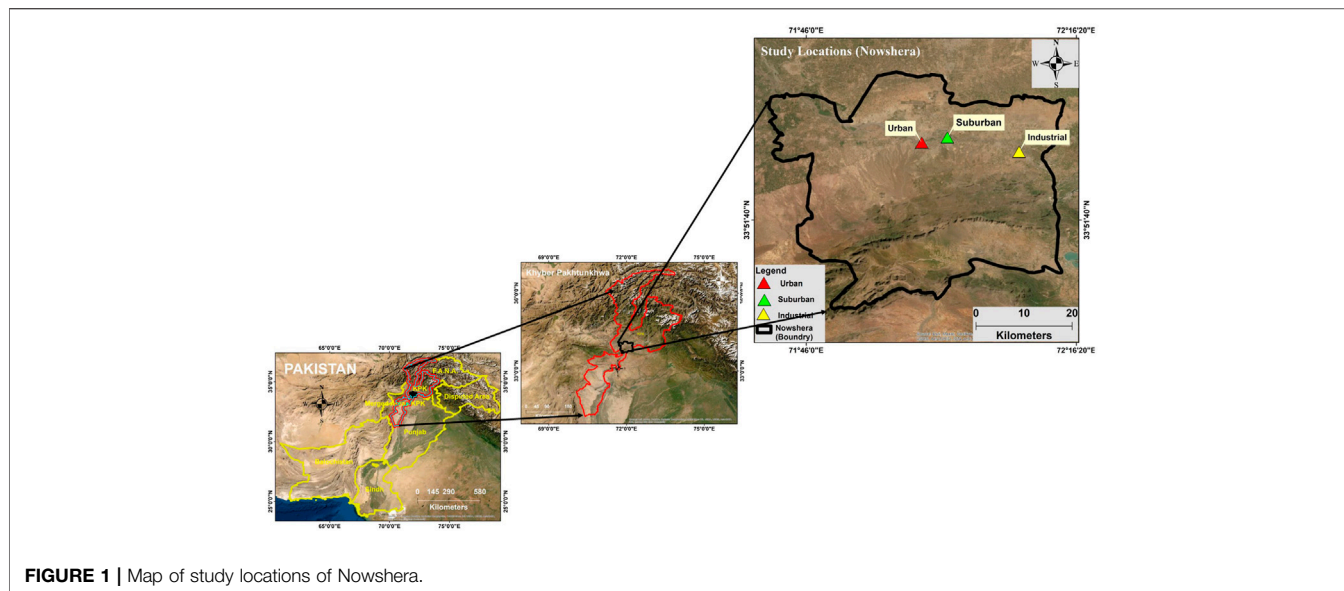


FIGURE 1 | Map of study locations of Nowshera.

Nowshera city. The morphology and chemical composition of samples were carried out through SEM and SEM–EDX, respectively. The current research will aid in the understanding of air quality, chemical composition, and sources of PM in Nowshera city. Furthermore, this study will help the researchers’ study the other types of PM in Nowshera.

MATERIALS AND METHODS

Descriptions of the Study Site

Nowshera lies on a sandy plain surrounded by small hills. The city is an important hub connected by rails and roads to various locations of the country, like Malakand, Mardan, Peshawar, and Islamabad. The city is also surrounded by national highways (N-5) and motorways (M-1). Nowshera city is the fastest growing city with a population of 1.394 million and lies between 34.01°N latitude and 71.97°E longitude having an altitude of 128 m, which covers an area of 1748 km³. The climate of Nowshera is a local steppe climate. The winters are chilly from November to February, and the summers are hot from June to August. The average annual temperature is 22.4°C, and the rainfall is 532 mm (Bibi et al., 2015). The study was carried out in three locations in Nowshera city, namely, 1) urban location, 2) industrial location, and 3) suburban location (Figure 1).

At the urban location, the sampling was performed at the top of a three-story building of Apostle Degree College located in the main city of Nowshera, near the main general bus stand and the main bazaar. In this area, there are many vehicle workshops and a huge traffic. Samplings were collected from the industrial location of Adamzai at the top of a two-story building surrounded by several factories, operating in the area, such as the Pakistan Tobacco Company, cotton, wool, newsprint, paper, marble, cement, and glass industries. The industrial location is also surrounded by mining areas. The sampling at a suburban location was performed in the city campus of Northern

University at the top of a two-story building situated in Hakimabad. In Hakimabad, there are many housing colonies. This site is 5 km away from the urban location. Agricultural lands, woodlands, tea plantations, riverine, and marshy areas make up the majority of the region.

Particulate Matter Sampling and Its Analysis

Sample Collection

During March, April, and May 2018, a low-volume sampler (LVS) with one head was used to collect PM₁₀ samples from 8 to 8 a.m. The LVS was run at a constant flow rate of 16 L/M for 24 h. PM₁₀ was collected on quartz fiber filter substrates with a diameter of 47 mm and a pore size of 0.4 μm. The PM filters were removed using forceps, kept in a petri dish, conditioned, weighed, and stored in the refrigerator at 4°C for further investigation to prevent thermal deterioration and evaporation of volatile components. Using a microbalance, the gravimetric PM masses were computed by subtracting the blank filter’s initial average mass from the sampled filter’s final average mass. Before and after sampling, each filter substrate was weighted three times, and the mean value was computed.

Scanning Electron Microscopy With Energy Dispersive X-Ray

Scanning electron microscopy (SEM) is an electron beam–based high-resolution surface imaging technique. The samples were analyzed by SEM (JSM-5910, JEOL Japan) equipped with energy-dispersive X-ray spectroscopy (INCA-200, United Kingdom) at the Centralized Resource Laboratory in the Department of Physics, University of Peshawar, to determine their shape and elemental composition. The PM produced from various sources may be recognized using this crucial technique, which is based on the shape and chemical content. Sections of a filter (1 mm by 1 mm) were cut from the center with scissors and put on an

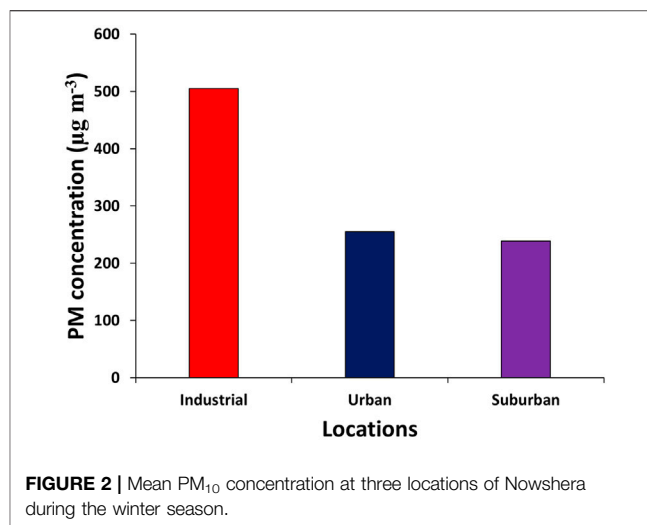
aluminum stub for analysis with double-sided tape. A very thin layer of gold (Au) was coated on the surface of each sample using a vacuum coating equipment called a gold sputter coater (SPI MODULE, United States) to improve conductivity and reduce electron charge. Six samples can be prepared at the same time using the sputter. Two photos of each sample were captured after they were placed in the corner of the SEM-EDX chamber. A microscope magnification of 550 was utilized to examine coarse particles, allowing detection areas of $60 \times 150 \text{ m}$. The backscattered electron model was utilized to examine the particle shape and position. Every signal has an embedded detector that recognizes its related signal and rejects others, effectively blocking undesirable or background transmissions. The morphological parameters such as particle shape and physical diameter were manually measured by using all the images recorded for each field and particle (minimum, maximum, and mean). The findings were obtained from three randomly selected fields for each filter substrate, which offered representative results and minimized subjectivity. Individual EDX spectra of PM particles were obtained after scanning an electron beam to determine the elemental composition of the particles. The distinct peaks were recognized, and the peak intensities were converted to weight using the computer program's quantifying function (Pipal and Satsangi, 2015). The Au data from EDX could not be used for quantitative purposes since the samples were Au-coated. As a result, the Au contribution was manually removed from the EDX spectra evaluation. SEM-EDX has also been used by other researchers to analyze the morphology and chemical composition (Chabas and Lefevre, 2000; Ma et al., 2001; Liu et al., 2005; Li et al., 2011).

Calculation of Element Percent Weight

The weight of individual elements was determined by the EDX examination of each particle. Blank quartz fiber filters were also subjected to EDX analysis, with the spectra results manually removed from those of the real individual aerosol particles. The weight of different elements was determined from the EDX spectra of each particle. For PM_{10} , the average weight percent of each constituent was computed after determining the total number of particles in each constituent.

Carbonaceous Species Analysis and Quality Control

A portion of the sample, 1.5 cm^2 , was cut and analyzed by using a thermal/optical carbon aerosol analyzer (Sunset Laboratory, Forest Grove, OR) using the NIOSH 5040 (National Institute of Occupational Safety and Health) protocol based on thermal optical transmittance for the measurement of the carbonaceous group in PM like OC, EC, CC, and TC (TOT). Satsangi et al. (2012) present a precise approach for the analysis of OC-EC. Sucrose solution (3.2 g L^{-1}) was used to standardize the instrument. A 10 L solution yields 32.0 1.8 g OC . Every day, the analyzer was calibrated for quality control using a blank punch of preheated quartz fiber filter and standard sucrose solutions. For blank filters, quartz filters were also sampled and evaluated in the same way. The overall blank concentrations for OC and EC were 0.5, 0.2, and 0.0, 0.02 g cm^{-2} , respectively, from the quartz filters. These were



removed from the aerosol samples' observed OC and EC values. With 3 mM MSA (methane sulfonic acid) as eluent, water-soluble K^+ was measured using a Dionex ICS 1100 Ion Chromatograph on an SCS 1 Cation Column and SCG 1 Guard Column. K^+ has a detection limit of 0.06 ppm, a precision of 3.7%, and an accuracy of 1.5%. The blank filter was retrieved and examined for blank adjustments and then subtracted from the observed K^+ concentrations in the PM samples for quality control.

RESULTS AND DISCUSSION

Particulate Matter Mass Concentration

The 24 h average PM mass concentrations at the three locations, that is, urban, industrial, and suburban, are investigated, as shown in **Figure 2**. The PM_{10} mass concentration at the industrial, urban, and suburban locations ranged from 412.52 to $597.60 \mu\text{g m}^{-3}$, 198.84 – $278.12 \mu\text{g m}^{-3}$ and 164.32 – $345.77 \mu\text{g m}^{-3}$, with an average value of 505.06 , 255.04 , and $238.48 \mu\text{g m}^{-3}$, respectively. The WHO permissible limit for 24 h PM_{10} mass concentrations is $50 \mu\text{g m}^{-3}$ and that of the National Ambient Air Quality Standards (NAAQS) is $150 \mu\text{g m}^{-3}$ (Davidson et al., 2005). In the three locations, that is, industrial, urban, and suburban locations, the measured values of PM mass concentration were 10.1, 5.1, and 4.8 times greater than the WHO standard and 3.4, 1.7, and 1.6 times higher than the NAAQS levels. There are heterogeneous sources of PM in the present study sites of Nowshera city. The urban location is surrounded by heavy transport and mechanical workshops; the industrial site is surrounded by Pakistan Tobacco Company, many small marble industries, vehicles, etc. At the suburban site, the emission sources were biomass burning and resuspended road dust. The high value of PM in these sites was due to the lack of vegetation cover area, bare earth, drought situation, unpaved roads, jammed traffic, and low-quality engines and fuels. The other reasons for high PM in the city were vehicular emissions, industrial emissions, re-suspension of road dust, emissions from brick kilns, and domestic combustion (Alam et al., 2011; Alam et al., 2015).

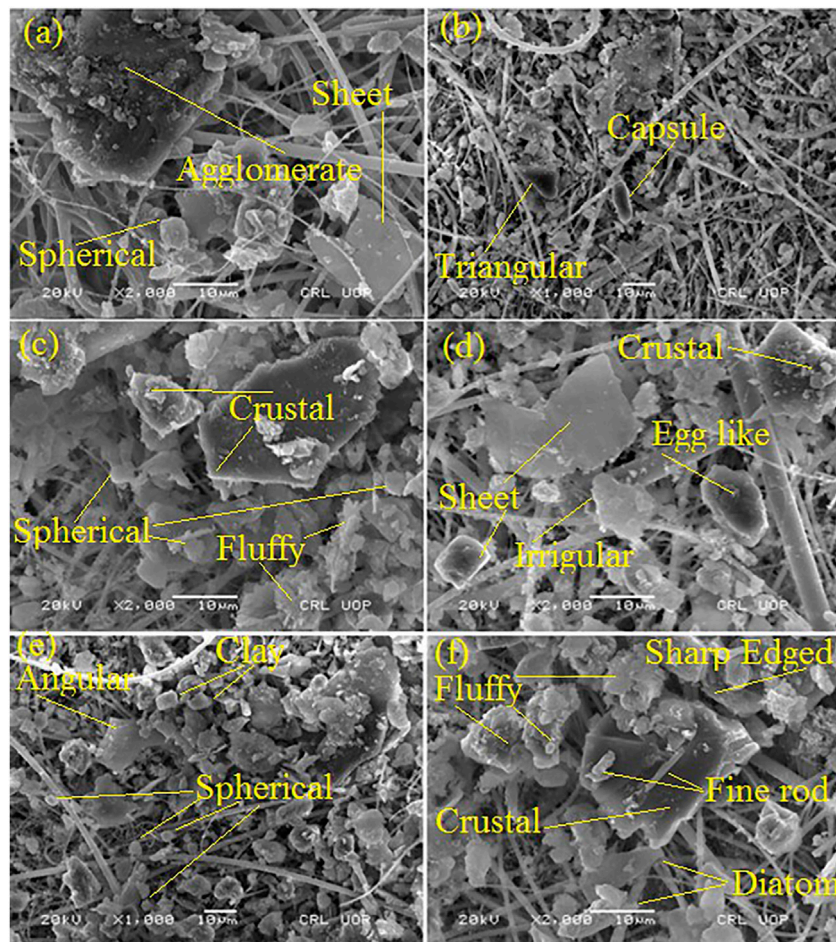
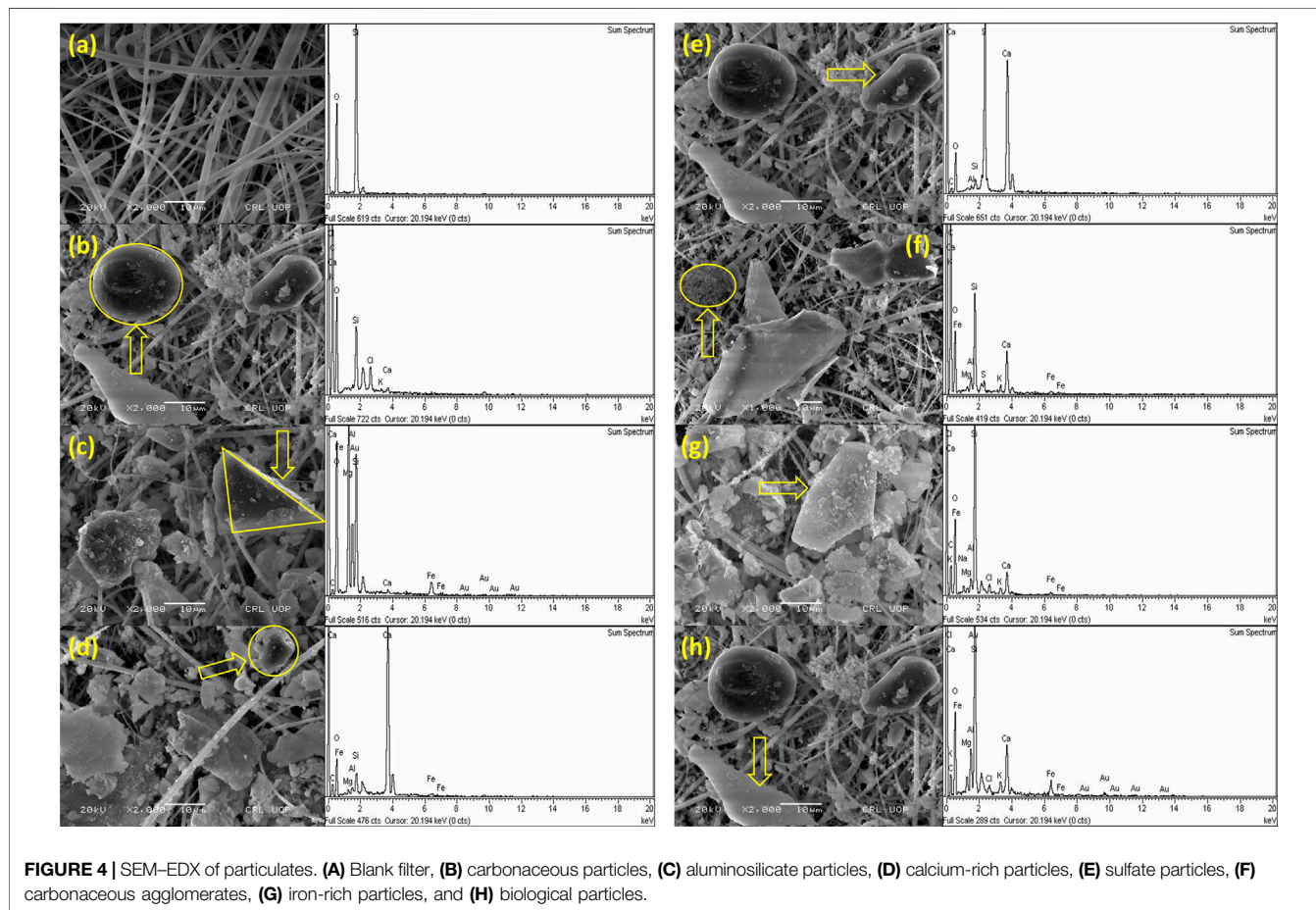


FIGURE 3 | SEM images of collected PM in three locations that is, industrial (A,B), urban (C,D), and suburban (E,F).

Various researchers have found different concentrations of PM under different scenarios. For example, Zeb et al. (2018) reported a PM_{10} concentration of $638 \mu g m^{-3}$ during the autumn season in Peshawar city, which was high as compared to that of the present study. However, Mehmood et al. (2018) reported the PM_{10} mass concentration of 152.4, 97.22, 86.52, and $40.45 \mu g m^{-3}$ during autumn, winter, spring, and summer seasons, respectively, in Islamabad city which is a low concentration in comparison to the present study. Khodeir et al. (2012) and Yadav et al. (2014) also reported a low concentration of PM_{10} . A PM_{10} concentration value of $480 \pm 83 \mu g m^{-3}$ was recorded in Peshawar city using the GRIMM instrument, which is similar to the present study (Alam et al., 2015). Nowshera had higher concentrations than other cities such as Lahore, Pakistan ($406 g m^{-3}$), Kolkata, India ($197 g m^{-3}$), Quetta, Pakistan ($331 g m^{-3}$), Karachi, Pakistan ($302 g m^{-3}$), Islamabad, Pakistan ($280 g m^{-3}$), Hangzhou, China ($119 g m^{-3}$), Punjab, India ($116 g m^{-3}$), Beijing-China ($140 \mu g m^{-3}$), Shanghai-China ($100 \mu g m^{-3}$), and Taipei-China ($60 \mu g m^{-3}$) (Karar and Gupta, 2007; Cao et al., 2009; Awasthi et al., 2011).

Morphological Characteristics of Coarse Particulate Matter in Three Locations

The scanning electron microscopic images of PM_{10} collected in three locations (industrial, urban, and suburban) in Nowshera city are shown in Figures 3A–F. SEM micrographs of PM at three locations show that particles were of various shapes and sizes, like triangular, oval, capsule, sharp-edged, spherical, fluffy, rod-like, angular, agglomerates, and diatom sheets. Other researchers also pointed out particles of various shapes (Prabhu et al., 2019). One of the key processes in the formation of irregular particles is the aggregation process (Rodríguez et al., 2009). The incomplete combustion of fossil fuels and biomass combustion are the primary causes of agglomeration particle emissions (Li et al., 2016). In the present study, carbonaceous chains and agglomerates were also observed, which are emitted from the combustion of fossil fuel in vehicles and industries (Mogo et al., 2005; Prabhu et al., 2019). Spherical particles were largely seen in micrographs, which were usually investigative of high-temperature processes or burning processes (Mogo et al., 2005). Particulate matter from sheet-like structures and



spherical carbonaceous particles were also observed. The existence of sheet-like structures related to clay is produced from geogenic sources, which are later resuspended from the surface of the earth by the wind. **Figures 3E,F** depict sharp-edged particles generated by anthropogenic activities such as construction and vehicular moments on the road. The figure also indicated biogenic particles, such as diatoms. The biological particles can be transported from 100 to 1,000 km from one place to another (Sofiev et al., 2006), which can also work as cloud condensation nuclei (Steiner et al., 2015). Fine PMs of size less than 1.5 μm are generated from fly ash emitted from the combustion process, and fine dust particles are produced from vehicular movement and construction activities (Mogo et al., 2005). The combustion of coal and biomass is one of the key sources of spherical soot particles. There are a large number of brick kilns surrounding the studied locations. The particles of irregular morphology are produced from crustal and construction activities. Therefore, the identified PM was grouped into geogenic, biogenic, and anthropogenic sources based on morphology.

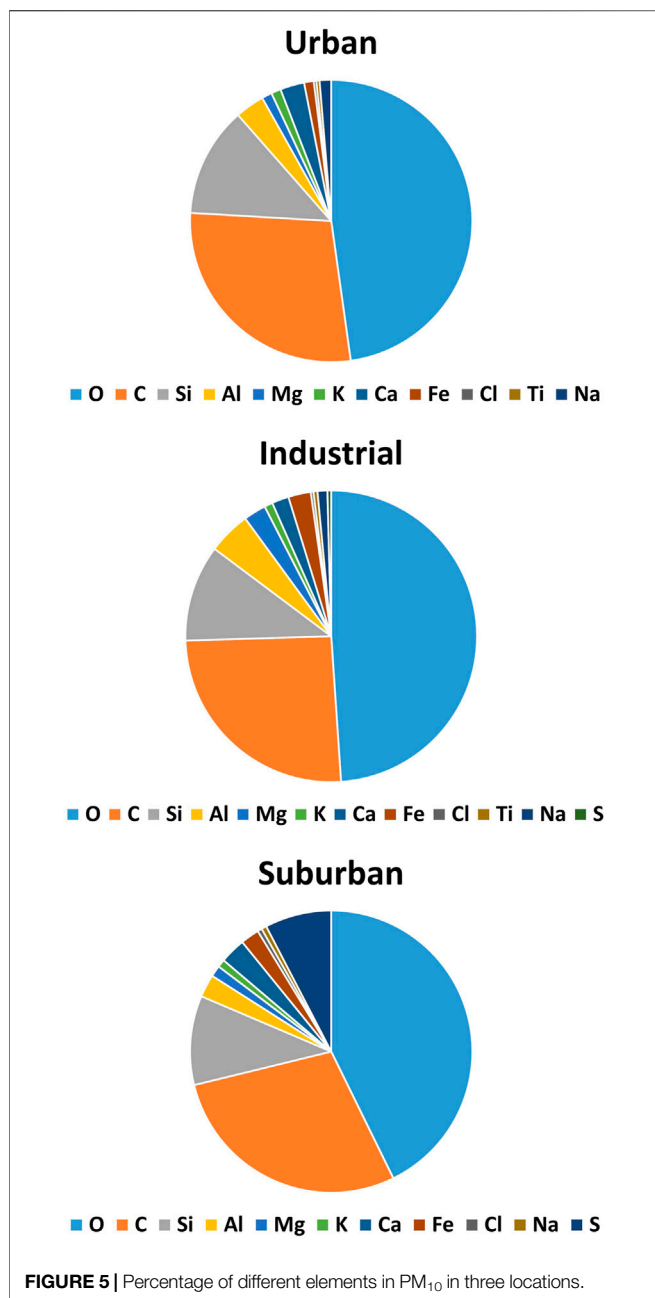
Energy Dispersive X-Ray Measurement Aluminosilicates

High concentrations of Si and Al, as well as varied proportions of Mg, K, Fe, and Ca, distinguish aluminosilicates (Cuadros et al., 2015) and have diverse sources and compositions.

Aluminosilicate particles of triangular shape are shown in **Figure 4C**. **Figure 4A** shows the presence of Si and O in the blank filter. Aluminosilicates, quartz, chlorides, calcium, and iron particles are found in irregular shapes (Pipal et al., 2011; Pachauri et al., 2013). The largest quantity of aerosol mass in the atmosphere is made up of these particles (Pipal and Satsangi, 2015). Soil dust, resuspended dust from roads and the earth's crust, and anthropogenic activities (constructions, combustion processes, agriculture fields, and moments of vehicles on roads) are the main sources of these particles (Satsangi and Yadav, 2014).

Aggregated Carbonaceous and Tarball Particle

Aggregate carbonaceous particles are shown in **Figure 4F**. Aggregate carbonaceous particles are generated from the combustion of fossil fuel and can be simply detected by their unique structure, varying from linear to complex forms (Tumolva et al., 2010). High-temperature process regions such as industries emit these types of particles (Gao and Ji, 2018). Near the study locations, the construction activities were the source of carbonaceous particles. **Figure 4B** shows tarball particles with spherical morphology. The special property of tarballs is recognized by their spherical and amorphous structures and generally is not aggregated with other particles (Cong et al., 2010). The typical source of tarballs is incomplete biomass and fossil fuel burning.



The existence of tarballs in the atmosphere has a key role in climate forcing and the regional haze because these affect the absorption and scattering of light efficiently (Hand et al., 2005). At Nowshera city, there was a mix of human activities including significant traffic congestion, which resulted in the emission of burnt or semi-burnt motor fuels, coal used in industries, generators used in suburban facilities, and tire abrasion (Satsangi and Yadav, 2014). Similar findings were reported in China by Cao et al. (2004).

Calcium-Rich Particles

The calcium (Ca) particles of irregular shape are shown in **Figure 4D**. These particles are rich in Ca and are found in various shapes and sizes. The sources of Ca-rich particles are

natural, as well as anthropogenic, that is, marble industries produce marble dust during their processes (Reist, 1993).

Iron (Fe)-Rich Particles

As indicated in **Figure 4G**, Fe-rich particles were also found, emitted from a variety of sources. Fe particles are released into the atmosphere by both natural and manmade activities. With the combination of mineral dust elements such as C, Al, K, Si, Ca, Na, and Mg, Fe-rich particles are found abundantly in Fe, Ti, and O. Fe oxides with irregular shapes make up the majority of this type of particles.

Sulfate Particles

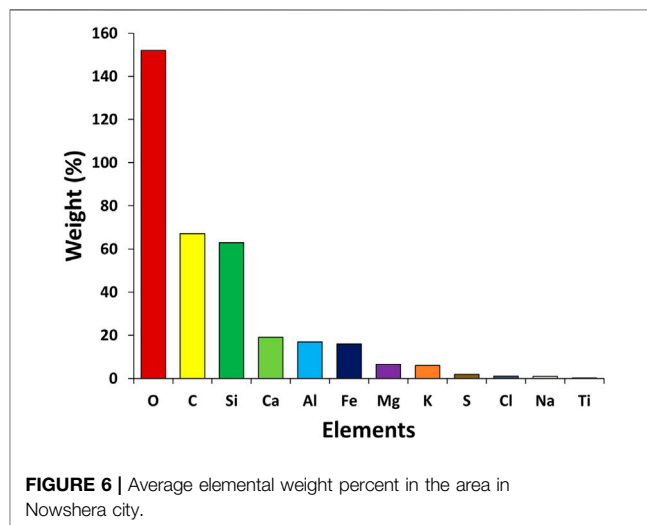
The existence of sulfur (S) particles confirms their source from the process of combustion. Generally, sulfate particles are generated from sulfur dioxide (SO₂) emitted through biomass and fossil fuel combustion (Pósfai et al., 1999). SO₂ is adsorbed on the surface of already existing particles and results in secondary minerals (Li and Shao, 2009). The high concentration of S is produced in the atmosphere through the combustion process of diesel inside heavy-duty vehicles and is very dangerous to human health. The sulfate particles having a rotundate rectangular shape of the size of the order of 8.4 μm are shown in **Figure 4E**.

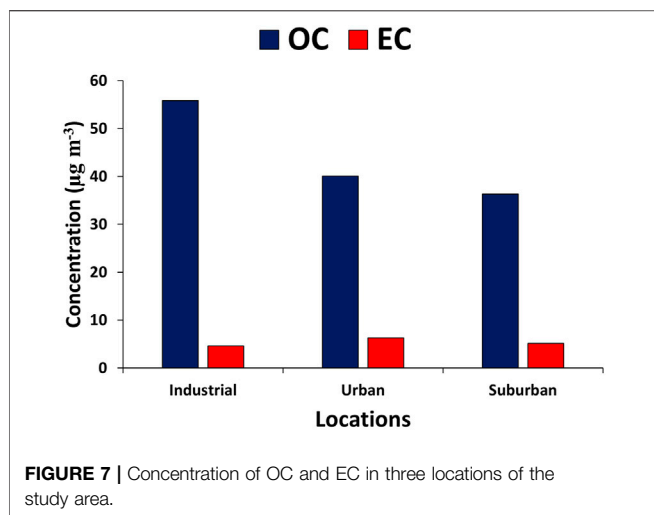
Biological Particles

Biological particles were detected, as shown in **Figure 4H**, containing >80% (C + O) and a small amount (<10%) of Na, Si, and Al. These particles are generated due to the existence of pollen, algae, protozoa, bacteria, and fragments of leaves. The excrements and fragments are also the sources of biological particles. Since there are a large number of trees, plants, and vegetation areas around the sampling sites, these particles are found in various shapes and sizes.

Elemental Contribution to Particulate Matter in Three Locations

Figures 5, 6 show the variation in the atomic percentage of thirteen elements, that is, carbon (C), oxygen (O), silicon (Si),





magnesium (Mg), sodium (Na), calcium (Ca), iron (Fe), aluminum (Al), potassium (K), sulfur (S), titanium (Ti), gold (Au), and chlorine (Cl) calculated from EDX spectra at three sites. The elemental arrangements at 1) industrial, 2) urban, and 3) suburban locations are $O > C > Si > Al > Mg > K > Ca > Fe > Cl > Ti > Na > S$, $O > C > Si > Al > Mg > K > Ca > Fe > Cl > Na > S$, and $O > C > Si > Al > Mg > K > Ca > Fe > Cl > Ti > Na$, respectively. The elemental compositions of PM_{10} in three locations are very similar. The difference between the two elements, that is, S and Ti, is observed in the locations. Sulfur was found only at industrial and urban sites, and titanium was found at urban and industrial sites. A large number of industries, such as brick kilns, pharmaceutical, steel and iron, and electroplating industries, were around the sampling locations. Oxygen was found in high amounts in all sites, which might be due to oxides of metals emitted from the burning of fossil fuels, vehicular exhaust, and different industries. After O, C was found to be the major element during the study.

The high amount of C after O in all three sites is a straightforward indicator of the various emission sources like fossil fuel and biomass burning, vehicular and coal burning emissions, energy sources in the industries, generator sets used in suburban facilities, and abraded vehicle tires (Shahid et al., 2013). Si is one of the most abundant elements in soil-derived mineral particles, according to the research. As a result, it could be caused by the transfer of airborne soil particles or fly ash created during industrial combustion. These particles were common in heavily industrialized areas and frequently contained Si (Herman et al., 1997; Baron and Willeke, 2001; Houghton et al., 2001). The anthropogenic sources of Al and Fe are industrial emissions, and their natural sources are mineral dust. The presence of soil-based elements like Ca, K, and Mg shows their existence from natural sources such as crustal resuspension and mineral dust (Kulshrestha et al., 2003; Kulshrestha et al., 2009; Tiwari et al., 2015). Regular open-air solid waste burning, as well as diverse industrial operations, can contribute to the elevated levels of Al (Pant et al., 2016; Mehra et al., 2020; Bangar et al., 2021) as found in the present study. From soil dust, Na was recognized as a crustal element. The presence of K has been linked to vegetation and biomass burning,

and Cl has been linked to industrial emissions because it is not found in ambient natural minerals but is commonly used in industrial processes (Hsu et al., 2016). Due to building and road renovations/construction, Ca and Al are produced, which cause mineral dust (Wang et al., 2013; Mustaffa et al., 2014; Rajput et al., 2018).

The Concentration of Carbonaceous Species in Coarse Particulate Matter, That Is, PM_{10}

Figure 7 depicts the mass concentrations of carbonaceous species such as OC and EC across the study period. In the industrial, urban, and suburban locations, OC ranged from 46.61 to 65.08, 29.01 to 43.69, and 38.19 to 41.91 $\mu\text{g m}^{-3}$, with an average value of 55.85 ± 13.06 , 36.35 ± 10.38 , and $40.05 \pm 2.64 \mu\text{g m}^{-3}$, respectively. Similarly, EC ranged from 4.18 to 5.06, 2.88 to 7.46, and 6.16 to 6.51 $\mu\text{g m}^{-3}$, with an average value of 4.62 ± 0.62 , 5.17 ± 3.23 , and $6.33 \pm 0.24 \mu\text{g m}^{-3}$ in the industrial, urban, and suburban locations, respectively. There are numerous sources of EC, including traffic pollutants, gasoline, diesel cars, and biomass burning (Ghauri et al., 2007; Zhang et al., 2008). OC is released directly from the combustion of fossil fuels or biomass, or by a chemical reaction known as secondary organic carbon (Satsangi et al., 2012). The high value of EC during the winter season might be due to less dispersion, wood burning, fuel burning, and coal burning in the region. Various researchers have also noted similar results regarding OC and EC values in different cities. For example, Satsangi et al. (2012) recorded OC and EC values during winter (37.4 ± 23.4 , $6.3 \pm 4.7 \mu\text{g m}^{-3}$), post-monsoon (33.0 ± 17.5 , $3.4 \pm 2.0 \mu\text{g m}^{-3}$), summer (29.4 ± 20.0 , $2.6 \pm 1.4 \mu\text{g m}^{-3}$), and monsoon (9.8 ± 9.6 , $1.7 \pm 1.3 \mu\text{g m}^{-3}$) in Dayalbagh, India. Similarly, Alam et al. (2012) reported the OC ($63 \pm 42 \mu\text{g m}^{-3}$) and EC ($21 \pm 15 \mu\text{g m}^{-3}$) values in the urban environment of Lahore, and these values were close to those observed in the present study.

The OC/EC, K^+/OC , and K^+/EC and Origin of Carbonaceous Aerosols

To validate the presence of primary and secondary organic aerosols, the ratios of organic carbon to elemental carbon (OC/EC) are utilized (Chow et al., 1996). OC/EC ratios have been reported by many researchers for various sources of emissions, such as biomass burning (3.8–13.2) (Zhang et al., 2007), coal smoke (2.5–10.5) (Chen et al., 2006), vehicular exhaust (2.5–5.00) (Schauer et al., 2002), kitchen emissions (4.3–7.7) (See and Balasubramaniam, 2008), and wood combustion (16.8–40.00) (Schauer et al., 2001). In the present study, the ratios (OC/EC) ranged from 11.13 to 12.86, 5.85 to 10.06, and 6.19 to 6.44, with average values of 12 ± 1.22 , 7.95 ± 2.96 , and 6.32 ± 0.17 in the industrial, urban, and suburban locations, respectively (Table 1). The larger values of the OC/EC ratio have a variety of reasons, including the combustion of wood and coal during the winter season, which results in large emissions of volatile organic precursors. During the winter season, stagnant meteorological circumstances produce a large amount of secondary organic aerosols (SOA). The high OC/EC values during the winter season are noted by different researchers

TABLE 1 | Ratio of OC/EC, K⁺/OC, and K⁺/EC in three locations of the study area.

	Industrial			Urban			Suburban		
	OC/EC	K ⁺ /OC	K ⁺ /EC	OC/EC	K ⁺ /OC	K ⁺ /EC	OC/EC	K ⁺ /OC	K ⁺ /EC
Min	11.13	0.05	0.55	5.85	0.04	0.31	6.19	0.03	0.22
Max	12.86	0.06	0.77	10.06	0.05	0.43	6.44	0.04	0.56
Mean	12.00	0.05	0.66	7.95	0.05	0.37	6.32	0.03	0.24
STD	1.22	0.007	0.15	2.96	0.007	0.38	0.17	0.004	0.022

TABLE 2 | Correlations between OC, EC, and water-soluble ions (K⁺) in study locations.

	Industrial		Urban		Suburban	
	EC	K ⁺	EC	K ⁺	EC	K ⁺
OC	0.92	0.78	0.89	0.71	0.85	0.81
EC		0.75		0.73		0.87

in different sites in China, including Shanghai (Feng et al., 2009), Tianjin (Li and Bai, 2009), Taiyuan (Meng et al., 2007), and Beijing (Dan et al., 2004). According to Alam et al., 2012, OC/EC ranged from 1.5 to 7.6, with an average value of 3.92 in the urban environment of Lahore, and was attributed to both biogenic and secondary aerosol formations. **Table 1** shows the K⁺/OC and K⁺/EC ratios for the three locations examined. In comparison to fossil fuels, K⁺/OC and K⁺/EC ratios can be utilized to classify carbonaceous aerosol emission sources such as biomass burning. The K⁺/EC ratio has been reported to be high (0.21–0.46) for biomass burning and low (0.025–0.09) for fossil fuel emissions (Ram and Sarin, 2010). In the present work, in industrial locations, K⁺/OC and K⁺/EC ranged from 0.05 to 0.06 and 0.55 to 0.77, with an average of 0.05 ± 0.007 and 0.66 ± 0.15, respectively. K⁺/OC and K⁺/EC varied from 0.04–0.05, 0.31–0.43, with an average value of 0.050.007 and 0.370.38, respectively, across urban locations. K⁺/OC varied from 0.03 to 0.04, with an average value of 0.03, 0.004. K⁺/EC ranged from 0.22 to 0.56, with an average value of 0.24 0.022 in suburban locations, respectively. The average K⁺/EC (0.560.32) value was discovered to be suggestive of biomass burning. According to Ram and Sarin (2010), the K⁺/EC ratio has varied between 0.30 and 0.69, 0.08 and 0.19, and 0.15 and 0.98 in Allahabad, Jaduguda, and Kanpur, respectively.

Correlation Between OC, EC, and Water-Soluble Ion (K⁺)

The correlation between OC, EC, and water-soluble ions (K⁺) is presented in **Table 2**. The origin of carbonaceous aerosols can be investigated by the correlation between OC and EC (Chow et al., 1996; Li and Bai, 2009). A very good correlation of OC to EC is found with a correlation coefficient of 0.92, 0.89, and 0.85 in industrial, urban, and suburban locations. OC has a good correlation with K⁺ in industrial (0.78), urban (0.71) and suburban (0.81) locations. Similarly, EC shows a strong relationship with K⁺ in three different settings: industrial

(0.75), urban (0.73), and domestic (0.73) (0.87). The OC/EC ratios are commonly employed to interpret carbonaceous aerosol emission and transformation features. Our findings reveal that OC and EC are proportionate, implying that the main sources of OC and EC are the same, that is, motor vehicle exhaust, coal combustion, and biomass burning (Li and Bai, 2009). Similarly, K⁺ produced during combustion is utilized as a marker of biomass burning (Deshmukh et al., 2010; Deshmukh et al., 2012). At the three locations, there is a strong association between water-soluble ion (K⁺), OC, and EC, indicating that biomass burning is a significant source of carbonaceous aerosols (Pachauri et al., 2013). Earlier, Stone et al. (2010) found a strong correlation between OC, coal combustion, and other unidentified sources, which is in line with the findings of the present study.

Water-Soluble Ion Concentration

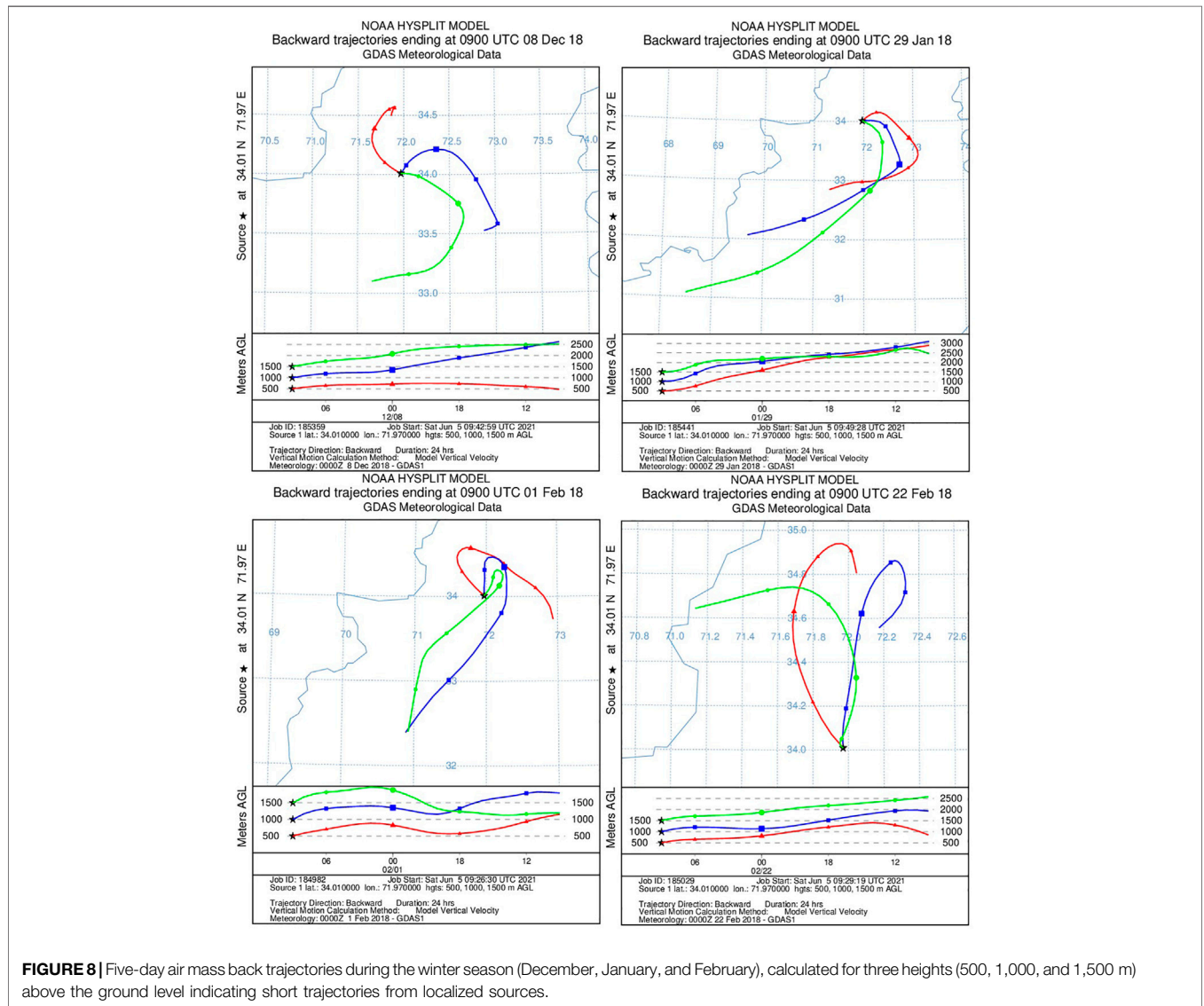
The number of water-soluble ions found in PM₁₀ in three locations is listed in **Table 3**. In each location, the water-soluble ions followed the trend, that is, Ca²⁺ > Na⁺ > K⁺ > Mg²⁺ > NH₄⁺. The values of the mass concentration of water-soluble ions in the industrial, urban and suburban locations were Ca²⁺ (30.05 ± 6.85, 18.31 ± 6.58, 17.87 ± 1.74 μg m⁻³) > Na⁺ (5.24 ± 0.17, 5.17 ± 0.40, 5.12 ± 0.09 μg m⁻³) > K⁺ (3.13 ± 1.13, 1.77 ± 0.77, 1.50 ± 0.84 μg m⁻³) > Mg²⁺ (2.50 ± 0.47, 1.27 ± 0.47, 1.26 ± 0.12 μg m⁻³). The values were high in the industrial region, followed by the urban and suburban regions. The concentration of NH₄⁺ (1.25 ± 0.72 μg m⁻³) was highest in urban areas, followed by suburban (0.77 ± 0.14) and industrial (0.71 ± 0.02 μg m⁻³) locations. NH₄⁺ was produced during the reactions of NH₃ with HNO₃, H₂SO₄, and their precursors. Generally, NH₃ is generated from anthropogenic sources, particularly from agricultural activities and fossil fuel burning (Meng et al., 2011). During winter season, NH₃ release from agricultural activities is decreased; therefore, NH₄⁺ in winter is associated with traffic, coal combustion, and biomass burning activities against cold weather. The concentration of different soluble ions is similar to that of Shahid et al. (2019) who found the concentration of different ions as Na⁺ (4.4 ± 0.4 μg m⁻³), Ca²⁺ (42.6 ± 15.4 μg m⁻³) K⁺ (2.5 ± 0.5 μg m⁻³), Mg²⁺ (0.8 ± 0.3 μg m⁻³), and NH₄⁺ (1.3 ± 0.7 μg m⁻³).

Analysis of the Air Mass Backward Trajectory

The results discussed in the previous sections show that biomass burning, cement, vehicular emissions, biological and brick kiln emissions were recognized to play a significant role in affecting the air quality of Nowshera city. However, it is suggested to have

TABLE 3 | Mass concentration of water-soluble ions in PM₁₀ in three locations.

Ion	Industrial			Urban			Suburban		
	Min	Max	AVG ± SD	Min	Max	AVG ± SD	Min	Max	AVG ± SD
Na ⁺	5.12	5.36	5.24 ± 0.17	4.89	5.46	5.17 ± 0.40	5.05	5.19	5.12 ± 0.09
NH ₄ ⁺	0.70	0.73	0.71 ± 0.02	0.74	1.77	1.25 ± 0.72	0.67	0.87	0.77 ± 0.14
K ⁺	2.33	2.93	3.13 ± 1.13	1.23	2.32	1.77 ± 0.77	1.44	1.56	1.50 ± 0.84
Mg ²⁺	2.16	2.83	2.50 ± 0.47	0.93	1.60	1.27 ± 0.47	1.17	1.34	1.26 ± 0.12
Ca ²⁺	25.20	34.89	30.05 ± 6.85	13.65	22.96	18.31 ± 6.58	16.57	19.02	17.87 ± 1.74



in-depth knowledge of the pathways traversed by the air masses converging in Nowshera city. Thus, to find out the source region of the air masses, analyses using the Hybrid-Single-Particle-Lagrangian Integrated-Trajectory (HYSPLIT) model were performed. The National Oceanic and Atmospheric Administration (NOAA) Air Resource Laboratory (ARL) Hybrid Single-Particle Lagrangian Integrated Trajectory (HYSPLIT) model was used to obtain these air mass back trajectories using the Global Data Assimilation System model's final run data archive. In **Figure 8**, five-day back trajectory studies of the winter months were simulated at the local standard time (noon) at three heights (500, 1,000, and 1,500 m) above ground level at three heights (500, 1,000, and 1,500 m) above ground level. The obtained data suggest that anthropogenic emission sources dominate short trajectories coming from local areas near Nowshera city.

CONCLUSION

The current study has investigated the atmospheric pollution in Nowshera city for the first time. The examined PM mass concentration in Nowshera city was confirmed to be above the Environmental Protection Agency's (EPA) and National Ambient Air Quality Standards (NAAQS) thresholds. This study revealed that the concentration of pollution was highest at the industrial site than at the other sites. The morphological and elemental characterization of the atmosphere of Nowshera has been investigated through SEM and EDX, respectively. Individual particulate matter characterizations through SEM–EDX have provided helpful insights corresponding to the PM sources based on the shape and size of particulates. Particulates produced from anthropogenic and geogenic sources were usually spherical and irregular in shape, but particles emitted from biogenic sources had a specific shape, according to the morphological investigation of particulate matter. The SEM results revealed various morphological shapes and sizes of pollutants, that is, sheets like rod-like, irregular, angular, triangular, spherical, agglomerates, and soot particles. The EDX results showed that O (oxygen), C (carbon), and Si (silicon) were the major pollutants in the study area. We have observed the values of OC and EC in industrial (55.85 ± 13.06 and 4.62 ± 0.62), urban (5.17 ± 3.23 and 36.35 ± 10.38), and suburban (40.05 ± 2.64 and 6.33 ± 0.24) locations. Similarly, the average values of OC/EC were 12 ± 1.22 , 7.95 ± 2.96 , and 6.32 ± 0.17 in industrial, urban, and suburban locations, respectively. In

industrial, urban, and suburban locations, the values of K^+/OC and K^+/EC are found to be (0.05 ± 0.007 and 0.66 ± 0.15), (0.05 ± 0.007 and 0.37 ± 0.38), and (0.03 ± 0.004 and 0.24 ± 0.022), respectively. In our study, we observed a good correlation of water-soluble ions (K^+), OC, and EC, which supports the idea that the sources of carbonaceous matter are biomass burning. In addition, Nowshera city has a large number of factories, that is, cement industries, brick kiln industries, and others that release a large number of emissions because of the poor technology used in the manufacturing processes. For that reason, it can be interesting to note that the air masses from various directions have a significant effect on the air quality in Nowshera city. The results of the HYSPLIT model show that the air mass originates from local sources.

DATA AVAILABILITY STATEMENT

The original contributions presented in the study are included in the article/supplementary material; further inquiries can be directed to the corresponding authors.

AUTHOR CONTRIBUTIONS

BZ contributed to conceptualization; BZ and SU helped with data curation; BZ, KA, AD, and SU assisted with formal analysis; HA, MI, and MS acquired funding; KA and SU framed the methodology; KA, HA, MI, and MS involved in project administration; KA, AD, and SU helped with software; AD, HA, MI, and MS validated the study; AD contributed to visualization; BZ and KA wrote the original draft; and BZ, AD, SU, HA, MI, and MS reviewed and edited the manuscript. All authors have read and agreed to the published version of the manuscript.

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